# WORKING GROUP ON WIDELY DISTRIBUTED STOCKS (WGWIDE) 

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## i Executive summary

WGWIDE reports on the status and considerations for management of the Northeast Atlantic mackerel, blue whiting, Western and North Sea horse mackerel, Northeast Atlantic boarfish, Norwegian spring-spawning herring, striped red mullet (subareas 6, 8, and divisions 7.a-c, e-k and 9.a), and red gurnard (subareas 3, 4, 5, 6, 7, and 8) stocks.
2024 catch advice was drafted for all eight stocks. Benchmark assessments are proposed for 2025 for mackerel, herring, and striped red mullet while both horse mackerel stocks and boarfish are scheduled to be benchmarked in 2024.

Northeast Atlantic mackerel. This migratory stock is widely distributed throughout the Northeast Atlantic with significant fisheries in several ICES subareas. The assessment conducted in 2023 is an update assessment, based on the configuration agreed during the 2019 interbenchmark and incorporates updates to the commercial catch, tagging, swept-area, egg survey, and recruitment index dataseries. Advice is given based on stock reference points which were updated during a management strategy evaluation carried out in 2020. SSB has been declining since 2014 but is above MSY $B_{\text {trigger }}$ in 2023. Fishing mortality has been increasing since 2016 and above $\mathrm{F}_{\text {mSY }}$ since 2021.

Blue whiting. This pelagic gadoid is widely distributed in the eastern part of the North Atlantic. The current assessment configuration (interbenchmark in 2016) uses preliminary catch and sampling data along with the acoustic survey data from the current year. The 2023 update assessment indicates that SSB is increasing following strong recent recruitment and is well above MSY $\mathrm{B}_{\text {trigger }}$. Fishing mortality has been above Fmsy since 2014.

Norwegian spring-spawning herring. This stock is migratory, spawning along the Norwegian coast and feeding throughout much of the Norwegian Sea. The 2023 update assessment is based on an implementation of the XSAM assessment model introduced following a benchmark in 2016 and is consistent with the 2022 assessment. SSB has been declining since 2009, except for an increase in 2021 which was due to the strong 2016 year class entering the SSB. Fishing mortality has been above FMSY since 2019. Recruitment since 2016 is estimated to be below average and SSB is forecast to fall below MSY Btrigger in 2024 and decrease further in 2025 even if the management plan is applied in 2024.

Western horse mackerel. The Western stock of horse mackerel is distributed throughout ICES subareas $4,6,7,8$, and 9 . Following a benchmark in 2017, the stock is assessed using the Stock Synthesis (SS) integrated assessment model. Stock reference points were revised in 2019. Following a period of declining SSB, there has been a modest rise since 2017, albeit from a low level. The 2023 assessment indicates that SSB is below $\mathrm{B}_{\mathrm{lim}}$ and will remain so in 2025, even under the scenario of zero catch in 2024. Based on the MSY approach, advice for 2024 is therefore for zero catch. The assessment has been displaying significant retrospective bias, but the 2023 assessment is consistent with the 2022 assessment.

North Sea horse mackerel. Catch advice for this stock is issued biennially on the basis of an assessment based on a combined index from groundfish surveys in the North Sea and the Channel. The 2023 value is an increase on the 2022 value for the exploitable stock. A length-based indicator continues to indicate that fishing mortality remains above Fmsy.

Northeast Atlantic boarfish. Boarfish is a small, pelagic, planktivorous, shoaling species found over much of the Northeast Atlantic shelf but primarily in ICES subareas 4, 6, 7, and 8 . The directed fishery occurs primarily in the Celtic Sea and developed during the early 2000s, initially unregulated before the introduction of a TAC in 2011. The stock is assessed using an exploratory

Bayesian surplus production model with catch and survey data from groundfish surveys and an acoustic survey. The current assessment indicates that, following a sharp decline after 2012, biomass has been increasing in recent years. The most recent acoustic surveys indicate a period of above-average recruitment in recent years. A length-based indicator indicates that fishing mortality is below Fmsy.

Northeast Atlantic red gurnard. This stock was first considered by WGWIDE in 2016 with advice issued biennially. The assessment was benchmarked in 2021 and a survey-based relative biomass indicator was developed. The 2023 update assessment continues to show the indicator fluctuating without trend since 2010. However, large uncertainties remain regarding landings data due to poor resolution at the species level and reported discarding levels vary widely.
Striped red mullet in Bay of Biscay, Southern Celtic Seas, Atlantic Iberian waters. No assessment is available for this stock and information on abundance and exploitation level is limited with advice given triennially on the basis of the precautionary approach. However, there are a number of research projects underway which will inform a future benchmark and potential upgrade of the assessment category.

## ii Expert group information

| Expert group name | Working Group on Widely Distributed Stocks (WGWIDE) |
| :--- | :--- |
| Expert group cycle | Annual |
| Year cycle started | 2022 |
| Reporting year in cycle | $1 / 1$ |
| Chair | Erling Kåre Stenevik. Norway |
| Meeting venue and dates | $23-29$ August 2023, Copenhagen, Denmark (38 participants) |

## 1 Introduction

## Working Group on Widely Distributed Stocks (WGWIDE)

### 1.1 Terms of References (ToRs)

The Working Group on Widely Distributed Stocks (WGWIDE), chaired by Erling Kåre Stenevik, met in ICES HQ in Copenhagen in hybrid format, 23-29 August 2023. The terms of reference for the meeting were the generic ToRs for Regional and Species Working Groups:
a) Consider and comment on Ecosystem and Fisheries Overviews with a focus on:
i) identifying and correcting mistakes and errors (both in the text, tables and figures), and
ii) proposing concrete evidence-based input that is considered essential to the advice but is currently underdeveloped or missing (with references and Data Profiling Tool entries, as appropriate).

The input will feed into the annual updates of the overviews. Delivery of contributions other than those outlined above is also welcomed but will be utilized during the revision process (around every 5 years).
b) Conduct an assessment on the stock(s) to be addressed in 2023 using the method (assessment, forecast or trends indicators) as described in the stock annex; - complete and document an audit of the calculations and results; and produce a brief report of the work carried out regarding the stock, providing summaries of the following where relevant:
i) Input data and examination of data quality; in the event of missing or inconsistent survey or catch information refer to the ACOM document for dealing with missing data and the linked template that formulates how deviations from the stock annex are to be reported.
ii) Where misreporting of catches is significant, provide qualitative and where possible quantitative information and describe the methods used to obtain the information;
iii) For relevant stocks (i.e. all stocks with catches in the NEAFC Regulatory Area), estimate the percentage of the total catch that has been taken in the NEAFC Regulatory Area in 2022.
iv) For category 3 and 4 stocks requiring new advice in 2023, implement the methods recommended by WKLIFE X (e.g. SPiCT, rfb, chr, rb rules) to replace the former 2 over 3 advice rule ( 2 over 5 for elasmobranchs). MSY reference points or proxies for the category 3 and 4 stocks (guidelines)
v) Evaluate spawning-stock biomass, total-stock biomass, fishing mortality, catches (projected landings and discards) using the method described in the stock annex;

1) for category 1 and 2 stocks, in addition to the other relevant model diagnostics, the recommendations and decision tree formulated by WKFORBIAS (see Annex 2) should be considered as guidance to determine whether an assessment remains sufficiently robust for providing advice.
2) If the assessment is deemed no longer suitable as basis for advice, provide advice using an appropriate Category 2-5 approach as described in ICES technical guidance for harvest control rules and stock assessments for stocks in categories 2 and 3 or ICES.
3) If the assessment has been moved to a Category 2-5 approach in the past year consider what is necessary to move back to a Category 1 and develop proposal for the appropriate benchmark process.
vi) Catch scenarios for the year(s) beyond the terminal year of the data for the stocks for which ICES has been requested to provide advice on fishing opportunities;
vii) Historical and analytical performance of the assessment and catch options with a succinct description of associated quality issues. For the analytical performance of category 1 and 2 age-structured assessments, report the mean Mohn's rho (assessment retrospective bias analysis) values for time-series of recruitment, spawningstock biomass, and fishing mortality rate. The WG report should include a plot of this retrospective analysis. The values should be calculated in accordance with the "Guidance for completing ToR viii) of the Generic ToRs for Regional and Species Working Groups - Retrospective bias in assessment" and reported using the ICES application for this purpose.
c) Produce a first draft of the advice on the stocks under considerations according to ACOM guidelines.
d) Review progress on benchmark issues and processes of relevance to the Expert Group.
i) update the benchmark issues lists for the individual stocks in SID;
ii) review progress on benchmark issues and identify potential benchmarks to be initiated in 2024 for conclusion in 2025;
iii) determine the prioritization score for benchmarks proposed for 2024-2025;
iv) as necessary, document generic issues to be addressed by the Benchmark Oversight Group (BOG)
e) Prepare the data calls for the next year's update assessment and for planned data evaluation workshops;
f) Identify research needs of relevance to the work of the Expert Group.
g) Review and update information regarding operational issues and research priorities on the Fisheries Resources Steering Group SharePoint site.
h) If not completed previously, complete the audit spread sheet 'Monitor and alert for changes in ecosystem/fisheries productivity' for the new assessments and data used for the stocks. Also note in the benchmark report how productivity, species interactions, habitat and distributional changes, including those related to climate-change, could be considered in the advice.
i) Deliver conservation status advice in accordance with the "Technical Guidelines on the conservation status advice". The advice is only to be given when conservation aspects were identified and where clear, demonstrable management action can be recommended for any non-catch anthropogenic pressure. It can also be used to highlight clear demonstrable sensitivity to climate change. The qualification required to show clear, demonstratable management action is high. Avoid generic statements that are of no specific application to management.
j) Update SAG and SID with final assessment input and output

### 1.1.1 The WG work 2023 in relation to the ToRs

The WG considered updates for all eight stocks within its remit. Based upon these assessments and associated short-term forecasts, the group produced draft advice sheets for Northeast Atlantic mackerel, Blue Whiting, Norwegian spring-spawning herring, Western horse mackerel, North Sea horse mackerel, Boarfish, red gurnard and striped red mullet. All draft advice sheets were agreed in plenary. Advice sheets, report sections and assessments were audited with 2-3 working group members assigned to each stock.

### 1.2 Participants at the meeting

WGWIDE 2023 was attended by 37 delegates (11 online) from the Netherlands, Ireland, Spain, Norway, Germany, Portugal, Iceland, UK (England and Scotland), Faroe Islands, France, Denmark, and Sweden. The full list of participants, all of whom are authors of this report is given in Annex 1.

All the participants were made aware of ICES Code of Conduct, which all abided by, and none had Conflicts of Interest that prevented them from acting with scientific independence, integrity, and impartiality.

### 1.3 Overview of stocks within the WG

Eight stocks are assessed by WGWIDE. In 2023, the group drafted 2024 advice sheets for all 8 stocks. A summary of the WGWIDE stocks, current data category and assessment method and advice frequency is given in the table below:

| Stock | ICES code | Data category | Assessment <br> method | Assessment fre- <br> quency | Last assessment |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Boarfish | boc.27.6-8 | 3.2 | Bayesian Schafer <br> surplus produc- <br> tion model | 2 | 2021 |
| Red gurnard | gur.27.3-8 | 3.2 | Survey trends <br> based | 2 | 2021 |
| Norwegian <br> spring-sp. Herring | her.27.1-24a514a | 1 | XSAM | 1 | 2022 |
| Western horse <br> mackerel | hom.27.2a4a5b6 <br> a7a-ce-k8 | 1 | Stock Synthesis | 1 | 2022 |
| North Sea horse <br> mackerel | hom.27.3a4bc7d | 3.2 | Survey trends | 2 | 2021 |
| NE-Atlantic <br> mackerel | mac.27.nea | 1 | SAM | 1 | 2022 |
| Striped red mul- <br> let | mur.27.67a-ce- | 5 | N89a | No assessment | 3 |

### 1.4 Quality and adequacy of fishery and sampling data

### 1.1.2 Sampling data from commercial fishery

Each year, the working group reviews available sampling data and the level of sampling on the commercial fisheries. Details are given in the relevant stock-specific sections of this report.

Generally, the amount and quality of available data to the WG has been unchanged in the most recent years. However, in 2022 no Russian data submissions were available (for 2021). In 2023, Russian catch data for 2022 on NEA Mackerel, Blue Whiting and Norwegian Spring-Spawning Herring were available and used in the assessments of the three stocks.

The WG identified issues associated with the formatting and availability of data from commercial catch sampling programmes such as the requirement for length frequency and age-length key data for the assessment of Western horse mackerel and the availability of data arising from the sampling of catches of North Sea horse mackerel from foreign flagged vessels. The issues have been included on the individual stock issue lists and the ICES data call has been updated such that future data submissions should provide data in the appropriate format.

### 1.1.3 Catch data

The WG has on a number of occasions discussed the accuracy of the catch statistics and the possibility of large-scale under reporting or species and area misreporting. The working group considers that the best estimates of catch it can produce are likely to be underestimates.

In the case of red gurnard catch data, the available information is of poor quality. Prior to 1977, red gurnard catches were not reported. Since this time, landings of gurnards have often been reported as mixed gurnards, or using the incorrect species code. With the exception of Portugal, there is no detail provided to the WG on the methodology used to estimate the proportion of red gurnards in mixed landings.

### 1.1.4 Discards

In 2015, the European Union introduced a landing obligation for fisheries directed on small pelagic fish including mackerel, horse mackerel, blue whiting and herring. The obligation was expanded over the following years in a stepwise fashion such that discarding of small pelagic species could still legally occur in other fisheries. From 2019 onwards the landing obligation is generally effective. A general discard ban is already in place for Norwegian, Faroese and Icelandic fisheries.

Historically, discarding in pelagic fisheries is more sporadic than in demersal fisheries. This is because the nature of pelagic fishing is to pursue schooling fish, creating hauls with low diversity of species and sizes. Consequently, discard rates typically show extreme fluctuation ( $100 \%$ or zero discards). High discard rates occurred especially during 'slippage' events, when the entire catch is released. The main reasons for 'slipping' are daily or total quota limitations, illegal size and mixture with unmarketable bycatch. Quantifying such discards at a population level is extremely difficult as they vary considerably between years, seasons, species targeted and geographical region.

Discard estimates of pelagic species from pelagic and demersal fisheries have been published by several authors. Discard percentages of pelagic species from demersal fisheries were estimated between $3 \%$ to $7 \%$ (Borges et al., 2005) of the total catch in weight, while from pelagic fisheries were estimated between $1 \%$ to $17 \%$ (Pierce et al. 2002; Hofstede and Dickey-Collas 2006, Dickey-

Collas and van Helmond 2007, Ulleweit and Panten 2007, Borges et al. 2008, van Helmond and van Overzee 2009, 2010, van Overzee and van Helmond 2011, Ulleweit et al. 2016, van Overzee et al. 2013, 2020). Slipping estimates have been published for the Dutch freezer trawler fleet only, with values at around $10 \%$ by number (Borges et al. 2008) and around $2 \%$ in weight (van Helmond et al. 2009, 2010 and 2011) over the period 2003-2010. In Iberian waters the discard composition of pelagic species, mainly blue whiting, in demersal fisheries were estimated between $20 \%$ and $30 \%$ of the total catch in weight (Fernandes et al. 2015). Nevertheless, the majority of these estimates were associated with very large variances and composition estimates of 'slippages' are liable to strong biases and are therefore open to criticism.

Because of the potential importance of significant discarding levels on pelagic species assessments, the Working Group again recommends that observers should be placed on board vessels in those areas in which discarding occurs, and existing observer programmes should be continued. Furthermore, agreement should be made on sampling methods and raising procedures to allow comparisons and merging of dataset for assessment purposes. The newest update on discards for the different stocks assessed by the WG is provided in the sections for each of the stocks.

### 1.1.5 Age reading

Reliable age data are an important prerequisite in the stock assessment process. The accuracy and precision of these data, for the various species, is kept under constant review by the Working Group. The most recent updates on this aspect for the different stocks are addressed below.

### 1.4.1.1 Mackerel

The most recent age calibration exercise for this stock was carried out in 2021 using the SmartDots platform under the remit of WGBIOP. The full exercise was completed by 37 readers from 12 countries across Europe. Otolith images ( $\mathrm{n}=237$ ) were provided by 12 of the participating laboratories with the aim to provide a set of images representative of the temporal and spatial coverage of otoliths read for stock assessment purposes (including the southern component, western component, North Sea component and the northern distribution).

Results show a slightly lower percentage of agreement and higher CV in the analysis taking all readers in account than the previous workshop (2018) and exchange (2014) which might be related to an increased number of readers ( 23 in WK 2018; 37 in Ex 2021 with 10 new (basic) readers). However, lower agreement (and higher CV) was also found for advanced readers. Here, numbers of readers also increased from 2018 to 2021 ( 15 to 21).

The overall conclusion was that the slightly worse results than in the prior workshops might be related to the increased number of readers. The image quality of otoliths from different areas was also discussed. However, the problem shown in previous workshops and exchanges persists: Agreement for otoliths with modal age 6 and older remains quite low. A new workshop was recommended.

### 1.4.1.2 Horse mackerel

The most recent workshop on the age reading of Trachurus trachurus (also T. mediterraneus and T. picturatus) was carried out in November 2018 and involved 15 age readers from 9 countries. This was followed by an otolith exchange which was conducted in 2021 for the three horse mackerel species T. Trachurus, T. mediterraneus, and T. picturatus and 28, 14 and 18 readers participated, respectively.

For T. trachurus both whole $(\mathrm{N}=249)$ and sliced $(\mathrm{N}=134)$ otoliths were read from 10 different areas ( 5 in the Mediterranean and 5 in the Atlantic). For whole otoliths and all readers, the weighted average percentage agreement (PA) based on modal ages for all readers is $46 \%$, with a
weighted average CV of $44 \%$ and an APE of $32 \%$. For sliced otoliths and all readers, the weighted average percentage agreement based on modal ages for all readers is $44 \%$, with a weighted average CV of $22 \%$ and an APE of $15 \%$. The readers had some difficulties in recognizing first ring and the edge nature, and with overlapped rings in old specimens. CV PA APE decrease for sliced oto-liths; however, high differences in the modal age is observed between same sample/different preparation.

### 1.4.1.3 Norwegian Spring-spawning Herring

For some years, there have been issues with age reading of herring. These issues were raised around 2010, and since then two scale/otolith exchanges and a workshop have been held; and a final workshop was planned after the second exchange. There were, however, concerns with the second scale/otolith exchange and the final workshop was postponed. It was therefore recommended to organize a new scale/otolith exchange and a follow up workshop.

There are several topics to cover in the recommended work.
First, age-error matrices are needed as input to the stock-assessment, to evaluate sensitivity to ageing errors, and such age-error matrices are an output of age-reading inter-calibrations.

Second, stock mixing is an issue. There are several herring stocks surrounding the distribution area of Norwegian spring-spawning (NSS) herring, e.g. North Sea herring, Icelandic summer spawning herring, local autumn-spawning herring in the Norwegian fjords, and Faroese autumn spawning herring. Mixing with these other stocks in the fringe areas of the NSS herring distribution area leads to confounding effects on the survey indices of NSS herring in the ecosystem surveys and potentially also in the catch data. Methods to separate the NSS herring stock from the other herring stocks are needed - both with regards to obtain more accurate age-readings as well as to reduce confounding effects on the survey indices.

Finally, the experience from earlier exchanges is that age of older fish is more prone to be underestimated when aged is read from otoliths as compared to being read from scales. Some of the institutes mainly sample and read scales, whereas other institutes use the otoliths.

In 2021, WGWIDE recommended to organize a scale/otolith exchange and workshop. This workshop, WKRANSSH was held in Bergen 17-21 April 2023 (ICES 2023a). Overall results for both exchanges and the combination were high and at acceptable levels. However, as described above, disagreement between scales and otoliths occurred especially for older individuals (age 8 and above), and disagreement between scales and otoliths occurred especially for older individuals. The next step is to investigate how these discrepancies impacts the stock assessment, based on the general age error matrix from the workshop. Stock mixing seems to be a minor issue that will have no direct impact on the assessment of NSS herring. However, this needs to be investigated in more detail in the near future.

### 1.4.1.4 Blue Whiting

The most recent workshop on the age reading of blue whiting (WKARBLUE3) took place in 2021 ( 31 May-4 June). The workshop was preceded by an inter-calibration age reading exchange, which was undertaken in 2020 using the SmartDots platform. In the exchange, the otolith collection included 407 otoliths from the entire stock distribution area, from which 190 otoliths where from the northern areas and 217 where from the southern areas of distribution. The otolith dataset allows a good coverage of samples by area and sex and took into account the differences in growth patterns by areas (northern and southern), and by sex due to the sexual dimorphism in blue whiting (Gonçalves et al. 2017).

The overall agreement of the pre-workshop exercise was $66 \%$ considering all readers and $70 \%$ for the assessment readers (advanced readers). Considering only the otoliths samples from the
northern areas and the readers from the northern that usually read the otoliths from those areas for the assessment, $69 \%$ of agreement was achieved. Otherwise, considering only the otoliths samples from the southern areas and the readers from the southern that usually read the otoliths from those areas for the assessment, $79 \%$ of agreement was achieved. During the workshop, a small exchange was also conducted with 55 otoliths in which $73 \%$ agreement between the advanced readers was achieved.

The main issues identified on blue whiting age reading are still: the fact that the otoliths from some areas revealed to be more difficult to read (e.g. 27.2.a, 27.5.b); the first ring identification; edge type interpretation and false or double rings identification (Gonçalves, 2021).

During the workshop some of the otoliths from the exercise were polished, to help readers in the cases were the first age ring were not so evident, completely absent, or showing a growth pattern different from the expected. The polishing results revealed to be useful on the ring interpretation and to help in cases here the visible first ring size presents a size higher than the expected and the readers have doubts if an inner first ring are there. The hypothesis of the existence of a nonvisible first ring has been described in the otoliths from the adult fish as the otolith becomes thicker and wider.

Although, during the WKARBLUE3 progresses have been made and objective and more clear age reading guidelines had been constructed. The recurrent age reading issues still remain the same, e.g. the identification of the position of the first annual growth ring, false rings and interpretation of the edge. In order to overcome those problems and increase the accuracy on age classifications, age validation studies on blue whiting otoliths to solve growth rings interpretation, were further recommended and should be conducted (ICES. 2023b).

### 1.4.1.5 Boarfish

Sampling of the commercial catch of boarfish has been included within the EU data collection framework since 2017. An age length key was produced in 2012 following increased sampling of a developing fishery. The age reading was conducted by DTU Aqua on samples from the three main fishery participants: Ireland, Denmark and UK (Scotland). No ageing has been carried out since 2012 although otoliths continue to be collected from the Irish fishery during routine catch sampling.
In November 2022, an ageing exercise was initiated via SmartDots with the aim to age samples less than 10 cm (all samples less than 10 cm used in the ALK were estimations), to compile otolith images for future training purposes and to potentially shed light on the possibility of adjusting the plus group designation. A total of 158 boarfish otoliths of varying difficulty were aged by three readers from Denmark and Ireland. The otoliths were sourced from both legs of the 2022 WESPAS survey ( $14^{\text {th }}$ June to $24^{\text {th }}$ July). The fish length distribution of the samples ranged from $55-170 \mathrm{~mm}$ with one third of the samples in the $55-95 \mathrm{~mm}$ length range. To image and read the otoliths, the protocol outlined in the boarfish age reading manual, created by DTU Aqua, was followed.

Results from the ageing exercise found that boarfish within the length range of $5.5 \mathrm{~cm}-9.5 \mathrm{~cm}$ were between the ages of 1 and 3 years old. Concerning the plus group designation, the exercise wasn't able to advise or provide useful information to support or oppose changes to the plus group. The full results from the exercise were compiled in a report and published on the SmartDots webpage (ID 509).

### 1.4.1.6 Striped red mullet

In 2011, an otolith exchange was carried out, the second such exercise for the striped red mullet. For details see section 10.5.

### 1.4.1.7 Red gurnard

Age data are available for red gurnard from the EVHOE and IGFS groundfish surveys. Improvements in the understanding of the age structure of this stock would be improved by reading otoliths from other surveys in the assessment area (e.g. NS-IBTS, SCO-WCS, CGFS) which also contribute information on stock status in term of their CPUE series.

### 1.1.6 Current methods of compiling fisheries assessment data

Information on official, area misreported, unallocated, discarded and sampled catches have again this year been recorded by the national laboratories on the WG-data exchange sheet (MS Excel; for definitions see text table below) and sent to the stock co-ordinators and uploaded through InterCatch. Co-ordinators collate data using the either the sallocl (Patterson, 1998) application which produces a standard output file (sam.out) or InterCatch.
There are at present no specified criteria on the selection of samples for allocation to unsampled catches. The following general process is implemented by the species co-ordinators. A search is made for appropriate samples by gear (fleet), area, and quarter. If an exact match is not available the search will extend to adjacent areas, should the fishery extend to this area in the same quarter. Should multiple samples be available, more than one sample may be allocated to the unsampled catch. A straight mean or weighted mean (by number of samples, aged or measured fish) of the observations may be used. If there are no samples available the search will move to the closest non-adjacent area by gear (fleet) and quarter, but not in all cases.

It is not possible to formulate a generic method for the allocation of samples to unsampled catches for all stocks considered by WGWIDE. However full documentation of any allocations made are stored each year in the data archives (see below). It should be noted that when samples are allocated the quality of the samples may not be examined (i.e. numbers aged) and that allocations may be made notwithstanding this. The Working Group again encourages national data submitters to provide an indication of what data could be used as representative of their unsampled catches.

Following the introduction of the landings obligations for EU fisheries new catch categories had to be introduced from 2015 onwards. The catch categories used by the WGWIDE are detailed below:

| Official Catch | Catches as reported by the official statistics to ICES |
| :--- | :--- |
| Unallocated Catch Adjustments (positive or negative) to the official catches made for any special knowledge of the <br> fishery, such as under- or over-reporting for which there is firm external evidence. <br> Area misreported <br> Catch To be used only to adjust official catches which have been reported from the wrong area (can be <br> negative). For any country the sum of all the area misreported catches should be zero. <br> BMS landing Landings of fish below minimum landing size according to landing obligation <br> Logbook registered <br> discards Discards which are registered in the logbooks according to landing obligation <br> Discarded Catch Catch which is discarded <br> Wa Catch The sum of the 6 categories above <br> The catch corresponding to the age distribution Catch  |  |

### 1.1.7 Quality of the Input data

Primary responsibility for the accuracy of national biological data lies with the national laboratories that submit such data. Each stock co-ordinator is responsible for combining, collating, and interpolating the national data where necessary to produce the input data for the assessments. A number of validation checks are already incorporated in the data submission spreadsheet currently in use, and these are checked by the co-ordinators who in the first instance report anomalies to the laboratory which provided the data.

Overall, data quality has improved and sampling deficiencies have been reduced compared to earlier years, partly due to the implementation of the EU sampling regulation for commercial catch data. However, some nations still have no (or inadequate) aged samples. Occasionally, no data are submitted such that only catch data from EuroStat is available, which are not aggregated quarterly but are yearly catch data per area.

The Working Group documents sampling coverage of the catches in two ways. National sampling effort is tabulated against official catches of the corresponding country (see stock specific sections). Furthermore, tables showing total catch in relation to numbers of aged and measured fish by area give a picture of the quality of the overall sampling programme in relation to where the fisheries are taking place. These tables are contained in the species sections of this report.

The national data on the amount and the structure of catches and effort are archived in the ICES InterCatch database. The data are provided directly by the individual countries and are highly aggregated for the use of stock assessments.

There exist gaps in some dataseries, in particular for historical periods. The WG has requested members to provide any national data reported to previous working groups (official catches, working group catches, catch-at-age and biological sampling data) not currently available to the WG. Furthermore, the WG recommends that national institutes increase national efforts to collate historic data.

A number of stock data problems relevant to data collections have been brought forward to the contact person in preceding years. Those that still apply are listed in table below for the information of ICES-Working Groups and RCGs as specified.

| Stock | Data Problem | How to be addressed in | By who |
| :---: | :---: | :---: | :---: |
| Northeast Atlantic Mackerel | Submission of data | Data submissions must include all the data outlined in the data call and be submitted by the deadline. Data should include length distributions split by area and quarter. <br> Should the data submitter be unavailable after the data has been submitted (e.g. vacation) an alternative contact should be available who can be contacted in the event of any queries. | National laboratories |
| Northeast Atlantic Mackerel | Discard and slippage information | Discard and slippage information is incomplete. All fleets, including demersal fleets should be monitored and sampled for discards and slipping. Data should be supplied to the coordinator by the submission deadline, accompanied by documentation describing the sampling protocol. | National laboratories, RCG NA, RCG NS\&EA |
| Northeast Atlantic Mackerel | Sampling deficien-cies- general | All countries involved should provide sampling information. Increased cooperation between countries would help reduce redundancy and increase coverage. | National laboratories, RCG NA, RCG NS\&EA |


| Stock | Data Problem | How to be addressed in | By who |
| :--- | :--- | :--- | :--- |
| Northeast At- <br> lantic Mackerel | Sampling of foreign <br> vessels | Any information available from the sampling of foreign <br> vessels should be forwarded to the appropriate person in <br> the national laboratory so that they may use this infor- <br> mation when compiling the data submission. | National labora- <br> tories; RCG NA, <br> RCG NS\&EA |
| Horse Macke- <br> rel - Western <br> Stock | Missing sampling <br> data for some parts <br> of the distribution <br> area (e.g. 27.2a, 7e) | Fishing nations to Sample age and length Distributions <br> from commercial fleets | National Insti- |
| Horse Macke- <br> rel - North Sea <br> Stock | Incomplete report of <br> discards by non-pe- <br> lagic fleet. | Reporting of discards by national institutes. | tutes |

### 1.1.8 Quality control of data and assessments, auditing

As a quality control of the data and the assessment, WG participants were appointed as auditors for each stock. The primary aim of the auditing process is to check that the assessment and forecast has been conducted as detailed in the relevant stock annex. Auditors conducted checks of the assessment input data, assessment code (time permitting), draft WG report and draft advice sheet. Auditors completed an audit report upon completion (annex 4). Issues identified in the audit reports were followed up by the appropriate stock coordinator/assessor with updates made where appropriate.

### 1.1.9 Information from stakeholders

The procedure for the submission of inputs from stakeholders into the scientific advice changed in 2020. Instead of contributing information directly into the Advice Drafting Groups, information from stakeholders is now submitted directly to the expert group for consideration and inclusion into the draft advice, if applicable.

For WGWIDE stocks there are several instances of strong cooperation between research institutes and fishing industry stakeholder in the collection of data that is used in the assessments, e.g. the acoustic survey for Norwegian Spring-spawning herring, the extension of the IESSNS survey into the North Sea and several cases where industry vessels are collecting samples for catch monitoring. In these cases, the research institutes are coordinating the activities and bringing the results directly to the expert group(s).

A recent development that started around 2014 involves fishing industry organizations taking initiatives on their own, to collect additional information that is contributed to the expert groups. In many cases these research activities are undertaken in close cooperation with research institutes. During WGWIDE 2022, the following contributions from fishing industry research activities were reported to the working group:

1. PFA self-sampling report 2015-2021
2. Gonad sampling for mackerel and horse mackerel in support of the 2022 egg survey
3. Horse mackerel genetics

### 1.1.9.1 PFA self-sampling report (WDO2)

The Pelagic Freezer-trawler Association (PFA) initiated a self-sampling programme in 2015, aimed at expanding and standardizing ongoing fish monitoring programmes by the vessel quality managers on board of the vessels. An overview of the self-sampling in widely distributed pelagic fisheries from 2016 onwards is presented in the text table below.

| Year | Number Vessels | Number <br> Trips | Number Days | Number Hauls | Catch (t) | Catch per Day ( t ) | Number <br> Length <br> Measurements |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | 12 | 62 | 842 | 1,783 | 178,162 | 212 | 0.25\% |
| 2018 | 16 | 86 | 1,220 | 2,679 | 253,474 | 208 | 0.22\% |
| 2019 | 16 | 97 | 1,233 | 2,668 | 225,059 | 183 | 0.29\% |
| 2020 | 17 | 113 | 1,434 | 3,051 | 306,172 | 214 | 0.35\% |
| 2021 | 19 | 119 | 1,401 | 2,881 | 282,898 | 202 | 0.52\% |
| 2022 | 18 | 114 | 1,259 | 2,736 | 237,438 | 189 | 0.69\% |
| 2023* | 18 | 68 | 785 | 1,934 | 189,832 | 242 | 0.18\% |
| (all) | 659 | 8,174 | 17,732 | 1,673,036 |  |  |  |

[^1]In the 2023 self-sampling report, a standardized CPUE calculation has been included for most of the stocks. The standardized CPUE is based on a GLM model with a negative binomial distribution. The response variable is catch by week and vessel, with an offset of the log effort (number of fishing days per week) and explanatory variables year, GT category, month, division and depth category. An assumed technical efficiency increase of $2.5 \%$ per year has been included in the fitting of the model (Rousseau et al 2019)

The trends in the 2023 Mackerel assessment are very similar to the standardized CPUE for the PFA vessels. The trends for Western horse mackerel CPUE are markedly different from the stock assessment where the CPUE suggests a further decline while the stock assessment points towards slow recovery of the stock. Limited quota may have resulted in a fishery not being representative of full stock exploitation. For North Sea horse mackerel, the assessment suggests a recent increase in stock size while the CPUE shows a continuous decline since 2017. It is suggested to investigate these differences at its next benchmark meeting. The CPUE for blue whiting showed a rather stable stock in the past 7 years while the stock assessment clearly shows a peak at around 2018, a decline followed by a peak again in 2022-2023. PFA will interact with the skippers to find an explanation for this difference in CPUE and stock trends as it is clear that the fishery was good this year.

### 1.4.1.8 Gonad sampling for mackerel and horse mackerel

During 2023, a dedicated PFA industry researcher carried out one sampling trip on-board of a commercial trawler with the aim to collect frozen gonad samples, all frozen in a different way to assess if the way of freezing has an impact on the fecundity estimates. Samples were delivered to Wageningen Marine Research and results are still pending.

### 1.4.1.9 Horse mackerel genetic stock identification

In an effort to address the longstanding questions regarding the stock identification of horse mackerel the northern European pelagic fishing industry commissioned a series of research projects to develop genetic methods for discriminating the stocks. Next Generation Sequencing (NGS) based approaches were used to identify informative genetic markers and to screen a larger number of samples, which indicated a clear separation of the southern North Sea spawning samples from other regions and further, less pronounced structure along the Northeast Atlantic continental shelf (Brunel et al., 2016; Farrell and Carlsson, 2018). However, it was concluded that further genetic analyses were required to increase the numbers and types of genetic markers available, which would improve the capacity for accurate genetic assignment.

In 2019 the Northern Pelagic Working Group (NPWG) of the European Association of Fish Producers Organisations (EAPO) commissioned a further project to develop a reference genome assembly for horse mackerel and to undertake whole genome sequencing (pool-seq) of population samples in order to identify informative single-base genetic markers (Single Nucleotide Polymorphisms - SNPs). Analyses of $\sim 12.8$ million SNPs indicated that the North Sea spawning samples were the most genetically differentiated group, whereas the structure among the Western, Southern and north African samples was less clear, though a north-south split was observed with a potential mixing zone between the Western and Southern populations in the southern part of Division 9.a, near Lisbon (Fuentes-Pardo et al., 2023). The analyses indicated that the samples from northern Portuguese waters were from the same biological population as those from the Western stock area i.e. the Western population. The samples from southern Portuguese waters comprised a mix of the 'Western population' type and a distinct southern population, which was more closely related to the North Africa population. Therefore, it was concluded that the southern limit of the Western stock was likely further south than the current definition of the border between Divisions 8.c and 9.a.

During this study a large number of informative SNPs were identified and c.4,000 of these, spread across the 24 chromosomes, were included on the multispecies genotyping array (IdentiGEN DNA TRACEBACK® Fisheries array FSHSTK1D) that is also being used for genotyping herring in divisions 6.a, 7.b-c, the Irish Sea and the Celtic Sea to allow population level genetic assignment.

In late 2022 the NPWG commissioned a further in-depth analysis of the genetic population structure of horse mackerel and the development of a genetic baseline and genetic assignment model using the new genotyping array. In total 35 samples, comprising 2,304 individuals were genotyped (Figure 1.4.8.3.1), including temporal replicates from all three stocks (See Farrell, 2023). Additional spawning samples from Division 9.a, which were not available for analysis by Fuentes-Pardo et al. (2023), were included and also non-spawning samples from potential mixing zones in divisions 4.a, 7.d and 7.e (Figure 1.4.8.3.1). The genotyping array yielded high quality data from 2,421 SNPs distributed across all twenty-four chromosomes.


Figure 1.4.8.3.1. The sample locations for the horse mackerel samples genotyped during the latest genetic project. Dark green indicates spawning baseline samples, light green indicates potential baseline samples and orange indicates nonspawning mixed samples.

Exploratory analyses of the data included filtering of the dataset to exclude linked SNPs and testing of different approaches to identify the most informative markers. The different panels of markers, including an 8-SNP panel (Figure 1.4.8.3.2), all indicated the same population structure, which agreed with Fuentes-Pardo et al. (2023). The additional spawning samples from Portuguese waters clustered with the Western population samples and only samples from the southern tip (Cabo de Sao Vicente) and south coast of Portugal were significantly differentiated (Figure 1.4.8.3.2), confirming that the Western population extends further south than the current Western stock limits.


Figure 1.4.8.3.2. Principle Coordinate Analysis (PCoa) of $F_{s t}$ illustrating the differentiation between the analysed baseline and potential baseline samples. NS = North Sea, WS = Western (west/southwest of Ireland), NSS = Northern Spanish Shelf, NPT = Northern Portugal, SPT = Southern Portugal, NAF = North Africa.

There were insufficient representative spawning samples available from the southern population to include in the development of the assignment model, therefore only the North Sea and Western populations were included. A support vector machine learning (svm) based approach was employed based on the available spawning baseline samples: 109 North Sea individuals ( 3 samples) and 383 Western individuals ( 10 samples). The assignment models were tested using two approaches, the first using all SNPs with dimensionality reduction through limiting the number of PCs based on the results of cross validation and the second with a reduced panel of 24 SNPs identified as the most informative to discriminate the two populations. The self-assignment rates were over $90 \%$ for each of the approaches, though the North Sea samples had a higher error rate than the Western samples.

Potentially mixed samples collected from commercial catches from Divisions $4 . a$ (quarters 1 and 2 and quarters 3 and 4), 7.d and 7.e were assigned to population of origin using the assignment models (Figure 1.4.8.3.3.). The samples from Division 4.a. assigned to primarily to the Western stock regardless of quarter, indicating that these Q 1and2 samples should not be allocated to the North Sea catches as they were from the Western population. The 7.e samples assigned primarily to the Western population though contained a small number of North Sea fish. The 7.d sample from September 2020 assigned primarily to the Western population whereas the November and December 2020 samples assigned primarily to the North Sea population though still contained a significant proportion of Western fish present. The analyses indicated that there was a significant level of mixing between the Western and North Sea populations in the Division 7.d samples analysed.

In summary the analyses undertaken indicated that the Southern stock area contained at least two populations, one which was the same as the western population and one in the south of the area which was a distinct Southern population. Analyses also indicated that the current delineation of the North Sea stock appears to be inappropriate in Division 4.a and that there was significant mixing with the Western population in Division 7.d. Given that the Western stock area extends into Division 9.a, 4.a in all quarters and 7.d it is likely that the recruitment, biomass and catch of the Western population are underestimated. A report with the detailed analyses described above is being prepared and will be submitted to the SIMWG for review in advance of the Horse Mackerel Benchmark (see Section 5.10).


Figure 1.4.8.3.3. The outputs of the exploratory North Sea vs. Western assignment model. Each pie represents a single sample but the number of individuals in the samples varies (see Figure 1.4.8.3.1.).

### 1.5 Comment on update and benchmark assessments

Updates were presented to the WG for the stocks in the group.
Western and North Sea horse mackerel were assessed on basis of a benchmark that took place in January 2017 (ICES, 2017) and NEA mackerel on an interbenchmark that took place in 2019 (ICES 2019a). Norwegian spring-spawning herring was assessed using the XSAM implementation benchmarked in 2016. The Blue whiting SAM assessment was introduced following a benchmark in 2012. Since this time, an interbenchmark in 2016 incorporated the use of preliminary in-year catch data with the stock weights in the assessment year estimated from catch sampling incorporated in 2019 (previously the average of the most recent three years was used). The acoustic survey time-series was updated in 2020 following recalculation by the StoX platform with minor updates to the historic index. The red gurnard assessment conducted at WGWIDE 2023 followed a benchmark in February 2021 (WKWEST) during which an index of abundance based on a number of bottom trawl surveys was developed. Boarfish was considered at the SPiCT benchmark in January 2023 (WKBMSYSPiCT) but unfortunately it was unsuccessful as no model fit met all the points of the SPiCT checklist.

The remaining stock addressed by the WG (striped red mullet) has not been benchmarked recently but was still assessed by the WG.

### 1.6 Planning future benchmarks

The striped red mullet stock is still to be benchmarked as there is no assessment in place. A full benchmark is proposed by WGWIDE 2023 to take place before the next advice is scheduled in 2026.

Since boarfish was unsuccessful at WKBMSYSPiCT this year, it is still on the list for a benchmark in 2024. Ongoing sampling of the commercial catch, an expanded acoustic survey time-series and exploring alternative assessment models e.g. Stock synthesis will be explored. Research projects underway for Striped red mullet are due to be completed in the near future and will inform the proposed benchmark for 2025.

Full benchmarks are scheduled for western horse mackerel and North Sea horse mackerel in 2024.

WGWIDE 2023 is proposing full benchmark to take place for Northeast Atlantic Mackerel and Norwegian Spring-spawning Herring.

The workshop to review the current assumptions with regard to stock structure (components) for mackerel (WKMACEVAL) were conducted in 2023. The Workshop rejected the current threecomponent structure and accepted a single NEA mackerel stock concept to advance towards assessment, management and advice. This conclusion will simplify data issues in a benchmark. Exploratory work is already underway or is planned on a number of issues related to the mackerel assessment including dealing with individual high catch rates in the swept-area survey, DEPMvs.AEPM methodologies for the egg survey time-series, inclusion of additional ages from the tagging dataset, increasing the assessment recruitment age and updating the SAM configuration.

The proposed benchmark of Norwegian Spring-spawning Herring will explore issues such as the splitting of exiting survey indices, inclusion of additional surveys (IESSNS and tagging data), assumptions on maturity in the most recent years.

The XSAM software is used for the assessment. However, the SAM software now supports the XSAM model as an optional model. A switch from the currently used code to the SAM platform should be done in order to make the model more publicly available and to ensure further development of the infrastructure. WGWIDE 2023 recommends conducting an expert group level benchmark in 2024 in order to switch from XSAM to SAM software from 2024 and onwards as the XSAM software is no longer maintained. A comparison between the two software has been made (WD09), resulting in a negligible difference between the two software.

Issue lists and benchmark scoring sheets for each of the stocks proposed for benchmarking by WGWIDE 2023 were reviewed and updated during the meeting.

The current status of the WGWIDE stocks with respect to benchmarking is summarized below:

| Stock | Benchmark History | WGWIDE 2023 Proposal |
| :--- | :--- | :--- |
| Boarfish | Benchmark scheduled for 2024 |  |
| Red gurnard | Full benchmark 2021 | Expert group level benchmark during WGWIDE 2024 and full <br> benchmark in 2025 |
| Norwegian Spring <br> Spawning herring | Full benchmark 2016 | Rull benchmark 2017 <br> Weference point interbenchmark horse <br> mackerel |


| Stock | Benchmark History | WGWIDE 2023 Proposal |
| :--- | :--- | :--- |
|  | Benchmark scheduled for 2024 |  |
| North Sea <br> horse mackerel | Full benchmark 2017 | Benchmark scheduled for 2024 |$\quad$| Northeast Atlantic <br> mackerel | Full benchmark 2014 |
| :--- | :--- |
| Full benchmark 2017 | Full benchmark in 2025 |
| Inter-benchmark 2019 |  |
| Blue whiting | Benchmarked 2012 |

### 1.7 Scientific advice and management of widely distributed and migratory pelagic fish

### 1.1.10 General overview of management system

The North East Atlantic Fisheries Commission (NEAFC) is the Regional Fisheries Management Organisation (RFMO) for the Northeast Atlantic. NEAFC is an end-user of ICES advice and provides a forum for its contracting parties (Coastal States and fishing parties) to manage the exploitation of straddling stocks that occur in several EEZs and international waters such as WGWIDE stocks Northeast Atlantic Mackerel, Blue Whiting and Norwegian Spring-spawning herring (also known as Atlanto-Scandian herring). There are 6 contracting parties to NEAFC: Denmark (in respect of the Faroe Islands and Greenland), European Union, Iceland, Norway, Russian Federation and the UK. The management of Western horse mackerel is not considered by NEAFC with sharing subject of separate agreements between EU, Norway and the UK.

### 1.1.11 Management plans

Catch advice in recent years for two stocks considered by WGWIDE has been given on the basis of an agreed long-term management strategy:

- A long-term management strategy for Norwegian spring-spawning herring was agreed by the European Union, the Faroe Islands, Iceland, Norway and Russian Federation in 2018 following an evaluation by ICES (WKNSSHMSE, ICES, 2018a) which found it to be precautionary. The plan is based on a target fishing mortality of 0.14 when the stock is above $\mathrm{B}_{\mathrm{pa}}$. Should SSB fall below $\mathrm{B}_{\mathrm{pa}}$, the target fishing mortality is linearly reduced to 0.05 at and below Blim. The plan incorporates TAC change limits of $-20 \%$ and $+25 \%$ which are suspended when below $\mathrm{B}_{\mathrm{pa}}$ and $10 \%$ interannual transfer which is suspended when below Blim. The plan is scheduled for review no later than 2023. Although the plan is agreed by the parties involved in the fishery and ICES advice is based on application of the management strategy, there has been no agreement on the relative catch share since 2013 with the total unilaterally declared quotas exceeding the management plan based catch advice since this time.
- A long-term management strategy for Blue Whiting was agreed by the European Union, the Faroe Islands, Iceland and Norway in 2016 following an evaluation by ICES (WKBWMS, ICES, 2016) in 2016 which found it to be precautionary. The plan is based on


#### Abstract

a target fishing mortality equivalent to $\mathrm{F}_{\text {MSY }}(0.32)$ when the stock is above $\mathrm{B}_{\text {pa }}$. Should SSB fall below $\mathrm{B}_{\mathrm{pa}}$, the target fishing mortality is linearly reduced to 0.05 at and below $B \lim$. The plan incorporates TAC change limits of $+/-20 \%$ which are suspended when below $\mathrm{B}_{\mathrm{pa}}$ and $10 \%$ interannual transfer. The Coastal States subsequently agreed a longterm management plan for the blue whiting stock with a TAC change limit of $-20 \% /+25 \%$ and a new clause (6.b.) stating that the TAC change limit is not applied when the catch advice deviates more than $40 \%$ from the TAC of the preceding year (clause 6.b., see Annex 1 in Anon, 2021). This clause was then evaluated by ICES and found precautionary in 2017 (Working Document in ICES, 2017). No agreement on quota shares has been reached since 2015 and catches have exceeded advice since this time. Since the management plan target fishing mortality is equivalent to $\mathrm{F}_{\text {MSY, }}$, the MSY approach results in the same advice as the LTMS.


There is no currently agreed management strategy for either Northeast Atlantic Mackerel or Western horse mackerel. Strategies have been proposed and evaluated but agreement has not yet been reached on their implementation such that catch advice has been given on the basis of the MSY approach.

### 1.7.1 Comparison of advice, TAC, and catches

This section presents an overview of the time-series (2010 to present) of ICES catch advice, TAC (either agreed between all fishing parties or a sum of unilaterally declared quotas) and ICES estimates of total catch for Norwegian spring-spawning herring, Western horse mackerel, Northeast Atlantic mackerel and blue whiting. The overviews are based on the history of advice, management and catch as reported in the ICES single-stock advice documents. The information is summarized in Tables 1.7.3.1-4 and figure 1.7.3.1. Figures 1.7.3.2-5 compare the TAC and advice, catch and advice and catch and TAC and catch and the sum of unilateral quotas respectively, each expressed as a percentage difference e.g. (TAC-advice)/advice.

For Norwegian spring-spawning herring some deviations between TAC and advice occurred between 2010-2013, but from 2014 on the sum of unilateral quotas has been in excess of the scientific catch advice which was based on the agreed management plan. Catches have likewise been in excess of the scientific advice and close to the sum of unilateral quotas.

Western horse mackerel: some deviations between TAC and advice have been occurring during the time-series presented, but there does not appear to be a clear trend. No management plan is applicable for western horse mackerel. Catches have generally been at or below the agreed TAC.

Northeast Atlantic mackerel has not had agreed TACs during the period presented. The sum of unilateral quota has always been higher than the scientific advice. Catches have on average been $44 \%$ above the scientific advice since 2010. In 2021 and 2022, catches were $10 \%$ below the sum of declared unilateral quotas.

Blue whiting: up to 2013, the agreed management plan has been followed. From 2014 onwards, the sum of unilateral quota has been in excess of the scientific advice and the agreed management plan. Catches have likewise been in excess of the scientific advice and close to the sum of unilateral quota.

In summary: although long-term management plans exist for Norwegian spring-spawning herring, Northeast Atlantic mackerel and Blue whiting, they have not been instrumental in limiting the TACs to the pre-agreed values. While the Coastal States may have agreed on the TACs for these stocks, there was no agreement on the distribution of quota between Coastal States. As a consequence, the sum of unilateral quota and the catches have been in excess of both the scientific advice and the rules of the management plans.

Table 1.7.3.1. Overview of scientific advice, agreed TAC, sum of unilateral quotas and catch for Norwegian Spring-spawning Herring.

| Year | Advice Basis | Advice (t) | TAC (t) | Unilateral Quota ( t ) | Catch (t) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | Do not exceed the harvest control rule | 1,483,000 | 1,483,000 |  | 1,457,000 |
| 2011 | See scenarios | 1,170,000 | 988,000 |  | 993,000 |
| 2012 | Follow the management plan | 833,000 | 833,000 |  | 826,000 |
| 2013 | Follow the management plan | 619,000 | 619,000 | 692,000 | 685,000 |
| 2014 | Follow the management plan | 418,000 | 418,487 | 436,000 | 461,000 |
| 2015 | Follow the management plan | 283,000 |  | 328,000 | 329,000 |
| 2016 | Follow the management plan | 317,000 |  | 377,000 | 383,174 |
| 2017 | Follow the management plan | 646,075 |  | 805,142 | 721,566 |
| 2018 | Follow the management plan | 384,197 |  | 546,448 | 592,899 |
| 2019 | Follow the management plan, $\mathrm{Fmgt}=0.14$ and Bmgt $=3.184 \mathrm{Mt}$ ) | 588,562 | 588,562 | 773,750 | 777,165 |
| 2020 | Follow the management plan, Fmgt $=0.14$ and Bmgt $=3.184 \mathrm{Mt}$ ) | 525,594 | 525,594 | 693,915 | 720,937 |
| 2021 | Follow the management plan, Fmgt $=0.14$ and Bmgt = 3.184 Mt) | 651,033 | 651,033 | 881,097 | 851,813 |
| 2022 | Follow the management plan, $\mathrm{Fmgt}=0.14$ and Bmgt $=3.184 \mathrm{Mt}$ ) | 598,588 | 598,588 | 827,963 | 813,834 |
| 2023 | Follow the management plan, $\mathrm{Fmgt}=0.14$ and $\mathrm{Bmgt}=3.184 \mathrm{Mt}$ ) | 511,171 | 511,171 | 692,942 |  |
| 2024 | Follow management strategy ( $\mathrm{Fmgt}=0.123$ $\mathrm{Bmgt}=3.184 \mathrm{Mt}$ ) | 390,010 |  |  |  |

Table 1.7.3.2. Overview of scientific advice, agreed TAC, sum of unilateral quotas and catch for Western Horse Mackerel.

| Year | Advice Basis | Advice (t) | TAC (t) | Unilateral <br> Quota (t) | Catch (t) |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 2010 | Follow proposed management plan | 180,000 | 185,000 | 203,112 |  |
| 2011 | See scenarios | 229,000 | 184,000 | 193,698 |  |
| 2012 | MSY framework | 211,000 | 183,000 | 169,858 |  |
| 2013 | MSY framework | 126,000 | 183,000 | 165,258 |  |
| 2014 | MSY approach | 110,546 | 135,000 | 136,360 |  |
| 2015 | MSY approach | 99,304 | 99,300 | 98,419 |  |
| 2016 | MSY approach | 126,000 | 126,000 | 98,811 |  |


| Year | Advice Basis | Advice (t) | TAC (t) | Unilateral <br> Quota (t) |
| :--- | :--- | :--- | :--- | :--- |
| 2017 | MSY approach | 69,186 | 95,500 | Catch (t) |

Table 1.7.3.3. Overview of scientific advice, agreed TAC, sum of unilateral quotas and catch for Northeast Atlantic Mackerel.

| Year | Advice Basis | Advice (t) | TAC ( t ) | Unilateral Quota (t) | Catch (t) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2010 | harvest control rule | 572,000 | 691,305 |  | 875,515 |
| 2011 | See scenarios | 672,000 | 929,943 |  | 946,661 |
| 2012 | Follow the management plan | 639,000 | 938,410 |  | 892,353 |
| 2013 | Follow the management plan | 542,000 | 857,319 |  | 931,732 |
| 2014 | Follow the management plan | 1,011,000 |  | 1,400,981 | 1,393,000 |
| 2015 | Follow the management plan | 906,000 | 1,054,000 | 1,208,719 | 1,208,990 |
| 2016 | MSY approach | 773,840 | 895,900 | 1,047,432 | 1,094,066 |
| 2017 | MSY approach | 857,000 | 1,020,996 | 1,191,970 | 1,155,944 |
| 2018 | MSY approach | 550,948 | 816,797 | 999,929 | 1,026,437 |
| 2019 | MSY approach | 770,358 | 653,438 | 864,000 | 840,021 |
| 2020 | MSY approach | 922,064 | 922,064 | 1,090,879 | 1,039,513 |
| 2021 | MSY approach | 852,284 | 852,284 | 1,199,103 | 1,081,540 |
| 2022 | MSY approach | 794,920 | 794,920 | 1,188,227 | 1,046,720 |
| 2023 | MSY approach | 782,066 | 782,066 | 1,188,265 |  |
| 2024 | MSY approach | 739,386 |  |  |  |

Table 1.7.3.4. Overview of scientific advice, agreed TAC, sum of unilateral quotas and catch for Blue Whiting.

| Year | Advice Basis | Advice (t) | TAC (t) | Unilateral <br> Quota (t) | Catch (t) |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 2010 | Follow the agreed management <br> plan | 540,000 | 548,000 | 540,000 |  |
| 2011 | See scenarios | 40,000 | 40,100 | 105,000 |  |
| 2012 | Follow the agreed management <br> plan | 391,000 | 391,000 | 1, | 1, |



Figure 1.7.3.1: Overview of scientific advice, agreed TAC (or sum of unilateral quota) and catch.


Figure 1.7.3.2: Relative deviations of TAC over advice. Red line indicates average relative deviation over the time-series shown.


Figure 1.7.3.3: Overview of catch over advice

## Catch over TAC



Figure 1.7.3.4: Overview of catch over TAC

Catch over Summed unilateral Quota


Figure 1.7.3.5: Overview of catch over sum of unilateral quotas.

### 1.8 General stock trends for widely distributed and migratory pelagic fish

WGWIDE 2023 has carried out the stock assessments of the following widely distributed and migratory pelagic species: boarfish, red gurnard, Norwegian spring-spawning herring, Western horse mackerel, North Sea horse mackerel, Northeast Atlantic mackerel, Striped red mullet and Blue whiting.

Analytical (category 1) assessments are available for the four species that make up the bulk of the biomass of pelagic species in the Northeast Atlantic:

- Northeast Atlantic mackerel
- Norwegian spring-spawning herring
- Blue whiting
- Western horse mackerel

The time-series of the combined catch of these four stocks since 1988 is shown in figure 1.8.1. The highest combined catch (approx. 4 million tonnes) for these four species was been taken in 2004 and 2005. In the most recent 6 years the total catch has been composed of $\sim 45 \%$ blue whiting, $\sim 33 \%$ mackerel, $\sim 18 \%$ herring and $\sim 3 \%$ horse mackerel.


Figure 1.8.1: Catch of blue whiting, mackerel, western horse mackerel and Norwegian spring-spawning herring.
An overview of the key variables for each of the stocks (SSB, fishing mortality and recruitment), is shown in Figure 1.8.2. Stock sizes of herring, mackerel and blue whiting have been declining from historical highs in the recent years, but remain above their respective MSY B trigger reference point values with the exception of Western Horse Mackerel which has been increasing from a
historic low in 2017 but is considered to be below Blim. The Norwegian Spring-spawning Herring SSB is forecast to fall below MSY B trigger in 2023. The Blue Whiting SSB has increased in the most recent years following strong recent recruitment.

Fishing mortality for herring, horse mackerel and mackerel has been around Fmsy in the most recent period. Fishing mortality for blue whiting has been above Fmsy for much of the time-series.

Recruitment estimates for blue whiting and herring are on a comparable scale (billions) and are substantially higher and more variable than those for horse mackerel (with the exception of the 1982 year class) and mackerel.





Figure 1.8.2: top - SSB (million tons), middle - fishing mortality and bottom - recruitment (billions) of Norwegian springspawning herring, western horse mackerel, Northeast Atlantic mackerel and blue whiting from the WGWIDE 2023 update assessments.

An overview of stock weight-at-age for mackerel and blue whiting is shown in figures 1.8.3 and 1.8.4.

For mackerel, a decline in weight at age started around 2005 for most ages. In more recent years, this has ceased with increases for younger fish noted since 2012.

Weight-at-age of blue whiting shows substantial fluctuations over time. For most ages, a decline in weight at age has been observed from 2010 although this appears to have ceased and, for some ages reversed in the most recent years.


Figure 1.8.3: Stock weight-at-age of NEA mackerel.


Figure 1.8.4: Stock weight at age of blue whiting.
WGWIDE (and its precursors WGMHSA and WGNPBW) have been publishing catch per statistical rectangle plots in their reports for many years. Catch by rectangle has been compiled by WG members and generally provide an estimate of total catch per rectangle (although catch by rectangle data do not represent the official catches and cannot be used for management purposes). In general, the total annual catches by rectangle are within $10 \%$ from the official catches. In the individual stock report sections, the catch by rectangle is been presented by quarter for the most recent year. For this overview, WGWIDE has collated all the catch by rectangle data that is available for herring, blue whiting, mackerel and horse mackerel. For horse mackerel and mackerel, a long time-series is available, starting in 2001 (horse mackerel) and 1998 (mackerel). The time-series for herring and blue whiting are shorter (from 2011) although additional information could still be derived from earlier WG reports.


Figure 1.8.5: Catch of mackerel (tonnes) by year and rectangle. Catch by rectangle data do not represent the official catches and cannot be used for management purposes. In general, the total annual catches by rectangle are within 10\% from the official catches.


Figure 1.8.6: Catch of horse mackerel (all stocks, tonnes) by year and rectangle. Catch by rectangle data do not represent the official catches and cannot be used for management purposes. In general, the total annual catches by rectangle are within $10 \%$ from the official catches.


Figure 1.8.7: Catch of blue whiting (tonnes) by year and rectangle. Catch by rectangle data do not represent the official catches and cannot be used for management purposes. In general, the total annual catches by rectangle are within $\mathbf{1 0 \%}$ from the official catches.


Figure 1.8.8: Catch of Norwegian spring-spawning (Atlanto-scandian) herring (tonnes) by year and rectangle. Catch by rectangle data do not represent the official catches and cannot be used for management purposes. In general, the total annual catches by rectangle are within $10 \%$ from the official catches.

### 1.9 Ecosystem considerations for widely distributed and migratory pelagic fish species

A number of studies demonstrate that environmental conditions (physical, chemical and biological) can significantly influence stock productivity by changing the level of recruitment, growth rates, survival rates, or inducing variations in their geographical distribution (e.g. Skjoldal et al., 2004, Sherman and Skjoldal 2002). It has been acknowledged that future development in stock assessment methods should take ecosystem considerations into account to reduce assessment uncertainty. Therefore, WGWIDE encourages further work to be carried out on ecosystem considerations linked to widely distributed fish stocks including NEA mackerel, Norwegian springspawning herring, blue whiting, and horse mackerel. A close collaboration with the Working Group on Integrated Assessment of Norwegian Sea (WGINOR; ICES 2018b; 2023c), and other relevant Integrated Assessment groups within ICES could help in operationalizing the ecosystem approach for widely distributed pelagic stocks assessed by WGWIDE. The text and figures below include summary of the Norwegian Sea ecosystem status on climate variability, circulation pattern, recent trends in oceanography, phytoplankton production, zooplankton biomass, marine mammals, pelagic fish biomass and pelagic fish spatial distribution in the Norwegian Sea. It was prepared by WGINOR (ICES, 2023c).

## Highlights

- Extent of Arctic Water has increased from 2017. This has been accompanied by a freshening and cooling, although heat content in the Norwegian Sea remains above the longterm average.
- Annual primary production has remained stable over the period 2003 to present. The seasonal timing of the peak of production has gradually shifted to a later date.
- Zooplankton spring biomass, measured since 1995, declined in the mid-2000s and has since remained relatively stable.
- Spawning biomass of Norwegian spring-spawning herring and mackerel declined slightly in 2022. After a few years of decline, biomass of blue whiting increased, driven by historically high recruitment.
- Long-term decrease in breeding numbers for Atlantic puffin and black-legged kittiwake continues. Common guillemot remains at high risk of extinction despite increasing abundance at some monitoring colonies over the last decade.
- Seals pup production is declining or at low level. Baleen whales' distribution has gradually moved from the Norwegian Sea towards the Barents Sea. Bycatch of harbour porpoise remains at unsustainable level.

Graphical summary

|  | Topic | Overall trend | Situation in 2022 | Certainty | Possible implications |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Ocean climate | Generally, warm and saline conditions prevailed from the early 2000s until 2016. Since 2012, temperature of the Atlantic inflow has been close to the long-term mean while salinity has been below the longterm mean since 2016. The extent of Arctic Water has increased from 2017. | The temperature of the Atlantic inflow is close to the longterm mean while salinity is below the long-term mean. The extent of Arctic Water continues to increase and relative heat content continues declining. | Highly certain: dedicated monitoring with good spatial coverage exists. | The recent increase of Arctic Water may lead to increased new production due to relative high winter nutrient concentration. |
|  | Primary production | There is no trend in the level of spring and summer primary production in the Norwegian Sea deep basins since $2003^{1}$. The timing of peak production has gradually shifted to a later date over the last two decades. | The primary production for 2021 (latest year of data available) is comparable to the mean over the whole period. The timing of the peak production is 2 weeks later than in 2003. | Certain: Phytoplankton estimates are based on satellite data covering the whole productive season with high geographic resolution. The production model is not calibrated for high latitudes and absolute estimates of primary production are uncertain. | Change in timing can lead to seasonal match/mismatch with reproduction and feeding of zooplankton. |
| $\overbrace{\text { KG }}^{\substack{0}}$ | Zooplankton biomass | The spring biomass of mesozooplankton was at a higher level from 1995 to mid-2000s and has been at a lower level in the following years. Summer biomass shows an increasing trend during the first ten years then a decline for the last two years. | Biomass in 2022 was at similar level as the previous year for all subareas and both seasons. | Moderately certain: plankton is patchily distributed, which leads to uncertain estimates. The timing of seasonal development relative to time of sampling can affect the level of biomass measured. | Reduced zooplankton biomass may have caused reduced food resources for planktivorous feeders, including pelagic fish. in the recent decade. |


|  | Topic | Overall trend | Situation in 2022 | Certainty | Possible implications |
| :---: | :---: | :---: | :---: | :---: | :---: |
| P | Zooplankton spatial distribution | The spring distribution of zooplankton has changed from higher biomasses in Arctic water in the west to become evenly distributed in the Norwegian Sea. | In 2022, the zooplankton was relatively evenly distributed in spring and summer, but with some confined high-concentration areas in southeast in spring. | Moderately certain: The spatial distribution reflects and is affected by the timing of the survey and the timing of the zooplankton seasonal development. | Changes in the spatial distribution of plankton can affect the spatial distribution of planktivorous fish |
|  | Pelagic fish biomass | Spawning biomass of Norwegian spring-spawning herring and mackerel remains close to the long-term average whereas blue whiting biomass is above. | Herring spawning-stock biomass decreased by $2 \%$ and mackerel by $3 \%$ whereas blue whiting increased by $11 \%$ compared to previous year. Estimated recruitment of blue whiting is at a historical high. Fishing remains above scientific advice for all stocks. | Highly certain for herring and blue whiting, moderately certain for mackerel: estimates are based on quantitative stock assessments. | Changes in pelagic fish biomass have direct implications for fisheries opportunities. |
|  | Pelagic fish spatial distribution | Since the mid-2000's, mackerel distribution expanded westward into Icelandic and Greenlandic waters, then retracted eastward from 2015. By 2020, most of the mackerel stock was feeding in the Norwegian Sea. In 2022, mackerel expanded westward again. | No mackerel in Greenlandic waters but increased presence and density in Icelandic Waters in 2022 compared to last two years. | Highly certain: based on ecosystem surveys in the Nordic Seas in spring (May) and summer (July) | Changes in pelagic fish spatial distribution have direct implications for fisheries opportunities. |
|  | Seabirds | Substantial long-term declines for most species, including common guillemot, Atlantic puffin, and black-legged kittiwake. | No clear signs of improvements, except common guillemot abundance appears stable in colonies which provide shelter from eagle predation. | Highly certain: Trends are derived from dedicated monitoring. | Many bird colonies are at risk of extinction, and some have already disappeared. |
|  | Marine mammals | Decline or sustained low levels of pup production in several seal species. Long-term shift in summer distribution of baleen whales from the Norwegian Sea to the Barents Sea. Unsustainable levels of harbour porpoise bycatch. | No new data on abundance and distribution. | Highly certain: Trends in pup production are based on dedicated surveys. <br> Moderately certain: Data are scarce on bycatch and produc-tivity-connectivity for harbour porpoises | Changes in marine mammals affect foodweb structure and long-term viability of marine mammal populations |



Figure 1.9.1. A subset of climate indicators for the Norwegian Sea: a) North Atlantic Oscillation Index (NAO), b) Subpolar Gyre index (SPG, note that strong gyre is represented by negative values and weak gyre by positive values), c) Norwegian Sea Gyre index, d) Atlantic Water Temperature at Svinøy section and East Icelandic Current Temperature, e) Atlantic Water Salinity at Svinøy section, f) Arctic Water amount in the Norwegian Sea, g) relative heat content (RHC) and h) Relative Freshwater Content (RFC).

## Pelagic Fish

## Current status

Three fish stocks dominate the pelagic ecosystem of the Norwegian Sea: Norwegian springspawning herring (NSSH, Clupea harengus), Northeast Atlantic (NEA) mackerel (Scomber scombrus), and blue whiting (Micromesistius poutassou). In 2022, estimated spawning-stock biomass (SSB) for all three stocks ranged from 3.7 to 5.0 million tonnes. Combined SSB for all three stocks was 12.5 million tonnes ${ }^{7}$ (Figure 4a).

Combined catch of the three stocks was 3.1 million tonnes in 2021, of which approximately 1.1 million tonnes was blue whiting, 1.1 million tonnes was mackerel, and 0.9 million tonnes was herring. Current exploitation levels, relative to biological reference points, show that fishing pressure on all three stocks is above that which leads to maximum sustainable yield ${ }^{7}$. Furthermore, herring and blue whiting exploitation is above management plan fishing targets and the $95 \%$ confidence interval for mackerel fishing mortality is higher than maximum sustainable yield targets ${ }^{7}$. Stock status, for all three stocks is above all biological reference points related to the risk of impaired reproductive capacity. However, herring SSB is very close to biological reference limits, as the $95 \%$ SSB confidence limits include the reference limits? ${ }^{7}$.

## Recent changes

The 2022 stock assessment results show an estimated $2 \%$ decrease in herring SSB in 2022 compared to 2021, after a brief increase in 2021, following a decade of continuous decline with an overall estimated decline of $52 \%^{7}$. After a slight increase in 2021, mackerel SSB is estimated to decline by $3 \%$ in 2022 and has declined $37 \%$ from peak stock size in 2014-2015 ${ }^{7}$. Blue whiting SSB increased an estimated $11 \%$ in 2022 compared to previous year and was estimated to be $20 \%$ lower than at the last peak size in $2017^{7}$.

Mackerel distribution in Nordic Seas in summer 2022 expanded further westward in 2022 compared to the last two years. The western boundary of the distribution was west of Iceland (longitude $\left.30^{\circ} \mathrm{W}\right)^{8}$. Distribution of blue whiting in 2022 was similar to the most recent years ${ }^{9}$. The distribution area of herring in May was comparable with the most recent period and the large 2016 year class is distributed throughout the geographical distribution range of the mature herring stock ${ }^{10}$. In July, the herring had shifted farther east and north ${ }^{8}$.

## Possible reasons for recent changes

Herring SSB is dominated by recruitment of large year classes at irregular intervals with many years of small year classes in between (Figure 4b). After the large 2002- and 2004-year classes, recruitment was below average until the 2016-year class. The 2016 year class appears to be fully recruited to the spawning stock and contributes approximately half of the SSB ${ }^{7}$. Fishing above the advised level has accelerated the stock decline during a period of low recruitment. Since 2013, unilaterally determined quotas have led to annual commercial catch being $31 \%$ higher than the advised total allowable catch (TAC) on average ${ }^{7}$.


Figure 4. a) estimated spawning-stock biomass (lines) including 95\% confidence intervals (shaded areas) for Norwegian spring-spawning herring (red filled circles), mackerel (purple filled triangles) and blue whiting (blue filled rectangles) from first stock assessment year, ranges from 1980 to 1988, to $2022^{7}$. b) estimated year-class size at recruitment for Norwegian spring-spawning herring (age 2; red filled circle), mackerel (age 2; purple filled triangles) and blue whiting (age 1; blue filled triangle) from first year of assessment, ranges from 1980 to 1988, to 2022 ${ }^{\text {. }}$. Note the different scale for mackerel recruitment.

The 2022 assessment changed the perception of the mackerel stock. This was due to changes in relative weights of data sources in the assessment model. Prior to 2017 SSB was revised upward and fishing mortality downward. After 2017, SSB was revised downward and fishing mortality upward. This resulted in current fishing pressure being above Fms ${ }^{7}$. Estimates of ages 0 and 1 are considered highly uncertain and year-class strength appears to be established at age 2-3 .

After mackerel abundance peaked in 2014-2015, the annual commercial catches have on average been $35 \%$ higher than the scientific advice ${ }^{7}$. Fishing above advised TAC is believed to have contributed to the observed decline in spawning stock size.

Blue whiting's sharp decline in SSB from 2018-2020 was caused by excessive fishing. Catches exceeded the advised TAC by $28 \%$ since 2018 along with low recruitment in 2017-2019. However, improved recruitment in recent years (2020-2022) explains increased SSB in 2022. Recruitment (age 1) in 2021 (i.e. the 2020 year-class) is estimated to be at a historical high and short-term forecasts project a high SSB in 2023 ( 6.7 million tonnes) ${ }^{7}$.

The blue whiting fishery mostly targets ages 3-5 years. Hence the stock can decline sharply when several years of poor recruitment coincide with excessive fishing. The stock can also recover quickly when recruitment is high, as seen in the stock fluctuations during the early 2000's and late 2010's.

The reasons why mackerel have retracted from the western area from 2015 onwards remain poorly understood. Since 2015, estimated mackerel stock size has declined by approximately a third, zooplankton abundance has remained within the range observed when mackerel were present, and the western area remains warm enough for mackerel $\left(>8-9{ }^{\circ} \mathrm{C}\right)^{8}$.

### 1.10 Future Research and Development Priorities

As part of the planning towards future benchmark assessments, the working group maintains, for each stock, a list of research and development priorities on topics including proposed research projects, improved sampling and data collection and development of stock assessment techniques. In addition to these individual stock issues, increased consideration should be given to integrated ecosystem assessments for the stocks within WGWIDE. Several WGWIDE members are also participants in the work of the Working Group on Integrated Assessment for Norwegian Sea (WGINOR). Improving linkages with other regional Integrated Ecosystem Assessment groups within ICES would be beneficial and should be considered in future.

### 1.10.1 NEA Mackerel

In 2019, the ICES Workshop on a Research Roadmap for Mackerel (WKRRMAC, (ICES, 2019b)) met to discuss the research needs for the provision of advice for the management of NEA Mackerel. The workshop involved a diverse range of stakeholders including industry representatives, managers and scientists and identified a number of priorities (see report of WGWIDE 2019 (ICES, 2019c) for details).

WGWIDE 2022 recommended the establishment of WKEVALMAC (A Workshop on the Evaluation of NEA Mackerel stock components and regional management measures) to review available information from appropriate methods to infer the stock structure of NEA Mackerel. This workshop took place in 2023 (ICES 2023 WKEVALMAC in prep). See section 8.11 for a summary of the conclusions of this workshop.

### 1.10.2 Blue Whiting

Numerous scientific studies have suggested that blue whiting in the North Atlantic consists of multiple stock units. The ICES Stock Identification Methods Working Group (SIMWG) reviewed this evidence in 2014 (ICES, 2014) and concluded that the perception of blue whiting in the NE Atlantic as a single-stock unit is not supported by the best available science. SIMWG further recommended that blue whiting be considered as two units. There is currently no information available that can be used as the basis for generating advice on the status of the individual stocks.

However, currently a project is going on with more data being collected (NEA and Mediterranean) to allow clarify the stock definition for this species. In future, this newly collected information on stock composition should be evaluated on the behalf of a benchmark of this stock.

### 1.10.3 NSS Herring

The Norwegian spawning ground survey was reintroduced in 2015 as part of the tuning series (fleet 1). However, changes were made to the survey compared to the older part of the series. At the 2016 assessment benchmark, the inclusion of the surveys from 2015 was accepted as an extension to the tuning series. It is now considered appropriate to investigate the splitting of this survey series since the time-series is now long enough to do this exercise. An interbenchmark exercise to explore this was proposed during WGWIDE 2020, but it was later decided to postpone such exploration for the next benchmark. Some exploratory work was presented in WGWIDE 2021.

- Consider the inclusion of a new tuning series (IESSNS) in the assessment.
- Consider the inclusion of a new tuning series (tagging data based on RFID) in the assessment.
- Consider the inclusion of a new Norwegian recruitment index into the assessment, as the current recruitment index has not been updated since 2021.
- Request and incorporate within the assessment information on the uncertainty in catches from all countries submitting catch data (currently only available from Norway).
The maturity ogive for NSSH is back-calculated but with a delay of 6 years, i.e. the 5 last years use one of two fixed maturity ogives scales (one for small cohort and the other for large cohort). The benchmark report has no objective criteria when to recognize a cohort as strong, and the current model is not optimal for medium-sized cohorts. This may result in deviation in SSB in intermediate year.

There is clear indication of a density-dependent effect on maturity-at-age. A more proper estimate of the maturity for the last 5 years (and for the forecast) should be made using the estimated cohort strength directly, and this should be evaluated through a peer-review process.

The XSAM software is used for the assessment. However, the SAM software now supports the XSAM model as an optional model. A switch from the currently used code to the SAM platform should be done in order to make the model more publicly available and to ensure further development of the infrastructure. The possibility to use the predicting the observation variance in SAM can then be used instead of including external variance from surveys. WGWIDE 2023 recommended to use SAM software from WGWIDE 2024 and onwards as the XSAM software is no longer maintained. A comparison between the two software has been made (WD09), resulting a negligible difference between the two software.

### 1.10.4 Western Horse Mackerel

Considering the potential of mixing between Western and North Sea horse mackerel occurring in division 7d and 7e, improved insight into the origin of catches from that area will be a major benefit for improvement of the quality of future scientific advice and thus management of the North Sea and Western horse mackerel stocks. A project addressing stock structure and boundaries of horse mackerel was initiated by the Northern Pelagic Working Group in collaboration with University College Dublin and Wageningen Marine Research. In 2018, the results of the genetic analysis have been published (Farrell et al 2018) which concluded that the spawners of North Sea and Western horse mackerel can be genetically identified as two distinct stocks. However, at that stage it was not yet possible to separate the two stocks when they occur in mixed
samples. Subsequently, a full genome sequencing on horse mackerel has been carried out (Fuentes-Pardo et al 2020), which confirmed the earlier results on separating western, North Sea and southern horse mackerel (see also text below on North Sea horse mackerel). In addition, this study concluded that it would also be possible to distinguish horse mackerel from different spawning populations in mixed samples.
The 2020 study also concluded that further analysis on the mixing between the Western stock and the Southern stock in area 8c should be carried out: the fishery in the area targets mainly juveniles, would be therefore be very important to understand the impact of this fishery on each of the two stocks.

The most recent results indicate that a further large-scale analysis of samples, with a greater temporal and spatial coverage, with the newly identified molecular markers was required to test and reassess the current stock delineations. Results were presented at WGWIDE 2023 and concluded that the stocks boundaries might need revision (see also section 1.4.8.3 for details).

### 1.10.5 North Sea horse mackerel

Studies on stock identity and the degree of connection and migrations between the North Sea and the Western Stock are considered particularly relevant. On behalf of the Pelagic Advisory Council and the EAPO Northern Pelagic Working Group, a research project on genetic composition of horse mackerel stocks was initiated. Genetic samples have been taken over the whole distribution area of horse mackerel during the years 2015-2017. The full genome of horse mackerel was sequenced and results indicated that the western horse mackerel stock is clearly genetically different from the North Sea stock (Farrell and Carlsson, 2019; Fuentes-Pardo et al., 2020). Markers were identified that are be able to reveal the stock identity of individual horse mackerel caught in potential mixing areas. Horse mackerel samples from division 7d and 7e have been collected by the PFA on board of commercial vessels in autumn 2020, while horse mackerel from division 4a have been collected during the NS-IBTS in Q3. Initial results of this comprehensive genetic research project on stock identification of horse mackerel suggest that the boundaries of the stocks might require revision. This relates to the Northern boundary in 4a to the Southern border in 7d. Additionally, the Institute of Marine Research in Norway sampled horse mackerel in coastal waters within 4a during all quarters in 2019. Preliminary results presented at WGWIDE 2021 showed that the genetic profile of individuals caught in all quarters matched well with the genetic profile of the Western HOM stock, with just one or two individuals matching better with North Sea HOM profile (Florian Berg, pers. comm.). Ongoing and new studies are being undertaken to investigate these issues. Potential changes in the perception of the stock distributions could impact the reliability of the assessments for the three current stocks of horse mackerel in the Northeast Atlantic (see also section 1.4.8.3 for details).

Efforts are required to upload historic age and length data to the InterCatch database. The current stock assessment method is based on length data and, with only data from 2016 onwards currently available in InterCatch, it is impossible to compare the F/FMSY proxy and the lengthbased indicators that the proxy is based on with information from earlier years. Furthermore, length data are only submitted by accessions to stock coordinators directly, and not through InterCatch. This makes the process of combining the data from different countries prone to error and lack transparency. Since 2020, national data submitters were requested to submit data both via the accessions as well as through InterCatch. A comparative analysis has to be carried out to evaluate the feasibility of using length data from InterCatch only in future. Moreover, it was discovered that several hundred Dutch age readings coming from foreign vessels (mainly UK) have not been uploaded to InterCatch in the past. Efforts will be made to ensure this historic information will be uploaded in order to increase (the currently low) confidence in the estimates
of catch-at-age. However, from 2021 onwards Dutch age samples were used in the raising procedure of UK and uploaded to InterCatch. The historic data will be dealt with in the upcoming benchmark.

### 1.10.6 Boarfish

From 2017, this stock has been included on the list of stocks sampled under the data collection framework (DCMAP). This permitted sampling of commercial catch for both length and age. However, age reading is difficult and expertise is limited. An increase in the number of age readers would help develop a time-series of commercial catch-at-age which would in turn allow the development of an age-based assessment methodology. The current ALK is static and is based on a limited number of age readings. Also investigating the viability of length based assessments should be explored.

Improvements in the survey data can be realized through a change in sampling protocol on groundfish surveys to ensure boarfish are measured to the 0.5 cm . The acoustic time-series should continue to be developed. The current survey does not contain the stock. The use of information from other acoustic surveys, for example, the Pélagiques GAScogne (PELGAS) survey should also be explored.

### 1.10.7 Striped red mullet

In the WGWIDE framework, the assessment of striped red mullet stock is currently absent. The limited and scarce availability of data within the ICES database, coupled with uncertainties regarding data quality, has precluded the establishment of any analytical assessment. The foremost development priorities encompass the enhancement of fishery-dependent information and the creation of biomass indices derived from EVHOE and SP-NSGFS. These indices will be meticulously tailored to the unique characteristics of the stock, which include its coastal distribution and potential variations in dynamics between the Bay of Biscay and the Celtic Sea. Ongoing research initiatives (MATO and ACOST, as detailed in the stock-related section of this report) are anticipated to furnish updated data and published findings pertaining to the sexual maturity of this species, along with reference fleet CPUE, by 2024. Armed with these inputs, an analytical assessment framework could be subjected to intersessional testing and subsequently presented to WGWIDE. This endeavor aims to bolster the potential for a benchmark request in the near future, pertaining to the assessment of this stock.

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# 2 Blue whiting in Northeast Atlantic and adjacent waters 

## Micromesistius poutassou in subareas 1-9, 12, and 14 (whb.27.1-91214)

Blue whiting (Micromesistius poutassou) is a small pelagic gadoid that is widely distributed in the eastern part of the North Atlantic. The highest concentrations are found along the edge of the continental shelf in areas west of the British Isles and on the Rockall Bank plateau, where it occurs in large schools at depths ranging between 300 and 600 metres and is also present in almost all other management areas between the Barents Sea and the Strait of Gibraltar and west to the Irminger Sea. Blue whiting reaches maturity at $2-7$ years of age. Adults undertake long annual migrations between the feeding and spawning grounds. Most of the spawning takes place between March and April, along the shelf edge and banks west of the British Isles. Juveniles are abundant in many areas, with the main nursery area believed to be the Norwegian Sea. See the Stock Annex for further details on stock biology.

### 2.1 ICES advice in 2022

Fishing mortality ( F ) is estimated to be above FMSY since 2014. Spawning-stock biomass (SSB) has been decreasing since 2018, but increased again in 2022. SSB is still estimated to remain above MSY $B_{\text {trigger. }}$ Recruitment (R) from 2017 to 2019 is estimated to be low, followed by a large increase in 2021 and another strong recruitment in 2022. ICES advises that when the long-term management strategy, agreed by the European Union, the Faroe Islands, Iceland, and Norway, is applied, catches in 2023 should be no more than 1359629 tonnes.

### 2.2 The fishery in 2022

As in previous years, the main fisheries on blue whiting were targeting spawning and postspawning fish (Figures 2.2.1-2). Most of the catches (89.2\%) were taken in the first two quarters of the year and the largest part of this was taken along the slopes of the Western European shelf and around the Faroes. Smaller quantities were taken in the Norwegian Sea, in the Norwegian Trench, in the Rockall Trough and along the coast of Spain and Portugal as well as in the Bay of Biscay.

The fishery in the second half of the year was mainly east of the Faroes and in the central Norwegian Sea, with smaller amounts in the Norwegian Trench and along the coast of Portugal and Spain.

The multinational fleet targeting blue whiting in 2022 consisted of several types of vessels from 16 countries. The bulk of the catch is caught by large pelagic trawlers, some with capacity to process or freeze on board. The remainder is caught by RSW vessels.

### 2.3 Input to the assessment

At the Inter-Benchmark Protocol on Blue Whiting, IBPBLW (ICES, 2016a), it was decided to use preliminary within year, quarter 1 and quarter 2, catch-at-age data in the assessment to get additional information to the within year IBWSS survey estimates. In recent years, $85-90 \%$ of the total annual catches of the age 3+ fish have been taken in the first half of the year, which makes
it reasonable to estimate the total annual catch-at-age from reported first semester (Q1 \& Q2) data and expected total catches for the remainder of the year. The catch data sections in this report contain a comprehensive description of the 2022 data as reported to ICES and a brief description of the 2023 preliminary catch data.

### 2.3.1 Officially reported catch data

Official catches in 2022 were estimated as 1038736 tonnes based on data provided by WGWIDE members (Table 2.3.1.1). Data provided as catch by rectangle represented $99.5 \%$ of the total WG catch in 2022.

In 2022, the majority of catches were caught on the spawning grounds with largest contribution from ICES divisions 27.5.b, 27.6.a, 27.7.c, 27.7.k, 27.2.a and 27.4.a (Figure 2.3.1.1; Tables 2.3.1.2, 2.3.1.3), caught respectively in quarter 1 and quarter 2 (Figure 2.3.1.3). In the first two quarters, catches are taken over a broad area, with the highest catches in 27.6.a, 27.5.b, 27.7.c and 27.7.k, while later in the year catches are mainly taken further north in division 27.2.a and in the North Sea (27.4.a) (Figures 2.3.1.6 and 2.3.1.7 and Table 2.3.1.3). The spatial and temporal distribution of catches in 2021 are similar to previous years (Figures 2.3.1.2, 2.3.1.3, 2.3.1.4; and Table 2.3.1.4 and Figure 1.10 .7 in Section 1). The majority of the blue whiting catch was caught by four nations - Faroe Islands, Norway, Iceland, and Russia, respectively (Figure 2.3.1.5).

Discards of blue whiting are small. Most of the blue whiting caught in directed fisheries are used for reduction to fish meal and fish oil. However, some discarding occurs in the fisheries for human consumption and as bycatch in fisheries targeting other species.

Reports on discarding from fisheries which catch blue whiting were available from the Netherlands for the years 2002-2007 and 2012-2014. A study carried out to examine discarding in the Dutch fleet found that blue whiting made a minor contribution to the total pelagic discards.
The blue whiting discards data provided by Portuguese vessels operating with bottom otter trawl within the Portuguese portions of ICES Division 27.9.a are available since 2004. The discards data are from two fisheries: the crustacean fishery and the demersal fishery. The blue whiting estimates of discards in the crustacean fishery for the period of 2004-2011 ranged between $23 \%$ and $40 \%$ (in weight). For the same period the frequency of occurrence in the demersal fishery was around zero for the most of the years, in the years where it was significant $(2004,2006,2010)$ discards ranged between $43 \%$ and $38 \%$ (in weight). In 2022, discards were $29.7 \%$ of the total catches for blue whiting along the Portuguese coast (Table 2.3.1.5). The total catch from Portugal is less than of one percent the total international catches.

Information on discards was available for Spanish fleets since 2006. Blue whiting is a bycatch in several bottom-trawl mixed fisheries. The estimates of discards in these mixed fisheries in 2006 ranged between $23 \%$ and $99 \%$ (in weight) as most of the catch is discarded and only the catch of the last day may be retained for marketing fresh. The catch rates of blue whiting in these fisheries are, however, low. In the directed fishery for blue whiting for human consumption with pair trawls, discards were estimated to be $7.8 \%$ (in weight) in 2022 (Table 2.3.1.5). Spanish catches are around $3 \%$ of the international catches.

In general, discards are assumed to be small in the blue whiting directed fishery. Discards data contributed to final catches of the following countries: Denmark, Ireland, Portugal, Spain, Sweeden, UK (England and Wales) and UK (Scotland). The total discards constituted 0.4\% of the total catches, 3641 tonnes. The largest fishing nations, , Faroe Islands, Norway, Iceland and Russia do not have discards on blue whiting.

The total estimated catches (tonnes) inside and outside the NEAFC regulatory area by country were reported on Table 2.3.1.6. The catches inside the NEAFC RA represent $8 \%$ of the total catches of blue whiting in 2022.

### 2.3.1.1 Sampling intensity

In 2022, $77 \%$ of catches were covered by the sampling program. In 2022, 1868 length samples and 1927 age samples were collected from the fisheries with 167650 fish measured and 15636 aged (Table 2.3.1.1.1 and Table 2.3.1.1.2). Sampling intensity for blue whiting with detailed information on catch, proportion of catch covered by the sampling program, the number of samples, number of fish measured, and number of fish aged per year from 2000 to 2022 is given in Table 2.3.1.1.1. Sampling intensity per country, quarter and ICES division for 2021 is listed in Tables 2.3.1.1.2, 2.3.1.1.3 and 2.3.1.1.4. The most intensive sampling, considering the age samples and the number of aged fish, took place in areas 27.2.a, 27.5.b, 27.6.b, 27.7.b, 27.7.c, 27.7.k, 27.8.c and 27.9.a (Figure 2.3.1.1.1). No sampling was carried out by Greenland, Poland and Sweden, which together represent $4 \%$ of the total catches. The sampled and estimated catch-at-age data are shown on Figure 2.3.1.1.1.

Sampling intensity for age and weight of blue whiting are made in proportion to landings according to CR 1639/2001 and apply to EU member states. The Fisheries Regulation 1639/2001, requires EU Member States to take a minimum of one sample for every 1000 tonnes landed in their country. Various national sampling programs are in force.

### 2.3.1.2 Age compositions

As an example of an age-length key from sampled catches in 2022, data from ICES area 27.6.a is presented by quarter and country (Figure 2.3.1.2.1). The mean length ( mm ) by age reveals that age classifications do present some differences between countries. A difference in mean length-at-age was observed in ages 0 and 1. Although, the differences in mean length-at-age increase in older ages, higher than age 9 .

The catch-at-age numbers reveal a higher proportion of age 2 individuals on quarters 1 and 2, which corresponds to the 2020 annual cohort, when a higher recruitment was observed (Figure 2.3.1.2.2).

The ICES InterCatch program was used to calculate the total international catch-at-age, and to document how it was done.

### 2.3.2 Preliminary 2022 catch data (Quarters 1 and 2)

The preliminary catches for 2023 as reported by the WGWIDE members are presented in Table 2.3.2.1.

The spatial distribution of these 2022 preliminary catches is similar to the distribution in 2022 with majority of catches taken in division 27.5.b, 27.6.a, 27.7.c and 27.7.k (Figure 2.3.2.1 and Table 2.3.2.2).

Sampling intensity for blue whiting from the preliminary catches by area with detailed information on the number of samples, number of fish measured, and number of fish aged is presented in Table 2.3.2.2.

WGWIDE estimated the expected total catch for 2022 from the sum of declared national quotas, corrected for expected national uptake and transfer of these quotas (Table 2.3.2.3).

For the period 2016 to 2022 , preliminary and final catch estimates are similar with maximum deviation in 2021 when the final catch was $8.3 \%$ higher than the preliminary catch (Table 2.3.2.4). Age compositions (Figure 2.3.2.2) are also similar between preliminary and final catch data with
the exception of an increase in age 1 in the final data from 2022 compared to the preliminary data. There is no clear pattern in the deviations; it is both the catch at age for young and older fish that change between preliminary and final data.
The estimation of catch at age and mean weight at age followed the method described in the Stock Annex.

### 2.3.3 Catch-at-age

The catch in numbers-at-age from 1981 to 2023 are presented in Table 2.3.3.1 and catch proportions at age shown in Figure 2.3.3.1. Strong year classes that dominated the catches can be clearly seen in the early 1980s, 1990, the late 1990s and early 2000's. More recently, the propagation of the large 2014 year class is also evident and the newly emerged strong year class of 2020. There is also an indication of the year class of 2021 in the catch data.

Catch curves for the international catch-at-age dataset (Figure 2.3.3.2) indicate a consistent decline in catch number by cohort in years with rather high landings (and probably similar high effort). The catch curves for year classes 2010-2015 show a consistent decline in the stock numbers with an estimated total mortality $(\mathrm{Z}=\mathrm{F}+\mathrm{M})$ around $0.6-0.7$ for the ages fully recruited to the fisheries. With an assumed natural mortality $(\mathrm{M}=0.2)$, the assessment F around $0.4-0.5$ fits well to the $Z$ values estimated from the catch curves.

### 2.3.4 Weight at age

Table 2.3.4.1 and Figure 2.3.4.1 show the mean weight-at-age for the total catch during 1981-2023 used in the stock assessment. Mean weight at ages 3-9 has generally decreased in the period 20102018, followed by an increase in the most recent years, for the most abundant ages in the catches. In 2021 and 2022, a decrease in mean weight in almost all ages was observed, which continued in 2023 for the younger ages (1-5 year) but increased again for the older ages ( $5-10+$ year).

The weight-at-age for the stock is assumed the same as the weight-at-age for the catch.

### 2.3.5 Maturity and natural mortality

Blue whiting natural mortality and proportion of maturation-at-age are shown in Table 2.3.5.1. See the Stock Annex for further details.

### 2.3.6 Fisheries independent data

Data from the International Blue Whiting spawning stock survey are used by the stock as-sessment model, while recruitment indices from several other surveys are used to qualita-tively adjust the most recent recruitment estimate by the assessment model and to guide the recruitments used in the forecast.

### 2.3.6.1 International Blue Whiting spawning stock survey

The Stock Annex gives an overview of the surveys available for the blue whiting. The International Blue Whiting Spawning Stock Survey (IBWSS) is the only survey used as input to the assessment model.

The full time series of IBWSS was recalculated in summer 2020, using the same software (StoX; Johnsen et al., 2019) and method as previously applied. The values are presented in Table 2.3.6.1.1 and Figure 2.3.6.1.1 A.

The survey time-series (2004-2023) show variable internal consistency ranging from 0.26 to 0.81 (Figure 2.3.6.1.1 B) The overall internal consistency for age-disaggregated year classes was slightly reduced compared to last year. There is a high internal consistency for the younger ages (1-5 years) and older ages (7-9 years) with correlation between 0.72 and 0.81 , but poor ( $0.2<\mathrm{r}<$ 0.3 ) between ages 5 to 7 . This may indicate age readings problems for this group of ages.

The distribution of acoustic backscattering densities for blue whiting for the period 2019-2023 is shown in Figure 2.3.6.1.2. The abundance estimate of blue whiting for IBWSS are presented in Table 2.3.6.1.1.

Length and age distributions for the period 2018 to 2023 are given in Figure 2.3.6.1.3.
Survey indices, (ages 1-8 years 2004-2023) as applied in the stock assessment are shown in Table 2.3.6.1.1.

### 2.3.6.2 Other surveys

The Stock Annex provides information and time-series from surveys covering parts of the stock area. A brief survey description and survey results are provided below.

The International ecosystem survey in the Nordic Seas (IESNS) in May which is aimed at observing the pelagic ecosystem with particular focus on Norwegian spring-spawning her-ring and blue whiting (mainly immature fish) in the Norwegian Sea (Table 2.3.6.2.1).

Norwegian bottom-trawl survey in the Barents Sea (BS-NoRu-Q1(Btr)) in February-March where blue whiting are regularly caught as a bycatch species. This survey gives the first reli-able indication of year class strength of blue whiting. The 1-group in this survey is defined as less than 19 cm (Table 2.3.6.2.2).

Icelandic bottom-trawl surveys on the shelf and slope area around Iceland. Blue whiting is caught as bycatch species and 1-group is defined as less than 22 cm in March (Table 2.3.6.2.3).

Faroese bottom-trawl survey on the Faroe plateau in spring where blue whiting is caught as bycatch species. The 1-group in this survey is defined as equal or less than 23 cm in March (Table 2.3.6.2.4).

The International Survey in Nordic Seas and adjacent waters in July-August (IESSNS). Blue whiting has been considered as a main target species in this survey since 2016 and as such methods were changed to ensure there was sampling for blue whiting. This was a recom-mendation from WGWIDE 2015 to try to have one more time-series for blue whiting. The estimate for 2023 was not completed before WGWIDE. Data for the survey are not used yet, due to the short time series.

### 2.4 Stock assessment

The IBWSS survey is the only survey used by the SAM assessment. The survey was cancelled in 2020 due to the COVID-19 pandemic, but conducted in 2021-2023.

The presented assessment in this report follows the recommendations from the Inter-Benchmark Protocol of Blue (ICES, 2016a) to use the SAM model. The configuration of the SAM model was kept unchanged in this year's assessment.

At WGWIDE in 2021 the time period for estimating recruitment for the short term forecast was changed from the full time series (minus terminal year) to the more recent period since 1996 (minus terminal year). This approach was again followed by WGWIDE 2023.

### 2.4.1 2023 stock assessment

For a model such as SAM, Berg and Nielsen (2016) pointed out that the so-called "One Step Ahead" (OSA) residuals should be used for diagnostic purposes. The OSA residuals (Figure 2.4.1.1) show a quite random distribution of residuals. There seems to be a year class effect showing in the catch residuals starting in the early 2000s. Similarly, there may be an indication of a "year effect" (too low index values) for the IBWSS 2015 observations which has also be seen in previous assessments. Moreover, a similar but weaker year effect of low index values is seen for the IBWSS 2023 observations.

The estimated parameters from the SAM model from this year's assessment and those from assessments conducted since 2019 are shown in Table 2.4.1.1. There are no abrupt changes in the estimated parameters over the time-series presented. The lowest observation noises, and therefore the largest weight in the assessment model, have in all years been from catches at ages 3-8, which constitute the largest proportion of the catch. The observation noise of age 1 (i.e. recruitment) in the IBWSS has decreased over the years, likely due to the increase in length of the time series. The observation noise of ages 4-6 and 7-8 is slightly higher compare to last year's assessment, indicating that the model slightly lowered the weight of these age groups in the survey in this year's assessment.

The process error residuals ("Joint sample residuals") (Figure 2.4.1.2) are reasonably well randomly distributed. Process noise within SAM is implemented as a "process mortality, Z"; these deviations in mortalities are shown in Figure 2.4.1.3. The deviations in mortality (plus or minus mortality) seems fairly randomly distributed without very pronounced clusters (as also seen in Figure 2.4.1.2).

The correlation matrix between ages for the catches and survey indices (Figure 2.4.1.4) shows a modest observation correlation for the younger ages and a stronger correlation for the older ages. This difference is more distinct for catches, probably because it includes older ages (1-10+) than the survey data (ages 1-8). The correlation of older ages (4-8 year) in the survey is stronger this year compared to last year's assessment (Figure 2.4.1.4). This may be a result of a potential year effect in this year's survey data for the older ages. An indication for this may be found in the low index values of ages 5-8 for 2023, being at their historical lowest or close to their historical lowest value (Table 2.3.6.1.1).

Figure 2.4.1.5 presents the exploitation pattern for the whole time-series. There are no abrupt changes in the exploitation pattern from 2010 to 2023, even though the landings in 2011 were just $19 \%$ of the landings in 2010, which might have given a change in exploitation pattern. The plateau in selection at age 6 and older seen since mid-2000s seems more realistic than the more linear selection estimated for the beginning of the time series. The estimated stable exploitation pattern might be influenced by the use of correlated random walks for $F$ at age with a high estimated correlation coefficient ( $\mathrm{Rho}=0.93$, Table 2.4.1.1).

The retrospective analysis (Figure 2.4.1.6) shows a reasonably stable assessment for the last 5 years, with the previous years within the $95 \%$ confidence interval for the current assessment. Mohn's rho by year and as the average value over the last five years are presented in (Table 2.4.1.2). The annual values are rather high (and negative) for recruitment such that the average Mohn's rho for recruitment becomes -0.188 . The average Mohn's rho for F and SSB indicates no bias. Yet, the most recent value for F is relatively high ( -0.188 ) compared to earlier years. This is because the removal of the first peel in the retrospective analysis leads to a quite lower $F$ value for 2022 (from 0.41 to 0.37 ).

Stock summary results with added $95 \%$ confidence limits (Figure 2.4.1.7 and Table 2.4.1.5) show a decrease in fishing mortality in the period 2004-2011, followed by a steep increase in F up to

2015 after which $F$ has decreased to around $0.40-0.45$ (above $\mathrm{F}_{\text {MSY }}$ at 0.32 ). In 2023, F is estimated to increase again to 0.515 . In Recruitment (age 1) was high in 2015 , followed by a series of lower recruitment in 2017-2020. The recruitment in both 2021 and 2022 is estimated to be historically high. After a period of descrease since 2016, SSB has been increasing again since 2022. A further increase in SSB is expected from 2023 to 2024, with both strong year classes being largely recruited to the fishery.

### 2.4.2 Alternative model runs

During WGWIDE 2023 an alternative configuraton of the Blue Whiting SAM assessment was prepared (available at stockassessment.org; BW2023_freeFRW). In this assessment, the number of parameters to estimate for the F-random walks was increased to allow estimation of these random walks for age 1-3 seperatly from ages $4-10$. At the same time, the imposed correlation structure in the catch-at-age data was turned off. The results are very similar to the final accepted model and show only minor differences when it comes to fitting catch-at-age. The overall negative log-likelihood was larger under the alternative configuration. The alternative configuration had no impact on the retrospective pattern either. Figure 2.4.2.1 shows the retrospective pattern as well as the final estimates of SSB for the alternative run.

### 2.5 Final assessment

Following the recommendations from Inter-Benchmark Protocol on Blue Whiting (ICES, 2016a) the SAM model is used for the final assessment. The model settings can be found in the Stock Annex.

Input data are catch numbers-at-age (Table 2.3.3.1), mean weight-at-age in the stock and in the catch (Table 2.3.4.1) and natural mortality and proportion mature in Table 2.3.5.1. Applied survey data are presented in Table 2.3.6.1.1.
The model was run for the period 1981-2023, with catch data up to 2022 and preliminary catch data for the first half-year (Q1 and Q2) of 2023 raised to expected annual catches, and survey data from March-April, 2004-2023. SSB 1 st January in 2023 is estimated from survivors and estimated recruits (for 2023 estimated outside the model, see short-term forecast section). $11 \%$ of age group 1 is assumed mature, thus recruitment influences the size of SSB. The key results are presented in Tables 2.4.1.3 and 2.4.1.4 and summarized in Table 2.4.1.5 and Figure 2.4.1.7. Residuals of the model fit are shown in Figures 2.4.1.1 and 2.4.1.2.

### 2.6 State of the Stock

Fishing pressure (2023) on the stock is above $\mathrm{F}_{\mathrm{mSy}}$ and between $\mathrm{F}_{\mathrm{pa}}$ and $\mathrm{F}_{\text {lim; }}$; spawning-stock size (2023) is above MSY $\mathrm{B}_{\text {trigger, }} \mathrm{B}_{\mathrm{pa}}$ and $\mathrm{B}_{\text {lim }}$.

After a small dip in F in 2019 (0.40), F is estimated to be 0.52 in 2023. F has been above Fmsy and Fpa 0.32 since 2014. SSB has increased from 2021 ( 4.0 million tonnes) to a close to historical high in 2023 ( 6.2 million tonnes). In 2024, SSB is estimated to reach 6.8 million tonnes under geometric recruitment (1996-2022) in 2024. SSB has been above MSY Btrigger since 1998.

Recruitment (age 1) in 2022 is estimated to be almost at the same level as the historical high of 2021. Recruitment in 2023 ( 16.7 billion) is estimated to be below the 1996-2022 geometric mean (23.0 billion).

### 2.7 Biological reference points

In spring of 2016, the Inter-Benchmark Protocol on Blue Whiting (IBPBLW) (ICES, 2016a) delegated the task of re-evaluating biological reference points of the stock to the ICES Workshop on Blue Whiting Long Term Management Strategy Evaluation (WKBWMSE) (ICES 2016b). During the WGWIDE meeting 2017, WKBWMSE concluded to keep Blim and $B_{p a}$ unchanged but revised $\mathrm{Flim}_{\mathrm{lim}} \mathrm{F}_{\mathrm{pa}}$, and $\mathrm{F}_{\mathrm{mS}}$.

ICES made in 2021 the decision to use $\mathrm{F}_{\mathrm{p} 05}$ as the value for $\mathrm{F}_{\mathrm{pa}}$. $\mathrm{F}_{\mathrm{p} 05}$ was estimated by WKBWMSE (ICES 2016b), where it was concluded that the EQSIM simulations showed that $\mathrm{Fp}_{0.05}(0.32)$ is less than the FMSY in the constant $F$ simulations, so FMSY was set to this lower value.

The table below summarises the currently used reference points.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | $2.25 \text { mil- }$ <br> lion $t$ | $\mathrm{B}_{\mathrm{pa}}$ | ICES (2013a, 2013b, 2016b) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.32 | Stochastic simulations with segmented regression stock-recruitment relationship | ICES (2016b) |
| Precautionary approach | $\mathrm{Bl}_{\text {lim }}$ | $1.50 \text { mil- }$ <br> lion $t$ | Approximately $\mathrm{B}_{\text {loss }}$ | ICES (2013a, 2013b, 2016b) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | $2.25 \text { mil- }$ <br> lion $t$ | $\mathrm{B}_{\text {lim }} \exp (1.645 \times \sigma)$, with $\sigma=0.246$ | ICES (2013a, 2013b, 2016b) |
|  | $\mathrm{F}_{\text {lim }}$ | 0.88 | Equilibrium scenarios with stochastic recruitment: F value corresponding to $50 \%$ probability of (SSB< $\mathrm{Blim}_{\mathrm{lim}}$ ) | ICES (2016b) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.32 | Fp05; the F that leads to SSB $\geq$ Blim with 95\% probability | ICES (2016b) and WGWIDE 2021 |

### 2.8 Short-term forecast

### 2.8.1 Recruitment estimates

The benchmark WKPELA in February 2012 concluded that the available survey indices should be used in a qualitative way to estimate recruitment, rather than using them in a strict quantitative model framework. The WGWIDE has followed this recommendation and investigated several survey time-series indices with the potential to give quantitative or semi-quantitative information of blue whiting recruitment. The investigated survey series were standardized by dividing with their mean and are shown in Figure 2.8.1.1.

The International Ecosystem Survey in the Nordic Seas (IESNS) only partially covers the known distribution of recruitment from this stock. The 1-group (2022 year class) is around the median and the 2 -group ( 2021 year class) index from the survey in 2023 was above the median of the historical range.

The 1-group (2022 year class) from The International Blue Whiting Spawning Stock Survey (IBWSS) was below the median in the time series and the 2-group (2021 year class) was the second highest in the time series (Table 2.3.6.1.1).

The Norwegian bottom-trawl survey in the Barents Sea (BS-NoRu-Q1(Btr)) in February-March 2023, showed that 1-group blue whiting was below the median in the time series (Ta-ble 2.3.6.2.2). This index should be used as a presence/absence index, in the way that when blue whiting is present in the Barents Sea, this is usually a sign of a strong year class, as all known strong year classes have been strong also in the Barents Sea.

The 1-group estimate in 2023 (2022 year class) from the Icelandic bottom-trawl survey showed a strong decrease compared to 2022 and was below the median in the time-series.

The 1-group estimate in 2023 (2022 year class) from the Faroese Plateau spring bottom-trawl survey showed a slight increase compared to 2022 but was below the median in the time-series. This is the only survey which doesn't pick up a strong signal from the 2020 and 2021 year classes.

In conclusion, the indices from available survey time-series indicate that the 2021 year class is among the strongest in the time series, which corresponds to the SAM assessment results. The 2022 year class estimated from surveys are much weaker and below the median, also seen in the SAM assessment. It was therefore decided not to change the SAM estimates of the 2021 and 2022 year classes for the purposes of the short term forecast.

No information is available for the 2023 and 2024 year classes and the geometric mean of the time-series from 1996-2022) was used for these year classes (23 billion at age 1 in 2023) (Table 2.8.2.1.2).

As described in the Stock Annex, WGWIDE decided in 2021 to change from using the geo-metric mean of the full time-series (since 1981) to use a shorter time-series (since 1996) for the calculations recruitment.

### 2.8.2 Short-term forecast

As decided at WGWIDE 2014, a deterministic version of the SAM forecast was applied. Details about specific implementation can be found in the Stock Annex.

### 2.8.2.1 Input

Table 2.8.2.1.1 lists the input data for the short-term predictions. Mean weight at age in the stock and mean weight in the catch are the same, and are calculated as three year averages (2021 2023) in accordance with the 2019 updated Stock Annex. Selection (exploitation pattern) is based on F in the most recent year. The proportion mature for this stock is assumed constant over the years and values are as used by the assessment.

Recruitment (age 1) in 2022 and 2023 are assumed as estimated by the SAM model, as additional survey information was not conflicting this result. Recruitment in 2024 and 2025 are assumed as the long-term average from the period with both high and low recruitments (geometric mean of the time-series since 1996, minus the terminal year, 1996-2022).

As the assessment uses preliminary catches for 2023 an estimate of stock size is available for the $1^{\text {st }}$ of January 2024. The normal use of an "intermediate year" calculation is not relevant in this case and F in the "intermediate year" (2023) is as calculated by the assessment model. Catches in 2023 are based on the preliminary catches based on declared national quotas and expected national uptake for 2023. Intermediate year assumptions are summarised in Table 2.8.2.1.2.

### 2.8.2.2 Output

A range of predicted catch and SSB options from the deterministic short-term forecast used for advice are presented in Table 2.8.2.2.1.

Following the ICES MSY framework for the target F from the LTMS implies fishing mortality to be at $\mathrm{F}_{\text {MSY }}=0.32$ which will give a TAC in 2024 at 1529754 tonnes. This corresponds to a $12.5 \%$ increase compared to the ICES advice last year, and a $8.5 \%$ decrease compared to the preliminary estimate of catches in 2023 (1 672378 tonnes).

SSB in 2025 is predicted to increase by $6.7 \%$ to 7258384 tonnes, if the advised catches are taken. The high recruitment estimated for 2021 and 2022 contributes to this increase in SSB.

### 2.8.2.3 Comparison of input with last year's forecast

Tables 2.8.2.3.1 and 2.8.2.3.2 compare the estimated catch numbers at age and catch times weight at age between WGWIDE 2023 and WGWIDE 2022 as they are estimated by both assessment for the last five year (2019-2023). For both the catch numbers and the catch numbers times weight, WGWIDE 2023 estimated lower values than WGWIDE 2022, except for the recruitment in 2022 and age 2 in 2023, which are from the same strong year class in 2021. Recruitment for 2023 (age 1) is estimated much lower in WGWIDE 2023 than WGWIDE 2022, as WGWIDE 2022 took the long-term geometric mean in the forecast (22.5 billion), while WGWIDE 2023 took the estimated value from SAM based on the in year survey and catch data ( 16.7 billion). A large difference is therefore to be expected for recruitment in the terminal year when catch and survey data point towards a much lower or higher recruitment than the geometric mean.

The mean weight-at-age in the forecast in this year's assessment (average of weight at age from 2021-2023) is slightly lower for age 1-4 compared to last year's assessment (average weight from 2020-2022; Figure 2.8.2.3.1). Vice versa, the weight at age 6 and 7 is slightly lower last year compared to this year. All other ages ( $8-10$ ) are very similar between the two assessments. Selectivity at age are also very similar between this year's and previous year's assessment (Figure 2.4.1.5).

Table 2.8.2.3.2 shows the intermediate year assumptions and recruitments for the current and previous assessment.

### 2.8.2.4 Medium term projections

Preliminary medium term projections were undertaken at the working group, progressing the stock another 3 years in time up to 2028. These analyses were undertaken to illustrate how stock biomass and catch would develop under expected exploitation regimes, being Fmsy, F status quo (2023) and F equal to that to realise the 2023 catches while assuming that recruitment would be in line with recent estimates of average recruitment. The results show (Figure 2.8.2.4.1) that SSB is likely to go down from 2024 onwards for the F2023 scenario and from 2025 onwards for the other scenarios. Catches are expected to increase up to 2025 and decline from 2026 onwards under all three scenarios. The stock is expected to remain well above MSY Btrigger under all scenarios up to 2028.

### 2.9 Comparison with previous assessment and forecast

Comparison of the assessments made in 2022 and 2023 (Figure 2.9.1) shows a revision of the historical values of F, SSB and recruitments. The 2022 recruitment is now estimated to be 68.9 billion, while last year's estimate was 43.2 billion. The 2021 recruitment is adjusted slightly downwards: from 71.56 billion to 69.63 . F for 2022 is now estimated to be 0.41 while the same value in last year's assessment was 0.37 . SSB for 2022 is now estimated to 4.58 million tonnes while last year's value was 4.96 million tonnes, whereas SSB for 2023 is now estimated to be 6.24 million tonnes compared to 6.62 million tonnes last year.

The effect on SSB and F of the final vs. preliminary catch data from 2022 used in the WGWIDE 2022 assessment are limited (Figure 2.9.2). Hence, the slight downscaling of SSB and upscaling of F in recent years as observed in this year's assessment (Figure 2.9.3) is likely due to the addition
of a new year of data rather than of having the final 2022 catch data included in the WGWIDE 2023 assessment. The revision in F and SSB may be resulting from the different weighting of the catch and survey data compared to WGWIDE 2022, with a lower weighting of the survey data for the older age classes in particular (see also discussion in Section 2.4.1).

The reasons for the revision of recruitment is linked to the high survey index for age 2 in 2023 (i.e. the 2021 year class and 2022 recruitment; Table 2.3.6.1.1) corroborated by high commercial catch at age of the same year class in 2022 and 2023 (Table 2.4.1.4), and 2) the use of (uncertain) preliminary catch data for 2022 in the 2022 assessment.

With respect to point 1, while the IBWSS index for age 1 in 2022 is historically high, the catch numbers at age 1 for 2022 in last year's assessment were not that high compared to earlier years in the time series (Table 2.4.1.4). This has likely led to a lower estimate of the 2022 recruitment last year compared to this year, where both catches and survey data indicated a strong year class.
With respect to point 2, catch numbers 'in year' are based on Q1 and Q2, whereas young blue whiting are caught at highest numbers in Q3 and Q4 (2.3.1.2.2). Hence, the final numbers at age in the catch are higher than the preliminary catch for age 1, while the final catch data are lower than the preliminary data for age 2-10+ (Figure 2.3.2.2). The final total catch weight is $6.2 \%$ lower than the preliminary values for 2022 (Table 2.3.2.4). Including the final catch data for 2022 in the WGWIDE 2022 assessment leads indeed to a much higher 2022 recruitment estimate (similar to that of WGWIDE 2023) than with the preliminary catch data for 2022 (Figure 2.9.2).

### 2.10 Quality considerations

Based on the confidence interval produced by the assessment model SAM there is a moderate to high uncertainty of the absolute estimate of F and SSB and the recruiting year classes (Figure 2.4.1.7). The retrospective analysis (Figure 2.4.1.6) shows a tendency to underestimate recruitment, but unbiased estimates of F and SSB.

There are several sources of uncertainty: age reading, stock identity, survey indices and the use of preliminary catch data. As there is only one survey (IBWSS) that covers the spawning stock, the quality of the survey influences the assessment result considerably. The Inter-Benchmark Protocol on Blue Whiting (IBPBLW 2016) introduced a configuration of the SAM model that includes the use of estimated correlation for catch and survey observations. This handles the "year effects" in the survey observation in a better way than assuming an uncorrelated variance structure as usually applied in assessment models. However, biased survey indices will still give a biased stock estimate with the new SAM configuration. The estimated correlation for catch at age observations might correspond to the age reading discrepancy as also estimated from intercalibration exercise. The use of additional survey data may be beneficial, especially in years without IBWSS data.

Utilization of preliminary catch data provides the assessment with information for the most recent year in addition to the survey information. This should give a less biased assessment, as potentially biased survey data in the final year are supplemented by additional catch data. The preliminary catch weight was $6.2 \%$ higher than the final data for 2022, although the differences are smaller for most years between 2016-2022 (Table 2.3.2.4).

### 2.11 Management considerations

The assessment this year estimates a lower $F(2022)$, a higher $\operatorname{SSB}(2023)$ and a considerably larger 2021 year class size than estimated last year. The 2021 year class will be fully recruited to the fishery in 2023 and contribute considerably to the SSB ( $82 \%$ mature at age 3). Estimated catch
proportions of the year classes 2020 and 2021 in 2024 are 44 and $31 \%$, respectively. SSB in 2024 is estimated to be well above MSY $B_{\text {trigger, }}$ but $F$ remains above FMSY. $^{\text {M }}$

The SSB is above 6 million tonnes, which is more than twice the $B_{\text {msy }}$ and $B_{\lim }(2250000$ tonnes and 1500000 tonnes respectively). Blue whiting was by IUCN in 2014 evaluated as a species of least concern (https://www.iucnredlist.org/) and it is neither found on OSPAR's list of threatened and declining species nor on the HELCOM Red List. Thus ICES has not identified any conservation aspects.

### 2.12 Ecosystem considerations

Blue whiting is one of the most abundant pelagic and mesopelagic fish stocks in the Northeast Atlantic, SSB estimated from 1.4-6.9 million tonnes during the period from 1981 to 2020 (ICES, 2020). The stock is widely distributed and highly migratory. Its distribution range is approximately from $30^{\circ} \mathrm{N}$ to $80^{\circ} \mathrm{N}$ latitude and from the coast of Europe to Greenland, into the Barents Sea and Mediterranean Sea (Trenkel et al., 2014). Spawning is in the spring and mostly occurs on the shelf and banks west of Ireland and Scotland. The major summer feeding area is in the Norwegian Sea. Blue whiting is most frequently observed at $100-600 \mathrm{~m}$ depth (Heino and Godo, 2002). Their most important prey are euphausiids, amphipods and copepods (Pinnegar et al., 2015, Bachiller et al., 2016) and they are prey for piscivorous fish (Dolgov et al., 2010) and cetaceans (Hátún et al., 2009a). Blue whiting is an important species in the NE Atlantic and it's best documented ecosystem interactions are listed below:
(a) Stock productivity - recruitment: blue whiting population dynamic is driven by large annual variability in recruitment (at age 1 in the assessment model) which is not linked to spawning stock size (ICES, 2020). Changes in recruitment have been correlated to changes in the North Atlantic subpolar gyre between strong and weak states (Hátún et al., 2009a,b). Two hypotheses have been suggested to explain a causal relationship between low gyre index and high recruitment (Payne et al., 2012). One suggests changes in marine climate where weak gyre results in increased flow of warm subtropical waters and increased abundance of important prey for juvenile blue whiting on their nursing grounds west of Ireland and Scotland. The other suggests increasing predation of mackerel on blue whiting larvae during years of weak index, but neither has been proven right (Payne et al., 2012).
(b) Changes in distribution: blue whiting spawning distribution varies between years. It has been linked to the North Atlantic subpolar gyre as a strong gyre (cold and fresh water masses on the Rockall Plateau) shrinks the spawning area compared to a weak gyre (increasing saline and warm waters at Rockall) which expands the spawning area northward and westward into Rockall Plateau (Hátún et al., 2009a,b; Miesner and Payne, 2018). Salinity appears specifically to impact spawning location of blue whiting (Miesner and Payne, 2018).
(c) It is still disputed whether there are one or two blue whiting populations in the Northeast Atlantic (Keating et al., 2014; Pointin and Payne, 2014; ICES, 2016c; Mahé et al., 2016). Currently blue whiting is considered a single population for management purposes.
(d) Trophic interactions in the Norwegian Sea: there appears to be limited prey competition between blue whiting and the two other abundant pelagic species, Norwegian spring-spawning herring and Atlantic mackerel, as studies show limited dietary overlap between blue whiting and the two other species (Bachiller et al., 2016; Pinnegar et al., 2015). Limited prey competition between blue whiting and mackerel can be explained by limited vertical spatial overlap, mackerel mostly feed in the surface layer and blue whiting deeper in the water column (Utne et al., 2012). Where distribution of blue whiting and herring overlap (Utne et al., 2012) they appear to feed on different species, herring mainly feed on copepods and blue whiting mainly on
euphausiids and amphipods, although juvenile blue whiting feed on copepods (Bachiller et al., 2016; Pinnegar et al., 2015).

An extensive overview of ecosystem considerations relevant for blue whiting can be found in the Stock Annex.

### 2.13 Regulations and their effects

There is a long-term management strategy agreed by the European Union, the Faroe Islands, Iceland and Norway. However there is no agreement between the Coastal States, i.e. EU, Norway, Iceland and the Faroe Islands on the share of the blue whiting TAC. The catch advice does not take into account consistent deviations from the long-term management strategy as evident from the sum of unilateral quotas since 2018. During the evaluation of the management strategy (ICES, 2016b), the implementation error in the form of a consistent overshoot of the TAC was not included. Therefore, the current implementation of the long-term management strategy may no longer be precautionary. See section 1.8 for a comparison of historic advice, TAC and catch.

WGWIDE estimates the total expected catch for 2023 to be 1672378 tonnes (Table 2.3.2.3), whereas ICES advised that when the long-term management strategy agreed by the European Union, the Faroe Islands, Iceland, and Norway is applied, catches in 2023 should be no more than 1359629 tonnes. This advice was followed by the Coastal States by setting a TAC at the ICES advice, however there was no agreement on the split of TAC between nations. The sum of unilateral quotas for 2023 exceeds the agreed TAC by $23 \%$.

### 2.13.1 Management plans and evaluations

A response to a NEAFC request to ICES to evaluate a long-term management strategy for the fisheries on the blue whiting ICES WKBWMSE was established in the fall of 2015. The ICES Advice September 2016, "NEAFC request to ICES to evaluate a long-term management strategy for the fisheries on the blue whiting (Micromesistius poutassou) stock" concluded:

- That the harvest control rule (HCR) proposed for the Long-Term Management Strategy (LTMS) for blue whiting, as described in the request, is precautionary given the ICES estimates of $B_{\lim }\left(1.5\right.$ million $t$ ), $B_{p a}\left(2.25\right.$ million $t$ ), and $F_{\text {msy }}(0.32)$.
- The HCR was found to be precautionary both with and without the $20 \%$ TAC change limits above $B_{\text {pa }}$. However, the $20 \%$ TAC change limits can lead to the TAC being lowered significantly if the stock is estimated to be below $B_{p a}$, while also limiting how quickly the TAC can increase once the stock is estimated to have recovered above $\mathrm{B}_{\mathrm{pa}}$.
- The evaluation found that including a $10 \%$ interannual quota flexibility ('banking and borrowing') in the LTMS had an insignificant effect on the performance of the HCR.

The managers subsequently agreed a long-term management plan for the blue whiting stock that included a new clause (6.b.) that stated that the $25 \%$ restriction in the increase of catch advice (TAC change limit) is not applied when the catch advice deviates more than $40 \%$ from the catch advice of the preceding year (see Annex 1 in Anon, 2021). This clause was then evaluated by ICES and found precautionary in 2017 (Working Document in ICES, 2017).

The management strategy evaluation did not take into account consistent deviations from the long-term management strategy as evident from the sum of unilateral quotas in recent years. During the evaluation of the management strategy (ICES, 2016b), the implementation error in the form of a consistent overshoot of the TAC was not included. Therefore, the current implementation of the long-term management strategy may no longer be precautionary.

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### 2.15 Tables

Table 2.3.1.1. Blue whiting. ICES estimated catches (tonnes) by country for the period 1988-2022.

| Country | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 18941 | 26630 | 27052 | 15538 | 34356 | 41053 | 20456 | 12439 | 52101 | 26270 | 61523 | 82935 | 89500 | 41450 | 54663 | 48659 | 18134 |
| Estonia |  |  |  |  | 6156 | 1033 | 4342 | 7754 | 10982 | 5678 | 6320 |  | ** |  |  |  |  |
| Faroe Islands | 79831 | 75083 | 48686 | 10563 | 13436 | 16506 | 24342 | 26009 | 24671 | 28546 | 71218 | 329895 | 322322 | 266799 | 321013 | 317859 | 225003 |
| France |  | 2191 |  |  |  | 1195 |  | 720 | 6442 | 12446 | 7984 | 14149 |  | 8046 | 18009 | 16638 | 11723 |
| Germany | 5546 | 5417 | 1699 | 349 | 1332 | 100 | 2 | 6313 | 6876 | 4724 | 17969 | 22803 | 15293 | 22823 | 36437 | 34404 | 25259 |
| Iceland |  | 4977 |  |  |  |  |  | 369 | 302 | 10464 | 68681 | 501493 | 379643 | 265516 | 309508 | 236538 | 159307 |
| Ireland | 4646 | 2014 |  |  | 781 |  | 3 | 222 | 1709 | 25785 | 45635 | 22580 | 75393 | 73488 | 54910 | 31132 | 22852 |
| Japan |  |  |  |  | 918 | 1742 | 2574 |  |  |  |  |  |  |  |  |  |  |
| Latvia |  |  |  |  | 10742 | 10626 | 2582 |  |  |  |  |  |  |  |  |  |  |
| Lithuania |  |  |  |  |  | 2046 |  |  |  |  |  |  |  |  | 4635 | 9812 | 5338 |
| Netherlands | 800 | 2078 | 7750 | 17369 | 11036 | 18482 | 21076 | 26775 | 17669 | 24469 | 27957 | 48303 | 95311 | 147783 | 102711 | 79875 | 78684 |
| Norway | 233314 | 301342 | 310938 | 137610 | 181622 | 211489 | 229643 | 339837 | 394950 | 347311 | 560568 | 834540 | 957684 | 738490 | 642451 | 539587 | 418289 |
| Poland | 10 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Portugal | 5979 | 3557 | 2864 | 2813 | 4928 | 1236 | 1350 | 2285 | 3561 | 2439 | 1900 | 2651 | 3937 | 5190 | 5323 | 3897 | 4220 |
| Spain | 24847 | 30108 | 29490 | 29180 | 23794 | 31020 | 28118 | 25379 | 21538 | 27683 | 27490 | 13825 | 15612 | 17643 | 15173 | 13557 | 14342 |
| Sweden** | 1229 | 3062 | 1503 | 1000 | 2058 | 2867 | 3675 | 13000 | 4000 | 4568 | 9299 | 65532 | 19083 | 2960 | 101 | 467 | 4 |
| UK (England + Wales)*** |  |  |  |  |  |  |  |  |  |  |  |  | 2593 | 7356 | 10035 | 12926 | 14147 |
| UK (Northern Ireland) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| UK (Scotland) | 5183 | 8056 | 6019 | 3876 | 6867 | 2284 | 4470 | 10583 | 14326 | 33398 | 92383 | 27382 | 57028 | 104539 | 72106 | 43540 | 38150 |
| USSR / Russia * | 177521 | 162932 | 125609 | 151226 | 177000 | 139000 | 116781 | 107220 | 86855 | 118656 | 130042 | 355319 | 346762 | 332226 | 329100 | 236369 | 225163 |
| Greenland** |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Unallocated |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| тоTAL | 557847 | 627447 | 561610 | 369524 | 475026 | 480679 | 459414 | 578905 | 645982 | 672437 | 1128969 | 2321406 | 2380161 | 2034309 | 1976176 | 1625259 | 1260615 |

* From 1992 only Russia.
** Estimates from Sweden and Greenland: are not included in the Catch at Age Number.
** Estonia (2004): Reported to the EU but not to the ICES WGNPBW. (Landings of 19,467 tonnes).
*** From 2012.

Table 2.3.1.1. (continued). Blue whiting. ICES estimated catches (tonnes) by country for the period 1988-2022.

| Country | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 248 | 140 | 165 | 340 | 2167 | 35256 | 45178 | 39395 | 60868 | 87348 | 68716 | 58997 | 40321 | 45644 |
| Estonia |  |  |  |  |  |  |  |  |  |  | 0 |  |  |  |
| Faroe Islands | 58354 | 49979 | 16405 | 43290 | 85768 | 224700 | 282502 | 282416 | 356501 | 349838 | 336569 | 343372 | 202415 | 217401 |
| France | 8831 | 7839 | 4337 | 9799 | 8978 | 10410 | 9659 | 10345 | 13369 | 16784 | 16095 | 13769 | 14612 | 14202 |
| Germany | 5044 | 9108 | 278 | 6239 | 11418 | 24487 | 24107 | 20025 | 45555 | 47708 | 38244 | 42362 | 35327 | 21667 |
| Iceland | 120202 | 87942 | 5887 | 63056 | 104918 | 182879 | 214870 | 186914 | 228934 | 292944 | 268356 | 243725 | 190146 | 191813 |
| Ireland | 8776 | 8324 | 1195 | 7557 | 13205 | 21466 | 24785 | 27657 | 43238 | 49903 | 38836 | 40135 | 39514 | 28972 |
| Lithuania |  |  |  |  |  | 4717 |  | 1129 | 5300 |  |  | 9543 | 21183 | 13149 |
| Netherlands | 35686 | 33762 | 4595 | 26526 | 51635 | 38524 | 56397 | 58148 | 81156 | 121864 | 75020 | 62309 | 62017 | 63249 |
| Norway | 225995 | 194317 | 20539 | 118832 | 196246 | 399520 | 489439 | 310412 | 399363 | 438426 | 351429 | 354033 | 233968 | 194973 |
| Poland |  |  |  |  |  |  |  |  | 15889 | 12152 | 27185 | 47616 | 26077 | 20948 |
| Portugal | 2043 | 1482 | 603 | 1955 | 2056 | 2150 | 2547 | 2586 | 2046 | 2497 | 3481 | 2819 | 2522 | 2784 |
| Spain | 20637 | 12891 | 2416 | 6726 | 15274 | 32065 | 29206 | 31952 | 28920 | 24718 | 22782 | 23676 | 25509 | 26310 |
| Sweden | 3 | 50 | 1 | 4 | 199 | 2 | 32 | 42 | 90 | 16** | 54 | 25 | 40 | 20 |
| UK (England + Wales) | 6176 | 2475 | 27 | 1590 | 4100 | 11 | 131 | 1374+ | 3447 | 1864 | 4062 | 7458 | 8783 | 7482 |
| UK (Northern Ireland) |  |  |  |  | 1232 | 2205 | 1119 |  |  | 4508 | 2899 | 2958 |  |  |
| UK (Scotland) | 173 | 5496 | 1331 | 6305 | 8166 | 24630 | 30508 | 37173 | 64724 | 66682 | 54040 | 41344 | 65085 | 42903 |
| Russia | 149650 | 112553 | 45841 | 88303 | 120674 | 152256 | 185763 | 173655 | 188449 | 170892 | 188006 | 181496 | 133605^ | 128002 |
| Greenland |  |  |  |  | 2133 |  |  |  | 20212 | 23333 | 19753 | 19611 | 20190 | 19218 |
| Unallocated |  |  |  | 3499 |  |  |  |  |  |  |  |  | 22137 |  |
| TOTAL | 641818 | 526357 | 103620 | 384021 | 628169 | 1155279 | 1396244 | 1181850 | 1558061 | 1711461 | 1515527 | 1495248 | 1143450 | 1038736 |

only landings (2018).

+ data updated in 2018.
^ Russia 2021 preliminary data (Q1+Q2) submitted to WGWIDE 2021.


## Table 2.3.1.2. Blue whiting. ICES estimated catches (tonnes) by country and ICES division for 2022

| ICES Division | Denmark | Faroe <br> Islands | France | Germany | Greenland | Iceland | Ireland | Lithuania | Netherlands | Norway | Poland | Portugal | Russia | Spain | Sweden | UK (England and Wales) | UK(Scotland) | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 27.2.a | 17 | 17599 |  | 4 | 2294 | 10624 | 0.837 |  | 244.83 | 12326 |  |  | 33256 |  |  |  | 2 | 76368 |
| 27.3.a | 103 |  |  |  |  |  |  |  |  | 0 |  |  |  |  | 7.192 |  |  | 110 |
| 27.4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 201 | 201 |
| 27.4.a | 96 | 11142 | 0 |  | 1750 | 20869 |  |  | 0.79 | 20300 |  |  | 4770 |  | 2 |  | 26 | 58956 |
| 27.4.b | 13 |  |  |  |  |  |  |  |  | 80 |  |  |  |  | 11 |  |  | 104 |
| 27.5.a |  | 1288 |  |  |  | 28094 |  |  |  |  |  |  |  |  |  |  |  | 29382 |
| 27.5.b |  | 161657 | 1170 | 5106 | 13774 | 116955 |  | 44 | 5475.36 | 11722 | 1743 |  | 62721 |  |  |  |  | 380366 |
| 27.6.a | 19808 | 22029 | 4581 | 583 | 1400 | 15269 | 12104.32 | 8555 | 16640.39 | 60427 | 5006 |  | 3943 | 21 |  | 7478 | 12063 | 189909 |
| 27.6.b | 1998 |  | 102 |  |  |  | 830 |  |  | 1875 |  |  | 6490 | 4 |  |  | 2139 | 13437 |
| 27.7.a |  |  | 31 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 31 |
| 27.7.b |  |  | 18 |  |  |  | 6.192 |  | 3.33 |  | 1365 |  |  | 8 |  |  | 268 | 1667 |
| 27.7.c | 16403 | 3686 | 2026 | 8957 |  |  | 12822.21 | 2895 | 7731.16 | 53612 | 1170 |  | 11822 | 35 |  |  | 17517 | 138676 |
| 27.7.e |  |  | 38 |  |  |  |  |  |  |  |  |  |  |  |  | 1 |  | 39 |
| 27.7.f |  |  | 0 |  |  |  |  |  | 0.38 |  |  |  |  |  |  | 0 |  | 1 |
| 27.7.g |  |  |  |  |  |  |  |  | 0.55 |  |  |  |  | 9 |  | 2 |  | 12 |
| 27.7.h |  |  | 132 | 111 |  |  |  |  | 235.24 |  |  |  |  | 20 |  | 2 |  | 501 |
| 27.7.j | 5 |  | 1319 | 269 |  |  | 18.71 |  | 2047.23 |  | 1857 |  |  | 143 |  |  |  | 5659 |
| 27.7.k | 7186 |  | 252 | 6619 |  |  | 3189.748 |  | 16547.78 | 34632 | 7108 |  | 5000 | 1 |  |  | 10686 | 91221 |
| 27.8.a | 15 |  | 834 | 18 |  |  |  |  | 623.49 |  |  |  |  | 58 |  |  |  | 1549 |
| 27.8.b |  |  | 5 |  |  |  |  |  |  |  |  |  |  | 516 |  |  |  | 521 |
| 27.8.c |  |  | 0 |  |  |  |  |  |  |  |  | 5 |  | 14953 |  |  |  | 14958 |
| 27.8.d |  |  | 3693 |  |  |  |  | 1655 | 13698.07 |  | 2700 |  |  | 9 |  |  |  | 21756 |
| 27.9.a |  |  |  |  |  |  |  |  |  |  |  | 2779 |  | 10531 |  |  |  | 13310 |
| 27.9.b |  |  |  |  |  |  |  |  |  |  |  |  |  | 0 |  |  |  | 0 |
| 27.14.a |  |  |  |  |  | 1 |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 27.14.b |  |  |  |  |  | 1 |  |  |  |  |  |  |  |  |  |  |  | 1 |
| Total | 45644 | 217401 | 14202 | 21667 | 19218 | 191813 | 28972 | 13149 | 63249 | 194973 | 20948 | 2784 | 128002 | 26310 | 20 | 7482 | 42903 | 1038736 |

Table 2.3.1.3. Blue whiting. ICES estimated catches (tonnes) by quarter and ICES division for 2022.

| ICES <br> Division | Quarter 1 | Quarter 2 | Quarter 3 | Quarter 4 | 2022* | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 27.2.a | 2132 | 50488 | 13752 | 9997 |  | 76368 |
| 27.3.a | 0 | 40 | 60 | 9 |  | 110 |
| 27.4 |  |  |  |  | 201 | 201 |
| 27.4.a | 6215 | 35454 | 6133 | 11154 |  | 58956 |
| 27.4.b | 0 | 80 | 20 | 3 |  | 104 |
| 27.5.a |  | 16 | 1267 | 28099 |  | 29382 |
| 27.5.b | 55090 | 306682 | 11 | 18583 |  | 380366 |
| 27.6.a | 36714 | 149333 | 1 | 3824 | 37 | 189909 |
| 27.6.b | 12367 | 1064 | 4 | 0 | 2 | 13437 |
| 27.7.a |  | 31 |  |  |  | 31 |
| 27.7.b | 1647 | 19 | 1 | 0 |  | 1667 |
| 27.7.c | 136741 | 1872 | 21 | 42 |  | 138676 |
| 27.7.e | 12 | 1 |  | 26 |  | 39 |
| 27.7.f | 0 |  |  | 0 |  | 1 |
| 27.7.g | 3 | 4 | 3 | 2 |  | 12 |
| 27.7.h | 241 | 8 | 9 | 243 |  | 501 |
| 27.7.j | 3247 | 55 | 219 | 2138 |  | 5659 |
| 27.7.k | 90816 | 252 |  | 153 |  | 91221 |
| 27.8.a | 411 | 40 | 2 | 1097 |  | 1549 |
| 27.8.b | 119 | 133 | 99 | 172 |  | 521 |
| 27.8.c | 3441 | 4417 | 4772 | 2327 |  | 14958 |
| 27.8.d | 21372 | 2 | 5 | 377 |  | 21756 |
| 27.9.a | 2123 | 3264 | 4800 | 3122 |  | 13310 |
| 27.9.b |  | 0 |  |  |  | 0 |
| 27.14.a |  |  | 1 |  |  |  |
| 27.14.b |  |  |  | 1 |  |  |
| Total | 372691 | 553255 | 31180 | 81370 | 240 | 1038736 |

*Discards data from UK(Scotland) were provided by year, due to sampling intensity.

Table 2.3.1.4. Blue whiting. ICES estimated catches (tonnes) from the main fisheries 1988-2022 by area.

| Year | Norwegian Sea fishery (SAs1+2;Divs. 5 .a,14a-b) | Fishery in the spawning area (SA 12.; Divs. 5.b, 6.ab, 7.a-c) | Directedand mixed fisheries in the North Sea (SA4; Div.3.a) | Total northern areas | Total southern areas (SAs8+9;Divs. 7.d-k) | Grand total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 55829 | 426037 | 45143 | 527009 | 30838 | 557847 |
| 1989 | 42615 | 475179 | 75958 | 593752 | 33695 | 627447 |
| 1990 | 2106 | 463495 | 63192 | 528793 | 32817 | 561610 |
| 1991 | 78703 | 218946 | 39872 | 337521 | 32003 | 369524 |
| 1992 | 62312 | 318018 | 65974 | 446367 | 28722 | 475026 |
| 1993 | 43240 | 347101 | 58082 | 448423 | 32256 | 480679 |
| 1994 | 22674 | 378704 | 28563 | 429941 | 29473 | 459414 |
| 1995 | 23733 | 423504 | 104004 | 551241 | 27664 | 578905 |
| 1996 | 23447 | 478077 | 119359 | 620883 | 25099 | 645982 |
| 1997 | 62570 | 514654 | 65091 | 642315 | 30122 | 672437 |
| 1998 | 177494 | 827194 | 94881 | 1099569 | 29400 | 1128969 |
| 1999 | 179639 | 943578 | 106609 | 1229826 | 26402 | 1256228 |
| 2000 | 284666 | 989131 | 114477 | 1388274 | 24654 | 1412928 |
| 2001 | 591583 | 1045100 | 118523 | 1755206 | 24964 | 1780170 |
| 2002 | 541467 | 846602 | 145652 | 1533721 | 23071 | 1556792 |
| 2003 | 931508 | 1211621 | 158180 | 2301309 | 20097 | 2321406 |
| 2004 | 921349 | 1232534 | 138593 | 2292476 | 85093 | 2377569 |
| 2005 | 405577 | 1465735 | 128033 | 1999345 | 27608 | 2026953 |
| 2006 | 404362 | 1428208 | 105239 | 1937809 | 28331 | 1966140 |
| 2007 | 172709 | 1360882 | 61105 | 1594695 | 17634 | 1612330 |
| 2008 | 68352 | 1111292 | 36061 | 1215704 | 30761 | 1246465 |
| 2009 | 46629 | 533996 | 22387 | 603012 | 32627 | 635639 |
| 2010 | 36214 | 441521 | 17545 | 495280 | 28552 | 523832 |
| 2011 | 20599 | 72279 | 7524 | 100401 | 3191 | 103592 |
| 2012 | 24391 | 324545 | 5678 | 354614 | 29402 | 384016* |
| 2013 | 31759 | 481356 | 8749 | 521864 | 103973 | 625837** |
| 2014 | 45580 | 885483 | 28596 | 959659 | 195620 | 1155279 |
| 2015 | 150828 | 895684 | 44661 | 1091173 | 305071 | 1396244 |
| 2016 | 59744 | 905087 | 55774 | 1020604 | 162583 | 1183187*** |
| 2017 | 136565 | 1284105 | 45474 | 1466144 | 91917 | 1558061 |
| 2018 | 143204 | 1445957 | 43484 | 1632646 | 78831 | 1711477 |
| 2019 | 68593 | 1271883 | 44856 | 1385333 | 130194 | 1515527 |
| 2020 | 92084 | 1059197 | 64327 | 1215608 | 279640 | 1495248 |
| 2021 | 112082 | 801768 | 39509 | 953359 | 190091 | 1143450 |
| 2022 | 105752 | 724086 | 59371 | 889209 | 149527 | 1038736 |

* Official catches by area from Sweden are not included (2012);
** Official catches by area from Sweden and Greenland are not included (2013);
*** Grand total includes only 1336 tonnes from UK(England + Wales) ( $\mathbf{2 0 1 6}$ total catch from UK(England + Wales) = $\mathbf{1 3 7 4}$ ton).

Table 2.3.1.5. Blue whiting. ICES estimates (tonnes) of catches, landings and discards by country for 2022.

| Country | BMS landing | Discards | Landings | Catches | \% discards |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  | 127 | 45517 | 45644 | 0.3 |
| Faroe Islands |  |  | 217401 | 217401 | 0.0 |
| France |  |  | 14202 | 14202 | 0.0 |
| Germany |  |  | 21667 | 21667 | 0.0 |
| Greenland |  |  | 19218 | 19218 | 0.0 |
| Iceland |  |  | 191813 | 191813 | 0.0 |
| Ireland |  | 390 | 28582 | 28972 | 1.3 |
| Lithuania |  |  | 13149 | 13149 | 0.0 |
| Netherlands |  | 0 | 63249 | 63249 | 0.0 |
| Norway |  |  | 194973 | 194973 | 0.0 |
| Poland |  |  | 20948 | 20948 | 0.0 |
| Portugal |  | 828 | 1956 | 2784 | 29.7 |
| Russia |  |  | 128002 | 128002 | 0.0 |
| Spain | 0 | 2047 | 24263 | 26310 | 7.8 |
| Sweden |  | 7 | 13 | 20 | 35.6 |
| UK (England \& Wales) | 0 | 2 | 7480 | 7482 | 0.0 |
| UK(Scotland) | 0 | 240 | 42662 | 42903 | 0.6 |
| Total | 1 | 3641 | 1035094 | 1038736 | 0.4 |

Table 2.3.1.6. Blue whiting. ICES estimated catches (tonnes) inside and outside NEAFC regulatory area for 2022 by country.

| Country | Outside NEAFC | Inside NEAFC | Total |
| :--- | ---: | ---: | ---: | ---: |
| Denmark | 45644 |  | 45644 |
| Faroe Islands | 210876 | 6525 | 217401 |
| France* | 14165 | 37 | 14202 |
| Germany | 16562 | 5106 | 21667 |
| Greenland | 13803 | 5415 | 19218 |
| Iceland | 191405 | 408 | 191813 |
| Ireland | 28972 |  | 28972 |
| Lithuania | 13149 | 5719 | 13149 |
| Netherlands | 57529 | 20572 | 63249 |
| Norway* | 174401 | 1743 | 194973 |
| Poland | 19205 |  | 20948 |
| Portugal | 2784 | 40093 | 2784 |
| Russia | 87909 |  | 128002 |
| Spain | 26310 | 26310 |  |
| Sweden | 20 |  | 20 |
| UK (England and Wales) | 7482 | 85618 | 7482 |
| UK(Scotland) | 42903 |  | 42903 |
| Total | 953119 |  | 1038736 |

[^2]Table 2.3.1.1.1. Blue whiting. ICES estimated catches (tonnes), the percentage of catch covered by the sampling programme, No. of age samples, No. of fish measured and No. of fish aged for 2000-2022.

| Year | Catch (tonnes) | \% catch covered by the sampling programme | No. Length Samples | No. Length measurements | No. Fish Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 1412928 | - | 1136 | 125162 | 13685 |
| 2001 | 1780170 | - | 985 | 173553 | 17995 |
| 2002 | 1556792 | - | 1037 | 116895 | 19202 |
| 2003 | 2321406 | - | 1596 | 188770 | 26207 |
| 2004 | 2377569 | - | 1774 | 181235 | 27835 |
| 2005 | 2026953 | - | 1833 | 217937 | 32184 |
| 2006 | 1966140 | - | 1715 | 190533 | 27014 |
| 2007 | 1610090 | 87 | 1399 | 167652 | 23495 |
| 2008 | 1246465 | 90 | 927 | 113749 | 21844 |
| 2009 | 635639 | 88 | 705 | 79500 | 18142 |
| 2010 | 524751 | 87 | 584 | 82851 | 16323 |
| 2011 | 103591 | 85 | 697 | 84651 | 12614 |
| 2012 | 373937 | 80 | 1143 | 173206 | 15745 |
| 2013 | 625837 | 96 | 915 | 111079 | 14633 |
| 2014 | 1155279 | 89 | 912 | 111316 | 39738 |
| 2015 | 1396244 | 94 | 1570 | 102367 | 29821 |
| 2016 | 1183187 | 89 | 1092 | 120329 | 13793 |
| 2017 | 1558061 | 91 | 1779 | 147297 | 15828 |
| 2018 | 1711078 | 87 | 1565 | 131779 | 16426 |
| 2019 | 1515527 | 84 | 1537 | 136604 | 17869 |
| 2020 | 1495248 | 81 | 672 | 89110 | 16641 |
| 2021 | 1143450 | 81 | 1676 | 129317 | 15215 |
| 2022 | 1038736 | 77 | 1868 | 167650 | 15636 |

Table 2.3.1.1.2. Blue whiting. ICES estimated catches (tonnes), the percentage of catch covered by the sampling programme (catch-at-age numbers), No. of length samples, No. of age samples, No. of fish measured, No. of fish aged, No. of fish aged by 1000 tonnes and No. of fish measured by 1000 tonnes by country for 2022.

| Country | Catch (tonnes) | \% catch <br> covered by the sampling programme | No. of Length Samples | No. of Length Measured | No. of Age Samples | No. Age Readings |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 45644.006 | 95 | 18 | 810 | 18 | 807 |
| Faroe Islands | 217401 | 95 | 16 | 1644 | 16 | 1526 |
| France | 14201.75117 | 0 | 56 | 3850 | 0 | 0 |
| Germany | 21667.404 | 2 | 9 | 2236 | 9 | 366 |
| Greenland | 19218 | 0 | 0 | 0 | 0 | 0 |
| Iceland | 191813 | 99 | 59 | 4734 | 59 | 1432 |
| Ireland | 28972.01355 | 93 | 45 | 11383 | 25 | 1400 |
| Lithuania | 13149.01716 | 0 | 0 | 0 | 0 | 0 |
| Netherlands | 63248.6 | 85 | 51 | 10362 | 51 | 1229 |
| Norway | 194973.044 | 94 | 75 | 2213 | 75 | 2213 |
| Poland | 20948.14 | 0 | 0 | 0 | 0 | 0 |
| Portugal | 2783.7031 | 100 | 29 | 2679 | 29 | 898 |
| Russia | 128002 | 14 | 224 | 66932 | 224 | 2980 |
| Spain | 26309.59696 | 99 | 1155 | 52028 | 1155 | 1830 |
| Sweden | 20.192 | 0 | 0 | 0 | 0 | 0 |
| UK (England) | 7482.381 | 100 | 21 | 3102 | 246 | 246 |
| UK(Scotland) | 42902.586 | 99 | 110 | 5677 | 20 | 709 |
| Total | 1038736 | 77 | 1868 | 167650 | 1927 | 15636 |

Table 2.3.1.1.3. Blue whiting. ICES estimated catches (tonnes), No. of Age samples, No. of fish measured and No. of fish aged by country and quarter for 2022 (cont.).

| Country | Catches (ton) | No. of Length Samples | No. of Length Measured | No. Age Readings |
| :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |
| Quarter 1 | 25640 | 10 | 487 | 485 |
| Quarter 2 | 19818 | 8 | 323 | 322 |
| Quarter 3 | 71 | 0 | 0 | 0 |
| Quarter 4 | 116 | 0 | 0 | 0 |
| Total | 45644 | 18 | 810 | 807 |
| Faroe Islands |  |  |  |  |
| Quarter 1 | 65336 | 4 | 441 | 400 |
| Quarter 2 | 139663 | 10 | 1036 | 1000 |
| Quarter 3 | 1836 | 0 | 0 | 0 |
| Quarter 4 | 10566 | 2 | 167 | 126 |
| Total | 217401 | 16 | 1644 | 1526 |
| France |  |  |  |  |
| Quarter 1 | 4773 | 31 | 2103 | 0 |
| Quarter 2 | 5778 | 0 | 0 | 0 |
| Quarter 3 | 2 | 0 | 0 | 0 |
| Quarter 4 | 3649 | 25 | 1747 | 0 |
| Total | 14202 | 56 | 3850 | 0 |
| Germany |  |  |  |  |
| Quarter 1 | 15594 | 0 | 0 | 0 |
| Quarter 2 | 601 | 0 | 0 | 0 |
| Quarter 3 | 28 | 1 | 324 | 84 |
| Quarter 4 | 5445 | 8 | 1912 | 282 |
| Total | 21667 | 9 | 2236 | 366 |
| Greenland |  |  |  |  |
| Quarter 2 | 19184 | 0 | 0 | 0 |
| Quarter 3 | 5 | 0 | 0 | 0 |
| Quarter 4 | 29 | 0 | 0 | 0 |
| Total | 19218 | 0 | 0 | 0 |
| Iceland |  |  |  |  |
| Quarter 2 | 157563 | 49 | 3993 | 1185 |
| Quarter 3 | 2627 | 1 | 75 | 25 |
| Quarter 4 | 31623 | 9 | 666 | 222 |
| Total | 191813 | 59 | 4734 | 1432 |
| Ireland |  |  |  |  |
| Quarter 1 | 17427 | 25 | 6332 | 1000 |
| Quarter 2 | 11350 | 10 | 3137 | 400 |
| Quarter 4 | 195 | 10 | 1914 | 0 |
| Total | 28972 | 45 | 11383 | 1400 |

Table 2.3.1.1.3. (continued) Blue whiting. ICES estimated catches (tonnes), No. of Age samples, No. of fish measured and No. of fish aged by country and quarter for 2022 (cont.).

| Country | Catches (ton) | No. of Length Samples | No. of Length Measured | No. Age Readings |
| :---: | :---: | :---: | :---: | :---: |
| Lithuania |  |  |  |  |
| Quarter 1 | 9534 | 0 | 0 | 0 |
| Quarter 4 | 3615 | 0 | 0 | 0 |
| Total | 13149 | 0 | 0 | 0 |
| Netherlands |  |  |  |  |
| Quarter 1 | 41359 | 31 | 6934 | 739 |
| Quarter 2 | 14815 | 20 | 3428 | 490 |
| Quarter 3 | 143 | 0 | 0 | 0 |
| Quarter 4 | 6932 | 0 | 0 | 0 |
| Total | 63249 | 51 | 10362 | 1229 |
| Norway |  |  |  |  |
| Quarter 1 | 111166 | 43 | 1264 | 1264 |
| Quarter 2 | 67106 | 25 | 739 | 739 |
| Quarter 3 | 10119 | 0 | 0 | 0 |
| Quarter 4 | 6582 | 7 | 210 | 210 |
| Total | 194973 | 75 | 2213 | 2213 |
| Poland |  |  |  |  |
| Quarter 1 | 18024 | 0 | 0 | 0 |
| Quarter 4 | 2924 | 0 | 0 | 0 |
| Total | 20948 | 0 | 0 | 0 |
| Portugal |  |  |  |  |
| Quarter 1 | 382 | 5 | 366 | 121 |
| Quarter 2 | 647 | 10 | 980 | 340 |
| Quarter 3 | 761 | 8 | 817 | 136 |
| Quarter 4 | 993 | 6 | 516 | 301 |
| Total | 2784 | 29 | 2679 | 898 |

Table 2.3.1.1.3. (continued) Blue whiting. ICES estimated catches (tonnes), No. of Age samples, No. of fish measured and No. of fish aged by country and quarter for 2022 (cont.).

| Country | Catches (ton) | No. of Length Samples | No. of Length Measured | No. Age Readings |
| :---: | :---: | :---: | :---: | :---: |
| Russia |  |  |  |  |
| Quarter 1 | 24906 | 46 | 11930 | 558 |
| Quarter 2 | 89610 | 158 | 49025 | 1879 |
| Quarter 3 | 6555 | 20 | 5977 | 543 |
| Quarter 4 | 6931 | 0 | 0 | 0 |
| Total | 128002 | 224 | 66932 | 2980 |
| Spain |  |  |  |  |
| Quarter 1 | 5363 | 295 | 18110 | 349 |
| Quarter 2 | 7262 | 315 | 11506 | 675 |
| Quarter 3 | 9006 | 350 | 14755 | 489 |
| Quarter 4 | 4679 | 195 | 7657 | 317 |
| Total | 26310 | 1155 | 52028 | 1830 |
| Sweden |  |  |  |  |
| Quarter 1 | 0 | 0 | 0 | 0 |
| Quarter 2 | 0 | 0 | 0 | 0 |
| Quarter 3 | 17 | 0 | 0 | 0 |
| Quarter 4 | 3 | 0 | 0 | 0 |
| Total | 20 | 0 | 0 | 0 |
| UK (England and Wales) |  |  |  |  |
| Quarter 2 | 7480 | 21 | 3102 | 246 |
| Quarter 4 | 2 | 0 | 0 | 0 |
| Total | 7482 | 21 | 3102 | 246 |
| UK(Scotland) |  |  |  |  |
| Quarter 1 | 33187 | 13 | 1841 | 484 |
| Quarter 2 | 9454 | 7 | 805 | 225 |
| Quarter 3 | 11 | 0 | $0$ | 0 |
| Quarter 4 | 10 | $0$ | 0 | 0 |
| 2022* | 240 | 90 | 3031 | 0 |
| Total | 42903 | 110 | 5677 | 709 |
| Total Geral | 1038736 | 1868 | 167650 | 15636 |

* discards data provided for the whole year, not raised by quarter due to sampling intensity.

Table 2.3.1.1.4. Blue whiting. ICES estimated catches (tonnes), the percentage of catch covered by the sampling programme, No. of length samples, No. of age samples, No. of fish measured, No. of fish aged, No. of fish aged by 1000 tonnes and No. of fish measured by 1000 tonnes by ICES division for 2022.

| ICES Division | Catch (tonnes) | No. Length samples | No. Age samples | No. Fish Measured | No. Aged | No <br> Aged/ <br> 1000 <br> tonnes | No Measured/ 1000 tonnes |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 27.4 | 201 | 62 | 0 | 2038 | 0 | 0 | 10150 |
| 27.2.a | 76368 | 78 | 78 | 21425 | 1551 | 20 | 281 |
| 27.3.a | 110 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.4.a | 58956 | 17 | 17 | 921 | 542 | 9 | 16 |
| 27.4.b | 104 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.5.a | 29382 | 7 | 7 | 515 | 174 | 6 | 18 |
| 27.5.b | 380366 | 164 | 157 | 37932 | 2988 | 8 | 100 |
| 27.6.a | 189909 | 123 | 333 | 14689 | 3239 | 17 | 77 |
| 27.6.b | 13437 | 12 | 10 | 2760 | 117 | 9 | 205 |
| 27.7.a | 31 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.7.b | 1667 | 5 | 5 | 591 | 50 | 30 | 354 |
| 27.7.c | 138676 | 133 | 127 | 9424 | 2182 | 16 | 68 |
| 27.7.e | 39 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.7.f | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.7.g | 12 | 1 | 1 | 0 | 0 | 0 | 0 |
| 27.7.h | 501 | 89 | 76 | 2399 | 134 | 268 | 4790 |
| 27.7.j | 5659 | 83 | 75 | 3158 | 325 | 57 | 558 |
| 27.7.k | 91221 | 75 | 61 | 14383 | 1429 | 16 | 158 |
| 27.8.a | 1549 | 49 | 27 | 1579 | 0 | 0 | 1019 |
| 27.8.b | 521 | 117 | 117 | 4839 | 0 | 0 | 9280 |
| 27.8.c | 14958 | 449 | 449 | 31324 | 877 | 59 | 2094 |
| 27.8.d | 21756 | 28 | 11 | 3433 | 257 | 12 | 158 |
| 27.9.a | 13310 | 376 | 376 | 16240 | 1771 | 133 | 1220 |
| 27.9.b | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.14.a | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 27.14.b | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| TOTAL (2022) | 1038736 | 1868 | 1927 | 167650 | 15636 | 15 | 161 |

Table 2.3.2.1. Blue whiting. ICES estimated preliminary landings (tonnes) in 2023 by quarter and ICES division. Data submitted to InterCatch.

| ICES Div. | Landings |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Quarter 1 | Quarter 2 | Quarter 3 | Total |
| 27.2.a | 13018 | 22651 |  | 35668 |
| 27.3.a |  | 4 | 0 | 4 |
| 27.4.a | 20983 | 49951 |  | 70935 |
| 27.4.b |  | 1 |  | 1 |
| 27.5.a | 0 |  |  | 0 |
| 27.5.b | 205409 | 361019 |  | 566428 |
| 27.6.a | 90371 | 70702 |  | 161072 |
| 27.6.b | 13366 | 3891 |  | 17257 |
| 27.7.b | 1532 |  |  | 1532 |
| 27.7.c | 174451 | 28653 |  | 203104 |
| 27.7.g | 0 |  |  | 0 |
| 27.7.j | 8933 | 4 |  | 8938 |
| 27.7.k | 375968 | 11 |  | 375979 |
| 27.8.a | 1414 | 2 |  | 1416 |
| 27.8.b | 42 | 5 |  | 47 |
| 27.8.c | 3615 | 4843 |  | 8458 |
| 27.8.d | 696 |  |  | 696 |
| 27.9.a | 2270 | 3234 |  | 5504 |
| 27.12 | 10670 |  |  | 10670 |
| 27.14.b | 3 | 0 |  | 3 |
| Total | 922741 | 544972 | 0 | 1467714 |

Table 2.3.2.2. Blue whiting. ICES estimated preliminary catches (tonnes), the percentage of catch covered by the sampling programme, No. of samples, No. of fish measured, No. of fish aged, No. of fish aged by 1000 tonnes and No. of fish measured by 1000 tonnes by ICES division for 2023 preliminary data (quarters 1 and 2). Data submitted to InterCatch.

| ICES Division | Catch (ton) | No. samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: |
| 27.2.a | 35668 | 4 | 388 | 320 |
| 27.3.a | 4 | 0 | 0 | 0 |
| 27.4.a | 70935 | 0 | 0 | 0 |
| 27.4.b | 1 | 0 | 0 | 0 |
| 27.5.a | 0 | 0 | 0 | 0 |
| 27.5.b | 566428 | 36 | 4760 | 1662 |
| 27.6.a | 161072 | 39 | 9650 | 1200 |
| 27.6.b | 17257 | 9 | 2211 | 251 |
| 27.7.b | 1532 | 0 | 0 | 0 |
| 27.7.c | 203104 | 30 | 4751 | 1215 |
| 27.7.g | 0 | 0 | 0 | 0 |
| 27.7.j | 8938 | 1 | 36 | 36 |
| 27.7.k | 375979 | 95 | 14008 | 2241 |
| 27.8.a | 1416 | 0 | 0 | 0 |
| 27.8.b | 47 | 0 | 0 | 0 |
| 27.8.c | 8458 | 0 | 0 | 0 |
| 27.8.d | 696 | 0 | 0 | 0 |
| 27.9.a | 5504 | 10 | 1577 | 461 |
| 27.12 | 10670 | 13 | 4036 | 142 |
| 27.14.b | 3 | 0 | 0 | 0 |
| Total | 1467714 | 237 | 41417 | 7528 |

Table 2.3.2.3. Blue whiting. ICES estimates of catches (tonnes) in 2023, based on (initial) declared quotas and expected uptake estimated by WGWIDE.

| Country | 2023 Q1 | 2023 Q2 | 2023 Q3 | $\begin{aligned} & 2023 \text { Total } \\ & \text { Catch (Q1+Q2) } \end{aligned}$ | Expected remained catch | 2023 Total <br> Catch |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 55666 | 20680 | 7 | 76353 | 1330 | 77683 |
| Faroe Islands | 171434 | 194673 |  | 366107 | 26435 | 392542 |
| France* | 10237 | 2547 |  | 12784 | 4261 | 17045 |
| Germany | 28916 | 5684 | 0 | 34600 | 0 | 34600 |
| Greenland | 4200 | 11222 |  | 15422 | 10378 | 25800 |
| Iceland | 78173 | 148260 |  | 226433 | 47009 | 273442 |
| Ireland | 38275 | 15363 |  | 53638 |  | 53638 |
| Lithuania** | 10065 | 4836 |  | 14901 | 5511 | 20412 |
| Netherlands | 51853 | 19766 | 652 | 72271 |  | 79500 |
| Norway | 283778 | 70478 |  | 354256 |  | 378000 |
| Poland | 19713 | 5934 |  | 25647 |  | 25647 |
| Portugal | 639 | 607 |  | 1246 | 2000 | 3246 |
| Russia*** | 58321 | 74710 |  | 133031 | 16442 | 149473 |
| Spain | 8869 | 8773 |  | 17643 |  | 35285 |
| Sweden | 0 | 0 |  | 0 |  | 30 |
| UK(England \& Wales) | 3719 | 3802 |  | 7521 | 895 | 8416 |
| UK(Scotland) | 61047 | 26284 |  | 87331 | 10287 | 97618 |
| Total | 884905 | 613619 | 659 | 1499183 | 124549 | 1672378 |

*Provisional 2023 data from the national landing information system for Q1 and Q2. Assumed this is 75\% of the total catch (based on final 2022 catches), so estimated $\mathbf{2 5 \%}$ still to be caught.
** Assumed this is $73 \%$ of the total catch (based on final 2022 catches), so estimated $23 \%$ still to be caught.
*** Assumed this is $89 \%$ of the total catch (based on final 2022 catches), so estimated $11 \%$ still to be caught.

Table 2.3.2.4. Blue whiting. Comparison of preliminary and final catches (in tonnes) calculated from sum of product of catch number and mean weight at age used in the assessment).

|  | Final | Preliminary | Change in \% * |
| :--- | :--- | :--- | :--- |
| 2016 | 1180786 | 1147000 | 2.9 |
| 2017 | 1555069 | 1559437 | -0.3 |
| 2018 | 1709856 | 1712874 | -0.2 |
| 2019 | 1512026 | 1444301 | 4.7 |


| $\mathbf{2 0 2 0}$ | 1460507 | 1478358 | -1.2 |
| :--- | :--- | :--- | :--- |
| $\mathbf{2 0 2 1}$ | 1139531 | 1242727 | -8.3 |
| $\mathbf{2 0 2 2}$ | 1038736 | 1107529 | -6.2 |
| * (fina |  |  |  |

* (final-preliminary)/preliminary*100

Table 2.3.3.1. Blue whiting. Catch-at-age numbers (thousands) by year. Discards included since 2014. Values for 2023 are preliminary.

| Year |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Age |


| Year <br> Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 2156261 | 4426323 | 6723748 | 6697923 | 3044943 | 1276412 | 649885 | 249097 | 75415 | 36805 |
| 2005 | 1427277 | 1518938 | 5083550 | 5871414 | 4450171 | 1419089 | 518304 | 249443 | 100374 | 55226 |
| 2006 | 412961 | 939865 | 4206005 | 6150696 | 3833536 | 1718775 | 506198 | 181181 | 67573 | 36688 |
| 2007 | 167027 | 306898 | 1795021 | 4210891 | 3867367 | 2353478 | 935541 | 320529 | 130202 | 88573 |
| 2008 | 408790 | 179211 | 545429 | 2917190 | 3262956 | 1919264 | 736051 | 315671 | 113086 | 126637 |
| 2009 | 61125 | 156156 | 231958 | 594624 | 1596095 | 1156999 | 592090 | 251529 | 88615 | 48908 |
| 2010 | 349637 | 222975 | 160101 | 208279 | 646380 | 992214 | 702569 | 256604 | 70487 | 43693 |
| 2011 | 162997 | 101810 | 63954 | 53863 | 69717 | 116396 | 120359 | 55470 | 25943 | 12542 |
| 2012 | 239667 | 351845 | 663155 | 141854 | 106883 | 203419 | 363779 | 356785 | 212492 | 157947 |
| 2013 | 228175 | 508122 | 848597 | 896966 | 462714 | 224066 | 321310 | 397536 | 344285 | 383601 |
| 2014 | 588717 | 584084 | 2312953 | 2019373 | 1272862 | 416523 | 386396 | 462339 | 526141 | 662747 |
| 2015 | 2944849 | 2852384 | 2427329 | 2465286 | 1518235 | 707533 | 329882 | 258743 | 239164 | 450046 |
| 2016 | 1239331 | 3518677 | 2933271 | 1874011 | 1367844 | 756824 | 339851 | 185368 | 131039 | 288635 |
| 2017 | 401947 | 1999011 | 7864694 | 4063916 | 1509651 | 777185 | 263007 | 110351 | 63945 | 149369 |
| 2018 | 418781 | 541041 | 3572357 | 7340084 | 2983975 | 1022883 | 424206 | 150753 | 90387 | 163289 |
| 2019 | 249923 | 433573 | 1288871 | 3778379 | 5037323 | 1645999 | 431925 | 145916 | 50622 | 81357 |
| 2020 | 1135859 | 834162 | 1106838 | 1797157 | 3072708 | 3041983 | 923392 | 235330 | 80440 | 64535 |
| 2021 | 2069387 | 830692 | 1266077 | 1214790 | 1438769 | 1404443 | 1360104 | 304891 | 100993 | 59441 |
| 2022 | 1927303 | 3265756 | 1640425 | 1370794 | 924112 | 746761 | 670201 | 616695 | 123772 | 85610 |
| 2023 | 391151 | 3847086 | 8551171 | 3024159 | 1356485 | 529259 | 527277 | 395306 | 376977 | 162922 |

Table 2.3.4.1. Blue whiting. Individual mean weight (kg) at age in the catch. Preliminary values for 2023.

| Year Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1981 | 0.052 | 0.065 | 0.103 | 0.125 | 0.141 | 0.155 | 0.170 | 0.178 | 0.187 | 0.213 |
| 1982 | 0.045 | 0.072 | 0.111 | 0.143 | 0.156 | 0.177 | 0.195 | 0.200 | 0.204 | 0.231 |
| 1983 | 0.046 | 0.074 | 0.118 | 0.140 | 0.153 | 0.176 | 0.195 | 0.200 | 0.204 | 0.228 |
| 1984 | 0.035 | 0.078 | 0.089 | 0.132 | 0.153 | 0.161 | 0.175 | 0.189 | 0.186 | 0.206 |
| 1985 | 0.038 | 0.074 | 0.097 | 0.114 | 0.157 | 0.177 | 0.199 | 0.208 | 0.218 | 0.237 |
| 1986 | 0.040 | 0.073 | 0.108 | 0.130 | 0.165 | 0.199 | 0.209 | 0.243 | 0.246 | 0.257 |
| 1987 | 0.048 | 0.086 | 0.106 | 0.124 | 0.147 | 0.177 | 0.208 | 0.221 | 0.222 | 0.254 |


| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0.053 | 0.076 | 0.097 | 0.128 | 0.142 | 0.157 | 0.179 | 0.199 | 0.222 | 0.260 |
| 1989 | 0.059 | 0.079 | 0.103 | 0.126 | 0.148 | 0.158 | 0.171 | 0.203 | 0.224 | 0.253 |
| 1990 | 0.045 | 0.070 | 0.106 | 0.123 | 0.147 | 0.168 | 0.175 | 0.214 | 0.217 | 0.256 |
| 1991 | 0.055 | 0.091 | 0.107 | 0.136 | 0.174 | 0.190 | 0.206 | 0.230 | 0.232 | 0.266 |
| 1992 | 0.057 | 0.083 | 0.119 | 0.140 | 0.167 | 0.193 | 0.226 | 0.235 | 0.284 | 0.294 |
| 1993 | 0.066 | 0.082 | 0.109 | 0.137 | 0.163 | 0.177 | 0.200 | 0.217 | 0.225 | 0.281 |
| 1994 | 0.061 | 0.087 | 0.108 | 0.137 | 0.164 | 0.189 | 0.207 | 0.217 | 0.247 | 0.254 |
| 1995 | 0.064 | 0.091 | 0.118 | 0.143 | 0.154 | 0.167 | 0.203 | 0.206 | 0.236 | 0.256 |
| 1996 | 0.041 | 0.080 | 0.102 | 0.116 | 0.147 | 0.170 | 0.214 | 0.230 | 0.238 | 0.279 |
| 1997 | 0.047 | 0.072 | 0.102 | 0.121 | 0.140 | 0.166 | 0.177 | 0.183 | 0.203 | 0.232 |
| 1998 | 0.048 | 0.072 | 0.094 | 0.125 | 0.149 | 0.178 | 0.183 | 0.188 | 0.221 | 0.248 |
| 1999 | 0.063 | 0.078 | 0.088 | 0.109 | 0.142 | 0.170 | 0.199 | 0.193 | 0.192 | 0.245 |
| 2000 | 0.057 | 0.075 | 0.086 | 0.104 | 0.133 | 0.156 | 0.179 | 0.187 | 0.232 | 0.241 |
| 2001 | 0.050 | 0.078 | 0.094 | 0.108 | 0.129 | 0.163 | 0.186 | 0.193 | 0.231 | 0.243 |
| 2002 | 0.054 | 0.074 | 0.093 | 0.115 | 0.132 | 0.155 | 0.173 | 0.233 | 0.224 | 0.262 |
| 2003 | 0.049 | 0.075 | 0.098 | 0.108 | 0.131 | 0.148 | 0.168 | 0.193 | 0.232 | 0.258 |
| 2004 | 0.042 | 0.066 | 0.089 | 0.102 | 0.123 | 0.146 | 0.160 | 0.173 | 0.209 | 0.347 |
| 2005 | 0.039 | 0.068 | 0.084 | 0.099 | 0.113 | 0.137 | 0.156 | 0.166 | 0.195 | 0.217 |
| 2006 | 0.049 | 0.072 | 0.089 | 0.105 | 0.122 | 0.138 | 0.163 | 0.190 | 0.212 | 0.328 |
| 2007 | 0.050 | 0.064 | 0.091 | 0.103 | 0.115 | 0.130 | 0.146 | 0.169 | 0.182 | 0.249 |
| 2008 | 0.055 | 0.075 | 0.100 | 0.106 | 0.120 | 0.133 | 0.146 | 0.160 | 0.193 | 0.209 |
| 2009 | 0.056 | 0.085 | 0.105 | 0.119 | 0.124 | 0.138 | 0.149 | 0.179 | 0.214 | 0.251 |
| 2010 | 0.052 | 0.064 | 0.110 | 0.154 | 0.154 | 0.163 | 0.175 | 0.187 | 0.200 | 0.272 |
| 2011 | 0.055 | 0.079 | 0.107 | 0.136 | 0.169 | 0.169 | 0.179 | 0.189 | 0.214 | 0.270 |
| 2012 | 0.041 | 0.072 | 0.098 | 0.141 | 0.158 | 0.172 | 0.180 | 0.185 | 0.189 | 0.203 |
| 2013 | 0.051 | 0.077 | 0.094 | 0.117 | 0.139 | 0.162 | 0.185 | 0.188 | 0.198 | 0.197 |
| 2014 | 0.049 | 0.078 | 0.093 | 0.112 | 0.128 | 0.155 | 0.178 | 0.190 | 0.202 | 0.217 |
| 2015 | 0.039 | 0.070 | 0.094 | 0.117 | 0.137 | 0.155 | 0.174 | 0.183 | 0.193 | 0.201 |
| 2016 | 0.047 | 0.066 | 0.084 | 0.107 | 0.125 | 0.142 | 0.152 | 0.167 | 0.184 | 0.206 |


| Year Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2017 | 0.056 | 0.072 | 0.080 | 0.094 | 0.113 | 0.131 | 0.148 | 0.172 | 0.190 | 0.212 |
| 2018 | 0.055 | 0.080 | 0.091 | 0.098 | 0.111 | 0.129 | 0.142 | 0.165 | 0.175 | 0.216 |
| 2019 | 0.068 | 0.085 | 0.099 | 0.109 | 0.118 | 0.130 | 0.144 | 0.167 | 0.167 | 0.228 |
| 2020 | 0.063 | 0.084 | 0.099 | 0.115 | 0.127 | 0.135 | 0.144 | 0.161 | 0.176 | 0.207 |
| 2021 | 0.058 | 0.086 | 0.099 | 0.119 | 0.133 | 0.143 | 0.150 | 0.166 | 0.181 | 0.209 |
| 2022 | 0.048 | 0.067 | 0.086 | 0.106 | 0.122 | 0.136 | 0.141 | 0.152 | 0.155 | 0.189 |
| 2023 | 0.045 | 0.061 | 0.076 | 0.092 | 0.120 | 0.144 | 0.155 | 0.169 | 0.182 | 0.207 |

Table 2.3.5.1. Blue whiting. Natural mortality and proportion mature.

| AGE | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7 - 1 0 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Proportion mature | 0.00 | 0.11 | 0.40 | 0.82 | 0.86 | 0.91 | 0.94 | 1.00 |
| Natural mortality | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 | 0.20 |

Table 2.3.6.1.1. Blue whiting. Time-series of StoX abundance estimates of blue whiting (millions) by age in the IBWSS. Total biomass in last column (1000 t). Shaded values (ages 1-8; years 2004-2023) are used as input to the assessment

| Year | Age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | TSB |
| 2004 | 1097 | 5538 | 13062 | 15134 | 5119 | 1086 | 994 | 593 | 164 | 0 | 3505 |
| 2005 | 2129 | 1413 | 5601 | 7780 | 8500 | 2925 | 632 | 280 | 129 | 23 | 2513 |
| 2006 | 2512 | 2224 | 10881 | 11695 | 4717 | 2719 | 923 | 352 | 198 | 39 | 3517 |
| 2007 | 468 | 706 | 5241 | 11244 | 8437 | 3155 | 1110 | 456 | 123 | 65 | 3274 |
| 2008 | 337 | 524 | 1455 | 6661 | 6747 | 3882 | 1719 | 1029 | 269 | 296 | 2647 |
| 2009 | 275 | 329 | 360 | 1292 | 3739 | 3458 | 1636 | 587 | 250 | 194 | 1599 |
| 2010* |  |  |  |  |  |  |  |  |  |  |  |
| 2011 | 312 | 1361 | 1135 | 930 | 1043 | 1713 | 2171 | 2423 | 1298 | 272 | 1827 |
| 2012 | 1140 | 1816 | 6454 | 1021 | 595 | 1415 | 2220 | 1777 | 1249 | 1085 | 2347 |
| 2013 | 582 | 1337 | 6175 | 7211 | 2938 | 1282 | 1308 | 1398 | 929 | 1807 | 3110 |
| 2014 | 4183 | 1491 | 5239 | 8420 | 10202 | 2754 | 772 | 577 | 899 | 2251 | 3761 |
| 2015 | 3255 | 4570 | 1891 | 3641 | 1797 | 466 | 174 | 108 | 206 | 365 | 1405 |
| 2016 | 2745 | 7893 | 10164 | 6274 | 4687 | 1539 | 413 | 133 | 235 | 361 | 2873 |


| Year | Age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ | TSB |
| 2017 | 262 | 2248 | 15682 | 10176 | 3762 | 1793 | 921 | 76 | 84 | 173 | 3135 |
| 2018 | 836 | 628 | 6615 | 21490 | 7692 | 2187 | 755 | 188 | 72 | 138 | 4035 |
| 2019 | 1129 | 1169 | 3468 | 9590 | 16979 | 3434 | 484 | 513 | 99 | 43 | 4198 |
| 2020** |  |  |  |  |  |  |  |  |  |  |  |
| 2021 | 1948 | 2095 | 2545 | 2275 | 3914 | 3197 | 3379 | 463 | 189 | 114 | 2357 |
| 2022 | 4461 | 9313 | 4830 | 5460 | 2587 | 1880 | 898 | 1764 | 71 | 178 | 2707 |
| 2023 | 873 | 8135 | 14771 | 2744 | 1352 | 711 | 520 | 202 | 508 | 67 | 2501 |

*Survey discarded. ${ }^{* *}$ No survey

Table 2.3.6.2.1. Blue whiting. Estimated abundance of 1 and 2 -year old blue whiting from the International Ecosystem Survey in Nordic Seas (IESNS), 2003-2023.

| Year\Age | Age 1 | Age 2 |
| :---: | :---: | :---: |
| 2003* | 16127 | 9317 |
| 2004* | 17792 | 11020 |
| 2005* | 19933 | 7908 |
| 2006* | 2512 | 5504 |
| 2007* | 592 | 213 |
| 2008 | 25 | 17 |
| 2009 | 7 | 8 |
| 2010 | 0 | 280 |
| 2011 | 1613 | 0 |
| 2012 | 9476 | 3265 |
| 2013 | 454 | 6544 |
| 2014 | 3937 | 2030 |
| 2015 | 8563 | 2796 |
| 2016 | 4223 | 8089 |
| 2017 | 1236 | 2087 |
| 2018 | 441 | 1491 |
| 2019 | 3157 | 215 |
| 2020 | 2822 | 481 |
| 2021 | 10264 | 1500 |
| 2022 | 17169 | 10575 |
| 2023 | 3873 | 4792 |

*Using the old TS-value. To compare the results all values were divided by approximately 3.1.

Table 2.3.6.2.2. Estimated abundance of 1-group of blue whiting from the Norwegian winter survey (late January-early March) in the Barents Sea. Blue whiting < 19 cm in total body length which most likely belong to 1-group. (Time series revised in 2023)

| Year | All | < 19 cm |
| :---: | :---: | :---: |
| 1994 | 65.43 | 0.04 |
| 1995 | 29.24 | 0.79 |
| 1996 | 1502.53 | 1461.18 |
| 1997 | 1671.57 | 1360.46 |
| 1998 | 85.99 | 17.28 |
| 1999 | 75.50 | 5.96 |
| 2000 | 644.59 | 554.44 |
| 2001 | 2365.41 | 1567.94 |
| 2002 | 1157.96 | 149.59 |
| 2003 | 826.79 | 192.15 |
| 2004 | 1834.75 | 676.71 |
| 2005 | 2072.79 | 722.10 |
| 2006 | 2322.05 | 26.96 |
| 2007 | 960.64 | 0.75 |
| 2008 | 162.36 | 0.17 |
| 2009 | 75.73 | 0.02 |
| 2010 | 29.63 | 0.44 |
| 2011 | 10.71 | 0.05 |
| 2012 | 768.68 | 752.68 |
| 2013 | 510.46 | 21.67 |
| 2014 | 227.63 | 172.10 |
| 2015 | 1227.69 | 1012.16 |
| 2016 | 701.67 | 95.35 |
| 2017 | 1299.24 | 6.36 |
| 2018 | 106.96 | 0.62 |
| 2019 | 110.02 | 55.22 |
| 2020 | 209.66 | 150.22 |
| 2021 | 964.32 | 828.96 |
| 2022 | 599.84 | 293.71 |
| 2023 | 343.02 | 23.59 |

Table 2.3.6.2.3. Blue whiting. 1-group indices of blue whiting from the Icelandic bottom-trawl surveys, 1-group (<22 cm in March).

| Catch Rate |  |
| :---: | :---: |
| Year | < 22 cm |
| 1996 | 6.5 |
| 1997 | 3.4 |
| 1998 | 1.1 |
| 1999 | 6.3 |
| 2000 | 9 |
| 2001 | 5.2 |
| 2002 | 14.2 |
| 2003 | 15.4 |
| 2004 | 8.9 |
| 2005 | 8.3 |
| 2006 | 30.4 |
| 2007 | 3.9 |
| 2008 | 0.1 |
| 2009 | 1.6 |
| 2010 | 0.2 |
| 2011 | 10.8 |
| 2012 | 29.9 |
| 2013 | 11.7 |
| 2014 | 66.3 |
| 2015 | 43.8 |
| 2016 | 6.3 |
| 2017 | 1.8 |
| 2018 | 0.4 |
| 2019 | 0.1 |
| 2020 | 9.8 |
| 2021 | 79.6 |
| 2022 | 91.2 |


| Catch Rate |  |
| :--- | :--- |
| Year | $<22 \mathrm{~cm}$ |
| 2023 | 6.8 |

$\qquad$

Table 2.3.6.2.4. Blue whiting. 1-group indices of blue whiting from Faroese bottom-trawl surveys, 1-group (<= $\mathbf{2 3} \mathbf{~ c m ~ i n ~}$ March).

| Catch Rate |  |
| :---: | :---: |
| Year | $<=\mathbf{2 3 ~ c m ~}$ |
| 1994 | 1401 |
| 1995 | 1162 |
| 1996 | 4821 |
| 1997 | 2307 |
| 1998 | 463 |
| 1999 | 1717 |
| 2000 | 863 |
| 2001 | 4424 |
| 2002 | 4480 |
| 2003 | 1038 |
| 2004 | 15749 |
| 2005 | 35159 |
| 2006 | 23105 |
| 2007 | 11568 |
| 2008 | 1268 |
| 2009 | 4362 |
| 2010 | 855 |
| 2011 | 23323 |
| 2012 | 8366 |
| 2013 | 13254 |
| 2014 | 70139 |
| 2015 | 34806 |
| 2016 | 21316 |
| 2017 | 4446 |
| 2018 | 1890 |
| 2019 | 286 |


| Catch Rate |  |
| :--- | :--- |
| Year | $<=23 \mathrm{~cm}$ |
| 2020 | 141 |
| 2021 | 2224 |
| 2022 | 1781 |
| 2023 | 3075 |

Table 2.4.1.1. Blue whiting. Parameter estimates, from final assessment (2023) and previous assessments (20192022).

| Parameter Year | 2019 | 2020 | 2021 | 2022 | 2023 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Random walk variance |  |  |  |  |  |
| -F Age 1-10 | 0.37 | 0.37 | 0.36 | 0.36 | 0.36 |
| Process error |  |  |  |  |  |
| $-\log (\mathrm{N})$ Age 1 | 0.61 | 0.61 | 0.60 | 0.62 | 0.66 |
| --- Age 2-10 | 0.18 | 0.18 | 0.18 | 0.18 | 0.18 |
| Observation variance |  |  |  |  |  |
| -Catch Age 1 | 0.45 | 0.43 | 0.43 | 0.43 | 0.42 |
| --- Age 2 | 0.31 | 0.27 | 0.28 | 0.27 | 0.27 |
| --- Age 3-8 | 0.19 | 0.19 | 0.19 | 0.18 | 0.18 |
| --- Age 9-10 | 0.39 | 0.38 | 0.38 | 0.37 | 0.37 |
| -IBWSS Age 1 | 0.81 | 0.75 | 0.71 | 0.74 | 0.69 |
| --- Age 2 | 0.38 | 0.34 | 0.32 | 0.33 | 0.32 |
| --- Age 3 | 0.41 | 0.41 | 0.39 | 0.39 | 0.36 |
| --- Age 4-6 | 0.37 | 0.37 | 0.37 | 0.35 | 0.37 |
| --- Age 7-8 | 0.54 | 0.55 | 0.53 | 0.53 | 0.56 |
| Survey catchability |  |  |  |  |  |
| -IBWSS Age 1 | 0.07 | 0.06 | 0.06 | 0.06 | 0.06 |
| --- Age 2 | 0.11 | 0.11 | 0.11 | 0.11 | 0.12 |
| --- Age 3 | 0.36 | 0.36 | 0.37 | 0.36 | 0.37 |
| --- Age 4 | 0.67 | 0.67 | 0.67 | 0.67 | 0.67 |
| --- Age 5-8 | 0.86 | 0.86 | 0.89 | 0.88 | 0.88 |


| Parameter Year | $\mathbf{2 0 1 9}$ | $\mathbf{2 0 2 0}$ | $\mathbf{2 0 2 1}$ | $\mathbf{2 0 2 2}$ | $\mathbf{2 0 2 3}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Rho |  |  |  |  |  |
| - | 0.93 | 0.94 | 0.93 | 0.93 | 0.93 |

Table 2.4.1.2. Blue whiting. Mohn's rho by year and average over the last five years ( $n=5$ ).

| Last data year | R(age 1) | SSB | Fbar(3-7) |
| :--- | :--- | :--- | :--- |
| 2018 | -0.165 | -0.075 | 0.051 |
| 2019 | -0.260 | 0.048 | -0.073 |
| 2020 | -0.124 | -0.004 | -0.013 |
| 2021 | -0.391 | -0.073 | 0.027 |
| 2022 | 0.003 | 0.110 | -0.188 |
| Rho mean | -0.188 | 0.001 | -0.039 |

Table 2.4.1.3. Blue whiting. Estimated fishing mortalities. Catch data for 2022 are preliminary.

| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1981 | 0.078 | 0.119 | 0.172 | 0.212 | 0.244 | 0.317 | 0.346 | 0.444 | 0.488 | 0.488 |
| 1982 | 0.068 | 0.103 | 0.149 | 0.183 | 0.208 | 0.270 | 0.293 | 0.372 | 0.406 | 0.406 |
| 1983 | 0.079 | 0.119 | 0.171 | 0.211 | 0.239 | 0.313 | 0.337 | 0.419 | 0.447 | 0.447 |
| 1984 | 0.096 | 0.144 | 0.212 | 0.265 | 0.304 | 0.397 | 0.418 | 0.508 | 0.530 | 0.530 |
| 1985 | 0.101 | 0.151 | 0.230 | 0.294 | 0.346 | 0.447 | 0.464 | 0.560 | 0.575 | 0.575 |
| 1986 | 0.113 | 0.169 | 0.268 | 0.357 | 0.431 | 0.552 | 0.572 | 0.690 | 0.704 | 0.704 |
| 1987 | 0.100 | 0.150 | 0.247 | 0.337 | 0.415 | 0.538 | 0.560 | 0.674 | 0.676 | 0.676 |
| 1988 | 0.098 | 0.148 | 0.253 | 0.349 | 0.440 | 0.578 | 0.590 | 0.695 | 0.678 | 0.678 |
| 1989 | 0.113 | 0.171 | 0.304 | 0.419 | 0.527 | 0.688 | 0.714 | 0.843 | 0.805 | 0.805 |
| 1990 | 0.105 | 0.159 | 0.291 | 0.407 | 0.511 | 0.666 | 0.714 | 0.852 | 0.817 | 0.817 |
| 1991 | 0.059 | 0.089 | 0.167 | 0.234 | 0.290 | 0.368 | 0.396 | 0.467 | 0.451 | 0.451 |
| 1992 | 0.049 | 0.073 | 0.140 | 0.195 | 0.233 | 0.286 | 0.311 | 0.370 | 0.363 | 0.363 |
| 1993 | 0.042 | 0.063 | 0.126 | 0.176 | 0.206 | 0.246 | 0.268 | 0.319 | 0.315 | 0.315 |
| 1994 | 0.036 | 0.054 | 0.113 | 0.160 | 0.186 | 0.219 | 0.242 | 0.292 | 0.287 | 0.287 |
| 1995 | 0.046 | 0.070 | 0.150 | 0.216 | 0.243 | 0.284 | 0.314 | 0.384 | 0.369 | 0.369 |
| 1996 | 0.056 | 0.085 | 0.186 | 0.272 | 0.297 | 0.347 | 0.383 | 0.474 | 0.451 | 0.451 |


| Year Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 0.054 | 0.084 | 0.188 | 0.280 | 0.300 | 0.348 | 0.381 | 0.475 | 0.453 | 0.453 |
| 1998 | 0.070 | 0.110 | 0.252 | 0.382 | 0.407 | 0.471 | 0.509 | 0.630 | 0.592 | 0.592 |
| 1999 | 0.064 | 0.101 | 0.238 | 0.371 | 0.397 | 0.457 | 0.482 | 0.594 | 0.558 | 0.558 |
| 2000 | 0.073 | 0.118 | 0.280 | 0.447 | 0.498 | 0.575 | 0.588 | 0.705 | 0.665 | 0.665 |
| 2001 | 0.069 | 0.111 | 0.265 | 0.431 | 0.494 | 0.571 | 0.572 | 0.679 | 0.643 | 0.643 |
| 2002 | 0.065 | 0.104 | 0.250 | 0.417 | 0.503 | 0.594 | 0.596 | 0.701 | 0.666 | 0.666 |
| 2003 | 0.067 | 0.107 | 0.262 | 0.441 | 0.545 | 0.634 | 0.628 | 0.709 | 0.668 | 0.668 |
| 2004 | 0.068 | 0.109 | 0.269 | 0.462 | 0.592 | 0.690 | 0.687 | 0.751 | 0.708 | 0.708 |
| 2005 | 0.059 | 0.095 | 0.238 | 0.420 | 0.557 | 0.650 | 0.655 | 0.702 | 0.665 | 0.665 |
| 2006 | 0.051 | 0.082 | 0.209 | 0.373 | 0.510 | 0.597 | 0.606 | 0.639 | 0.604 | 0.604 |
| 2007 | 0.048 | 0.078 | 0.198 | 0.359 | 0.508 | 0.606 | 0.631 | 0.663 | 0.630 | 0.630 |
| 2008 | 0.042 | 0.068 | 0.172 | 0.310 | 0.447 | 0.532 | 0.566 | 0.593 | 0.571 | 0.571 |
| 2009 | 0.027 | 0.046 | 0.113 | 0.200 | 0.290 | 0.343 | 0.373 | 0.388 | 0.376 | 0.376 |
| 2010 | 0.019 | 0.033 | 0.081 | 0.139 | 0.202 | 0.237 | 0.261 | 0.265 | 0.258 | 0.258 |
| 2011 | 0.006 | 0.010 | 0.024 | 0.041 | 0.058 | 0.067 | 0.074 | 0.075 | 0.075 | 0.075 |
| 2012 | 0.012 | 0.021 | 0.052 | 0.086 | 0.122 | 0.141 | 0.160 | 0.166 | 0.166 | 0.166 |
| 2013 | 0.020 | 0.035 | 0.091 | 0.152 | 0.216 | 0.244 | 0.278 | 0.292 | 0.291 | 0.291 |
| 2014 | 0.036 | 0.067 | 0.177 | 0.298 | 0.417 | 0.469 | 0.535 | 0.564 | 0.560 | 0.560 |
| 2015 | 0.047 | 0.087 | 0.233 | 0.394 | 0.550 | 0.621 | 0.695 | 0.727 | 0.718 | 0.718 |
| 2016 | 0.040 | 0.075 | 0.202 | 0.347 | 0.486 | 0.555 | 0.619 | 0.642 | 0.633 | 0.633 |
| 2017 | 0.038 | 0.071 | 0.194 | 0.337 | 0.468 | 0.532 | 0.583 | 0.595 | 0.587 | 0.587 |
| 2018 | 0.037 | 0.071 | 0.195 | 0.344 | 0.480 | 0.544 | 0.598 | 0.603 | 0.596 | 0.596 |
| 2019 | 0.033 | 0.064 | 0.179 | 0.320 | 0.447 | 0.502 | 0.551 | 0.546 | 0.539 | 0.539 |
| 2020 | 0.037 | 0.072 | 0.200 | 0.364 | 0.510 | 0.567 | 0.627 | 0.616 | 0.603 | 0.603 |
| 2021 | 0.033 | 0.063 | 0.179 | 0.330 | 0.463 | 0.510 | 0.565 | 0.556 | 0.542 | 0.542 |
| 2022 | 0.032 | 0.062 | 0.176 | 0.329 | 0.464 | 0.508 | 0.567 | 0.561 | 0.546 | 0.546 |
| 2023 | 0.040 | 0.077 | 0.220 | 0.416 | 0.586 | 0.638 | 0.717 | 0.713 | 0.693 | 0.693 |

Table 2.4.1.4. Blue whiting. Estimated stock numbers-at-age (thousands). Preliminary catch data for 2023 have been used

| Year <br> Age | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | 8 | 9 | $10+$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 198 | 3911082 | 3473729 | 4859801 | 2075296 | 2621128 | 214370 | 164493 | 174031 | 122185 | 296825 |
| 1 |  |  |  |  |  |  |  |  |  |  |


| Year <br> Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 200 | 39473578 | 1520756 | 1652709 | 1576909 | 4333565 | 110960 | 472780 | 323599 | 153264 | 313137 |
| 0 |  | 5 | 8 | 0 |  | 8 |  |  |  |  |
| 200 | 56306875 | 3170288 | 1204835 | 1070345 | 7436363 | 169332 | 489373 | 227283 | 163176 | 177924 |
| 1 |  | 1 | 5 | 4 |  | 2 |  |  |  |  |
| 200 | 48903240 | 4541472 | 2045210 | 8301987 | 5446779 | 338789 | 688508 | 255107 | 102940 | 154399 |
| 2 |  | 4 | 4 |  |  | 4 |  |  |  |  |
| 200 | 52805994 | 3905564 | 3501275 | 1357364 | 5072028 | 296933 | 120454 | 345754 | 88733 | 106793 |
| 3 |  | 7 | 5 | 3 |  | 4 | 4 |  |  |  |
| 200 | 28725410 | 4160331 | 3002806 | 2084339 | 7251492 | 246609 | 131334 | 500605 | 151357 | 80302 |
| 4 |  | 0 | 0 | 8 |  | 8 | 1 |  |  |  |
| 200 | 22392243 | 2175353 | 2828792 | 1809670 | 1073500 | 322431 | 111004 | 513923 | 191656 | 98590 |
| 5 |  | 5 | 5 | 0 | 2 | 8 | 3 |  |  |  |
| 200 | 8996712 | 1546680 | 2225552 | 1929181 | 9477173 | 445716 | 135547 | 482997 | 217912 | 119959 |
| 6 |  | 9 | 2 | 5 |  | 6 | 5 |  |  |  |
| 200 | 4795411 | 5934075 | 1311100 | 1589823 | 1028752 | 469070 | 183264 | 607844 | 228608 | 162811 |
| 7 |  |  | 9 | 1 | 4 | 8 | 2 |  |  |  |
| 200 | 5770397 | 3374375 | 4312576 | 1101624 | 9133410 | 489786 | 184847 | 750221 | 233866 | 199331 |
| 8 |  |  |  | 1 |  | 2 | 4 |  |  |  |
| 200 | 5610707 | 3935760 | 2371835 | 3683213 | 6911864 | 469433 | 218302 | 851727 | 322225 | 187523 |
| 9 |  |  |  |  |  | 5 | 4 |  |  |  |
| 201 | 15216679 | 4959540 | 2332092 | 1837257 | 3344578 | 432193 | 282361 | 119386 | 409841 | 263848 |
| 0 |  |  |  |  |  | 2 | 5 | 6 |  |  |
| 201 | 19324947 | 1312054 | 3292693 | 1650381 | 1605558 | 260092 | 268103 | 134117 | 806710 | 388065 |
| 1 |  | 3 |  |  |  | 5 | 7 | 6 |  |  |
| 201 | 19332266 | 1532573 | 1243344 | 2295475 | 1185709 | 161329 | 232615 | 211141 | 107109 | 891741 |
| 2 |  | 9 | 5 |  |  | 8 | 3 | 4 | 0 |  |
| 201 | 15784350 | 1598911 | 1160320 | 7330439 | 2218277 | 109261 | 137599 | 163711 | 134072 | 137247 |
| 3 |  | 4 | 0 |  |  | 0 | 2 | 7 | 7 | 7 |
| 201 | 36984108 | 1250942 | 1384294 | 7962259 | 4344040 | 134320 | 936231 | 100704 | 101930 | 148600 |
| 4 |  | 5 | 8 |  |  | 5 |  | 6 | 8 | 5 |
| 201 | 64187156 | 3215032 | 1069213 | 8438945 | 4190162 | 173860 | 740021 | 525537 | 488437 | 105980 |
| 5 |  | 3 | 9 |  |  | 3 |  |  |  | 2 |
| 201 | 34652405 | 5649762 | 2124528 | 7590563 | 4288098 | 180179 | 707388 | 355953 | 225142 | 598518 |
| 6 |  | 8 | 9 |  |  | 6 |  |  |  |  |
| 201 | 11587354 | 2765076 | 4494335 | 1492751 | 4486338 | 213721 | 733011 | 285910 | 163047 | 378334 |
| 7 |  | 0 | 9 | 4 |  | 3 |  |  |  |  |
| 201 | 12227217 | 8989382 | 2196133 | 2890624 | 8628995 | 244784 | 935525 | 314034 | 143534 | 266617 |
| 8 |  |  | 7 | 2 |  | 2 |  |  |  |  |
| 201 | 13717621 | 9033416 | 8505214 | 1458845 | 1604701 | 451827 | 112041 | 403736 | 138260 | 197105 |
| 9 |  |  |  | 6 | 1 | 0 | 5 |  |  |  |


| Age |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 202 | 23895318 | 1127174 | 6621636 | 6480692 | 8349859 | 777349 | 212907 | 559040 | 194769 | 159464 |
| 0 |  | 1 |  |  |  | 7 | 5 |  |  |  |
| 202 | 69630575 | 1722605 | 8478662 | 4466807 | 4174926 | 386502 | 358203 | 809847 | 269137 | 161491 |
| 1 |  | 2 |  |  |  | 3 | 4 |  |  |  |
| 202 | 68911606 | 6171130 | 1216460 | 5753449 | 2678693 | 217778 | 176057 | 167461 | 368745 | 223031 |
| 2 |  | 3 | 1 |  |  | 9 | 8 | 7 |  |  |
| 202 | 16657741 | 6126895 | 4712376 | 8654318 | 3192184 | 126877 | 111591 | 802669 | 750286 | 298302 |
| 3 |  | 9 | 8 |  |  | 9 | 1 |  |  |  |
| 202 | 22997749 | 1310944 | 4643722 | 3095019 | 4676343 | 145502 | 548909 | 446069 | 322198 | 429264 |
| 4 | * | 8 | 9 | 0 |  | 7 |  |  |  |  |

*assuming GM(1996-2022) recruitment in 2024.

Table 2.4.1.5. Blue whiting. Estimated recruitment (R) in thousands, spawning-stock biomass (SSB) in tonnes, average fishing mortality for ages 3 to $\mathbf{7}$ (Fbar 3-7) and total-stock biomass (TSB) in tonnes. Preliminary catch data for 2022 are included. Low and High refer to the 95\% confidence limits

| Yea <br> r | R(age 1) | Low | High | SSB | Low | High | Fba <br> r <br> (3- <br> 7) | Low | Hig h | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 198 | 3911082 | 254255 | 6016225 | 284503 | 22510 | 35958 | 0.25 | 0.18 | 0.35 | 334113 | 26914 | 414761 |
| 1 |  | 2 |  | 6 | 18 | 09 | 8 | 9 | 2 | 3 | 67 | 6 |
| 198 | 4602198 | 295428 | 7169319 | 230009 | 18400 | 28751 | 0.22 | 0.16 | 0.29 | 276579 | 22501 | 339962 |
| 2 |  | 7 |  | 1 | 37 | 69 | 0 | 4 | 6 | 3 | 33 | 7 |
| 198 | 1851194 | 121469 | 2821223 | 185606 | 15167 | 22713 | 0.25 | 0.19 | 0.33 | 289560 | 23659 | 354384 |
| 3 | 4 | 33 | 0 | 6 | 23 | 32 | 4 | 2 | 6 | 5 | 42 | 4 |
| 198 | 1822499 | 120576 | 2754697 | 175736 | 14590 | 21167 | 0.31 | 0.24 | 0.41 | 310004 | 25134 | 382349 |
| 4 | 9 | 09 | 1 | 9 | 01 | 55 | 9 | 5 | 6 | 7 | 87 | 1 |
| 198 | 9520388 | 631688 | 1434850 | 209493 | 17355 | 25286 | 0.35 | 0.27 | 0.46 | 323360 | 26513 | 394373 |
| 5 |  | 0 | 7 | 2 | 90 | 72 | 6 | 6 | 0 | 5 | 42 | 7 |
| 198 | 7170043 | 478833 | 1073639 | 227326 | 18867 | 27390 | 0.43 | 0.33 | 0.56 | 311099 | 25877 | 373998 |
| 6 |  | 9 | 7 | 2 | 03 | 21 | 6 | 9 | 1 | 7 | 94 | 1 |
| 198 | 9163882 | 610537 | 1375456 | 193194 | 16058 | 23242 | 0.41 | 0.32 | 0.54 | 281823 | 23471 | 338380 |
| 7 |  | 1 | 7 | 9 | 69 | 41 | 9 | 5 | 1 | 0 | 89 | 0 |
| 198 | 6359063 | 423532 | 9547720 | 163748 | 13722 | 19539 | 0.44 | 0.34 | 0.56 | 242330 | 20265 | 289774 |
| 8 |  | 3 |  | 6 | 85 | 38 | 2 | 3 | 9 | 4 | 41 | 8 |
| 198 | 8473071 | 562028 | 1277390 | 154491 | 12987 | 18377 | 0.53 | 0.41 |  | $238782$ | 19872 | 286906 |
| 9 |  | 1 | 4 | 3 | 03 | 99 | 0 | 4 | 0 | 0 | 96 |  |
| 199 | 1920506 | 125774 | 2932498 | 135892 | 11324 | 16306 | 0.51 | 0.39 | 0.67 | 251743 | 20188 | 313919 |
| 0 | 6 | 84 | 7 | 5 | 65 | 70 | 8 | 7 | 5 | 1 | 15 | 8 |
| 199 | 8990754 | 580797 | 1391768 | 178375 | 14380 | 22124 | 0.29 | 0.21 | 0.39 | 323418 | 25447 | 411037 |
| 1 |  | 9 | 9 | 1 | 91 | 93 | 1 | 6 | 3 | 9 | 73 | 8 |


| Yea <br> r | R(age 1) | Low | High | SSB | Low | High | Fba <br> r <br> (3- <br> 7) | Low | $\mathrm{Hig}$ h | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 199 2 | 6672861 | $\begin{aligned} & 436724 \\ & 3 \end{aligned}$ | $\begin{aligned} & 1019569 \\ & 4 \end{aligned}$ | $246107$ | $\begin{aligned} & 19583 \\ & 96 \end{aligned}$ | $\begin{aligned} & 30927 \\ & 71 \end{aligned}$ | 0.23 3 | 0.17 3 | $\begin{aligned} & 0.31 \\ & 4 \end{aligned}$ | $\begin{aligned} & 353003 \\ & 5 \end{aligned}$ | $\begin{aligned} & 28118 \\ & 33 \end{aligned}$ | $\begin{aligned} & 443168 \\ & 1 \end{aligned}$ |
| 199 3 | 4894647 | $\begin{aligned} & 316831 \\ & 8 \end{aligned}$ | 7561606 | $\begin{aligned} & 253906 \\ & 7 \end{aligned}$ | $\begin{aligned} & 20295 \\ & 21 \end{aligned}$ | $\begin{aligned} & 31765 \\ & 45 \end{aligned}$ | 0.20 4 | $\begin{aligned} & 0.15 \\ & 2 \end{aligned}$ | $\begin{aligned} & 0.27 \\ & 5 \end{aligned}$ | $\begin{aligned} & 341131 \\ & 7 \end{aligned}$ | $\begin{aligned} & 27452 \\ & 55 \end{aligned}$ | $\begin{aligned} & 423898 \\ & 1 \end{aligned}$ |
| 199 4 | 8099362 | $\begin{aligned} & 528944 \\ & 5 \end{aligned}$ | $\begin{aligned} & 1240199 \\ & 4 \end{aligned}$ | $\begin{aligned} & 253022 \\ & 1 \end{aligned}$ | $\begin{aligned} & 20441 \\ & 10 \end{aligned}$ | $\begin{aligned} & 31319 \\ & 34 \end{aligned}$ | 0.18 4 | $\begin{aligned} & 0.13 \\ & 6 \end{aligned}$ | 0.24 8 | $\begin{aligned} & 340859 \\ & 6 \end{aligned}$ | $\begin{aligned} & 27787 \\ & 48 \end{aligned}$ | $\begin{aligned} & 418120 \\ & 9 \end{aligned}$ |
| 199 5 | 9198605 | $\begin{aligned} & 607212 \\ & 0 \end{aligned}$ | $\begin{aligned} & 1393489 \\ & 1 \end{aligned}$ | $\begin{aligned} & 230471 \\ & 4 \end{aligned}$ | $\begin{aligned} & 19033 \\ & 50 \end{aligned}$ | $\begin{aligned} & 27907 \\ & 15 \end{aligned}$ | $\begin{aligned} & 0.24 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0.18 \\ & 3 \end{aligned}$ | $\begin{aligned} & 0.31 \\ & 9 \end{aligned}$ | $\begin{aligned} & 334275 \\ & 9 \end{aligned}$ | $\begin{aligned} & 27621 \\ & 81 \end{aligned}$ | $\begin{aligned} & 404536 \\ & 9 \end{aligned}$ |
| 199 6 | $\begin{aligned} & 2826655 \\ & 6 \end{aligned}$ | $\begin{aligned} & 186901 \\ & 53 \end{aligned}$ | $\begin{aligned} & 4274968 \\ & 8 \end{aligned}$ | $\begin{aligned} & 220531 \\ & 2 \end{aligned}$ | $\begin{aligned} & 18386 \\ & 29 \end{aligned}$ | $\begin{aligned} & 26451 \\ & 23 \end{aligned}$ | $\begin{aligned} & 0.29 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0.22 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.38 \\ & 9 \end{aligned}$ | $\begin{aligned} & 372885 \\ & 7 \end{aligned}$ | $\begin{aligned} & 30436 \\ & 50 \end{aligned}$ | $\begin{aligned} & 456832 \\ & 2 \end{aligned}$ |
| 199 7 | $\begin{aligned} & 4554457 \\ & 1 \end{aligned}$ | $\begin{aligned} & 301967 \\ & 97 \end{aligned}$ | $\begin{aligned} & 6869297 \\ & 9 \end{aligned}$ | $\begin{aligned} & 246935 \\ & 3 \end{aligned}$ | $\begin{aligned} & 20543 \\ & 31 \end{aligned}$ | $\begin{aligned} & 29682 \\ & 20 \end{aligned}$ | $\begin{aligned} & 0.29 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.22 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.39 \\ & 1 \end{aligned}$ | $\begin{aligned} & 547122 \\ & 7 \end{aligned}$ | $\begin{aligned} & 43172 \\ & 44 \end{aligned}$ | $\begin{aligned} & 693366 \\ & 5 \end{aligned}$ |
| 199 8 | $\begin{aligned} & 2660511 \\ & 9 \end{aligned}$ | $\begin{aligned} & 177287 \\ & 79 \end{aligned}$ | $\begin{aligned} & 3992561 \\ & 5 \end{aligned}$ | $\begin{aligned} & 368283 \\ & 8 \end{aligned}$ | $\begin{aligned} & 30223 \\ & 28 \end{aligned}$ | $\begin{aligned} & 44876 \\ & 99 \end{aligned}$ | 0.40 4 | $\begin{aligned} & 0.31 \\ & 3 \end{aligned}$ | $\begin{aligned} & 0.52 \\ & 2 \end{aligned}$ | $\begin{aligned} & 682878 \\ & 3 \end{aligned}$ | $\begin{aligned} & 54828 \\ & 56 \end{aligned}$ | $\begin{aligned} & 850510 \\ & 7 \end{aligned}$ |
| 199 9 | $\begin{aligned} & 2005874 \\ & 5 \end{aligned}$ | $\begin{aligned} & 133175 \\ & 35 \end{aligned}$ | $\begin{aligned} & 3021229 \\ & 2 \end{aligned}$ | $\begin{aligned} & 443409 \\ & 9 \end{aligned}$ | $\begin{aligned} & 36265 \\ & 79 \end{aligned}$ | $\begin{aligned} & 54214 \\ & 28 \end{aligned}$ | $\begin{aligned} & 0.38 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.30 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0.50 \\ & 3 \end{aligned}$ | $\begin{aligned} & 714399 \\ & 2 \end{aligned}$ | $\begin{aligned} & 58329 \\ & 94 \end{aligned}$ | $\begin{aligned} & 874964 \\ & 2 \end{aligned}$ |
| 200 0 | $\begin{aligned} & 3947357 \\ & 8 \end{aligned}$ | $\begin{aligned} & 261148 \\ & 54 \end{aligned}$ | $\begin{aligned} & 5966578 \\ & 7 \end{aligned}$ | $\begin{aligned} & 422297 \\ & 3 \end{aligned}$ | $\begin{aligned} & 35201 \\ & 88 \end{aligned}$ | $\begin{aligned} & 50660 \\ & 65 \end{aligned}$ | 0.47 8 | $\begin{aligned} & 0.37 \\ & 3 \end{aligned}$ | $\begin{aligned} & 0.61 \\ & 2 \end{aligned}$ | $\begin{aligned} & 745750 \\ & 4 \end{aligned}$ | $\begin{aligned} & 60995 \\ & 69 \end{aligned}$ | $\begin{aligned} & 911775 \\ & 4 \end{aligned}$ |
| 200 1 | $\begin{aligned} & 5630687 \\ & 5 \end{aligned}$ | $\begin{aligned} & 375812 \\ & 10 \end{aligned}$ | $\begin{aligned} & 8436301 \\ & 4 \end{aligned}$ | $\begin{aligned} & 456986 \\ & 5 \end{aligned}$ | $\begin{aligned} & 38241 \\ & 37 \end{aligned}$ | $\begin{aligned} & 54610 \\ & 15 \end{aligned}$ | $\begin{aligned} & 0.46 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0.36 \\ & 4 \end{aligned}$ | $\begin{aligned} & 0.59 \\ & 9 \end{aligned}$ | $\begin{aligned} & 902780 \\ & 7 \end{aligned}$ | $\begin{aligned} & 73084 \\ & 40 \end{aligned}$ | $\begin{aligned} & 111516 \\ & 69 \end{aligned}$ |
| 200 2 | $\begin{aligned} & 4890324 \\ & 0 \end{aligned}$ | $\begin{aligned} & 326299 \\ & 77 \end{aligned}$ | $\begin{aligned} & 7329232 \\ & 5 \end{aligned}$ | $\begin{aligned} & 540545 \\ & 1 \end{aligned}$ | $\begin{aligned} & 45164 \\ & 18 \end{aligned}$ | $\begin{aligned} & 64694 \\ & 87 \end{aligned}$ | 0.47 2 | $\begin{aligned} & 0.36 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0.60 \\ & 7 \end{aligned}$ | $\begin{aligned} & 103444 \\ & 00 \end{aligned}$ | $\begin{aligned} & 84093 \\ & 24 \end{aligned}$ | $\begin{aligned} & 127247 \\ & 58 \end{aligned}$ |
| 200 3 | $\begin{aligned} & 5280599 \\ & 4 \end{aligned}$ | $\begin{aligned} & 357381 \\ & 39 \end{aligned}$ | $\begin{aligned} & 7802513 \\ & 1 \end{aligned}$ | $\begin{aligned} & 686560 \\ & 1 \end{aligned}$ | $\begin{aligned} & 57206 \\ & 99 \end{aligned}$ | $\begin{aligned} & 82396 \\ & 37 \end{aligned}$ | 0.50 2 | $\begin{aligned} & 0.39 \\ & 6 \end{aligned}$ | 0.63 7 | $\begin{aligned} & 118350 \\ & 00 \end{aligned}$ | $\begin{aligned} & 97486 \\ & 35 \end{aligned}$ | $\begin{aligned} & 143678 \\ & 81 \end{aligned}$ |
| 200 4 | $\begin{aligned} & 2872541 \\ & 0 \end{aligned}$ | $\begin{aligned} & 194269 \\ & 22 \end{aligned}$ | $\begin{aligned} & 4247452 \\ & 0 \end{aligned}$ | $\begin{aligned} & 675721 \\ & 4 \end{aligned}$ | $\begin{aligned} & 56980 \\ & 44 \end{aligned}$ | $\begin{aligned} & 80132 \\ & 65 \end{aligned}$ | 0.54 0 | 0.42 8 | $\begin{aligned} & 0.68 \\ & 1 \end{aligned}$ | $\begin{aligned} & 103590 \\ & 31 \end{aligned}$ | $\begin{aligned} & 86947 \\ & 42 \end{aligned}$ | $\begin{aligned} & 123418 \\ & 88 \end{aligned}$ |
| 200 5 | $\begin{aligned} & 2239224 \\ & 3 \end{aligned}$ | $\begin{aligned} & 151684 \\ & 75 \end{aligned}$ | $\begin{aligned} & 3305622 \\ & 6 \end{aligned}$ | $\begin{aligned} & 601333 \\ & 7 \end{aligned}$ | $\begin{aligned} & 50811 \\ & 92 \end{aligned}$ | $\begin{aligned} & 71164 \\ & 85 \end{aligned}$ | 0.50 4 | 0.39 7 | 0.64 0 | $\begin{aligned} & 849232 \\ & 9 \end{aligned}$ | $\begin{aligned} & 71652 \\ & 56 \end{aligned}$ | $\begin{aligned} & 100651 \\ & 88 \end{aligned}$ |
| 200 6 | 8996712 | $\begin{aligned} & 602916 \\ & 8 \end{aligned}$ | $\begin{aligned} & 1342487 \\ & 4 \end{aligned}$ | $\begin{aligned} & 588879 \\ & 0 \end{aligned}$ | $\begin{aligned} & 49606 \\ & 96 \end{aligned}$ | $\begin{aligned} & 69905 \\ & 22 \end{aligned}$ | $\begin{aligned} & 0.45 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.35 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.58 \\ & 7 \end{aligned}$ | $\begin{aligned} & 773039 \\ & 1 \end{aligned}$ | $\begin{aligned} & 65183 \\ & 75 \end{aligned}$ | $\begin{aligned} & 916776 \\ & 7 \end{aligned}$ |
| 200 7 | 4795411 | $\begin{aligned} & 321072 \\ & 0 \end{aligned}$ | 7162245 | $\begin{aligned} & 466584 \\ & 3 \end{aligned}$ | $\begin{aligned} & 39191 \\ & 53 \end{aligned}$ | $\begin{aligned} & 55547 \\ & 95 \end{aligned}$ | $\begin{aligned} & 0.46 \\ & 0 \end{aligned}$ | $\begin{aligned} & 0.35 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.59 \\ & 4 \end{aligned}$ | $\begin{aligned} & 569443 \\ & 1 \end{aligned}$ | $\begin{aligned} & 47949 \\ & 95 \end{aligned}$ | $\begin{aligned} & 676258 \\ & 1 \end{aligned}$ |
| 200 8 | 5770397 | $\begin{aligned} & 381186 \\ & 5 \end{aligned}$ | 8735220 | $\begin{aligned} & 357864 \\ & 3 \end{aligned}$ | $\begin{aligned} & 29660 \\ & 20 \end{aligned}$ | $\begin{aligned} & 43178 \\ & 02 \end{aligned}$ | $\begin{aligned} & 0.40 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.30 \\ & 6 \end{aligned}$ | $\begin{aligned} & 0.53 \\ & 7 \end{aligned}$ | $\begin{aligned} & 438884 \\ & 9 \end{aligned}$ | $\begin{aligned} & 36548 \\ & 92 \end{aligned}$ | $\begin{aligned} & 527019 \\ & 5 \end{aligned}$ |
| 200 9 | 5610707 | $\begin{aligned} & 359888 \\ & 4 \end{aligned}$ | 8747165 | $\begin{aligned} & 273582 \\ & 2 \end{aligned}$ | $\begin{aligned} & 22130 \\ & 93 \end{aligned}$ | $\begin{aligned} & 33820 \\ & 19 \end{aligned}$ | $\begin{aligned} & 0.26 \\ & 4 \end{aligned}$ | $\begin{aligned} & 0.19 \\ & 4 \end{aligned}$ | $\begin{aligned} & 0.35 \\ & 9 \end{aligned}$ | $\begin{aligned} & 343932 \\ & 4 \end{aligned}$ | $\begin{aligned} & 28037 \\ & 07 \end{aligned}$ | $\begin{aligned} & 421903 \\ & 9 \end{aligned}$ |
| 201 0 | $\begin{aligned} & 1521667 \\ & 9 \end{aligned}$ | $\begin{aligned} & 999626 \\ & 9 \end{aligned}$ | $\begin{aligned} & 2316337 \\ & 5 \end{aligned}$ | 266579 7 | $\begin{aligned} & 21167 \\ & 73 \end{aligned}$ | $\begin{aligned} & 33572 \\ & 19 \end{aligned}$ | 0.18 4 | 0.13 2 | 0.25 6 | $\begin{aligned} & 372947 \\ & 5 \end{aligned}$ | $\begin{aligned} & 29896 \\ & 65 \end{aligned}$ | $\begin{aligned} & 465235 \\ & 6 \end{aligned}$ |


| Yea r | R(age 1) | Low | High | SSB | Low | High | Fba <br> r <br> (3- <br> 7) | Low | $\mathrm{Hig}$ h | TSB | Low | High |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 201 1 | $\begin{aligned} & 1932494 \\ & 7 \end{aligned}$ | $\begin{aligned} & 128171 \\ & 12 \end{aligned}$ | $\begin{aligned} & 2913710 \\ & 9 \end{aligned}$ | $268713$ | $\begin{aligned} & 21494 \\ & 71 \end{aligned}$ | $\begin{aligned} & 33592 \\ & 77 \end{aligned}$ |  |  |  | $440752$ | $\begin{aligned} & 35291 \\ & 58 \end{aligned}$ | $\begin{aligned} & 550450 \\ & 1 \end{aligned}$ |
| 201 2 | $\begin{aligned} & 1933226 \\ & 6 \end{aligned}$ | $\begin{aligned} & 130514 \\ & 26 \end{aligned}$ | $\begin{aligned} & 2863568 \\ & 3 \end{aligned}$ | $\begin{aligned} & 342813 \\ & 8 \end{aligned}$ | $\begin{aligned} & 28131 \\ & 03 \end{aligned}$ | $\begin{aligned} & 41776 \\ & 40 \end{aligned}$ |  |  |  | $\begin{aligned} & 510017 \\ & 9 \end{aligned}$ | $\begin{aligned} & 41827 \\ & 29 \end{aligned}$ | $\begin{aligned} & 621886 \\ & 6 \end{aligned}$ |
| 201 3 | $\begin{aligned} & 1578435 \\ & 0 \end{aligned}$ | $\begin{aligned} & 106952 \\ & 54 \end{aligned}$ | $\begin{aligned} & 2329497 \\ & 8 \end{aligned}$ | $\begin{aligned} & 375426 \\ & 1 \end{aligned}$ | $\begin{aligned} & 31411 \\ & 04 \end{aligned}$ | $\begin{aligned} & 44871 \\ & 09 \end{aligned}$ |  |  |  | $\begin{aligned} & 555946 \\ & 3 \end{aligned}$ | $\begin{aligned} & 46355 \\ & 90 \end{aligned}$ | $\begin{aligned} & 666746 \\ & 3 \end{aligned}$ |
| 201 4 | $\begin{aligned} & 3698410 \\ & 8 \end{aligned}$ | $\begin{aligned} & 248588 \\ & 89 \end{aligned}$ | $\begin{aligned} & 5502354 \\ & 6 \end{aligned}$ | $\begin{aligned} & 399114 \\ & 6 \end{aligned}$ | $\begin{aligned} & 33792 \\ & 63 \end{aligned}$ | $\begin{aligned} & 47138 \\ & 23 \end{aligned}$ | $\begin{aligned} & 0.37 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.29 \\ & 3 \end{aligned}$ | $\begin{aligned} & 0.49 \\ & 0 \end{aligned}$ | $\begin{aligned} & 660456 \\ & 2 \end{aligned}$ | $\begin{aligned} & 54885 \\ & 38 \end{aligned}$ | $\begin{aligned} & 794751 \\ & 5 \end{aligned}$ |
| 201 5 | $\begin{aligned} & 6418715 \\ & 6 \end{aligned}$ | $\begin{aligned} & 434808 \\ & 22 \end{aligned}$ | $\begin{aligned} & 9475421 \\ & 0 \end{aligned}$ | $\begin{aligned} & 415757 \\ & 8 \end{aligned}$ | $\begin{aligned} & 35173 \\ & 03 \end{aligned}$ | $\begin{aligned} & 49144 \\ & 05 \end{aligned}$ | $\begin{aligned} & 0.49 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.39 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0.63 \\ & 6 \end{aligned}$ | $\begin{aligned} & 813452 \\ & 8 \end{aligned}$ | $\begin{aligned} & 66298 \\ & 40 \end{aligned}$ | $\begin{aligned} & 998071 \\ & 4 \end{aligned}$ |
| 201 6 | $\begin{aligned} & 3465240 \\ & 5 \end{aligned}$ | $\begin{aligned} & 235374 \\ & 36 \end{aligned}$ | $\begin{aligned} & 5101614 \\ & 1 \end{aligned}$ | $\begin{aligned} & 488626 \\ & 3 \end{aligned}$ | $\begin{aligned} & 40709 \\ & 05 \end{aligned}$ | $\begin{aligned} & 58649 \\ & 30 \end{aligned}$ |  | $\begin{aligned} & 0.34 \\ & 4 \end{aligned}$ |  | $\begin{aligned} & 906462 \\ & 8 \end{aligned}$ | $\begin{aligned} & 73908 \\ & 61 \end{aligned}$ | $\begin{aligned} & 111174 \\ & 42 \end{aligned}$ |
| 201 7 | $\begin{aligned} & 1158735 \\ & 4 \end{aligned}$ | $\begin{aligned} & 779095 \\ & 5 \end{aligned}$ | $\begin{aligned} & 1723367 \\ & 5 \end{aligned}$ | $\begin{aligned} & 603018 \\ & 6 \end{aligned}$ | $\begin{aligned} & 49921 \\ & 40 \end{aligned}$ | $\begin{aligned} & 72840 \\ & 78 \end{aligned}$ | $\begin{aligned} & 0.42 \\ & 3 \end{aligned}$ | $\begin{aligned} & 0.33 \\ & 0 \end{aligned}$ | $\begin{aligned} & 0.54 \\ & 2 \end{aligned}$ | $\begin{aligned} & 871243 \\ & 2 \end{aligned}$ | $\begin{aligned} & 72057 \\ & 61 \end{aligned}$ | $\begin{aligned} & 105341 \\ & 36 \end{aligned}$ |
| 201 8 | $\begin{aligned} & 1222721 \\ & 7 \end{aligned}$ | $\begin{aligned} & 823991 \\ & 1 \end{aligned}$ | $\begin{aligned} & 1814398 \\ & 6 \end{aligned}$ | $\begin{aligned} & 587096 \\ & 3 \end{aligned}$ | $\begin{aligned} & 48851 \\ & 16 \end{aligned}$ | $\begin{aligned} & 70557 \\ & 62 \end{aligned}$ | $\begin{aligned} & 0.43 \\ & 2 \end{aligned}$ | $\begin{aligned} & 0.33 \\ & 6 \end{aligned}$ |  | $\begin{aligned} & 776412 \\ & 7 \end{aligned}$ | $\begin{aligned} & 64830 \\ & 86 \end{aligned}$ | $\begin{aligned} & 929829 \\ & 9 \end{aligned}$ |
| 201 9 | $\begin{aligned} & 1371762 \\ & 1 \end{aligned}$ | $\begin{aligned} & 899182 \\ & 5 \end{aligned}$ | $\begin{aligned} & 2092713 \\ & 4 \end{aligned}$ | $\begin{aligned} & 502857 \\ & 3 \end{aligned}$ | $\begin{aligned} & 41936 \\ & 76 \end{aligned}$ | $\begin{aligned} & 60296 \\ & 85 \end{aligned}$ | $\begin{aligned} & 0.40 \\ & 0 \end{aligned}$ | $\begin{aligned} & 0.30 \\ & 7 \end{aligned}$ | $\begin{aligned} & 0.52 \\ & 0 \end{aligned}$ | $\begin{aligned} & 689043 \\ & 0 \end{aligned}$ | $\begin{aligned} & 57519 \\ & 63 \end{aligned}$ | $\begin{aligned} & 825422 \\ & 9 \end{aligned}$ |
| 202 0 | $\begin{aligned} & 2389531 \\ & 8 \end{aligned}$ | $\begin{aligned} & 152882 \\ & 65 \end{aligned}$ | $\begin{aligned} & 3734800 \\ & 5 \end{aligned}$ | $\begin{aligned} & 414696 \\ & 4 \end{aligned}$ | $\begin{aligned} & 34365 \\ & 47 \end{aligned}$ | $\begin{aligned} & 50042 \\ & 41 \end{aligned}$ | $\begin{aligned} & 0.45 \\ & 4 \end{aligned}$ | $\begin{aligned} & 0.34 \\ & 5 \end{aligned}$ | $\begin{aligned} & 0.59 \\ & 6 \end{aligned}$ | $\begin{aligned} & 644057 \\ & 0 \end{aligned}$ | $\begin{aligned} & 52271 \\ & 27 \end{aligned}$ | $\begin{aligned} & 793570 \\ & 7 \end{aligned}$ |
| 202 1 | $\begin{aligned} & 6963057 \\ & 5 \end{aligned}$ | $\begin{aligned} & 427265 \\ & 73 \end{aligned}$ | $\begin{aligned} & 1134754 \\ & 44 \end{aligned}$ | $\begin{aligned} & 396372 \\ & 6 \end{aligned}$ | $\begin{aligned} & 31666 \\ & 61 \end{aligned}$ | $\begin{aligned} & 49614 \\ & 17 \end{aligned}$ | $\begin{aligned} & 0.40 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.30 \\ & 1 \end{aligned}$ | $\begin{aligned} & 0.55 \\ & 7 \end{aligned}$ | $\begin{aligned} & 877310 \\ & 0 \end{aligned}$ | $\begin{aligned} & 64308 \\ & 44 \end{aligned}$ | $\begin{aligned} & 119684 \\ & 58 \end{aligned}$ |
| 202 2 | $\begin{aligned} & 6891160 \\ & 6 \end{aligned}$ | $\begin{aligned} & 391446 \\ & 19 \end{aligned}$ | $\begin{aligned} & 1213144 \\ & 89 \end{aligned}$ | $\begin{aligned} & 457998 \\ & 3 \end{aligned}$ | $\begin{aligned} & 33595 \\ & 98 \end{aligned}$ | $\begin{aligned} & 62436 \\ & 78 \end{aligned}$ | $\begin{aligned} & 0.40 \\ & 9 \end{aligned}$ | $\begin{aligned} & 0.27 \\ & 8 \end{aligned}$ | $\begin{aligned} & 0.60 \\ & 1 \end{aligned}$ | $\begin{aligned} & 103333 \\ & 19 \end{aligned}$ | $\begin{aligned} & 70916 \\ & 84 \end{aligned}$ | $\begin{aligned} & 150567 \\ & 17 \end{aligned}$ |
| 202 3 | $\begin{aligned} & 1665774 \\ & 1 \end{aligned}$ | $\begin{aligned} & 725164 \\ & 3 \end{aligned}$ | $\begin{aligned} & 3826447 \\ & 6 \end{aligned}$ | $\begin{aligned} & 623619 \\ & 1 \end{aligned}$ | $\begin{aligned} & 41623 \\ & 99 \end{aligned}$ | $\begin{aligned} & 93431 \\ & 88 \end{aligned}$ | $\begin{aligned} & 0.51 \\ & 5 \end{aligned}$ | 0.30 7 | 0.86 5 | $\begin{aligned} & 995475 \\ & 9 \end{aligned}$ | $\begin{aligned} & 64707 \\ & 27 \end{aligned}$ | $\begin{aligned} & 153146 \\ & 98 \end{aligned}$ |
| 202 4 | $\begin{aligned} & 2299774 \\ & 9^{*} \end{aligned}$ |  |  | $\begin{aligned} & 679998 \\ & 5^{\wedge} \end{aligned}$ |  |  |  |  |  |  |  |  |

[^3]$\wedge$ ^SSB calculated from the survivors age 2-10 and GM(1996-2022) recruitment in 2024

Table 2.4.6. Blue whiting. Model estimate of total catch weight (in tonnes) and Sum of Product of catch number and mean weight at age for ages 1-10+ (Observed catch). Preliminary catch data for 2022 are included.

| Year | Estimate | Low | High | SOP <br> catch |
| :---: | :---: | :---: | :---: | :---: |
| 1981 | 788159 | 568629 | 1092442 | 922980 |
| 1982 | 544325 | 415970 | 712287 | 550643 |
| 1983 | 513140 | 398675 | 660470 | 553344 |
| 1984 | 564938 | 438740 | 727436 | 615569 |
| 1985 | 639830 | 504897 | 810823 | 678214 |
| 1986 | 759908 | 599871 | 962641 | 847145 |
| 1987 | 638144 | 504074 | 807873 | 654718 |
| 1988 | 569734 | 450658 | 720273 | 552264 |
| 1989 | 618654 | 492608 | 776953 | 630316 |
| 1990 | 554539 | 438604 | 701117 | 558128 |
| 1991 | 408698 | 319299 | 523127 | 364008 |
| 1992 | 438842 | 347459 | 554257 | 474592 |
| 1993 | 439924 | 346651 | 558294 | 475198 |
| 1994 | 424170 | 332565 | 541008 | 457696 |
| 1995 | 507570 | 404279 | 637251 | 505176 |
| 1996 | 597406 | 475912 | 749915 | 621104 |
| 1997 | 641969 | 507116 | 812682 | 639681 |
| 1998 | 1080685 | 849227 | 1375226 | 1131955 |
| 1999 | 1247277 | 975503 | 1594767 | 1261033 |
| 2000 | 1502683 | 1183719 | 1907595 | 1412449 |
| 2001 | 1560764 | 1229375 | 1981480 | 1771805 |
| 2002 | 1710847 | 1348172 | 2171086 | 1556955 |
| 2003 | 2203545 | 1744202 | 2783857 | 2365319 |
| 2004 | 2317253 | 1841339 | 2916173 | 2400795 |
| 2005 | 1996202 | 1588667 | 2508281 | 2018344 |
| 2006 | 1856505 | 1477663 | 2332475 | 1956239 |
| 2007 | 1560030 | 1239967 | 1962708 | 1612269 |


| Year | Estimate | Low | High | SOP <br> catch |
| :---: | :---: | :---: | :---: | :---: |
| 2008 | 1168540 | 921892 | 1481176 | 1251851 |
| 2009 | 656960 | 517335 | 834270 | 634978 |
| 2010 | 477260 | 370094 | 615457 | 539539 |
| 2011 | 136359 | 101155 | 183814 | 103771 |
| 2012 | 325859 | 259332 | 409452 | 375692 |
| 2013 | 588691 | 467970 | 740553 | 613863 |
| 2014 | 1102584 | 872097 | 1393988 | 1147650 |
| 2015 | 1342585 | 1070110 | 1684438 | 1390656 |
| 2016 | 1243089 | 987995 | 1564046 | 1180786 |
| 2017 | 1479682 | 1174544 | 1864094 | 1555069 |
| 2018 | 1705935 | 1347983 | 2158940 | 1709856 |
| 2019 | 1538458 | 1214042 | 1949564 | 1512026 |
| 2020 | 1427786 | 1131907 | 1801008 | 1460507 |
| 2021 | 1148634 | 916696 | 1439255 | 1139531 |
| 2022 | 1085570 | 861688 | 1367621 | 1035891 |
| 2023 | 1641728 | 1276219 | 2111918 | 1672378 |

Table 2.8.2.1.1. Blue whiting. Input to short-term projection (median values for exploitation pattern and stock numbers). Mean weight in 2024+ is the average weight for 2021-2023.

| Age | Mean weight in <br> the stock and catch <br> (kg) in 2023 | Mean weight in the <br> stock and catch (kg) <br> in 2024+ | Proportion <br> mature | Natural mor- <br> tality | Exploitation <br> pattern | Stock number <br> (2024) <br> (thousands) |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | 0.045 | 0.050 | 0.11 | 0.20 | 0.077 | 22997749 |
| 2 | 0.061 | 0.072 | 0.40 | 0.20 | 0.150 | 13109448 |
| 3 | 0.076 | 0.087 | 0.82 | 0.20 | 0.428 | 46437229 |
| 4 | 0.092 | 0.106 | 0.91 | 0.20 | 0.806 | 30950190 |
| 5 | 0.120 | 0.144 | 0.141 | 0.94 | 0.20 | 1.137 |
| 7 | 0.155 | 0.149 | 1.00 | 0.20 | 1.398 | 4676343 |
| 8 | 0.169 | 0.162 |  | 0.20 | 1.383 | 1455027 |


| Age | Mean weight in <br> the stock and catch <br> (kg) in 2023 | Mean weight in the <br> stock and catch (kg) <br> in 2024+ | Proportion <br> mature | Natural mor- <br> tality | Exploitation <br> pattern | Stock number <br> (2024) <br> (thousands) |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 9 | 0.182 | 0.173 | 1.00 | 0.20 | 1.345 | 322198 |
| $10+$ | 0.207 | 0.201 | 1.00 | 0.20 | 1.345 | 429264 |

Table 2.8.2.1.2. Blue whiting. Deterministic forecast, intermediate year assumptions and recruitments.

| Variable | Value | Notes |
| :---: | :---: | :---: |
| $\mathrm{F}_{\text {ages 3-7 }}$ (2023) | 0.515 | From the assessment (based on assumed 2023 catches) |
| SSB (2024) | 6799985 | From the forecast; in tonnes |
| $\mathrm{R}_{\text {age } 1}$ (2023) | 16657741 | From the assessment; in thousands |
| $\mathrm{R}_{\text {age } 1}$ (2024-2025) | 22997749 | GM (1996-2022); in thousands |
| Total catch (2023) | 1672378 | As estimated by ICES, based on declared national quotas and expected uptake; in tonnes (see also Table 2.3.2.3) |

Table 2.8.2.2.1. Blue whiting. Deterministic forecast (weights in tonnes).

| Basis | Catch(2024) | F(2024) | SSB(2025) | \% SSB <br> change* | \% Catch <br> change** | \% Advice <br> change*** |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Long-term management strategy <br> (F=FMSY) | 1529754 | 0.320 | 7258384 | 6.7 | -8.5 | 12.5 |
| MSY approach: FMSY | 1529754 | 0.320 | 7258384 | 6.7 | -8.5 | 12.5 |
| F = 0 | 5 | 0.000 | 8709263 | 28 | -100 | -100 |
| Fpa | 1529754 | 0.320 | 7258384 | 6.7 | -8.5 | 12.5 |
| Flim | 3507334 | 0.880 | 5411744 | -20 | 110 | 158 |
| SSB (2025) = Blim | 7997215 | 4.455 | 1500006 | -78 | 378 | 488 |
| SSB (2025 = Bpa | 7066101 | 3.043 | 2250000 | -67 | 323 | 420 |
| SSB (2025) = MSY Btrigger | 7066101 | 3.043 | 2250000 | -67 | 323 | 420 |
| F = F (2023) | 2306302 | 0.515 | 6528681 | -4.0 | 38 | 70 |
| SSB (2025) = SSB (2024) | 2016931 | 0.439 | 6799982 | 0 | 21 | 48 |
| Catch (2024) = Catch (2023) | 1672363 | 0.354 | 7123994 | 4.8 | 0 | 23 |


| Basis | Catch(2024) | F(2024) | SSB(2025) | $\begin{aligned} & \text { \% SSB } \\ & \text { change* } \end{aligned}$ | \% Catch change** | \% Advice change*** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $F=0.05$ | 263038 | 0.050 | 8458687 | 24 | -84 | -81 |
| $F=0.10$ | 516613 | 0.100 | 8217528 | 21 | -69 | -62 |
| $F=0.15$ | 761129 | 0.150 | 7985381 | 17.4 | -55 | -44 |
| $F=0.16$ | 808979 | 0.160 | 7939998 | 16.8 | -52 | -41 |
| $F=0.17$ | 856485 | 0.170 | 7894957 | 16.1 | -49 | -37 |
| $F=0.18$ | 903651 | 0.180 | 7850255 | 15.4 | -46 | -34 |
| $F=0.19$ | 950478 | 0.190 | 7805889 | 14.8 | -43 | -30 |
| $F=0.20$ | 996970 | 0.200 | 7761855 | 14.1 | -40 | -27 |
| $F=0.21$ | 1043130 | 0.210 | 7718152 | 13.5 | -38 | -23 |
| $F=0.22$ | 1088960 | 0.220 | 7674775 | 12.9 | -35 | -19.9 |
| $F=0.23$ | 1134464 | 0.230 | 7631723 | 12.2 | -32 | -16.6 |
| $F=0.24$ | 1179644 | 0.240 | 7588992 | 11.6 | -30 | -13.2 |
| $F=0.25$ | 1224503 | 0.250 | 7546580 | 11 | -27 | -9.9 |
| $F=0.26$ | 1269044 | 0.260 | 7504483 | 10.4 | -24 | -6.7 |
| $F=0.27$ | 1313269 | 0.270 | 7462700 | 9.7 | -22 | -3.4 |
| $F=0.28$ | 1357181 | 0.280 | 7421227 | 9.1 | -18.8 | -0.2 |
| $F=0.29$ | 1400783 | 0.290 | 7380061 | 8.5 | -16.2 | 3.0 |
| $F=0.30$ | 1444078 | 0.300 | 7339201 | 7.9 | -13.7 | 6.2 |
| $F=0.31$ | 1487067 | 0.310 | 7298643 | 7.3 | -11.1 | 9.4 |
| $F=0.32$ | 1529754 | 0.320 | 7258384 | 6.7 | -8.5 | 12.5 |
| $F=0.33$ | 1572141 | 0.330 | 7218423 | 6.2 | -6.0 | 15.6 |
| $F=0.34$ | 1614231 | 0.340 | 7178756 | 5.6 | -3.5 | 18.7 |
| $F=0.35$ | 1656026 | 0.350 | 7139381 | 5 | -1.0 | 22 |
| $F=0.45$ | 2058300 | 0.450 | 6761149 | -0.6 | 23 | 51 |
| $F=0.50$ | 2249216 | 0.500 | 6582141 | -3.2 | 35 | 65 |

* SSB 2025 relative to SSB 2024.
** Catch 2024 relative to expected catch in 2023 (1 672378 tonnes).
*** Catch 2024 relative to advice for 2023 (1 359629 tonnes).

Table 2.8.2.3.1. Blue whiting. Catch numbers at age estimated by WGWIDE 2023 as a ratio of catch numbers at age estimated by WGWIDE 2023.

| Age | 2019 | 2020 | 2021 | 2022 | 2023 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.90 | 0.89 | 0.97 | 1.59 | 0.39 |
| 2 | 0.93 | 0.88 | 0.89 | 0.94 | 1.79 |
| 3 | 0.94 | 0.91 | 0.85 | 0.88 | 0.93 |
| 4 | 0.95 | 0.92 | 0.87 | 0.76 | 0.89 |
| 5 | 0.96 | 0.93 | 0.90 | 0.80 | 0.69 |
| 6 | 0.97 | 0.94 | 0.90 | 0.88 | 0.70 |
| 7 | 0.97 | 0.95 | 0.89 | 0.85 | 0.88 |
| 8 | 0.98 | 0.96 | 0.92 | 0.76 | 0.79 |
| 9 | 0.99 | 0.97 | 0.93 | 0.93 | 0.70 |
| 10 | 0.99 | 0.97 | 0.94 | 0.90 | 0.94 |

Table 2.8.2.3.2. Blue whiting. Catch numbers * weight at age estimated by WGWIDE 2023 as a ratio of catch numbers * weight at age estimated by WGWIDE 2023. Weight at age in the forecast in WGWIDE 2023 is the average weight at age of 2021-2023, whereas for WGWIDE 2022 this is 2020-2022.

| Age | 2019 | 2020 | 2021 | 2022 | 2023 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.90 | 0.89 | 0.97 | 1.80 | 0.36 |
| 2 | 0.93 | 0.88 | 0.89 | 0.98 | 1.63 |
| 3 | 0.94 | 0.91 | 0.85 | 0.90 | 0.86 |
| 4 | 0.95 | 0.92 | 0.87 | 0.78 | 0.84 |
| 5 | 0.96 | 0.93 | 0.90 | 0.82 | 0.68 |
| 6 | 0.97 | 0.94 | 0.90 | 0.88 | 0.72 |
| 7 | 0.97 | 0.95 | 0.89 | 0.86 | 0.91 |
| 8 | 0.98 | 0.96 | 0.92 | 0.73 | 0.79 |
| 9 | 0.99 | 0.97 | 0.93 | 0.94 | 0.71 |
| 10 | 0.99 | 0.97 | 0.94 | 0.85 | 0.92 |

Table 2.8.2.3.2. Blue whiting. Deterministic forecast, intermediate year assumptions and recruitments used in current and previous assessment.

|  |  | WGWIDE 2023 |  | WGWIDE 2022 |
| :--- | :--- | :--- | :--- | :--- |
| F | Year | Value | Value |  |
| SSB | 2023 | 0.515 | 2022 | 0.371 |
| Assumed recruitment | 2023 | 6799985 | 2023 | 6621207 |
|  | 2024 | 16657741 | 2022 | 43220294 |
| Catch | $2024-2025$ | 22997749 | $2023-2024$ | 2237250 |

### 2.16 Figures



Figure 2.2.1. Blue whiting catches in 2022. The catches on the map constitute $99.5 \%$ of the ICES estimated catches. The 200 m and 1000 m depth contours are indicated in blue.

First quarter
371542 tonnes, $35.9 \%$


Third quarter 30396 tonnes, 2.9\%


Second quarter 551124 tonnes, $53.3 \%$


Fourth quarter
80467 tonnes, 7.8\%

> 10000 tonnes

- 1000-10000 tonnes
- 100-1000 tonnes
- 10-100 tonnes

Figure 2.2.2. Blue whiting catches per quarter 2022. The catches on the map constitute $99.5 \%$ of the ICES estimated catches. The $\mathbf{2 0 0} \mathbf{~ m}$ and $\mathbf{1 0 0 0} \mathbf{~ m}$ depth contours are indicated in blue.


Figure 2.3.1.1. Blue whiting. ICES estimated catches (' 1000 tonnes) in 2022 by ICES division and country.

A


B


Figure 2.3.1.2. Blue whiting.(A) ICES estimated catches (tonnes) of blue whiting by fishery subareas from 1988-2022 and $(B)$ the percentage contribution to the overall catch by fishery subarea over the same period.


Figure 2.3.1.3. Blue whiting. Distribution of 2022 ICES estimated catches (in percentage) by quarter.


Figure 2.3.1.4. Blue whiting. Distribution of 2022 ICES estimated catches (in percentage) by ICES division area.


Figure 2.3.1.5. Blue whiting. ICES estimated catches (' 1000 tonnes) in 2022 by country.


Figure 2.3.1.6. Blue whiting. Distribution of 2022 ICES estimate ed catches (' 1000 tonnes) by ICES division and by quarter.


Figure 2.3.1.1.1. Blue whiting. 2022 ICES catches (' 1000 tonnes) based on sampled or estimated distribution by ICES division.


Figure 2.3.1.2.1. Blue whiting. Mean length ( mm ) by age ( $0-10$ year), by quarter ( 1,2 ), by country for ICES division area 27.6.a. These data only comprises the 2022 ICES catch-at-age sampled estimates for ICES division 27.6.a.


Figure 2.3.1.2.2. Blue whiting. Catch-at-age numbers (CANUM) distribution by quarter and ICES division for 2022.


Figure 2.3.2.1. Blue whiting. 2023 ICES preliminary catches (' 1000 tonnes) (Quarter 1 + Quarter 2) based on sampled or estimated distribution by ICES division.


Figure 2.3.2.2 Preliminary and final estimates of catch at age number by age and year (2017-2022).


Figure 2.3.3.1. Blue whiting. Catch proportion at age, 1981-2023. Preliminary values for 2023 have been used.


Figure 2.3.3.2. Blue whiting. Age disaggregated catch (numbers) plotted on log scale. The labels for each panel indicate year classes. The grey dotted lines correspond to $Z=0.6$. Preliminary catch-at-age data for 2023 have been used.


Figure 2.3.4.1. Blue whiting. Mean catch (and stock) weight ( kg ) at age by year. Preliminary values for 2023 have been used


Figure 2.3.6.1.1. Blue whiting. (A) Estimate of total biomass from the International blue whiting spawning stock survey. The black dots and error bands are StoX estimates with $90 \%$ confidence intervals. (B) Internal consistency within the International blue whiting spawning stock survey. The upper left part of the plots shows the relationship between log index-at-age within a cohort. Linear regression line shows the best fit to the log-transformed indices. The lower-right part of the plots shows the correlation coefficient ( $r$ ) for the two ages plotted in that panel. The background colour of each panel is determined by the $r$ value, where red equates to $r=1$ and white to $r<0$.

## NO SURVEY

2020


2022
2021


2023

Figure 2.3.6.1.2. Blue whiting. NASC distribution of the blue whiting stock in the area to the west of the British Isles, spring 2020 (upper panel) to 2023 (lower panel).

| 2023 |   |
| :---: | :---: |
| 2022 |   |
| 2021 |   |
| 2020 | NO SURVEY |
| 2019 |  |

Figure 2.3.6.1.3. Blue whiting. Length (line) and age (bars) distribution of the blue whiting stock in the area to the west of the British Isles, spring 2019 (lower panel) to 2023 (upper panel). Spawning-stock biomass and numbers are given.


Figure 2.4.1.1. Blue Whiting. OSA (One Step Ahead) residuals (see Berg and Nielsen, 2016) from catch-at-age and the IBWSS survey 2004-2023 (no survey in 2010 and 2020). Red (lighter) bubbles show that the observed value is less than the expected value. Preliminary catch data for 2023 have been used.


Figure 2.4.1.2 Blue whiting. Joint sample residuals (Process errors) for stock number and F at age. Red (lighter) bubbles show that the observed value is less than the expected value. Preliminary catch data for 2023 have been used.

Process error expressed as deviations in mortality


Figure 2.4.1.3. Blue whiting. Process errors expressed as deviation in instantaneous mortality at age by age and year.


Figure 2.4.1.4. Blue whiting. The correlation matrix between ages for the catches and survey index of this year's assessment (WGWIDE 2023; left) and last year's assessment (WGWIDE 2022; right). Each ellipse represents the level curve of a bivariate normal distribution with the corresponding correlation. Hence, the sign of a correlation corresponds to the sign of the slope of the major ellipse axis. Increasingly darker shading is used for increasingly larger absolute correlations, while uncorrelated pairs of ages are depicted as circles with no shading. Preliminary catch data for 2023 have been used in this year's assessment and preliminary catch data for 2022 in last year's assessment.


Figure 2.4.1.5. Blue whiting. Exploitation pattern by 5-years' time blocks. Preliminary catch data for 2023 have been used.


Figure 2.4.1.6. Blue whiting. Retrospective analysis of recruitment (age 1), F, SSB (tonnes) and total catch using the SAM model. The 95\% confidence interval is shown for the most recent assessment.


Figure 2.4.1.7. Blue whiting. SAM final run: total catches, recruitment (age 1), F and SSB. The graphs show the median value and the $95 \%$ confidence interval. Catches for 2023 are preliminary (shaded).


Figure 2.4.2.1. Blue whiting. Retrospective in SSB for the alternative configuration of the SAM model.


Figure 2.8.1.1. Blue whiting young fish indices from five different surveys and recruitment index from the assessment, standardized by dividing each series by their mean. BarSea - Norwegian bottom-trawl survey in the Barents Sea, IESNS: International Ecosystem Survey in the Nordic Seas in May (1 and 2 is the age groups), IBWSS (Not updated in 2020): International Blue Whiting Spawning Stock survey (1 and 2 is the age groups), FO: the Faroese bottom-trawl surveys in spring, IS: the Icelandic bottom-trawl survey in spring, SAM: recruits from the assessment.


Figure 2.8.2.3.1. Blue whiting. Mean weight at age used in the forecast of current and last year's assessment. The mean is taken as the last three years of each assessment, including the terminal year. Weight-at-age values of the terminal year are based on preliminary catch data.


Figure 2.8.2.4.1 Blue whiting. Medium term projection under three F scenarios: Fmsy, F=F2023 and F=catches in 2023. Left-hand panel shows the catch (in tonnes), the middle panel shows the $F$ (per year) and the right-hand panel shows the trajectory of SSB up to 2028 (in tonnes).


Figure 2.9.1. Blue whiting. SAM final run: Comparison of the 2022 (light blue line) and 2023 (black line) stock assessments, shown with $95 \%$ confidence intervals. Preliminary catch data for 2023 have been used.


Figure 2.9.2. Blue whiting. Comparison of assessment runs with the configuration from WGWIDE 2022 with preliminary data for 2022 (black line) and the WGWIDE 2022 assessment with final data for 2022 (blue line).


Figure 2.9.3. Blue whiting. Comparison of the 2019-2023 assessments (historical retrospective).

## 3 Boarfish in Celtic Seas, English Channel, and Bay of Biscay)

## Capros aper in subareas 6-8 (boc.27.6-8)

### 3.1 Biological features and population parameters

The boarfish (Capros aper, Linnaeus) is a deep bodied, laterally compressed, pelagic shoaling species distributed from Norway to Senegal, including the Mediterranean, Azores, Canaries, Madeira and Great Meteor Seamount (Blanchard \& Vandermeirsch 2005).

Mean weight-at-age was obtained from the ageing studies of Hüssy et al. (2012b). These mean weights are presented in the text table below. The variation in weight-at-age is due to the small sample size and the seasonal variation in weight and maturity stage.

| Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mean <br> Weight (g) | 0.84 | 6.65 | 14.6 | 19.5 | 23.7 | 26.8 | 33.3 | 37.7 | 40 | 47.1 |


| Age | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mean <br> Weight (g) | 50.2 | 51.2 | 62.8 | 56.4 | 62.2 | 68.9 | 50.5 | 86.7 | 77.9 | 64.6 |


| Age | 20 | 21 | 22 | $\mathbf{2 3}$ | $\mathbf{2 4}$ | $\mathbf{2 5}$ | $\mathbf{2 6}$ | $\mathbf{2 7}$ | $\mathbf{2 8}$ | $\mathbf{2 9}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Mean Weight <br> (g) | 63.5 | 75 | 86 | 71 | 77 | 84.4 | 79.4 | - | 67.6 | 52.8 |

Maturity-at-age was obtained from the ageing studies of Hüssy et al. (2012a; b) and the reproductive study by Farrell et al. (2012).

| Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Prop mature | 0 | 0 | 0.07 | 0.25 | 0.81 | 0.97 | 1 |

Natural mortality (M) was estimated over the life span of the stock using the method described by King (1995). This method assumed that M was the mortality that would reduce a population to $1 \%$ of its initial size over the lifespan of the stock. Based on a maximum age of $31, \mathrm{M}$ was calculated as follows

$$
M=-\ln (0.01) / 31
$$

Following this procedure, $M=0.16$ year ${ }^{-1}$ was considered a good estimate of natural mortality over the life span of the boarfish stock, as it was similar to the total mortality estimate from 2007, ( $Z=0.18$ ). Given that catches in 2007 were relatively low, this estimate of total mortality was considered a good estimate of natural mortality, assuming negligible fishing mortality in
previous years. Similarly, total mortality was estimated from age-structured IBTS data from 2003 to 2006 (years from which data was available for all areas). The total mortality was considered a good estimate of natural mortality as fishing mortality was assumed to be negligible during this period. Total mortality ranged from $0.09-0.2$ with a mean of 0.16 .

The special review in 2012 questioned the validity of a single estimate of M across the entire age range. If an age based assessment is possible in the future, age specific estimates of natural mortality will be required. However, the current estimate of $M$, which covers the whole age range, is considered appropriate in the context of the current situation where age data are used as an indicator approach, rather than as a full assessment method. Given that Z and F are also calculated over the entire (fully selected) range, a single value of $M$ was considered appropriate.

### 3.1.1 Stock Identity

An analysis of bottom trawl survey data suggests a continuity of distribution spanning ICES Subareas 27.4, 6, 7, 8 and 9 (Figure 3.1.1.1). Isolated occurrences appear in the North Sea (ICES Subarea 27.4) in some years indicating spill-over into this region. A hiatus in distribution was suggested between ICES Divisions 27.8.c and 9.a as boarfish were considered very rare in northern Portuguese waters but abundant further south (Cardador \& Chaves 2010). Results from a dedicated genetic study on the stock structure of boarfish within the Northeast Atlantic and Mediterranean Sea suggests that this hiatus represents a true stock separation (Farrell et al. (2016); see Figure 3.1.1.2). Based on these data, a single stock is considered to exist in ICES Subareas $27.4,6,7,8$ and the northern part of 9 .a. This distribution is slightly broader than the current EC TAC area (27.6, 7 and 8 ) and for the purposes of this year's assessment only data from these areas were utilized.

### 3.2 The fishery

Boarfish is targeted in a pelagic trawl fishery for fish meal, to the south and southwest of Ireland and Northern Biscay. The boarfish fishery is conducted in shelf waters with the first landings reported in 2001. Landings were at very low levels from 2001-2005. The main expansion period of the fishery took place between 2006 and 2010 when unrestricted landings increased from 2772 t to 137503 t . A restrictive TAC of 33000 t was implemented in 2011. In 2011, ICES was asked by the European Commission to provide catch advice for 2012 for the first time.

### 3.2.1 Advice and management

In 2010, an interim management plan was proposed by Ireland, which included a number of measures to mitigate potential bycatch of other TAC species in the boarfish fishery. The plan was enacted in legislation by Ireland but not in other countries. The details of the measures can be found in the Annex. In August 2012 the Pelagic RAC proposed a long term management plan for boarfish. The management plan was not fully evaluated by ICES; however, in 2013 ICES advised that Tier 1 of the plan could be considered precautionary if a Category 1 assessment was available. For 2014, ICES advised that, based on Fmsy $^{(0.23)}$, catches of boarfish should not be more than 133957 t , or 127509 t when the average discard rate of the previous ten years (6 448 $\mathrm{t})$ is taken into account. For 2014 the TAC was set at 133957 t by the Council of the European Union. This advice was based on a Schaefer state space surplus production model (see section 3.4 for further details). There was also concern about the use of the production model (see stock annex). ICES considered that the model was no longer suitable for providing category 1 advice and further model development was required. The model is still considered suitable for category 3 advice.

A revised draft management strategy was proposed by the Pelagic AC in July 2015. This management strategy aimed to achieve exploitation of boarfish in line with the precautionary approach to fisheries management, FAO guidelines for new and developing fisheries, and the ICES form of advice. ICES evaluated the plan and considered it to be precautionary, in that it followed the rationale for TAC setting enshrined in the ICES advice, but with additional caution (ICES, 2015a).

The advised catch for 2015 of 53296 t was based on the data limited stock HCR and an index calculated (method 3.1; ICES, 2012) using the total stock biomass trends from the model. Further work has been undertaken in 2015 to address the issues with the surplus production model and this work has continued since.

For 2017, ICES advised based on the precautionary approach that catches should be no more than 27288 t . For the first time, the precautionary buffer was applied resulting in a $36 \%$ reduction compared to the year before. The acoustic survey suggested that the stock abundance was at an historic low. The Advice Drafting Group decided the advice of 21830 proposed ( $20 \%$ reduction) would stand for 2 years. The update assessments in 2018 and 2019 confirmed that the biomass was rather stable and at a low level.

In 2021, with the stock showing a strong recruitment signal, advice of 22791 t was given for 2022 and 2023 based on the precautionary approach.

Since 2011, there has been a provision for bycatch of boarfish (also whiting, haddock and mackerel) to be taken from the Western and North Sea horse mackerel EC quotas. These provisions are shown in the table below. The effect of this is that a quantity not exceeding the value of these 4 species combined may be landed legally and subtracted from quotas for horse mackerel.

| Year | North Sea (t) | Western (t) |
| :---: | :---: | :---: |
| 2011 | 2031 | 7779 |
| 2012 | 2148 | 7829 |
| 2013 | 1702 | 7799 |
| 2014 | 1392 | 5736 |
| 2015 | 583 | 4202 |
| 2016 | 760 | 5443 |
| 2017 | 912 | 4191 |
| 2018 | 759 | 5053 |
| 2019 | 759 | 5956 |
| 2020 | 688 | 3531 |
| 2021 | 701 | 3513 |
| 2022 | 173 | 2459 |
| 2023 | 448 | 658 |

### 3.2.2 The fishery in recent years (2019-2021)

A total of 11312 t of boarfish was caught in 2019. This represents $52 \%$ of the 2019 quota of 21830 t (Table 3.2.2.1). The main participant in the fishery, Ireland, landed $9910 \mathrm{t}(75 \%$ of its national quota). The Irish catch represents $88 \%$ of the total boarfish catch in 2019 . Other countries reporting boarfish catches in 2019 were Denmark ( 757 t ), the Netherlands ( 317 t ), England ( 19 t ) and Spain ( 2.5 t ). Discards accounted for 306 t overall (Table 3.2.2.2). Tables 3.2.2.5 and 3.2.2.7 shows that about $87 \%$ of Irish landings were taken in ICES divisions 7.h and 8.a respectively.

In 2020, the total catch was 15649 t which represented $82 \%$ of the quota (19 152 t ). Ireland was the main partaker in the fishery ( 14666 t ) and landed more than its national quota ( 13234 t ) for the first time since TAC and quota regulations were established. The Irish landings accounted for $94 \%$ of the total catch. The other countries reporting catches are Denmark ( 196 t ), the Netherlands ( 416 t ), England ( 62 t ), Poland ( 109 t ) and Spain ( 1 t ). The total discards for this year were 198 t . The majority of landings were taken in ICES divisions 7.b and 7.h (Tables 3.2.2.4 and 3.2.2.5).

In 2021, 17693 t of boarfish was caught, representing $92 \%$ of the total allowable catch (19 152 t ) for the year (Table 3.2.2.1). Ireland was the main contributor to the fishery landing $11830 \mathrm{t}, 89 \%$ of their quota (13 234 t ). Other countries reporting landings for 2021 were Denmark ( 4322 t ), the Netherlands (781 t), England (45 t), Poland (45 t), Spain (11 t) and Scotland (9t). Total discards were 651 t . ICES divisions 7.j and 7.b had the highest landings of 10466 t and 3984 t respectively (Tables 3.2.2.5 and 3.2.2.4).

### 3.2.3 Catch distribution

The fishery operates during the first and fourth quarters of the year. In recent years (2019-2021) a larger proportion of the total catch has been taken in the fourth quarter. In 2022, the majority of the total catch was in the first quarter. The bulk of the catches were taken in ICES divisions 7.j and 7.b (Tables 3.2.2.5 and 3.2.2.4). Historically, the largest catches have been taken in 7.j although between 2016 and 2019, an increased proportion of the catch was taken in Northern Biscay (8.a).

Results from a questionnaire with Irish fishers found that the majority observed good marks of boarfish along the Irish west coast even up to $57^{\circ} \mathrm{N}$ which is a change from 2009/2010 when fishable marks were mainly found south from the Pistola Bank. In 2022, some decent catches were taken west of Aran and west of the Blasket islands. The fishers also remarked that since 2020, they have noticed a return of boarfish to the banks of the southern Celtic sea region where they had been absent in previous years. Also, the schools of boarfish are more dispersed than they were in high catch period of the fishery $(2009 / 2010)$ but they are larger than in previous few years. Some behavioural changes observed are that boarfish are often encountered close to the sea bed instead of dense shoals in the water column. Also, they commented that they can fish both day and night for boarfish in the west of Ireland.

### 3.2.4 Discards

It is to be expected that discarding occurred before 2003, particularly in demersal fisheries, however it is difficult to predict what the levels may have been. Since 2003, the major sources of discard estimates are the Dutch pelagic freezer trawlers (Borges et al. 2008) and both the Irish and Spanish demersal fleets. Sporadic discards were observed in German pelagic freezer trawlers before 2015 and the UK demersal fleet beginning in 2015. In 2016, Lithuania declared discards for the first time but hasn't since then. Discard estimates are not obtained from French freezer
trawlers, though discard patterns in these fleets are likely to be similar to the Dutch fleet. Discard data from the Portuguese bottom otter trawl fleet in ICES Division 9.a are also available but are not included in the assessment as they are outside the TAC area. Discard sampling data have improved in recent years with Ireland, Spain, England and Scotland providing sampling data in 2022. Table 3.2.2.2 show the total annual discards and estimates from the demersal and nontarget fisheries respectively.

### 3.2.5 Sampling data

Length-frequencies of the international commercial landings by year are presented in Table 3.2.5.1. Sampling in the early years of the fishery (2006-2009) was sparse as there was no dedicated sampling programme in place. The sampling programme was initiated in 2010 and good coverage of the landings has been achieved since then. Full details of the sampling programme in the earlier years are presented in the stock annex. Before 2017, boarfish was not included in the list of species to be sampled by the Data Collection Multi Annual Programme (DCMAP). In 2022, 21 samples, comprising of 4756 fish, were collected and measured for length from Ireland and Denmark (Tables 3.2.5.2 and 3.2.5.3). These samples covered the most heavily fished areas of 7.j and 7.b as well as 7.h. and 6.a and in quarters 1,3 and 4 .

Following standard protocols, 9 samples from the Irish catch were randomly selected for biological data collection i.e. otolith extraction, sex and maturity determination. Samples are collected on-board during fishing operations and are frozen until the vessel returns to port, which ensures high quality samples. The established sampling target is one sample per 1000 t of landings per ICES Division, which is also standard in other pelagic fisheries such as mackerel. Since 2017, all fish in each sample are be measured to the 0.5 cm below for length frequency.

### 3.3 Input data

### 3.3.1 Catch data

In 2022, 16860 t of boarfish was caught, representing $74 \%$ of the total allowable catch ( 22791 t ) for the year (Table 3.2.2.1). Ireland was the main contributor to the fishery landing $14055 \mathrm{t}, 89 \%$ of their quota ( 15749 t ). Other countries reporting landings for 2022 were Denmark ( 1859 t ), the Netherlands (858 t), England (73 t), Spain ( 10 t ) and Portugal (5t). Total discards were 1037 t (Table 3.2.2.2). ICES divisions 7.j and 7.b had the highest landings of 13242 t and 3279 t respectively (Tables 3.2.2.5 and 3.2.2.4). Landings from Denmark were revised from 1859 t to 4305 t ; however, the revision was made after the assessment was completed and the new figure will be included in next year's assessment.
A general ALK was constructed based on 814 aged fish from Irish, Danish and Scottish caught samples from 2012 (Table 3.3.1.1). A description of the ALKs prior to 2012 can be found in the stock annex. The results of the application of the ALK to commercial length-frequency data (available for the years 2007-2022) produced proxy catch numbers-at-age values which are available in Table 3.3.1.2. In the last couple of years, there has been the appearance of strong year classes in the catch numbers (Figure 3.3.1.1). A high number of 2-5 year olds were present in the 2022 data. The modal age from 2007-2011 was 6, 7 in 2012-2018, 2 in 2019-2020 and 4 in 2021. It should be noted that in WGWIDE 2011 and 2012 the plus group for boarfish was 20+. This was reduced to $15+$ in WGWIDE 2013 due to potential inaccuracy of the age readings of older fish. Ageing was based on the method that has been validated for ages 0-7 by Hüssy et al. (2012a; b). The age range is similar to the published growth information presented by White et al. (2011).

### 3.3.2 Acoustic Survey

The Boarfish Acoustic Survey (BFAS) series was initiated in 2011 in partnership with industry. The 2011 survey collected data over 24 hours. In 2012, the protocol was changed to exclude the hours between 00:00 and 04:00 as aggregations break up during the hours of darkness. The 2011 data was reworked in 2015 to exclude the data between 00:00 and 04:00. An acoustic target strength model of (-66.2dB) was developed in 2013 (Fässler et al. (2013)) and is applied to all surveys in the time series. Over the time series of the survey total biomass has been estimated in the range 863 kt (in 2012) to 70 kt (2016) with CV estimates ranging 0.11 to 0.31 . Total biomass estimates declined sharply between 2012 and 2016 after which an increasing trend is seen (Figure 3.3.2.1).

An annual index of biomass and abundance has been reported for boarfish since 2011. The 2023 survey was carried out onboard the RV Celtic Explorer during the WESPAS survey (Western European Shelf Pelagic Acoustic Survey). The survey was carried out over a 42-day period beginning on the 10 June in the south $\left(47^{\circ} 30 \mathrm{~N}\right)$ and working northwards to $59^{\circ} 30 \mathrm{~N}$ ending on 20 July.

## Calculation of acoustic abundance

The StoX software package (Johnsen et. al., 2019) was used to calculate acoustic abundance from survey data (StoX V3.6.1 and R-StoX V3.6.1). Aggregated survey data are available for download from the ICES Trawl Acoustic database. Survey design and analysis procedures adhere to guidelines laid out in the Manual for International Pelagic Surveys (ICES, 2015b).

## Survey results 2023

The estimate of boarfish biomass is presented in Table 3.3.2.1 and the spatial distribution of the echotraces attributed to boarfish in 2023 are presented in Figure 3.3.2.2. Overall, the WESPAS survey provided continuous synoptic coverage from south to north over 42 days, relating to an area coverage of over $59,632 \mathrm{nmi}^{2}$ (boarfish strata) and transect mileage of over 4,597 nmi. In total, 53 trawl stations were undertaken with 29 hauls containing boarfish providing 4,377 individual lengths, 1,800 length and weight measurements and 876 otoliths for use during the analysis.

Survey effort was comparable to 2022, with a $2 \%$ increase in both acoustic sampling effort (survey miles) and area coverage. Boarfish TSB (total stock biomass) and abundance (TSN) estimates were $525,701 \mathrm{t}$ and $15,938,301,000$ individuals (CV 0.11) respectively. The 2023 estimate saw an increase of $16 \%$ in TSB compared to 2022 , while TSN saw a reduction of $14 \%$. Spawning stock biomass (SSB) increased by $18 \%$ as compared to 2022 . The reported increase in biomass (TSB \& SSB) and reduction in abundance (TSN) is largely driven by the reduced numbers of immature fish observed during the survey.

The geographical distribution of boarfish was comparable to earlier years in the time series in terms of latitudinal range. Within this range, clusters of individual schools were most frequently found towards the shelf margin in the northern and western strata and more widespread across the shelf in the Celtic Sea. The centre of gravity of schools during this year's survey in the Celtic Sea stratum was more towards the shelf edge than in recent years. It is possible this more westward distribution was linked to the hydrographic conditions encountered this year compared to recent years; with a cool pool of water extending across the seafloor in the Celtic Sea.

Over $65 \%$ of the standing stock biomass was observed in the Celtic Sea stratum ( $50 \% \mathrm{TSN}$ ), followed by the Irish west coast ( $28 \%$ TSB \& $22 \%$ TSN), the southern Hebrides stratum ( $13 \%$ TSB \& $9 \%$ TSN ), the western Hebrides stratum ( $8 \%$ TSB \& 5\% TSN) and the Porcupine Bank stratum ( $2 \%$ TSB \& $1 \%$ TSN) (Figure 3.3.2.3).

The 4 -year age class dominated the 2023 estimate contributing $25 \%$ of TSB and $29.4 \%$ of TSN. Ranked second and third were the 3-year old ( $16 \%$ TSB \& $24.9 \%$ TSN) and 7-year old fish ( $14.1 \%$ TSB \& $11.5 \% \mathrm{TSN}$ ) respectively. Ranked fourth is the 6 -year old fish ( $12.2 \%$ TSB \& $11.4 \%$ TSN). Combined, these four age classes represented $67 \%$ of TSB and $77 \%$ of TSN. The survey has successfully tracked these four cohorts from recruitment. The 15+ age class represented $10.4 \%$ of TSB and $4.3 \%$ of TSN (Figure 3.3.2.4).

Maturity analysis indicated over $99.6 \%$ of TSB and $99 \%$ of TSN was represented by mature fish. The proportion of immature fish in the stock estimate has decreased from a high in 2021 ( $41 \%$ TSN) over recent years (2022;5\% and 2023; 1\%). This differs with observations from the PELGAS survey where the overall abundance of boarfish, notably immature fish, has increased over the same period.

### 3.3.3 International bottom trawl survey (IBTS) Indices

An index of abundance was constructed from the following surveys:

- EVHOE, French Celtic Sea and Biscay Survey, (Q4) 1997 to 2022
- IGFS, Irish Groundfish Survey, (Q4) 2003 to 2022
- WCSGFS, West of Scotland, (Q1 and Q4) 1986 to 2009 (survey design changed in 2010)
- SPPGFS, Spanish Porcupine Bank Survey, (Q3) 2001 to 2022
- SPNGFS, Spanish North Coast Survey, (Q3/Q4) 1991 to 2022
- ECSGFS, CEFAS English Celtic Sea Groundfish Survey, (Q4) 1982 to 2003

The spatial extent and haul positions of each survey is shown in Figure 3.3.3.1 and the length frequencies are presented in Table 3.3.3.1 for each survey. These surveys cover the majority of the observed range of boarfish in the ICES area (Figure 3.3.3.2). The index of abundance, CPUE, was computed as the number of boarfish per 30 min haul. The abundance of boarfish per year per ICES statistical rectangle (used for visualisation only) was then calculated by summing the boarfish in a given rectangle and dividing by the total number of hauls in that rectangle. The West of Scotland survey index had been increasing between 2000 and 2009 but is uncertain and was stopped soon after. The English Celtic Sea survey showed an upward trend in the last couple of years before the survey ended in 2003. Of the four current bottom trawl surveys, the French, Irish and Spanish Porcupine groundfish surveys experienced an increase in CPUE, particularly the French survey.

Some of the IBTS CPUE indices displayed marked variability with a large proportion of zero tows and occasionally very large tows (e.g. West of Scotland survey, Figure B.4.7 stock annex). More southern surveys displayed a consistently higher proportion of positive tows. The variability of the data is reflected in the estimated mean CPUE indices (Figure 3.3.3.3).

A detailed analysis of the IBTS data was carried out in 2012 to investigate the main areas of abundance of boarfish in these surveys. This analysis included GAM modelling based on the probability of occurrence of boarfish. The full details of this work are presented in the stock annex. The IBTS appears to give a relative index of abundance, with good resolution between periods of high and low abundance. The main centres of abundance in the survey correspond to main fishing grounds. Figures 3.3.3.4a and $b$ shows the signal in abundance and biomass, increasing gradually in the 1990s, slowly declining in the early 2000s, before increasing again with a strong increase in the most recent period. Much of this increase which is stronger in terms of abundance is due to increased recruitment since 2017. The low estimates for the 2017 survey are partly explained by issues with the execution of the EVHOE survey. Due to mechanical breakdown, the majority of the survey stations could not be completed. The missed stations would
have covered the area in North Biscay typically associated with the highest catch rates of boarfish.

For subsequent surplus production modelling (see Section 3.4), biomass indices were extracted from each of the IBTS surveys using a delta-lognormal model (Stefánsson 1996); details of which can be found in the annex.

When the indices were recalculated in 2021, (following a refresh of the input data from DATRAS and national data submitters), the following issues were encountered

- An error with the coding of the EVHOE 2018 data in DATRAS was corrected, revising upwards the estimates from 2018 for this survey
- The truncated EVHOE 2017 dataset was removed from the analysis. In previous years, this data was retained but, because the available data only corresponds to a small fraction of the total survey area (where boarfish are not usually encountered in significant quantities) a very low survey estimate resulted. It was considered appropriate to remove this data from the analysis. In future, explicit modelling of spatial and temporal correlations may permit this data to be considered again.
- An error in the analysis was discovered whereby hauls with more than one catch category were underrepresented as only a single catch category was included during the model fitting. Multiple catch categories are usually the result of splitting the catch into adult and juvenile portions and using an appropriate subsampling strategy for each. This issue is particularly relevant for the IGFS which, over the most recent 4 years has 2 catch categories for boarfish recorded for approximately $20 \%$ of hauls. The outcome is an increase in CPUE for these hauls and a subsequent increase in the survey index for the IGFS in recent years (2016 onwards).


### 3.4 Biomass dynamic model

In 2012 an exploratory biomass dynamic model was developed for the assessment of boarfish. The model is a Bayesian state space surplus production model (Meyer \& Millar 1999), incorporating the catch data, delta-lognormal estimated IBTS survey indices and acoustic biomass estimates. Following the initial development of the model, the assessment was peer-reviewed by two independent experts on behalf of ICES. In 2013 a new assessment was provided, which was based on the previous year's work and the reviewers' comments and formed the basis of a category 1 assessment. Details of the review and the associated changes can be found in the stock annex.

In 2014 the Bayesian state space surplus production model was fit using the catch data, deltalognormal estimated IBTS survey indices, and the acoustic survey estimates. However, the inclusion of the low 2014 acoustic biomass estimate changed the perception on the stock, which raised concerns over the sensitivity and process error of the model and the stock assessment was moved from ICES category 1 to category 3 with the results of the surplus production model being used to calculate an index for the data limited stock approach.

Since 2014, the procedure used to run the model has not changed with annual updates to the input data only. The details of the progression of the production model can be found in the Annex.

In the Bayesian state space surplus production model the biomass dynamics are given by a difference form of a Schaefer biomass dynamic model:

$$
B_{t}=B_{t-1}+r B_{t-1}\left(1-\frac{B_{t-1}}{K}\right)-C_{t-1}
$$

where $B_{t}$ is the biomass at time $\mathrm{t}, \mathrm{r}$ is the intrinsic rate of population growth, $K$ is the carrying capacity, and $C_{t}$ is the catch, assumed known exactly. To assist estimation, the biomass is scaled by the carrying capacity, denoting the scaled biomass $P_{t}=B_{t} / K$. A lognormal error structure is assumed giving the scaled biomass dynamics (process) model:

$$
P_{t}=\left(P_{t-1}+r P_{t-1}\left(1-P_{t-1}\right)+\frac{C_{t-1}}{K}\right) e^{\mu_{t}}
$$

where the logarithm of process deviations are assumed normal $u_{t}=N\left(0, \sigma_{2}^{\mu}\right)$ with $\sigma_{2}^{\mu}$ the process error variance.

The starting year biomass is given by $a K$, where a is the proportion of the carrying capacity in the first year. The biomass dynamics process is related to the observations on the indices through the measurement error equation:

$$
I_{j, t}=q_{j} P_{t} K e^{\varepsilon_{j, t}}
$$

where $I_{j, t}$ is the value of abundance index $j$ in year $t, q_{j}$ is survey-specific catchability, $B_{t}=P_{t} K$, and the measurement errors are assumed log-normally distributed with $u_{t}=N\left(0, \varepsilon_{e, j, t}^{2}\right)$ where $\varepsilon_{e, j, t}^{2}$ is the index-specific measurement error variance. $\operatorname{Var}\left(I_{j, t}\right)$ is obtained from the delta-lognormal survey fits. That is, the variance of the mean annual estimate per survey is inputted directly from the delta-lognormal fits (Figure 3.3.3.3) as opposed to estimating a measurement error within the assessment. The measurement error is obtained from:

$$
\sigma_{e, j, t}^{2}=\ln \left(1+\frac{\operatorname{Var}\left(I_{j, t}\right)}{\left(I_{j, t}\right)^{2}}\right)
$$

For the acoustic survey, the CV of the survey was transformed into a lognormal variance via

$$
\sigma_{\varepsilon, \text { acoustic }, t}^{2}=\ln \left(C V_{\text {acoustic }, t}^{2}+1\right)
$$

Prior assumptions on the parameter distributions were:

- Intrinsic rate of population growth: $r \sim U(0.001,2)$
- Natural logarithm of the carrying capacity: $\ln (K) \sim U(\ln (\max (C), \ln (10 . \operatorname{sum}(C))=$ $U(\ln (144047), \ln (4450407))$
- $\quad$ Proportion of carrying capacity in first year of assessment: $a \sim U[0.001,1.0]$
- $\quad$ Natural logarithm of the survey-specific catchabilities $\ln \left(q_{i}\right) \sim U(-16,0)$ (for IBTS only). The acoustic survey prior is discussed below.
- $\quad$ Process error precision $\frac{1}{\sigma_{u}^{2}} \sim \operatorname{gamma}(0.001,0.001)$


### 3.4.1 Specification

The specifications for the final boarfish assessment model runs are:

## Acoustic survey

Years: 2011-2023
No substantial evidence exists for removing any of the survey points from the time series although 2016 may be considered an outlier (Table 3.3.2.1).

In the index value from 2011 - 2015 is calculated using a bespoke protocol as outlined below Index value (Iacoustic,y): 'total' in tonnes (i.e. Definitely Boarfish + Probably Boarfish + Boarfish in a Mix)

For the years 2016 - present, the StoX software package (Johnsen et. al., 2019) was used to calculate the acoustic abundance index.

Catchability ( $q_{\text {acoustic }}$ ): A free, but strong prior (i.e. the acoustic survey is treated as a relative index but is strongly informed, this allows the survey to cover $<100 \%$ of the stock).

## IBTS surveys

6 delta log normal indices (WCSGFS, SPPGFS, IGFS, ECSGFS, SPNGFS, EVHOE)
The final run is based on a run that includes all surveys with the omission of the first 5 years of the WCSGFS and first 9 years of the ECSGFS as it was unclear whether boarfish were consistently recorded in the early part of the ECSGFS. The WCSGFS is thought to be at the northern extreme of the distribution and may not be an appropriate index for the whole stock. The initial data year was set at 1991 when 3 groundfish survey indices are available (SPNGFS, ECSGFS and WCSGFS). The survey indices are weighted such that highly uncertain values receive lower weight in the fitting.

## Catches

2003-2022 time-series

## Priors

The final run assumes a strong prior for the acoustic survey catchability with $\ln \left(q_{\text {acoustic }}\right) \sim N(1$, 1/4) (mean 1, standard deviation 0.25 ), which has $95 \%$ of the density between 0.5 and 2 . Given the relatively short acoustic series it is not possible to estimate this parameter freely (i.e. using an uninformative prior). The prescription of a strong prior removes the assumption of an absolute index from the acoustic survey. This assumption will be continually updated as additional data accrue.

### 3.4.2 Results

## Run convergence

Parameters for the 2023 model run converged with good mixing of the chains and Rhat values lower than 1.1 indicating convergence and acceptable autocorrelation (Figures 3.4.2.1-3).

Diagnostic plots are provided in Figure 3.4.2.4 showing residuals about the model fit. A fairly balanced residual pattern is evident. In some cases, outliers are apparent, for instance in the later years of WCSGFS survey. Large residuals tend to be observed where the model has ignored data points that had high uncertainties. However, these points are down weighted according to the inverse of their variance and hence do not contribute much to the model fit. The acoustic survey (WESPAS) overestimates the stock in the first 3 years, then underestimates it for the next 4 years before again overestimating it slightly until the present survey. This suggests that this index is perhaps not representative of the whole stock. Figure 3.4.2.5 shows the prior and posterior distributions of the parameters of the biomass dynamic model. The estimate of $q$ for the acoustic survey is 0.79 , leading to a higher estimate of final stock biomass than the acoustic survey result.

Trajectories of observed and expected indices are shown in Figure 3.4.2.6, along with the stock size over time and a harvest ratio (total catch divided by estimated biomass). Parameter estimates from the model run are summarized in Table 3.4.2.1. Biomass in 2023 is estimated to be 734 kt , continuing the increasing trend in stock size since 2016. The extremely low biomass estimate from the 2016 acoustic survey appears to be largely considered as an outlier by the model. This is also the case for the high survey estimate in 2012 although the drop in biomass between these points is seen in a number of the input data series. Retrospective plots of TSB and F, presented in Figure 3.4.2.6, show that the perception of the stock is stable over the most recent 5 years.

Diagnostics from the positive component of the delta-lognormal fits indicate relatively good agreement with a normal distribution on the natural logarithmic scale (Figure 3.4.2.8). There is an indication of longer tails in some of the surveys (e.g. WCSGFS, SPPGFS).

Pair-wise correlation between the annual mean survey indices varied. The updates described in section 3.3.3 with respect to data and analysis code corrections have resulted in increased correlation between the surveys most affected i.e. IGFS and EVHOE (Figure 3.4.2.9). The WCSGFS displayed positive correlations with all five surveys except the Spanish north coast survey (SPNGFS). The SPPGFS displayed a small negative correlation with EVHOE. Weighting the correlations by the sum of the pair-wise variances resulted in a largely similar correlation structure (Figure 3.4.2.9). Note that though some surveys displayed weak or no correlation, no surveys were excluded a-priori from the assessment.

### 3.4.3 State of the stock

This year's assessment indicates that total stock biomass increased from a low to average level from the early to mid-1990s (Figure 3.4.2.7). The stock fluctuated around this level until 2009, before increasing until 2012. A sharp decline is seen between 2013 and 2014. Since 2014, the abundance has increased and is close to the historic maximum from the assessment (2012) in 2023. There was concern in 2014 that this decline was exaggerated by an unusually low acoustic biomass estimate that led to a downward revision in stock trajectory. However, the 2014 survey is considered satisfactory in terms of containment. The comparably low 2014 biomass estimate was supported by results of the 2015 survey. The 2016 biomass estimate, the lowest of the time series is considered likely an outlier and has little influence on stock abundance estimates. The assessment is characterised by relatively high uncertainty reflecting the uncertainty in the survey indices, and short exploitation history of the stock and the treatment of the acoustic survey as a relative biomass index.

Catch data are available from 2001, the first year of commercial landings, and reasonably comprehensive discard data are available from 2003. Peak catches were recorded in 2010, when over 140000 t were taken. Elevated fishing mortality was observed, associated with the highest recorded catch in 2010. Fishing mortality, expressed as a harvest ratio (catch divided by total biomass), was first recorded in 2003. Before that time, it is to be expected that some discarding took place, and there were some commercial landings. Fishing mortality increased measurably from 2006, reaching a peak in 2009-2010. F declined in 2011 as catches became regulated by the precautionary TAC but increased year on year until 2015 when reduced catches resulted in a reduction.

MSY reference points can be estimated from the production model assessment parameter values. In 2023, $\mathrm{F}_{\text {MSY }}(\mathrm{r} / 2)$ is estimated to be 0.16 , similar to estimates from previous assessments and MSY Btrigger $(\mathrm{K} / 4) 173 \mathrm{kt}$. Throughout the history of the fishery, estimates of total biomass have remained above MSY $B_{\text {trigger. Fishing mortality ( }}^{\text {( }) ~ w a s ~ b r i e f l y ~ l a r g e r ~ t h a n ~ t h e ~ e s t i m a t e ~ o f ~ F M S Y ~}$ between 2009 and 2010 and again in 2014, but has decreased since. In 2023, the stock is in the green area of the Kobe plot (Figure 3.4.3.1).

## Recruitment

Estimates of recruitment are not available from the stock assessment. However, all available data sources (catch, acoustic survey and IBTS surveys) indicate above average recruitment over the period 2019-2021. The common ALK (Table 3.3.1.1.) was applied to the IBTS number-at-length data. The length-frequency is presented in Table 3.3.3.1. and the age-structured index in Table 3.3.3.2. and Figure 3.3.3.5.

The EVHOE, IGFS and SPNGFS surveys provide the best indices of recruitment as this is where the juveniles appear to be most abundant (Table 3.3.3.1). It appears that recruitment was high in
the late 1990s in the EVHOE survey with 2010 and 2015 also indicating above average recruitment. Particularly strong recruitment has been noted in each of 2018-2020, especially for the EVHOE survey but also the IGFS in 2020.

The PELGAS survey is conducted annually in waters to the south of the boarfish (WESPAS) survey. In 2021, PELGAS recorded a significant increase in biomass on its most northerly transects (immediately south of the WESPAS southern limit) compared to 2019 (no survey was conducted in 2020), in broad agreement with increases noted on WESPAS. In 2022 and 2023, boarfish was encountered as far south as $45^{\circ} \mathrm{N}$, primarily on the western extent of the survey transects. Estimates of biomass in 2022 and 2023 were approximately 70kt. The PELGAS survey takes place approximately 1 month prior to the boarfish survey.

### 3.5 Short Term Projections

As the assessment is exploratory, no short term projections were conducted.

### 3.6 Long term simulations

No long term simulations were conducted.

### 3.7 Candidate precautionary and yield based reference points

### 3.7.1 Yield per Recruit

A yield per recruit analysis was conducted in 2011 (Minto et al. 2011) and $F 0.1$ was estimated to be 0.13 whilst $\mathrm{Fmax}_{\text {mas }}$ wastimated in the range 0.23 to 0.33 (Figure 3.7.1.1). F0.1 was considered to be well estimated (Figure 3.7.1.2). No new yield per recruit analyses were performed in subsequent years.

### 3.7.2 Precautionary reference points

No reference points have been defined for boarfish.

### 3.7.3 Other yield based reference points

Yield per recruit analysis, following the method of Beverton \& Holt (1957), found F0.1 to be robustly estimated at 0.13 (ICES 2011; Minto et al. 2011).

### 3.8 Quality of the assessment

ICES considers the current basis for the advice on this stock to be an interim measure. Extensive sampling of the commercial catch and information from surveys will permit assessment development as the available data improves. The acoustic survey has undergone several developments to improve its suitability with updates to methodology in 2012, a change in direction in 2017 and extension of transects at the boundaries to improve containment. The current assessment was downgraded from Category 1 to Category 3 in 2014, and it has remained in this category since. The model is still considered suitable for category 3 advice, because it provides the best means of combining the available survey series. The assessment is sensitive to the acoustic
series. In addition, a substantial part of the year to year variations in the stock abundance is linked to the process error. The use of some priors (like ratio to virgin biomass in the first year of the assessment) and survey (e.g. WCSGFS for instance) may require revision.

The bottom trawl survey data are considered to be a good index of abundance given that boarfish aggregate near the bottom at this time of year. The trawl surveys record high abundances of the species, but with many zero hauls. The delta-lognormal error structure used in the analyses is considered to be an appropriate means of dealing with such data. The biomass dynamic model used in the stock assessment is based on the assessment of megrim in Sub-divisions 4 and 6 with the model further developed by including acoustic survey biomass estimates. A drawback of the current assessment model is that it does not provide estimates of recruitment although estimates of recruitment strength are available from the Spanish and French bottom trawl surveys.

### 3.9 Basis for 2024 and 2025 advice

Advice for boarfish in 6-8 has been issued biennially since 2018 and is based on the trend in the relative TSB from the assessment. Advice for 2024 and 2025 is based on the updated assessment presented in this report and the application of the updated ICES guidance for stocks in category 3, based on the work of the ICES WKLIFE series of workshops (ICES 2022).

Although considered by the SPiCT benchmark workshop (WKBMSYSPiCT2) in 2023, an accepted SPiCT assessment could not be achieved for this stock. However, with an index of abundance available from the current assessment model, comprehensive length sampling of the commercial catch and available estimates of the von Bertalanffy growth parameter (k), method 2.1, an rfb rule with $\mathrm{m}=0.95$ can be used to derive category 3 MSY advice for this stock.

Estimates of k are available from White et al (2011) (0.186 year ${ }^{-1}$ ) and the more comprehensive study of Hüssy et al (2012b) ( 0.145 year $^{-1}$ for females, 0.181 year $^{-1}$ for males).

The calculations for the individual rfb components are described below:

## Biomass ratio (r)

$r$ is based on the relative TSB from the 2023 update assessment over the most recent 5 years. Index A, the mean for the most recent 2 years $(2022,2023)$ is 1.82 . Index $B$, the mean over the 3 years prior to this (2019-2021) is 1.23 such that the resultant r parameter value is 1.48 .

## Fishing pressure proxy (f)

Based on the analyses carried out during WKBMSYSPiCT2, Linf is estimated as 153.5 mm the average of the individual male $(160 \mathrm{~mm})$ and female $(147 \mathrm{~mm})$ estimates.

Extensive length sampling of the commercial catch of boarfish has been carried out by Ireland since 2007. Irish catches account for by far the greatest proportion of the total catch and the sampling programme covers the entire fleet exploiting boarfish.

For the purposes of calculating the target reference length ( $\mathrm{L}_{\mathrm{f}=\mathrm{m})}$, length at first capture $\left(\mathrm{L}_{\mathrm{c}}\right)$ is based on the most recent 4-year period, following strong recent recruitment to the stock and is estimated at 70 mm , the smallest length class for which the estimate of total catch in numbers exceeds half of the estimate at the modal length. Identical estimates of $\mathrm{L}_{\mathrm{c}}$ result when extending the historic period to include the most recent 10 years of length sampling information. When combined with the above estimate of $\mathrm{Linff} \mathrm{LF}=\mathrm{M}=0.75 \mathrm{~L}_{\mathrm{c}}+0.25 \mathrm{Linf}=91 \mathrm{~mm}$. The mean length in the catch for the most recent year (2022) is estimated as 103 mm resulting in a fishing pressure proxy of 1.13.

## Biomass safeguard (b)

The minimum relative biomass occurred in the first year of the assessment (0.52) although similar levels were reached in $2003(0.63)$ and $2016(0.62)$, leading to an $I_{\text {trigger }}$ of $1.4^{*} 0.52=0.73$. With the most recent estimate of relative biomass in 2023 of 1.88 , the $b$ parameter value for application of the rfb rule is set to 1 .

The selected rfb rule specifies an additional precautionary multiplier (m) of 0.95 which, when applied in conjunction with the most recent advice $(22,791 \mathrm{t})$, the $\mathrm{r}, \mathrm{f}$ and b values outlined above leads to an advice for 2024 and 2025 more than $20 \%$ above that for 2022-23. As such, the stability rule is invoked and the increase limited to $20 \%(27,349 t)$.

### 3.10 Management considerations

This stock is placed in category 3 with advice is based on the harvest control rules developed by ICES WKLIFE (ICES 2022). The biomass estimate from the Bayesian model is considered reliable for trends and is used as an index of stock development. Advice is based on rfb method 2.1 (as the von Bertalanffy k parameter for boarfish is estimated to be less than 0.2 ). The advice is calculated based on a comparison of the average of the two most recent index values with the average of the three preceding values multiplied by the most recent catch, a biomass safeguard, a lengthbased fishing mortality proxy, a precautionary multiplier and the most recent advice. Table 3.9.1.1 shows the biomass estimates from the model from which the index was calculated. Although not currently accepted as the basis for an analytic assessment, the surplus production model still provides the best unified view of this stock (Figure 3.4.2.5).

### 3.11 Stock structure

A dedicated study on the stock structure of boarfish within the Northeast Atlantic and Mediterranean Sea was carried out in 2013. Details can be found in the stock annex. Ecosystem considerations

### 3.12 Ecosystem considerations

The main ecosystem aspects for boarfish in the NE Atlantic are outlined in the stock annex. Management plan.

### 3.13 Proposed management plan

In 2012 and 2015 the Pelagic Advisory Council submitted a proposed management strategy for Northeast Atlantic boarfish. Details of the plans and their evaluations can be found in the stock annex.

### 3.14 References

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### 3.15 Tables

Table 3.2.2.1. Boarfish in ICES Subareas 27.6, 7, 8. Landings by country, total discards and TAC by year (t), 2001-2021. (Data provided by Working Group members)

|  | Denmark | Germany | Ireland | Neth-er- | England | Poland | Scot- <br> land | Spain | Por- <br> tugal | Discards | Total | TAC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 |  |  | 120 |  |  |  |  |  |  |  | 120 |  |
| 2002 |  |  | 91 |  |  |  |  |  |  |  | 91 |  |
| 2003 |  |  | 458 |  |  |  |  |  |  | 10929 | 11387 |  |
| 2004 |  |  | 675 |  |  |  |  |  |  | 4476 | 5151 |  |
| 2005 |  |  | 165 |  |  |  |  |  |  | 5795 | 5960 |  |
| 2006 |  |  | 2772 |  |  |  |  |  |  | 4365 | 7137 |  |
| 2007 |  |  | 17615 |  |  |  | 772 |  |  | 3189 | 21576 |  |
| 2008 | 3098 |  | 21585 |  |  |  | 0 |  |  | 10068 | 34751 |  |
| 2009 | 15059 |  | 68629 |  |  |  |  |  |  | 6682 | 90370 |  |
| 2010 | 39805 |  | 88457 |  |  |  | 9241 |  |  | 6544 | 144047 |  |
| 2011 | 7797 |  | 20685 |  |  |  | 2813 |  |  | 5802 | 37097 | 33000 |
| 2012 | 19888 |  | 55949 |  |  |  | 4884 |  |  | 6634 | 87354 | 82000 |
| 2013 | 13182 |  | 52250 |  |  |  | 4380 |  |  | 5598 | 75410 | 82000 |
| 2014 | 8758 |  | 34622 |  |  |  | 38 |  |  | 1813 | 45231 | 133957 |
| 2015 | 29 | 4 | 16325 | 375 | 104 |  |  |  |  | 929 | 17766 | 53296 |
| 2016 | 337 | 7 | 17496 | 171 | 21 |  |  |  |  | 1283 | 19315 | 47637 |
| 2017 | 548 |  | 15485 | 182 | 0 |  |  |  |  | 1173 | 17388 | 27288 |
| 2018 | 94 |  | 9513 | 172 | 0 |  | 0 | 148 |  | 1359 | 11286 | 21830 |
| 2019 | 757 |  | 9910 | 318 | 19 |  |  | 3 |  | 306 | 11312 | 21830 |
| 2020 | 196 |  | 14666 | 416 | 62 | 109 |  | 1 |  | 198 | 15649 | 19152 |
| 2021 | 4322 |  | 11830 | 781 | 45 | 45 | 9 | 11 |  | 651 | 17693 | 19152 |
| 2022 | 1859 |  | 14055 | 858 | 73 |  |  | 10 | 5 | 1037 | 17898 | 22791 |
| $0=<0.5 \mathrm{t}$ |  |  |  |  |  |  |  |  |  |  |  |  |

Table 3.2.2.2. Boarfish in ICES Subareas 27.6, 7, 8. Discards in demersal and non-target pelagic fisheries by year (data provided by Working Group members)
\(\left.\begin{array}{lllllll}\hline Year \& Denmark \& Germany \& Ireland \& Nether- <br>

lands\end{array}\right]\) Ppain | Lithuania |
| :--- |
| 2003 |

Table 3.2.2.3 Boarfish in ICES Subareas 27.6

| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 37 | 67 | 172 | 10 | 23 | 0 |
| England |  |  |  |  |  |  |  |  |  |  |  |  | 9 |  |  |  | 9 | 7 |  | 9 |
| Ireland | 65 | 292 | 10 | 21 | 99* | 28 | 45 | 1356 | 26 | 125 | 538 | 182 | 116 | 377 | 907 | 269 | 568 | 1214 | 378 | 114 |
| Netherlands |  |  |  |  |  |  |  |  |  |  |  |  | 128 | 45 | 34 | 78 | 79 | 108 | 52 | 1 |
| Scotland |  |  |  |  |  |  |  | 10 |  |  | 15 | 30 |  |  |  |  |  |  | 6 | 2 |

*6t in $5 \mathrm{~b}, 0=0-0.5 \mathrm{t}$

Table 3.2.2.4 Boarfish in ICES Subareas 27.7bc

| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |  |  |  |  |  |  | 80 | 12 | 8 | 21 |  |  |  | 85 | 13 | 1002 |
| England |  |  |  |  |  |  |  |  |  |  |  |  | 85 | 1 |  |  | 0 | 32 | 10 | 3 |
| Germany |  |  |  |  |  |  |  |  |  |  |  |  | 4 | 5 |  |  |  |  |  |  |
| Ireland | 214 | 224 | 105 | 15 | 1259 | 3 | 74 | 2293 | 283 | 4609 | 10405 | 3262 | 2829 | 1198 | 124 | 163 | 241 | 6818 | 3732 | 2262 |
| Netherlands |  |  |  |  |  |  |  |  |  |  |  |  | 33* | 35 | 138 | 10 | 150 | 212 | 228 | 12 |
| Scotland |  |  |  |  |  |  |  | 4 |  | 1745 | 100 |  |  |  |  |  |  |  | 2 |  |
| *Division 7, 0=0-0.5t |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

## Table 3.2.2.5 Boarfish in ICES Subareas 27.7h-k



## Table 3.2.2.7 Boarfish in ICES Subarea 8

| Country | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  |  |  |  |  | 18 |  | 1354 |  | 6 | 7 | 271 |  | 315 |  | 111 | 81 |
| England |  |  |  |  |  |  |  |  |  |  |  |  |  | 5 |  |  |  |  |  |  |
| Germany |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 1 |  |  |  |  |  |  |
| Ireland |  | 38 | 38 | 1 | 5 |  |  |  |  | 93 | 1140 | 119 | 682 | 7297 | 11458 | 5336 | 2876 | 269 |  |  |
| Netherlands |  |  |  |  |  |  |  |  |  |  |  |  | 2014 |  |  | 14 | 0 | 17 | 48 | 33 |
| Spain |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 148* | 2 | 1 | 11 | 11 |
| Portugal |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 5 |
| *94t in 9a, 0=0-0.5t |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |

Table 3.2.5.1. Boarfish in ICES Subareas 27.6, 7, 8. Length-frequency distributions of the international catches (raised numbers in ‘000s) for the years 2007-2022

| Length | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4.5 |  |  |  |  |  |  |  |  | 14 |  |  |  |  |  |  |  |
| 5.0 |  |  |  |  |  |  |  |  | 878 |  |  |  |  |  |  | 1 |
| 5.5 |  |  |  |  |  |  |  |  | 515 |  |  |  |  | 2746 |  |  |
| 6.0 |  |  |  | 156 |  |  |  |  | 810 |  | 765 |  | 15868 | 37073 | 537 | 5058 |
| 6.5 |  |  |  | 439 |  |  |  |  | 14 |  | 4607 | 203 | 70362 | 150810 | 2147 | 1453 |
| 7.0 |  |  |  | 1090 | 522 | 56 | 52 |  | 513 | 417 | 5250 | 405 | 80160 | 233347 | 13936 | 26532 |
| 7.5 |  |  | 1354 | 1574 |  |  | 551 |  | 10598 | 1684 | 12616 | 2635 | 85420 | 147915 | 25740 | 21853 |
| 8.0 |  |  | 677 | 375 | 1345 | 185 | 1419 |  | 80716 | 8685 | 11473 | 4703 | 115154 | 38949 | 30699 | 30617 |
| 8.5 |  |  |  | 1082 |  | 555 | 3592 | 1064 | 49508 | 6412 | 10115 | 3559 | 67471 | 43556 | 45234 | 16174 |
| 9.0 |  |  | 677 | 5382 | 851 | 555 | 7263 | 327 | 10219 | 7104 | 3874 | 6554 | 16504 | 101918 | 107121 | 50374 |
| 9.5 |  | 7473 | 17367 | 7883 | 7012 | 641 | 47509 | 4916 | 213 | 23065 | 14047 | 6196 | 3147 | 115103 | 191656 | 121067 |
| 10.0 | 9609 | 11209 | 54130 | 29410 | 33243 | 2791 | 94702 | 31649 | 1211 | 46010 | 32346 | 5559 | 9173 | 100550 | 177751 | 173811 |
| 10.5 |  | 52308 | 174796 | 130889 | 15848 | 6132 | 59833 | 71344 | 3865 | 39071 | 36242 | 4450 | 10144 | 55049 | 98863 | 135187 |
| 11.0 | 84555 | 63517 | 343283 | 361774 | 70615 | 24571 | 18359 | 108261 | 12226 | 14181 | 32445 | 17658 | 5796 | 9475 | 72207 | 104658 |
| 11.5 |  | 59781 | 321637 | 655875 | 93487 | 81928 | 20938 | 82470 | 28142 | 18249 | 31589 | 22826 | 22722 | 3172 | 44227 | 64098 |
| 12.0 | 44199 | 119561 | 297737 | 739025 | 189434 | 264888 | 98564 | 84288 | 41613 | 30975 | 33618 | 24070 | 22353 | 2396 | 14710 | 29920 |
| 12.5 |  | 70990 | 207739 | 564347 | 114904 | 398772 | 204868 | 112826 | 42461 | 51110 | 41650 | 24514 | 17521 | 3251 | 5711 | 8411 |
| 13.0 | 82633 | 52308 | 147965 | 353484 | 133539 | 419060 | 315063 | 172416 | 59990 | 57000 | 46495 | 30665 | 28815 | 9494 | 6738 | 4477 |


| Length | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 13.5 |  | 29890 | 149314 | 246146 | 51235 | 307533 | 285688 | 153742 | 52625 | 58696 | 43121 | 38698 | 16688 | 13707 | 8599 | 6754 |
| 14.0 | 117224 | 22418 | 105782 | 224611 | 50857 | 176710 | 210137 | 138549 | 50139 | 76872 | 45353 | 34080 | 20053 | 16381 | 8468 | 9886 |
| 14.5 |  | 14945 | 71273 | 127711 | 25309 | 89726 | 105571 | 74059 | 28771 | 37755 | 39524 | 29908 | 13809 | 14913 | 7389 | 8555 |
| 15.0 | 65338 | 33627 | 47816 | 125463 | 25569 | 52791 | 62175 | 43347 | 16087 | 23137 | 21854 | 15561 | 5710 | 12563 | 7222 | 18872 |
| 15.5 |  | 11209 | 13082 | 81386 | 5473 | 25065 | 31122 | 22629 | 8572 | 7841 | 4932 | 5778 | 1513 | 4304 | 2880 |  |
| 16.0 | 13452 | 11209 | 19397 | 24256 | 4181 | 13149 | 14990 | 7672 | 4331 | 625 | 1020 | 1948 | 143 | 1041 | 633 |  |
| 16.5 |  | 3736 | 4061 | 6209 | 2280 | 2738 | 4918 | 2134 | 2081 | 128 |  | 54 | 143 | 353 | 457 |  |
| 17.0 |  | 3736 | 677 | 1913 | 456 | 827 | 1109 | 1361 | 289 |  |  |  |  |  |  |  |
| 17.5 |  |  |  |  |  |  | 407 |  | 23 |  |  |  |  | 353 |  |  |
| 18.0 |  |  |  | 283 |  |  | 296 |  |  |  |  |  |  |  |  |  |
| 18.5 |  |  |  |  |  |  |  |  | 592 |  |  |  |  |  |  |  |

Table 3.2.5.2 Boarfish in ICES Subareas 27.6, 7, 8. Number of samples collected from the catch per year

| Year | Landings | Percent landings covered by sampling | No. samples | No. measured | No. aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 120 | 0 | 0 | 0 | 0 |
| 2002 | 91 | 0 | 0 | 0 | 0 |
| 2003 | 458 | 0 | 0 | 0 | 0 |
| 2004 | 675 | 0 | 0 | 0 | 0 |
| 2005 | 165 | 0 | 0 | 0 | 0 |
| 2006 | 2772 | 0 | 0 | 0 | 0 |
| 2007 | 18387 | NA | 3 | 217 | 0 |
| 2008 | 24683 | NA | 1 | 152 | 0 |
| 2009 | 83688 | NA | 9 | 1475 | 0 |
| 2010 | 137503 | NA | 95 | 10675 | 403* |
| 2011 | 31295 | NA | 27 | 4066 | 704 |
| 2012 | 80720 | NA | 80(68)*** | 9656(8565)*** | 814** |
| 2013 | 69812 | NA | 76 | 9392 | 0**** |
| 2014 | 43418 | NA | 54 | 7008 | 0**** |
| 2015 | 16837 | NA | 32 | 3356 | 0**** |
| 2016 | 18031 | NA | 27 | 3861 | 0**** |
| 2017 | 16215 | NA | 18 | 1140 | 0**** |
| 2018 | 9927 | NA | 12 | 556 | 0**** |
| 2019 | 11006 | NA | 8 | 371 | 0**** |
| 2020 | 15451 | NA | 10 | 534 | 0**** |
| 2021 | 17042 | NA | 12 | 564 | 0**** |
| 2022 | 16860 | NA | 21 | 4756 | 0**** |

* A common ALK was developed from fish collected from both commercial and survey samples. This comprehensive ALK was used to produce catch numbers at age data for pseudo-cohort analyses.
** A common ALK was developed from fish collected from Danish, Irish and Scottish commercial landings. This comprehensive ALK was used for all métiers to produce catch numbers-at-age for the pseudo-cohort analysis.

Only aged fish measured to the 0.5 cm were included in the ALK.
*** Only Irish collected samples were used for the length frequency, see stock annex.
**** 2012 ALK was used.

Table 3.2.5.3. Boarfish in ICES Subareas 27.6, 7, 8. Catch per country and corresponding number of samples collected in 2021

| Official catch | Country | No. samples | No. measured | No. aged |
| :---: | :---: | :---: | :---: | :---: |
| 1859 | DK | 1 | 263 | 0 |
| 10 | ES | 0 | 0 | 0 |
| 14055 | IE | 20 | 4493 | 0 |
| 858 | NL | 0 | 0 | 0 |
| 5 | PT | 0 | 0 | 0 |
| 73 | UKE | 0 | 0 | 0 |

Table 3.3.1.1. Boarfish in ICES Subareas 27.6, 7, 8. General boarfish age length key produced from 2012 commercial samples. Figures highlighted in grey are estimated

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7.25 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 7.75 | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 8.25 |  | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 8.75 |  | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 9.25 |  | 1 | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 9.75 |  |  | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 10.25 |  |  | 1 |  |  |  |  |  |  |  |  |  |  |  |  |
| 10.75 |  |  | 2 | 10 | 3 |  |  |  |  |  |  |  |  |  |  |
| 11.25 |  |  | 1 | 29 | 14 | 2 | 2 |  |  |  |  |  |  |  |  |
| 11.75 |  |  |  | 9 | 21 | 21 | 18 | 2 | 2 | 1 |  |  |  |  |  |
| 12.25 |  |  |  | 4 | 17 | 22 | 38 | 12 | 8 |  |  |  |  |  | 1 |
| 12.75 |  |  |  |  | 5 | 9 | 42 | 37 | 14 | 6 | 2 |  | 1 | 1 | 1 |
| 13.25 |  |  |  |  | 2 | 4 | 31 | 28 | 24 | 12 | 6 | 2 | 3 | 1 | 5 |
| 13.75 |  |  |  |  | 1 | 3 | 25 | 22 | 21 | 14 | 6 | 5 | 4 | 2 | 11 |
| 14.25 |  |  |  |  |  |  | 6 | 8 | 18 | 22 | 8 | 3 | 7 | 1 | 20 |
| 14.75 |  |  |  |  |  | 1 | 1 | 2 | 3 | 8 | 1 | 6 | 6 | 6 | 30 |
| 15.25 |  |  |  |  |  |  | 1 | 1 |  | 2 | 2 | 2 | 5 | 2 | 19 |
| 15.75 |  |  |  |  |  |  |  |  |  | 2 |  |  |  | 2 | 19 |
| 16.25 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 8 |
| 16.75 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 17.25 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 17.75 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 18.25 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |
| 18.75 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 |

Table 3.3.1.2. Boarfish in ICES Subareas 27.6, 7, 8. Proxy catch numbers-at-age of the international catches (raised numbers in '000s) for the years 2007-2022

| Age | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0 | 0 | 1575 | 2415 | 0 | 28 | 301 | 0 | 5556 | 218 | 1862 | 314 | 17427 | 40397 | 4147 | 16995 |
| 2 | 352 | 5488 | 15043 | 11229 | 2894 | 893 | 7148 | 695 | 116135 | 2385 | 4387 | 1736 | 37620 | 57719 | 21195 | 71996 |
| 3 | 2114 | 21140 | 65744 | 72709 | 41913 | 5467 | 156680 | 49503 | 32248 | 10737 | 8830 | 2628 | 9737 | 37192 | 56256 | 359182 |
| 4 | 40851 | 105575 | 338931 | 294382 | 28148 | 41278 | 58522 | 127520 | 16588 | 25114 | 34448 | 13610 | 9944 | 26433 | 78892 | 170175 |
| 5 | 48915 | 141300 | 475619 | 567689 | 30116 | 110272 | 59797 | 93705 | 24564 | 20263 | 27266 | 15570 | 12682 | 10162 | 41988 | 85079 |
| 6 | 62713 | 195339 | 543707 | 878363 | 175696 | 146582 | 68949 | 67275 | 26566 | 18025 | 21103 | 14731 | 12716 | 2583 | 16995 | 31560 |
| 7 | 26132 | 104031 | 307333 | 522703 | 143967 | 492078 | 302967 | 193061 | 74115 | 61229 | 55189 | 38686 | 29513 | 9113 | 22437 | 39024 |
| 8 | 29766 | 66570 | 172783 | 293719 | 107126 | 365840 | 250341 | 139124 | 52052 | 47573 | 38229 | 26821 | 18819 | 7487 | 8077 | 11792 |
| 9 | 56075 | 53159 | 155477 | 276672 | 77861 | 271916 | 212318 | 121042 | 44615 | 42478 | 32258 | 23670 | 15875 | 7897 | 7021 | 9566 |
| 10 | 44875 | 46893 | 130148 | 232122 | 60022 | 173486 | 160137 | 94225 | 34264 | 35150 | 25716 | 19395 | 11359 | 8164 | 5266 | 7206 |
| 11 | 14019 | 15289 | 42521 | 78588 | 46079 | 69396 | 63025 | 36078 | 12999 | 13297 | 9560 | 7148 | 4272 | 3049 | 1818 | 2218 |
| 12 | 32359 | 21178 | 61350 | 114600 | 40468 | 40968 | 41490 | 24895 | 9114 | 9132 | 7564 | 5846 | 2937 | 2786 | 1532 | 2085 |
| 13 | 4848 | 11854 | 39609 | 59932 | 24352 | 58888 | 59380 | 36309 | 13362 | 13774 | 10922 | 8183 | 4256 | 4152 | 2316 | 3359 |
| 14 | 16837 | 13570 | 31569 | 59060 | 19724 | 30277 | 30355 | 19064 | 7152 | 6682 | 5924 | 4554 | 2156 | 2333 | 1314 | 2379 |
| 15+ | 109481 | 112947 | 196967 | 349320 | 157707 | 217260 | 239366 | 150688 | 59139 | 49589 | 40797 | 32130 | 14864 | 17663 | 10006 | 20904 |

Table 3.3.2.1. Boarfish in ICES Subareas 27.6. 7, 8. Acoustic survey abundance and biomass estimates

| $\begin{aligned} & \text { Ag } \\ & \mathrm{e} \end{aligned}$ | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 |  |  |  |  |  |  |  |  |  | 1084 | 259 |  |  |
| 1 | 5 | 22 |  |  | 199 | 5 | 111 | 77 | 782 | 897 | 9523 | 587 | 14 |
| 2 | 12 | 11 | 78 |  | 319 | 36 | 127 | 31 | 389 | 1157 | 3392 | 3234 | 299 |
| 3 | 58 | 174 | 1843 | 15 | 17 | 46 | 345 | 115 | 97 | 967 | 2955 | 7537 | 4011 |
| 4 | 187 | 65 | 696 | 98 | 34 | 44 | 367 | 68 | 93 | 113 | 1315 | 4259 | 4737 |
| 5 | 437 | 95 | 382 | 102 | 80 | 6 | 156 | 107 | 88 | 157 | 463 | 619 | 1143 |
| 6 | 1166 | 736 | 254 | 105 | 112 | 10 | 209 | 166 | 106 | 183 | 150 | 509 | 1834 |
| 7 | 1184 | 974 | 1057 | 415 | 437 | 169 | 493 | 321 | 446 | 913 | 953 | 752 | 1857 |
| 8 | 704 | 759 | 879 | 344 | 363 | 113 | 463 | 198 | 183 | 885 | 207 | 266 | 554 |
| 9 | 1095 | 849 | 801 | 342 | 354 | 118 | 397 | 293 | 288 | 721 | 378 | 302 | 470 |
| 10 | 1032 | 956 | 704 | 332 | 360 | 97 | 286 | 625 | 290 | 331 | 249 | 122 | 191 |
| 11 | 333 | 651 | 264 | 130 | 132 | 17 | 121 | 339 | 50 | 81 | 151 | 41 | 19 |
| 12 | 653 | 1100 | 203 | 105 | 113 | 32 | 82 | 264 | 192 | 195 | 188 | 23 | 14 |
| 13 | 336 | 857 | 297 | 166 | 174 | 49 | 74 | 198 | 79 | 299 | 81 | 127 | 161 |
| 14 | 385 | 656 | 170 | 89 | 108 | 18 | 220 | 117 | 57 | 267 | 327 | 90 | 100 |
| $\begin{aligned} & 15 \\ & + \end{aligned}$ | 3519 | 6354 | 1464 | 855 | 1195 | 400 | 931 | 302 | 759 | 1641 | 1213 | 148 | 685 |
| $\begin{aligned} & \mathrm{TS} \\ & \mathrm{~N} \end{aligned}$ | 11104 | 14257 | 9091 | 3098 | 3996 | 1157 | 4387 | 3221 | 3899 | 9888 | 21805 | 18614 | 16091 |
| TS B | $\begin{aligned} & 67017 \\ & 6 \end{aligned}$ | $\begin{aligned} & 86344 \\ & 6 \end{aligned}$ | $\begin{aligned} & 43989 \\ & 0 \end{aligned}$ | 18777 9 | $\begin{aligned} & 23263 \\ & 4 \end{aligned}$ | 6969 0 | $\begin{aligned} & 23006 \\ & 2 \end{aligned}$ | $\begin{aligned} & 18625 \\ & 2 \end{aligned}$ | $\begin{aligned} & 17915 \\ & 6 \end{aligned}$ | $\begin{aligned} & 39987 \\ & 2 \end{aligned}$ | $\begin{aligned} & 44377 \\ & 7 \end{aligned}$ | $\begin{aligned} & 45141 \\ & 5 \end{aligned}$ | $\begin{aligned} & 52570 \\ & 1 \end{aligned}$ |
| SS B | 66939 2 | 86154 4 | 42315 8 | 18765 4 | 22665 9 | 6910 3 | $\begin{aligned} & 21881 \\ & 0 \end{aligned}$ | $\begin{aligned} & 18462 \\ & 4 \end{aligned}$ | $\begin{aligned} & 16921 \\ & 3 \end{aligned}$ | $\begin{aligned} & 35787 \\ & 1 \end{aligned}$ | $\begin{aligned} & 35195 \\ & 5 \end{aligned}$ | $\begin{aligned} & 42272 \\ & 2 \end{aligned}$ | $\begin{aligned} & 52358 \\ & 8 \end{aligned}$ |
| CV | 21.2 | 10.6 | 17.5 | 15.1 | 17.0 | 19 | 21.9 | 19.9 | 25.4 | 34.8 | 31.0 | 24.0 | 11.0 |

3.3.3.1. Boarfish in ICES Subareas 27.6, 7, 8. IBTS length-frequency data

EVHOE

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1997 | 0 | 5 | 12 | 7 | 17 | 195 | 2645 | 5006 | 3691 | 3570 | 4422 | 12054 | 16633 | 7200 | 3472 | 503 | 18 | 1 | 0 | 0 |
| 1998 | 0 | 1 | 4 | 25 | 70 | 2083 | 18263 | 8566 | 6117 | 5961 | 7082 | 11828 | 14363 | 9600 | 5261 | 971 | 8 | 0 | 0 | 1 |
| 1999 | 0 | 0 | 13 | 52 | 33 | 245 | 10949 | 25911 | 23235 | 6484 | 2818 | 4632 | 7780 | 6151 | 1357 | 268 | 8 | 0 | 0 | 0 |
| 2000 | 0 | 17 | 79 | 120 | 8 | 1508 | 26901 | 17725 | 9864 | 22076 | 16424 | 29584 | 36849 | 16508 | 5399 | 988 | 76 | 0 | 0 | 0 |
| 2001 | 0 | 1 | 45 | 687 | 490 | 916 | 21328 | 37173 | 13322 | 28492 | 31640 | 18378 | 12315 | 6507 | 3193 | 1272 | 81 | 4 | 0 | 0 |
| 2002 | 0 | 2 | 18 | 23 | 11 | 547 | 9634 | 29844 | 17728 | 13175 | 9280 | 9513 | 9615 | 6185 | 2458 | 642 | 37 | 1 | 1 | 0 |
| 2003 | 0 | 0 | 17 | 47 | 17 | 57 | 426 | 1663 | 7155 | 20073 | 24977 | 21358 | 21939 | 15004 | 7355 | 1599 | 35 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 33 | 534 | 397 | 123 | 1248 | 1420 | 1308 | 1083 | 3102 | 7308 | 7224 | 6353 | 7866 | 3630 | 241 | 5 | 0 | 0 |
| 2005 | 0 | 2 | 94 | 964 | 1264 | 146 | 1097 | 2302 | 1225 | 1551 | 3182 | 13394 | 15782 | 9879 | 6012 | 1658 | 117 | 70 | 0 | 0 |
| 2006 | 1 | 26 | 111 | 77 | 74 | 15506 | 37545 | 10729 | 3611 | 2128 | 1518 | 1960 | 4165 | 4024 | 2601 | 940 | 93 | 2 | 12 | 0 |
| 2007 | 0 | 7 | 188 | 473 | 234 | 1511 | 22812 | 127331 | 65589 | 6442 | 6823 | 5477 | 6110 | 6003 | 4268 | 1411 | 118 | 11 | 0 | 0 |
| 2008 | 0 | 3 | 432 | 2795 | 823 | 5487 | 54355 | 256210 | 169633 | 163128 | 69199 | 38406 | 18310 | 17213 | 9157 | 3486 | 745 | 6 | 1 | 0 |
| 2009 | 0 | 6 | 128 | 194 | 69 | 1482 | 19663 | 35649 | 5260 | 3906 | 9562 | 12271 | 9402 | 10835 | 6722 | 775 | 39 | 1 | 0 | 0 |
| 2010 | 0 | 21 | 529 | 116 | 154 | 5774 | 46490 | 74999 | 27177 | 12168 | 37971 | 59369 | 38501 | 37683 | 15699 | 1555 | 248 | 8 | 1 | 0 |
| 2011 | 0 | 61 | 95 | 214 | 5 | 536 | 2232 | 8210 | 14905 | 32671 | 29788 | 50316 | 56963 | 36588 | 11723 | 3058 | 572 | 159 | 47 | 0 |
| 2012 | 0 | 9 | 146 | 594 | 142 | 2913 | 28823 | 26800 | 6124 | 11739 | 13607 | 22370 | 37138 | 44084 | 19963 | 4893 | 127 | 1 | 0 | 0 |
| 2013 | 0 | 3 | 48 | 92 | 10 | 305 | 2187 | 2141 | 2558 | 13769 | 9938 | 15006 | 37563 | 40266 | 20130 | 6888 | 686 | 0 | 3 | 0 |
| 2014 | 0 | 2 | 693 | 1386 | 508 | 84 | 1440 | 885 | 3074 | 8732 | 28586 | 39397 | 74122 | 69736 | 26871 | 3908 | 59 | 433 | 0 | 0 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 0 | 5 | 183 | 5898 | 4143 | 607 | 19075 | 179269 | 119004 | 15765 | 18014 | 61575 | 62024 | 59904 | 21525 | 5487 | 541 | 429 | 8 | 0 |
| 2016 | 5 | 31 | 379 | 846 | 115 | 733 | 10284 | 14280 | 17251 | 42132 | 25304 | 68583 | 130633 | 131220 | 48538 | 11611 | 1358 | 26 | 0 | 0 |
| 2018 | 0 | 14 | 4957 | 193861 | 173779 | 210 | 10910 | 76288 | 48343 | 29096 | 45773 | 85164 | 132174 | 157883 | 48603 | 14951 | 592 | 18 | 0 | 0 |
| 2019 | 2 | 997 | 6467 | 589 | 10688 | 531908 | 561517 | 329850 | 59733 | 4505 | 3418 | 8451 | 32547 | 61582 | 30031 | 7468 | 962 | 204 | 0 | 0 |
| 2020 | 3 | 283 | 1280 | 657 | 21381 | 408706 | 595107 | 142947 | 218153 | 421028 | 220190 | 54726 | 70612 | 97364 | 74415 | 30606 | 4736 | 1 | 0 | 0 |
| 2021 | 0 | 35 | 166 | 27 | 32861 | 954046 | 852223 | 313053 | 640456 | 208802 | 106995 | 57674 | 96633 | 65504 | 12047 | 3416 | 387 | 53 | 0 | 0 |
| 2022 | 0 | 25 | 78 | 66 | 15 | 6041 | 88521 | 309670 | 433131 | 585482 | 250402 | 80362 | 41907 | 18504 | 4317 | 885 | 68 | 0 | 0 | 0 |

IGFS

| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2003 | 0 | 1 | 33 | 22 | 7 | 22 | 129 | 172 | 879 | 2942 | 2322 | 1325 | 3823 | 4629 | 2898 | 896 | 163 | 38 | 0 | 0 |
| 2004 | 0 | 23 | 63 | 34 | 8 | 117 | 628 | 1444 | 423 | 397 | 464 | 2276 | 4325 | 4709 | 3972 | 1019 | 90 | 5 | 1 | 0 |
| 2005 | 0 | 8 | 59 | 52 | 20 | 203 | 1024 | 585 | 288 | 636 | 341 | 3463 | 11457 | 11348 | 7955 | 1744 | 382 | 2 | 1 | 0 |
| 2006 | 5 | 60 | 68 | 48 | 35 | 212 | 969 | 621 | 2046 | 4190 | 8044 | 7946 | 24208 | 42119 | 32168 | 12296 | 2454 | 532 | 0 | 0 |
| 2007 | 1 | 6 | 44 | 18 | 31 | 501 | 923 | 1251 | 1638 | 1166 | 2510 | 3581 | 8275 | 10740 | 7093 | 1934 | 92 | 0 | 0 | 0 |
| 2008 | 0 | 0 | 26 | 18 | 23 | 127 | 672 | 531 | 2095 | 13780 | 17664 | 19268 | 16980 | 19484 | 15953 | 8789 | 1747 | 76 | 1 | 0 |
| 2009 | 0 | 3 | 80 | 76 | 25 | 94 | 228 | 486 | 1000 | 1139 | 9081 | 7749 | 5138 | 6921 | 5592 | 1084 | 68 | 1 | 0 | 0 |
| 2010 | 0 | 6 | 42 | 3 | 18 | 199 | 272 | 463 | 920 | 393 | 7914 | 34236 | 28611 | 16063 | 8161 | 1974 | 433 | 0 | 0 | 0 |
| 2011 | 0 | 7 | 17 | 5 | 4 | 189 | 772 | 592 | 556 | 669 | 2600 | 20246 | 22121 | 10851 | 5319 | 2218 | 269 | 9 | 6 | 0 |
| 2012 | 0 | 7 | 36 | 20 | 10 | 130 | 271 | 378 | 702 | 2143 | 1183 | 11104 | 34005 | 22731 | 10905 | 3901 | 525 | 4 | 0 | 0 |
| 2013 | 1 | 3 | 9 | 9 | 20 | 127 | 352 | 340 | 1320 | 2833 | 3971 | 15572 | 51637 | 52868 | 20485 | 6560 | 492 | 20 | 0 | 0 |


| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2014 | 0 | 10 | 68 | 54 | 4 | 18 | 13 | 25 | 60 | 130 | 1127 | 3251 | 19125 | 23016 | 10355 | 2988 | 284 | 18 | 0 | 0 |
| 2015 | 0 | 3 | 11 | 16 | 24 | 193 | 1008 | 3708 | 848 | 105 | 713 | 6315 | 29727 | 48220 | 33024 | 17350 | 1885 | 531 | 0 | 0 |
| 2016 | 4 | 31 | 121 | 63 | 7 | 67 | 187 | 1515 | 4057 | 2891 | 1349 | 4111 | 32753 | 57753 | 40907 | 15527 | 3670 | 85 | 0 | 0 |
| 2017 | 0 | 0 | 37 | 131 | 48 | 132 | 460 | 652 | 11411 | 20321 | 5909 | 5520 | 16426 | 33117 | 29972 | 15815 | 3194 | 369 | 0 | 0 |
| 2018 | 4 | 51 | 247 | 139 | 32 | 45 | 286 | 585 | 1194 | 6107 | 17005 | 15168 | 48895 | 61833 | 36519 | 10722 | 2030 | 63 | 0 | 0 |
| 2019 | 4 | 19 | 117 | 47 | 52 | 262 | 583 | 173 | 106 | 487 | 2677 | 4967 | 6863 | 12080 | 10480 | 5125 | 772 | 71 | 4 | 0 |
| 2020 | 9 | 388 | 233 | 21 | 16 | 1772 | 2052 | 13941 | 65121 | 24505 | 7709 | 17859 | 12157 | 17223 | 9125 | 2499 | 110 | 2 | 0 | 0 |
| 2021 | 2 | 7 | 98 | 36 | 293 | 16275 | 125036 | 87742 | 210710 | 171970 | 67893 | 20086 | 16044 | 22040 | 23112 | 4589 | 816 | 7 | 1 | 0 |
| 2022 | 0 | 8 | 43 | 133 | 59 | 611 | 2778 | 26375 | 45984 | 58006 | 106343 | 51790 | 21705 | 25100 | 20712 | 7111 | 2614 | 318 | 70 | 0 |

SPNGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 0 | 0 | 8 | 0 | 16 | 317 | 1817 | 2496 | 260 | 141 | 154 | 314 | 632 | 613 | 689 | 97 | 0 | 0 | 0 | 0 |
| 1991 | 0 | 1 | 0 | 0 | 31 | 690 | 1311 | 313 | 49 | 9 | 6 | 7 | 7 | 4 | 0 | 0 | 0 | 6 | 0 | 0 |
| 1992 | 0 | 57 | 38 | 9 | 178 | 3290 | 2743 | 282 | 48 | 10 | 8 | 69 | 162 | 390 | 779 | 246 | 95 | 0 | 0 | 0 |
| 1993 | 0 | 57 | 1206 | 488 | 97 | 3730 | 3753 | 421 | 105 | 54 | 7 | 4 | 8 | 3 | 2 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 1 | 40 | 33 | 0 | 342 | 4789 | 10162 | 8920 | 3195 | 53 | 106 | 20 | 9 | 12 | 1 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 0 | 84 | 108 | 4 | 342 | 3063 | 2157 | 220 | 84 | 65 | 58 | 105 | 105 | 90 | 20 | 4 | 0 | 0 | 0 | 0 |
| 1996 | 0 | 218 | 537 | 143 | 245 | 4457 | 4449 | 267 | 820 | 722 | 82 | 145 | 126 | 219 | 96 | 39 | 2 | 0 | 0 | 0 |
| 1997 | 2 | 102 | 809 | 441 | 235 | 3458 | 6824 | 2189 | 1923 | 534 | 156 | 353 | 161 | 88 | 3 | 0 | 0 | 0 | 0 | 0 |
| 1998 | 3 | 2 | 7 | 4 | 49 | 1920 | 4685 | 2217 | 337 | 153 | 125 | 88 | 147 | 135 | 86 | 13 | 2 | 3 | 0 | 0 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 0 | 6 | 59 | 13 | 134 | 2736 | 3010 | 193 | 106 | 83 | 109 | 143 | 390 | 645 | 402 | 69 | 0 | 0 | 0 | 0 |
| 2000 | 0 | 7 | 3729 | 2046 | 17 | 554 | 1947 | 489 | 277 | 486 | 756 | 1252 | 999 | 1021 | 199 | 34 | 13 | 0 | 0 | 0 |
| 2001 | 0 | 68 | 4 | 1 | 153 | 3241 | 5085 | 659 | 225 | 206 | 205 | 236 | 692 | 407 | 120 | 22 | 9 | 0 | 0 | 0 |
| 2002 | 0 | 4 | 20 | 0 | 133 | 2333 | 2013 | 284 | 50 | 58 | 54 | 60 | 231 | 314 | 72 | 9 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 4 | 950 | 567 | 4 | 77 | 221 | 57 | 39 | 28 | 16 | 22 | 17 | 23 | 16 | 5 | 1 | 0 | 0 | 0 |
| 2004 | 0 | 6 | 22 | 4 | 43 | 2289 | 3808 | 443 | 110 | 83 | 58 | 219 | 931 | 776 | 303 | 2 | 1 | 0 | 0 | 0 |
| 2005 | 0 | 16 | 451 | 25 | 9 | 754 | 1007 | 207 | 85 | 102 | 30 | 54 | 257 | 218 | 90 | 44 | 2 | 0 | 0 | 0 |
| 2006 | 0 | 14 | 156 | 160 | 50 | 2238 | 8913 | 4507 | 175 | 94 | 9 | 36 | 229 | 419 | 169 | 9 | 2 | 0 | 0 | 0 |
| 2007 | 0 | 49 | 40 | 1 | 111 | 3025 | 6620 | 1099 | 129 | 260 | 81 | 7 | 93 | 215 | 89 | 21 | 3 | 0 | 0 | 0 |
| 2008 | 7 | 4 | 92 | 247 | 1 | 936 | 1561 | 1326 | 234 | 1483 | 304 | 537 | 11 | 833 | 201 | 186 | 11 | 0 | 0 | 0 |
| 2009 | 1 | 17 | 62 | 119 | 11 | 2587 | 3893 | 4070 | 119 | 250 | 45 | 142 | 59 | 819 | 120 | 17 | 1 | 1 | 0 | 0 |
| 2010 | 0 | 55 | 102 | 5 | 232 | 13090 | 22032 | 3169 | 1160 | 1056 | 89 | 82 | 179 | 1007 | 1981 | 518 | 9 | 0 | 0 | 0 |
| 2011 | 0 | 29 | 260 | 105 | 46 | 2805 | 5511 | 1278 | 148 | 340 | 145 | 100 | 144 | 591 | 724 | 134 | 3 | 1 | 0 | 0 |
| 2012 | 0 | 29 | 132 | 35 | 556 | 7550 | 7844 | 1364 | 88 | 53 | 59 | 170 | 1051 | 2394 | 1553 | 432 | 21 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 2 | 11 | 126 | 2163 | 4664 | 854 | 302 | 609 | 251 | 61 | 113 | 134 | 156 | 81 | 8 | 0 | 0 | 0 |
| 2014 | 0 | 75 | 117 | 6 | 12 | 263 | 465 | 79 | 1083 | 1175 | 1174 | 1266 | 998 | 2444 | 3623 | 817 | 31 | 1 | 0 | 0 |
| 2015 | 0 | 13 | 67 | 3 | 58 | 1889 | 4248 | 534 | 75 | 465 | 750 | 970 | 695 | 1173 | 1473 | 453 | 70 | 1 | 0 | 0 |
| 2016 | 0 | 17 | 99 | 5 | 41 | 922 | 2423 | 473 | 925 | 746 | 346 | 548 | 452 | 561 | 169 | 22 | 4 | 0 | 0 | 0 |
| 2017 | 1 | 23 | 20 | 1 | 16 | 641 | 1947 | 755 | 134 | 165 | 285 | 405 | 579 | 967 | 936 | 177 | 13 | 3 | 0 | 0 |
| 2018 | 0 | 0 | 2 | 0 | 45 | 708 | 1635 | 258 | 43 | 99 | 230 | 605 | 1370 | 3324 | 3865 | 949 | 3 | 0 | 0 | 2 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2019 | 0 | 12 | 2 | 1 | 259 | 4128 | 3887 | 379 | 18 | 83 | 273 | 329 | 717 | 4200 | 8402 | 2215 | 202 | 0 | 0 | 0 |
| 2020 | 0 | 8 | 33 | 2 | 33 | 1218 | 2123 | 525 | 387 | 314 | 75 | 225 | 705 | 2518 | 4751 | 1603 | 10 | 0 | 0 | 0 |
| 2021 | 1 | 10 | 11 | 0 | 42 | 803 | 2654 | 562 | 127 | 1367 | 3149 | 1102 | 2200 | 4773 | 6485 | 1175 | 118 | 1 | 0 | 0 |
| 2022 | 0 | 12 | 38 | 2 | 35 | 1640 | 3395 | 550 | 220 | 465 | 1945 | 3197 | 3705 | 9321 | 7926 | 3200 | 39 | 0 | 0 | 0 |

SPPGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 0 | 1 | 0 | 1 | 1 | 2 | 0 | 44 | 5 | 52 | 133 | 162 | 667 | 1129 | 230 | 40 | 0 | 0 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 90 | 212 | 791 | 843 | 313 | 60 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 15 | 22 | 21 | 62 | 268 | 426 | 249 | 51 | 2 | 1 | 0 | 0 |
| 2004 | 0 | 1 | 0 | 0 | 0 | 6 | 3 | 0 | 5 | 6 | 23 | 124 | 385 | 592 | 390 | 52 | 1 | 0 | 0 | 0 |
| 2005 | 0 | 1 | 0 | 1 | 8 | 1 | 20 | 11 | 10 | 16 | 8 | 118 | 628 | 1118 | 833 | 272 | 23 | 0 | 0 | 0 |
| 2006 | 0 | 0 | 1 | 1 | 8 | 120 | 118 | 26 | 43 | 95 | 34 | 58 | 431 | 863 | 716 | 252 | 13 | 1 | 0 | 0 |
| 2007 | 0 | 0 | 0 | 0 | 4 | 5 | 12 | 20 | 16 | 12 | 37 | 34 | 96 | 202 | 191 | 34 | 5 | 0 | 0 | 0 |
| 2008 | 0 | 1 | 0 | 0 | 0 | 1 | 17 | 10 | 23 | 19 | 79 | 156 | 349 | 666 | 442 | 113 | 7 | 0 | 0 | 0 |
| 2009 | 0 | 8 | 7 | 0 | 3 | 10 | 11 | 1 | 0 | 2 | 220 | 457 | 1333 | 1746 | 1698 | 474 | 11 | 0 | 0 | 0 |
| 2010 | 2 | 0 | 0 | 1 | 6 | 17 | 4 | 1 | 6 | 3 | 43 | 390 | 710 | 976 | 620 | 164 | 13 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 20 | 22 | 6 | 180 | 815 | 960 | 522 | 151 | 17 | 0 | 2 | 0 |
| 2012 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 2 | 2 | 1 | 10 | 87 | 456 | 570 | 267 | 79 | 4 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 1 | 0 | 8 | 24 | 7 | 10 | 0 | 1 | 48 | 500 | 1032 | 564 | 163 | 15 | 1 | 0 | 0 |
| 2014 | 0 | 10 | 9 | 0 | 1 | 0 | 3 | 17 | 62 | 11 | 6 | 85 | 2453 | 6703 | 3168 | 2115 | 162 | 82 | 0 | 0 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2015 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 1 | 1 | 0 | 0 | 32 | 300 | 471 | 316 | 151 | 43 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 3 | 0 | 0 | 0 | 1 | 0 | 13 | 7 | 0 | 9 | 157 | 336 | 220 | 84 | 19 | 0 | 0 | 0 |
| 2017 | 0 | 67 | 19 | 0 | 0 | 0 | 10 | 0 | 0 | 1 | 18 | 26 | 148 | 498 | 529 | 268 | 17 | 0 | 0 | 0 |
| 2018 | 0 | 2 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 37 | 1159 | 3574 | 2449 | 1131 | 159 | 0 | 0 | 0 |
| 2019 | 5 | 36 | 4 | 0 | 0 | 0 | 0 | 0 | 3 | 4 | 0 | 15 | 426 | 952 | 796 | 192 | 15 | 0 | 0 | 0 |
| 2020 | 0 | 5 | 1 | 0 | 0 | 4 | 1 | 1 | 2 | 4 | 0 | 26 | 250 | 616 | 851 | 661 | 111 | 0 | 0 | 1 |
| 2021 | 1 | 20 | 0 | 0 | 5 | 12 | 0 | 5 | 34 | 38 | 24 | 39 | 129 | 916 | 768 | 357 | 147 | 3 | 0 | 0 |
| 2022 | 1 | 0 | 3 | 6 | 4 | 6 | 3 | 9 | 16 | 79 | 102 | 45 | 74 | 395 | 409 | 222 | 42 | 3 | 0 | 0 |

WCSGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 0 | 0 | 3 | 2 | 0 | 3 | 24 | 42 | 62 | 172 | 210 | 1286 | 856 | 450 | 52 | 17 | 0 | 0 | 0 | 0 |
| 1991 | 0 | 0 | 0 | 2 | 0 | 31 | 138 | 80 | 183 | 644 | 683 | 848 | 226 | 89 | 12 | 1 | 2 | 4 | 0 | 0 |
| 1992 | 0 | 0 | 0 | 1 | 0 | 8 | 12 | 14 | 44 | 478 | 1160 | 4028 | 1674 | 502 | 5 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 0 | 0 | 0 | 0 | 0 | 1 | 109 | 2 | 670 | 2078 | 1074 | 4904 | 2753 | 2882 | 28 | 2 | 0 | 0 | 0 | 0 |
| 1994 | 0 | 0 | 2 | 0 | 0 | 0 | 15 | 30 | 30 | 205 | 283 | 312 | 454 | 388 | 147 | 0 | 0 | 0 | 0 | 0 |
| 1995 | 8 | 12 | 18 | 4 | 2 | 10 | 40 | 30 | 94 | 162 | 640 | 1485 | 1770 | 1139 | 318 | 14 | 2 | 4 | 6 | 0 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1996 | 0 | 0 | 0 | 4 | 0 | 10 | 48 | 27 | 49 | 48 | 64 | 188 | 920 | 1888 | 416 | 18 | 1 | 0 | 0 | 0 |
| 1997 | 0 | 0 | 4 | 0 | 0 | 1 | 17 | 42 | 120 | 64 | 116 | 249 | 436 | 301 | 91 | 8 | 4 | 0 | 0 | 0 |
| 1998 | 0 | 0 | 0 | 1 | 0 | 1 | 7 | 6 | 7 | 16 | 47 | 69 | 105 | 171 | 78 | 8 | 2 | 0 | 0 | 0 |
| 1999 | 0 | 0 | 1 | 0 | 0 | 2 | 6 | 8 | 189 | 221 | 312 | 458 | 346 | 221 | 69 | 0 | 0 | 0 | 0 | 0 |
| 2000 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 3 | 42 | 118 | 230 | 303 | 206 | 108 | 54 | 8 | 0 | 0 | 0 | 0 |
| 2001 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 12 | 27 | 54 | 90 | 233 | 414 | 242 | 80 | 15 | 1 | 0 | 0 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 1 | 8 | 2 | 1 | 82 | 759 | 3243 | 5711 | 5896 | 1558 | 189 | 1 | 0 | 0 | 0 |
| 2003 | 0 | 0 | 1 | 0 | 0 | 0 | 3 | 52 | 9 | 107 | 326 | 1536 | 3294 | 5409 | 3553 | 413 | 37 | 0 | 0 | 0 |
| 2004 | 0 | 0 | 0 | 1 | 0 | 0 | 6 | 2 | 45 | 83 | 744 | 4576 | 8611 | 9526 | 5698 | 954 | 84 | 0 | 0 | 0 |
| 2005 | 0 | 2 | 0 | 0 | 0 | 9 | 38 | 15 | 30 | 31 | 113 | 442 | 1115 | 1747 | 818 | 141 | 9 | 3 | 2 | 0 |
| 2006 | 0 | 1 | 2 | 1 | 0 | 2 | 9 | 4 | 22 | 256 | 311 | 508 | 1524 | 2964 | 2104 | 449 | 73 | 2 | 0 | 0 |
| 2007 | 0 | 0 | 3 | 2 | 0 | 8 | 14 | 65 | 118 | 182 | 795 | 2938 | 5220 | 6953 | 5332 | 1538 | 116 | 0 | 0 | 0 |
| 2008 | 0 | 1 | 3 | 0 | 0 | 16 | 37 | 38 | 200 | 482 | 1406 | 3218 | 9904 | 22777 | 18407 | 6293 | 575 | 71 | 0 | 0 |
| 2009 | 0 | 0 | 1 | 0 | 1 | 1 | 4 | 6 | 64 | 2460 | 2246 | 694 | 505 | 416 | 338 | 136 | 12 | 0 | 0 | 0 |
| 2010 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 530 | 1443 | 1384 | 1357 | 828 | 149 | 29 | 0 | 0 | 0 |

Table 3.3.3.2 Boarfish in ICES Subareas 27.6, 7, 8. IBTS length-frequency data converted to age-structured indices by application of the 2012 common ALK rounded down to 1 cm length classes
EVHOE

| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 4}$ |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1997 | 1323 | 5891 | 4835 | 3829 | 3369 | 3053 | 9614 | 6955 | 5556 | 3779 | 1521 | 973 | 1456 | 828 | 6235 |
| 1998 | 9132 | 16881 | 8109 | 6147 | 4527 | 3452 | 9545 | 6632 | 5452 | 4058 | 1597 | 1312 | 1733 | 1022 | 8419 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1999 | 5474 | 30494 | 25366 | 5015 | 2592 | 1427 | 4373 | 3215 | 2887 | 2276 | 855 | 564 | 888 | 491 | 3675 |
| 2000 | 13450 | 28555 | 16758 | 19454 | 12310 | 8420 | 23424 | 16159 | 12783 | 8538 | 3354 | 1885 | 3099 | 1722 | 12485 |
| 2001 | 10664 | 39887 | 26874 | 27998 | 16428 | 8946 | 15285 | 7816 | 5688 | 3538 | 1301 | 863 | 1271 | 750 | 6396 |
| 2002 | 4817 | 30622 | 24313 | 11299 | 6215 | 3393 | 7688 | 4838 | 3852 | 2716 | 1035 | 726 | 1060 | 611 | 4928 |
| 2003 | 213 | 3707 | 9293 | 20716 | 13365 | 8409 | 18107 | 11109 | 8937 | 6448 | 2467 | 1932 | 2635 | 1547 | 12700 |
| 2004 | 624 | 2006 | 1574 | 1777 | 1923 | 1842 | 5376 | 3816 | 3078 | 2541 | 1075 | 1423 | 1434 | 932 | 11369 |
| 2005 | 549 | 2492 | 1901 | 2205 | 2758 | 2983 | 9853 | 7261 | 5865 | 4310 | 1727 | 1437 | 1869 | 1110 | 9951 |
| 2006 | 18772 | 27129 | 6395 | 1838 | 1086 | 692 | 2217 | 1683 | 1593 | 1407 | 557 | 586 | 688 | 416 | 4256 |
| 2007 | 11406 | 118156 | 87434 | 6252 | 3796 | 2250 | 4968 | 3140 | 2686 | 2208 | 861 | 923 | 1067 | 657 | 6591 |
| 2008 | 27177 | 254528 | 229646 | 124210 | 54539 | 19047 | 30818 | 15021 | 10954 | 7348 | 2618 | 2251 | 2934 | 1795 | 16959 |
| 2009 | 9832 | 35351 | 16200 | 5643 | 4832 | 3830 | 8969 | 5783 | 4721 | 3809 | 1459 | 1524 | 1806 | 1110 | 9216 |
| 2010 | 23245 | 82303 | 45710 | 20517 | 19648 | 16749 | 39369 | 25075 | 19324 | 14156 | 5280 | 4343 | 5906 | 3511 | 26732 |
| 2011 | 1116 | 11557 | 19043 | 30617 | 20479 | 14495 | 39161 | 26846 | 21792 | 15613 | 5980 | 3928 | 6016 | 3404 | 27139 |
| 2012 | 14412 | 34320 | 15329 | 11984 | 8843 | 6877 | 21882 | 16580 | 15805 | 14165 | 5382 | 5221 | 6581 | 3893 | 34397 |
| 2013 | 1093 | 3373 | 5082 | 11975 | 7436 | 5156 | 18526 | 14722 | 14572 | 13248 | 5121 | 5049 | 6254 | 3703 | 35819 |
| 2014 | 720 | 2334 | 4216 | 15081 | 14776 | 13252 | 40953 | 30549 | 28568 | 24182 | 9208 | 7776 | 10517 | 6071 | 49039 |
| 2015 | 9537 | 168718 | 142196 | 16589 | 15129 | 14025 | 43805 | 31952 | 26892 | 21239 | 8025 | 6461 | 8982 | 5218 | 43843 |
| 2016 | 5142 | 20412 | 24368 | 35467 | 23775 | 18507 | 68150 | 53795 | 50979 | 44038 | 16743 | 14289 | 19326 | 11149 | 95082 |
| 2018 | 5455 | 72428 | 63489 | 33998 | 28889 | 24760 | 79148 | 59901 | 56898 | 49999 | 18526 | 15688 | 21690 | 12453 | 106474 |
| 2019 | 280759 | 520569 | 150645 | 4035 | 3104 | 2844 | 14950 | 13581 | 15700 | 16891 | 6358 | 7404 | 8669 | 5219 | 49538 |


| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 4}$ |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2020 | 297553 | 465569 | 273832 | 332726 | 148543 | 51435 | 79125 | 38909 | 36296 | 32676 | 12326 | 15407 | 16693 | 10460 | 118335 |
| 2021 | 426111 | 848299 | 571349 | 164881 | 76916 | 31315 | 65603 | 40367 | 35579 | 26598 | 9833 | 5812 | 9725 | 5289 | 39566 |
| 2022 | 44260 | 395084 | 503808 | 445388 | 190241 | 59875 | 81650 | 32581 | 22992 | 12777 | 4165 | 1890 | 3565 | 2040 | 12932 |

IGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 64 | 472 | 1214 | 2586 | 1401 | 743 | 2065 | 1523 | 1556 | 1484 | 578 | 653 | 750 | 456 | 4672 |
| 2004 | 314 | 1418 | 842 | 434 | 493 | 543 | 2252 | 1838 | 1732 | 1603 | 653 | 802 | 864 | 541 | 5422 |
| 2005 | 512 | 998 | 509 | 567 | 717 | 908 | 4790 | 4166 | 4162 | 3867 | 1557 | 1730 | 1973 | 1201 | 11568 |
| 2006 | 484 | 1580 | 2423 | 5269 | 4211 | 3388 | 12623 | 10487 | 11436 | 12263 | 4853 | 6606 | 6952 | 4368 | 50651 |
| 2007 | 462 | 1842 | 1748 | 1576 | 1408 | 1235 | 4362 | 3474 | 3496 | 3378 | 1326 | 1557 | 1754 | 1076 | 10509 |
| 2008 | 336 | 1388 | 4302 | 14466 | 9811 | 6581 | 15265 | 9859 | 8231 | 6912 | 2728 | 3247 | 3553 | 2238 | 28119 |
| 2009 | 114 | 772 | 1117 | 3682 | 3665 | 2967 | 5991 | 3553 | 2883 | 2398 | 928 | 1136 | 1233 | 783 | 7266 |
| 2010 | 136 | 752 | 906 | 3336 | 6161 | 7220 | 21721 | 15262 | 11417 | 7656 | 3025 | 2151 | 3055 | 1795 | 14845 |
| 2011 | 386 | 966 | 715 | 1598 | 3198 | 4038 | 13856 | 10232 | 7932 | 5384 | 2159 | 1453 | 2121 | 1224 | 10962 |
| 2012 | 136 | 622 | 1006 | 1911 | 2306 | 2843 | 13844 | 11639 | 10956 | 8966 | 3576 | 2903 | 3900 | 2242 | 21003 |
| 2013 | 176 | 843 | 1557 | 3292 | 3917 | 4545 | 21801 | 18670 | 19029 | 17278 | 6613 | 5870 | 7777 | 4484 | 40599 |
| 2014 | 6 | 43 | 82 | 492 | 927 | 1262 | 7300 | 6613 | 7255 | 7083 | 2717 | 2714 | 3384 | 1986 | 18529 |
| 2015 | 504 | 3259 | 1827 | 403 | 1251 | 1945 | 12476 | 11625 | 13072 | 13999 | 5512 | 7082 | 7697 | 4765 | 58017 |
| 2016 | 93 | 2456 | 3763 | 2302 | 1775 | 1846 | 13082 | 12553 | 14753 | 16394 | 6464 | 8634 | 9226 | 5742 | 65723 |
| 2017 | 230 | 4468 | 11683 | 14642 | 6277 | 2402 | 9024 | 7578 | 8395 | 9474 | 3824 | 5785 | 5766 | 3703 | 49915 |


| Year | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 4}$ |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2018 | 143 | 930 | 2275 | 9391 | 8194 | 6861 | 23782 | 19030 | 19873 | 19320 | 7511 | 8412 | 9756 | 5903 | 59025 |
| 2019 | 292 | 442 | 242 | 1229 | 1449 | 1419 | 4664 | 3618 | 3540 | 3626 | 1453 | 2058 | 2107 | 1346 | 16899 |
| 2020 | 1026 | 32027 | 52719 | 18043 | 8761 | 4356 | 11714 | 8061 | 6664 | 5578 | 2105 | 2193 | 2649 | 1618 | 14790 |
| 2021 | 62518 | 191249 | 202522 | 128995 | 53951 | 16137 | 23800 | 10942 | 9297 | 7968 | 3069 | 4310 | 4329 | 2815 | 28141 |
| 2022 | 1389 | 34301 | 51195 | 70323 | 46874 | 28049 | 43180 | 20071 | 14649 | 10314 | 3757 | 4203 | 4709 | 3015 | 32808 |

SPNGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 909 | 2660 | 1033 | 142 | 110 | 93 | 335 | 263 | 243 | 224 | 95 | 128 | 129 | 83 | 770 |
| 1991 | 656 | 880 | 138 | 8 | 4 | 2 | 6 | 3 | 3 | 2 | 1 | 0 | 1 | 0 | 8 |
| 1992 | 1371 | 1575 | 128 | 10 | 13 | 16 | 97 | 89 | 92 | 122 | 57 | 124 | 102 | 71 | 965 |
| 1993 | 1877 | 2192 | 220 | 36 | 13 | 2 | 5 | 3 | 2 | 2 | 1 | 0 | 1 | 0 | 3 |
| 1994 | 5081 | 12093 | 5114 | 66 | 43 | 23 | 28 | 9 | 7 | 5 | 1 | 1 | 1 | 1 | 5 |
| 1995 | 1079 | 1254 | 142 | 61 | 41 | 29 | 78 | 54 | 44 | 33 | 12 | 8 | 13 | 7 | 53 |
| 1996 | 2225 | 2676 | 772 | 479 | 175 | 40 | 109 | 77 | 70 | 65 | 24 | 25 | 31 | 18 | 181 |
| 1997 | 3412 | 5512 | 2113 | 389 | 183 | 84 | 198 | 123 | 82 | 47 | 17 | 6 | 14 | 8 | 43 |
| 1998 | 2343 | 3933 | 993 | 137 | 76 | 41 | 96 | 64 | 58 | 49 | 19 | 19 | 23 | 14 | 125 |
| 1999 | 1505 | 1669 | 151 | 88 | 66 | 53 | 202 | 168 | 181 | 188 | 73 | 89 | 100 | 61 | 556 |
| 2000 | 973 | 1392 | 445 | 562 | 447 | 351 | 877 | 582 | 475 | 359 | 130 | 88 | 138 | 78 | 577 |
| 2001 | 2542 | 3057 | 410 | 197 | 130 | 93 | 311 | 237 | 219 | 170 | 66 | 43 | 66 | 36 | 286 |
| 2002 | 1006 | 1212 | 139 | 54 | 35 | 26 | 103 | 87 | 95 | 92 | 33 | 28 | 40 | 22 | 172 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 110 | 162 | 50 | 23 | 12 | 7 | 16 | 11 | 9 | 8 | 3 | 3 | 4 | 2 | 25 |
| 2004 | 1904 | 2236 | 237 | 74 | 66 | 71 | 359 | 310 | 313 | 273 | 106 | 88 | 120 | 68 | 508 |
| 2005 | 504 | 670 | 145 | 74 | 36 | 21 | 99 | 85 | 86 | 76 | 30 | 25 | 34 | 19 | 191 |
| 2006 | 4457 | 7519 | 1636 | 62 | 27 | 14 | 93 | 89 | 106 | 114 | 42 | 46 | 56 | 33 | 268 |
| 2007 | 3310 | 4086 | 502 | 187 | 74 | 19 | 50 | 39 | 50 | 56 | 20 | 24 | 28 | 17 | 155 |
| 2008 | 781 | 1743 | 878 | 1031 | 419 | 134 | 290 | 185 | 174 | 186 | 60 | 69 | 89 | 53 | 594 |
| 2009 | 1947 | 4700 | 1483 | 173 | 75 | 31 | 113 | 100 | 138 | 174 | 56 | 59 | 81 | 46 | 363 |
| 2010 | 11016 | 13516 | 2029 | 689 | 234 | 34 | 167 | 157 | 182 | 283 | 134 | 313 | 253 | 178 | 2099 |
| 2011 | 2756 | 3657 | 590 | 260 | 117 | 46 | 134 | 106 | 121 | 158 | 67 | 127 | 114 | 77 | 791 |
| 2012 | 3922 | 4860 | 523 | 54 | 58 | 68 | 465 | 450 | 551 | 640 | 247 | 337 | 361 | 225 | 2268 |
| 2013 | 2332 | 3002 | 602 | 460 | 194 | 59 | 100 | 54 | 51 | 48 | 19 | 28 | 28 | 18 | 238 |
| 2014 | 232 | 646 | 978 | 1123 | 697 | 431 | 1071 | 739 | 675 | 751 | 325 | 610 | 539 | 367 | 3971 |
| 2015 | 2124 | 2505 | 322 | 542 | 409 | 300 | 726 | 482 | 406 | 388 | 162 | 260 | 245 | 163 | 1874 |
| 2016 | 1211 | 1835 | 917 | 584 | 300 | 157 | 397 | 267 | 226 | 184 | 67 | 55 | 77 | 45 | 347 |
| 2017 | 974 | 1522 | 374 | 199 | 161 | 129 | 397 | 301 | 291 | 298 | 121 | 178 | 178 | 115 | 1130 |
| 2018 | 817 | 1004 | 135 | 145 | 163 | 171 | 810 | 719 | 786 | 945 | 398 | 690 | 641 | 424 | 4531 |
| 2019 | 1943 | 2202 | 156 | 143 | 137 | 120 | 669 | 645 | 749 | 1182 | 560 | 1325 | 1065 | 752 | 9058 |
| 2020 | 1062 | 1540 | 492 | 224 | 113 | 68 | 460 | 447 | 505 | 731 | 341 | 759 | 623 | 436 | 5435 |
| 2021 | 1327 | 1744 | 554 | 1855 | 1300 | 818 | 1784 | 1197 | 1245 | 1445 | 616 | 1116 | 1005 | 675 | 7033 |
| 2022 | 1698 | 2138 | 433 | 955 | 1015 | 934 | 2952 | 2295 | 2350 | 2592 | 1026 | 1553 | 1560 | 1004 | 11458 |

SPPGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 0 | 31 | 29 | 77 | 73 | 68 | 300 | 262 | 304 | 308 | 110 | 94 | 135 | 76 | 596 |
| 2002 | 0 | 0 | 2 | 34 | 58 | 71 | 330 | 283 | 294 | 270 | 103 | 92 | 122 | 70 | 584 |
| 2003 | 0 | 7 | 15 | 21 | 20 | 21 | 115 | 105 | 117 | 123 | 48 | 57 | 65 | 39 | 366 |
| 2004 | 1 | 3 | 5 | 13 | 25 | 34 | 177 | 158 | 169 | 175 | 69 | 85 | 94 | 58 | 515 |
| 2005 | 10 | 21 | 14 | 14 | 25 | 38 | 264 | 251 | 288 | 319 | 126 | 172 | 182 | 114 | 1218 |
| 2006 | 59 | 91 | 56 | 71 | 39 | 28 | 184 | 176 | 209 | 242 | 97 | 142 | 145 | 92 | 1021 |
| 2007 | 6 | 25 | 20 | 20 | 18 | 15 | 54 | 46 | 50 | 58 | 23 | 36 | 36 | 23 | 230 |
| 2008 | 8 | 23 | 23 | 40 | 47 | 48 | 193 | 163 | 176 | 188 | 73 | 95 | 104 | 64 | 636 |
| 2009 | 6 | 7 | 3 | 78 | 127 | 147 | 639 | 540 | 550 | 561 | 232 | 325 | 329 | 210 | 2203 |
| 2010 | 2 | 5 | 5 | 22 | 61 | 85 | 379 | 317 | 313 | 301 | 118 | 138 | 156 | 96 | 930 |
| 2011 | 0 | 9 | 19 | 19 | 35 | 52 | 320 | 290 | 310 | 301 | 118 | 125 | 149 | 89 | 861 |
| 2012 | 0 | 2 | 3 | 5 | 18 | 28 | 176 | 161 | 177 | 174 | 67 | 68 | 84 | 50 | 466 |
| 2013 | 12 | 20 | 9 | 1 | 12 | 22 | 197 | 197 | 244 | 277 | 105 | 132 | 148 | 90 | 899 |
| 2014 | 2 | 33 | 49 | 11 | 45 | 89 | 992 | 1044 | 1403 | 1685 | 624 | 783 | 898 | 543 | 6669 |
| 2015 | 0 | 1 | 1 | 1 | 7 | 14 | 112 | 109 | 126 | 137 | 54 | 68 | 75 | 46 | 564 |
| 2016 | 1 | 5 | 10 | 5 | 4 | 6 | 61 | 62 | 78 | 91 | 35 | 48 | 51 | 32 | 360 |
| 2017 | 5 | 5 | 0 | 7 | 10 | 12 | 80 | 80 | 100 | 132 | 54 | 96 | 90 | 59 | 786 |
| 2018 | 0 | 0 | 0 | 1 | 19 | 41 | 501 | 534 | 718 | 906 | 349 | 516 | 536 | 337 | 4050 |
| 2019 | 0 | 1 | 3 | 3 | 8 | 15 | 167 | 172 | 215 | 260 | 104 | 157 | 158 | 101 | 1040 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2020 | 0 | 2 | 2 | 3 | 7 | 11 | 113 | 115 | 136 | 177 | 77 | 146 | 129 | 87 | 1519 |
| 2021 | 0 | 15 | 32 | 32 | 20 | 14 | 104 | 109 | 154 | 219 | 86 | 149 | 144 | 94 | 1290 |
| 2022 | 2 | 13 | 29 | 82 | 50 | 28 | 76 | 60 | 74 | 100 | 41 | 75 | 69 | 46 | 655 |

## WCSGFS

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1986 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1987 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1988 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1989 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 12 | 61 | 90 | 197 | 233 | 248 | 736 | 509 | 363 | 224 | 85 | 38 | 74 | 41 | 261 |
| 1991 | 69 | 184 | 275 | 631 | 405 | 256 | 482 | 257 | 153 | 72 | 25 | 8 | 19 | 12 | 63 |
| 1992 | 6 | 30 | 133 | 733 | 849 | 840 | 2097 | 1321 | 823 | 409 | 155 | 41 | 112 | 63 | 301 |
| 1993 | 54 | 279 | 846 | 1723 | 1227 | 981 | 2777 | 1908 | 1446 | 1017 | 359 | 177 | 351 | 191 | 1165 |
| 1994 | 8 | 38 | 71 | 222 | 157 | 112 | 292 | 202 | 179 | 143 | 54 | 43 | 60 | 35 | 250 |
| 1995 | 20 | 71 | 109 | 328 | 387 | 385 | 1141 | 811 | 665 | 480 | 184 | 116 | 183 | 102 | 718 |
| 1996 | 24 | 59 | 51 | 53 | 58 | 67 | 398 | 375 | 458 | 490 | 174 | 160 | 222 | 126 | 953 |
| 1997 | 8 | 76 | 107 | 81 | 76 | 71 | 233 | 174 | 154 | 119 | 46 | 31 | 47 | 26 | 197 |
| 1998 | 4 | 10 | 10 | 26 | 25 | 22 | 68 | 52 | 52 | 50 | 19 | 20 | 24 | 15 | 121 |
| 1999 | 3 | 71 | 173 | 244 | 182 | 134 | 315 | 199 | 150 | 100 | 38 | 24 | 37 | 21 | 141 |
| 2000 | 2 | 18 | 53 | 151 | 122 | 93 | 205 | 125 | 90 | 56 | 22 | 14 | 21 | 12 | 92 |


| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2001 | 0 | 5 | 14 | 35 | 33 | 30 | 122 | 103 | 112 | 118 | 45 | 55 | 62 | 38 | 397 |
| 2002 | 4 | 6 | 23 | 347 | 634 | 778 | 3010 | 2402 | 2269 | 1942 | 725 | 559 | 813 | 459 | 3480 |
| 2003 | 2 | 39 | 46 | 196 | 311 | 380 | 1730 | 1482 | 1545 | 1585 | 619 | 774 | 853 | 528 | 4647 |
| 2004 | 3 | 19 | 52 | 367 | 802 | 1054 | 4442 | 3641 | 3470 | 3148 | 1237 | 1315 | 1553 | 939 | 8289 |
| 2005 | 19 | 39 | 32 | 63 | 97 | 118 | 547 | 472 | 504 | 506 | 191 | 207 | 250 | 149 | 1307 |
| 2006 | 4 | 15 | 67 | 266 | 208 | 177 | 781 | 680 | 760 | 834 | 326 | 442 | 470 | 294 | 2900 |
| 2007 | 7 | 90 | 141 | 415 | 626 | 727 | 2893 | 2356 | 2285 | 2205 | 881 | 1104 | 1195 | 746 | 7600 |
| 2008 | 18 | 110 | 248 | 798 | 948 | 1026 | 5180 | 4696 | 5396 | 6246 | 2479 | 3677 | 3739 | 2381 | 26466 |
| 2009 | 2 | 27 | 524 | 2249 | 1182 | 537 | 771 | 336 | 263 | 187 | 68 | 70 | 81 | 51 | 531 |
| 2010 | 0 | 0 | 4 | 191 | 315 | 347 | 1030 | 738 | 612 | 492 | 192 | 191 | 231 | 140 | 1236 |

Table 3.4.2.1 Boarfish in ICES Subareas 27.6, 7, 8. Key parameter estimates from the exploratory Schaeffer state space surplus production model. Posterior parameter distributions are provided in Figure 3.6.3.5

| Parameter | Mean | SD | 2.5 | 25 | 50 | 75 | 97.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $r$ | 0.32 | 0.16 | 0.05 | 0.21 | 0.31 | 0.43 | 0.67 |
| K | 691791 | 434944 | 321000 | 457800 | 573100 | 760400 | 1890000 |
| Fmsy | 0.16 | 0.08 | 0.03 | 0.10 | 0.16 | 0.21 | 0.34 |
| Bmsy | 345896 | 217472 | 160500 | 228900 | 286550 | 380200 | 945000 |
| TSB | 734324 | 352743 | 393000 | 540200 | 656900 | 823300 | 1517000 |

Table 3.4.2.2 2Boarfish in ICES Subareas 27.6, 7, 8. Estimates of total stock biomass and $F$

| Year | TSB.2.5 | TSB. 50 | TSB.97.5 | F2.5 | F. 50 | F.97.5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1991 | 95650 | 182100 | 445100 |  |  |  |
| 1992 | 155000 | 280000 | 663500 |  |  |  |
| 1993 | 189400 | 341900 | 811997 |  |  |  |
| 1994 | 222800 | 404800 | 967800 |  |  |  |
| 1995 | 194100 | 351100 | 836697 |  |  |  |
| 1996 | 194600 | 350400 | 840500 |  |  |  |
| 1997 | 166700 | 295600 | 700997 |  |  |  |
| 1998 | 219500 | 386500 | 911995 |  |  |  |
| 1999 | 163600 | 291400 | 690297 |  |  |  |
| 2000 | 143600 | 254700 | 603597 |  |  |  |
| 2001 | 160300 | 278900 | 655097 |  |  |  |
| 2002 | 138100 | 239600 | 561897 |  |  |  |
| 2003 | 127400 | 220700 | 514000 | 0.02 | 0.05 | 0.09 |
| 2004 | 174200 | 300300 | 702892 | 0.01 | 0.02 | 0.03 |
| 2005 | 168800 | 292100 | 685900 | 0.01 | 0.02 | 0.04 |
| 2006 | 214400 | 365300 | 850797 | 0.01 | 0.02 | 0.03 |
| 2007 | 193400 | 333300 | 775900 | 0.03 | 0.06 | 0.11 |
| 2008 | 237000 | 402600 | 931097 | 0.04 | 0.09 | 0.15 |
| 2009 | 241800 | 407700 | 933800 | 0.10 | 0.22 | 0.37 |
| 2010 | 357702 | 600400 | 1389975 | 0.10 | 0.24 | 0.40 |
| 2011 | 316600 | 536100 | 1256000 | 0.03 | 0.07 | 0.12 |
| 2012 | 453300 | 745000 | 1704000 | 0.05 | 0.12 | 0.19 |
| 2013 | 304000 | 509400 | 1188000 | 0.06 | 0.15 | 0.25 |
| 2014 | 144600 | 241300 | 559700 | 0.08 | 0.19 | 0.31 |
| 2015 | 171902 | 288600 | 668997 | 0.03 | 0.06 | 0.10 |
| 2016 | 127400 | 215600 | 502200 | 0.04 | 0.09 | 0.15 |
| 2017 | 223000 | 376400 | 875397 | 0.02 | 0.05 | 0.08 |
| 2018 | 239500 | 404200 | 937197 | 0.01 | 0.03 | 0.05 |
| 2019 | 203400 | 342200 | 795597 | 0.01 | 0.03 | 0.06 |


| Year | TSB.2.5 | TSB.50 | TSB.97.5 | F2.5 | F.50 | F.97.5 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2020 | 238300 | 408200 | 950695 | 0.02 | 0.04 | 0.07 |
| 2021 | 320500 | 543800 | 1259000 | 0.01 | 0.03 | 0.06 |
| 2022 | 367402 | 617500 | 1419000 | 0.01 | 0.03 | 0.05 |
| 2023 | 393000 | 656900 | 1517000 |  |  |  |

### 3.16 Figures

Mean Catch Rate (Kg/30min) Boarfish 2003-2022


Figure 3.1.1.1. Boarfish in ICES Subareas 4, 27.6, 7, 8 and 9. Distribution and mean catch rates of boarfish in the NE Atlantic area from IBTS surveys (all years).


Figure 3.1.1.2. Boarfish in ICES Subareas 27.6, 7, 8. Boarfish samples included in the genetic stock identification study are indicated in green. Population clusters identified by the STRUCTURE analyses are indicated by colour coded circles.


Figure 3.3.1.1. Boarfish in ICES Subareas 27.6, 7, 8. Catch numbers-at-age standardised by yearly mean. 15+ is the plus group.

Boarfish: SSB, TSB Acoustic Survey Estimates 2011-2023


Figure 3.3.2.1. Boarfish in ICES Subareas 27.6, 7, 8. Boarfish acoustic survey track and haul positions 2023 (left), estimates of biomass at length by stratum (right).


Figure 3.3.2.2. Boarfish in ICES Subareas 27.6, 7, 8. Boarfish acoustic survey track and haul positions 2023


Figure 3.3.2.3. Boarfish in ICES Subareas 27.6, 7, 8. Boarfish acoustic survey biomass estimate by stratum, 2023.


Figure 3.3.2.4. Boarfish in ICES Subareas 27.6, 7, 8. Boarfish acoustic survey time series of acoustic estimates of abundance at age, 2011-2023.


Figure 3.3.3.1. Boarfish in ICES Subareas 27.6, 7, 8. Spatial coverage and haul positions of bottom trawl surveys analysed as an index for boarfish abundance.


Figure 3.3.3.2. Boarfish in ICES Subareas 27.6, 7, 8. Distribution of boarfish in the NE Atlantic from the 6 IBTS surveys.


Figure 3.3.3.3 Boarfish in ICES Subareas 27.6, 7, 8. Boarfish IBTS survey CPUE fitted delta-lognormal mean (solid line) and 95\% credible intervals (grey region).


Figure 3.3.3.4a. Boarfish in ICES Subareas 27.6, 7, 8. CPUE in number per 30-minute haul of boarfish per rectangle in the western IBTS survey 1982 to 2022.


Figure 3.3.3.4b. Boarfish in ICES Subareas 27.6, 7, 8 . CPUE in kg per 30-minute haul of boarfish per rectangle in the western IBTS survey 1982 to 2022.


Figure 3.3.3.5. Boarfish in ICES Subareas 27.6, 7, 8. Abundance-at-age in EVHOE, IGFS and SPNGFS surveys. Yearly mean standardised abundance-at-age.


Figure 3.4.2.1. Boarfish in ICES Subareas 27.6, 7, 8. MCMC chain autocorrelation for final run.


Figure 3.4.2.2. Boarfish in ICES Subareas 27.6, 7, 8. Rhat values lower than 1.01 indicating convergence.


Figure 3.4.2.3. Boarfish in ICES Subareas 27.6, 7, 8. Parameters for final run converged with good mixing of the chains.


Figure 3.4.2.4. Boarfish in ICES Subareas 27.6, 7, 8. Residuals around the model fit for the final assessment run.


Figure 3.4.2.5. Boarfish in ICES Subareas 27.6, 7, 8. Prior (red) and posterior (black) distributions of the parameters of the biomass dynamic model.


Figure 3.4.2.6. Boarfish in ICES Subareas 27.6, 7, 8. Trajectories of observed and expected indices for the final assessment run. The stock size over time and a harvest ratio (total catch divided by estimated biomass) are also shown.


Figure 3.4.2.7. Boarfish in ICES Subareas 27.6, 7, 8. Retrospective plot of total stock biomass (above) and fishing mortality (below) from the surplus production model in 2013-2022.


Figure 3.4.2.8. Boarfish in ICES Subareas 27.6, 7, 8. Diagnostics from the positive component of the delta-lognormal fits


Figure 3.4.2.9. Boarfish in ICES Subareas 27.6, 7, 8. Pair-wise correlation between the annual mean survey indices.


Figure 3.4.2.10. Boarfish in ICES Subareas 27.6, 7, 8. Weighted correlation between the annual mean survey indices. Correlations are weighted by the sum of the pair-wise variances.


Figure 3.4.3.1. Boarfish in ICES Subareas 27.6, 7, 8. Ratios 'B / MSYBtrigger' and 'F / FMSY' through time and corresponding Kobe plot. Confidence intervals ( $\mathbf{5 0}$ and $95 \%$ ) are given for the first two panels, the third displays median estimates only with the pink point representing the first point of the time series and the purple point the last.


Figure 3.7.1.1. Boarfish in ICES Subareas 27.6, 7, 8. Sensitivity of estimation of F0.1.


Figure 3.7.1.2. Boarfish in ICES Subareas 27.6, 7, 8. Results of exploratory yield per recruit analysis. Beverton and Holt model applied to various fits of the VBGF and for comparison with the VBGF parameters provided by White et al. 2011.

# 4 Norwegian spring-spawning herring in Northeast Atlantic and Arctic Ocean 

## Clupea harengus in subareas 1, 2, 5 and divisions 4.a and 14.a (her.27.124a514a)

### 4.1 ICES advice in 2022

ICES advised that when the long-term management strategy agreed by the European Union, the Faroe Islands, Iceland, Norway, and the Russian Federation is applied, catches in 2023 should be no more than 511171 tonnes. The advice for 2023 was $15 \%$ lower than that for 2022.

### 4.2 The fishery in 2022

### 4.2.1 Description and development of the fisheries

The distribution of the Norwegian spring-spawning herring (NSSH) fishery in 2022 for all countries by ICES rectangles is shown in Figure 4.2.1.1. The catches by ICES statistical rectangle and quarter are seen in Figure 4.2.1.2. The 2022 herring fishing pattern was similar to recent years. The fishery began in January on the Norwegian shelf and focused on overwintering, pre-spawning, spawning and post-spawning fish (Figure 4.2.1.2, quarter 1). In the second quarter, the fishery was insignificant (Figure 4.2.1.2, quarter 2). In summer, the fishery moved into mainly Icelandic, Faroese and International waters and in early autumn started in the overwintering area off Lofoten (Figure 4.2.1.2, quarter 3). In autumn and winter, the fishery continued in Icelandic and Faroese waters but also in the overwintering area in the fjords and oceanic areas off Lofoten (Figure 4.2.1.2, quarter 4). $66 \%$ of the catches were taken in the fourth quarter. Catches of Norwegian spring-spawning herring inside the NEAFC regulatory area was estimated by the working group to be 65015 tonnes in 2022, which represents $10 \%$ of the total catch. Note though that this does not include catches from the Russian Federation.

### 4.3 Stock description and management units

### 4.3.1 Stock description

A description of the stock is given in the Stock Annex.

### 4.3.2 Changes in migration

Generally, it is not clear what drives the variability in migration of the stock, but the biomass and production of zooplankton are likely factors, as well as feeding competition with other pelagic fish species (e.g. mackerel and to a lesser extent blue whiting) and oceanographic conditions (e.g. limitations due to cold areas). Besides environmental factors, the age distribution in the stock will also influence the migration. Changes in the migration pattern of NSSH, as well as that of other herring stocks, are often linked to large year classes entering the stock initiating a different migration pattern, which subsequent year classes will follow. This pattern does not appear to be prevalent for NSS herring during the feeding season, where changes in distribution occur for all
age-classes simultaneously (Eliasen et al., 2021). The large 2016 year-class has now entered the adult stock. The distribution of the 2016 year-class in the feeding area in 2023 as observed in the ecosystem survey in May appeared to be primarily distributed in the western part of the survey area. In 2017/2018 there was a shift in wintering areas. While wintering has been observed in fjords west of Tromsø (Norway) for several years, the 2013 year-class wintered in fjords farther north (Kvænangen) since 2017/2018 while the older fish seemed to have had an oceanic wintering area. A different pattern was observed during the winter 2022/2023. The wintering of old fish in the Norwegian Sea was not observed in 2023 catches The oldest and largest fish move farthest south and west during feeding, and the older year classes were in May-July 2023 concentrated in the south-western areas during the feeding season.

### 4.4 Input data

### 4.4.1 Catch data

Catches in tonnes by ICES division, ICES rectangle and quarter in 2022 were available from Denmark, Faroe Islands, Germany, Greenland, Iceland, Ireland, The Netherlands, Norway, Sweden, United Kingdom and Russia. The total working group catch in 2022 was 813834 tonnes (Table 4.4.1.1) compared to the ICES-recommended catch of a maximum of 598588 tonnes. The majority of the catches (around $83 \%$ ) were taken in division 2.a as in previous years. Age samples were not provided by Greenland, Germany, The Netherlands or Sweden. Sampled catches accounted for $97 \%$ of the total catches, which is higher compared to last year. The sampling levels of catches in 2022 in total, by country and by ICES division are shown in Tables 4.4.1.2, 4.4.1.3 and 4.4.1.4. Catch by nation, ICES division and quarter are shown in Table 4.4.1.5. The software SALLOC (ICES, 1998) was used to calculate total catches in numbers-at-age and mean weight at age representing the total catch. Samples allocated (termed fill-in in SALLOC) to cells (nation, ICES division and quarter) without sampling information are shown in Table 4.4.1.5. This year Russian age sample information and catch by quarter and ICES division and ICES rectangle for 2022 were officially delivered to IMR Norway, and ICES decided to use this information. This information from Russia were not available for 2021 and this and the implications are described in the WGWIDE report from 2022.

### 4.4.2 Discards

In 2008, the Working Group noted that in this fishery an unaccounted mortality caused by fishing operations and underreporting probably exists (ICES, 2008). It has not been possible to assess the magnitude of these extra removals from the stock, and considering the large catches taken after the recovery of the stock, the relative importance of such additional mortality is probably low. Therefore, no extra mortality to account for these factors has been added since 1994. In previous years, when the stock and the quotas were much smaller, an estimated number of fish was added to the catches.

The Working Group has not had access to comprehensive data to estimate discards of herring. Although discarding may occur on this stock, it is considered to be low and a minor problem for the assessment. This is confirmed by estimates from sampling programmes carried out by some EU countries in the Data Collection Framework. Estimates of discarding in 2008 and 2009 of about $2 \%$ in weight were provided for the trawl fishery carried out by the Netherlands. In 2010 and 2012, this métier was sampled by Germany. No discarding of herring was observed (0\%) in either of the two years. An investigation on fisheries induced mortality carried out by IMR with EU partners on fisheries induced and unreported mortality in mackerel and herring fisheries in the North Sea concluded with an estimated level of discarding at around $3 \%$.

To provide information on unaccounted mortality caused by fishing operations in the Norwegian fishery, Ipsos Public Affairs, in cooperation with IMR and the fishing industry, conducted a survey in January/February 2016. The survey was done by phoning skippers and interviewing them. A total of 146 herring skippers participated in the survey, 31 skippers representing the bigger vessel group and 115 skippers representing the smaller vessel group. The data provided an indication that there have been periods of increased occurrence of net bursting. This was seen especially in the period 2007-2010. There was, however, no trend in the size of catches where bursting has occurred.

When it comes to slipping, the data showed a steady increase in the percentage that has slipped herring from 2004-2012, and then a significant decline in recent years. The variations in the proportion that have slipped herring were largely driven by the skippers on smaller coastal purseseiners. Average size of purse-seine hauls slipped seems to be relatively steady over the period. However, the average size of net hauls slipped was lowest in the recent period.

### 4.4.3 Age composition of the catch

The estimated catch-at-age in numbers by year are shown in Table 4.4.3.1. The numbers are calculated using the SALLOC software. In 2022, catches (in numbers) were dominated by the 2016 year-class which comprised around $53 \%$ of the catch followed by the 2013 year class which comprises around $10 \%$ of the catch. Catch curves were made on the basis of the international catch-at-age (Figure 4.4.3.1). For comparison, lines corresponding to $\mathrm{Z}=0.3$ are drawn in the background. The big year classes, in the periods of relatively constant effort, show a consistent decline in catch number by cohort, indicating a reasonably good quality of the catch-at-age data. Catch curves for year classes 2005 onwards show a flatter curve than for previous year classes indicating a lower F or a changed exploitation pattern.

### 4.4.4 Weight-at-age in catch and in the stock

The weight-at-age in the catches in 2022 was computed from the sampled catches using SALLOC. Trends in weight-at-age in the catch are presented in Figure 4.4.4.1 and Table 4.4.4.1. The mean weights at age for most of the age groups generally increased in 2010-2013 but levelled off around 2014. Since around 2015 the weight-at-age have decreased for most ages - earlier for the younger ages than for the older ages. From 2020 to 2021 all ages decreased in weight and the decrease was generally larger than the preceding years and from 2021 to 2022 weight increased for all ages. A similar pattern is observed in weight-at-age in the stock, however, the weight trend in 2021 and 2022 is more in line with the cohort trends (Figure 4.4.4.2 and Table 4.4.4.2). The mean weight-at-age in the stock was based on the survey in the wintering area until 2008. Since then, the mean weight-at-age in the stock was derived from samples taken in the fishery in the same area and at the same time as the wintering surveys were conducted in.

### 4.4.5 Maturity-at-age

In 2010 the method for estimating maturity-at-age in the stock assessment of NSSH was changed based on work done by the "workshop on estimation of maturity ogive in Norwegian springspawning herring" (WKHERMAT; ICES, 2010a). The method which was adopted by WGWIDE in 2010 (ICES, 2010b) is based on work by Engelhard et al. (2003) and Engelhard and Heino (2004). They developed a method to back-calculate age-at-maturity for individual herring based on scale measurements, and used this to construct maturity ogives for the year classes 1930-1992.
The NSSH has irregular recruitment pattern with a few large year classes dominating in the stock when it is on a high level. Most of the year classes are, however, relatively small and referred to
as "normal" year classes. The back-calculation dataset indicates that maturation of the large year classes is slower than for "normal" year classes.

WKHERMAT and WGWIDE considered the dataset derived by back calculation as a suitable candidate for use in the assessment because it is conceived in a consistent way over the whole period and can meet standards required in a quality-controlled process. However, the back-calculation estimates cannot be used for the most recent years since all year classes have to be fully matured before the calculation can be made. Therefore, assumptions have to be made for the recent year classes. For recent year classes, WGWIDE (ICES, 2010b) decided to use average backcalculated maturity for "normal" and "big"year classes thereby reducing maturity-at-age for ages 4,5 and 6 when strong year classes enter the spawning stock. The default maturity ogives used for "normal" and "big" year-classes are given in the text table below.

| age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| normal <br> year <br> class 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |  |  |
| strong <br> year <br> class | 0 | 0 | 0 | 0 | 0.1 | 0.6 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |

Assumed values should be replaced by back-calculated values in the annual assessments for each year where updated values are available. In 2023 the year 2018 was updated with back-calculated values used in the present assessment. Assumed (old) and updated values for 2018 are shown in figure 4.4.5.1. The 2016 year-class was considered a strong year-class by the working group based on the assessment where several survey indices of this year-class are included. The resulting maturity ogives used in the present assessment are presented in Table 4.4.5.1.

### 4.4.6 Natural mortality

In this year's assessment, the natural mortality $M=0.15$ was used for ages 3 and older and $M=0.9$ was used for ages $0-2$. These levels of natural mortality are in accordance to previous years and their justification is provided in the stock annex. Information about deviations from these levels in the time-series, e.g. due to diseases, are also provided in the stock annex.

### 4.4.7 Survey data

The surveys available for the assessment are described in the stock annex. Only two of the available surveys are used in the final assessment and will therefore be dealt with in this section:

1. The International Ecosystem Survey in the Nordic Seas (IESNS) in May. This survey covers the entire stock during its migration on the feeding grounds, the adults in the Norwegian Sea and adjacent waters ("Fleet 5") and the juveniles in the Barents Sea ("Fleet 4")
2. The Norwegian acoustic survey on the spawning grounds in February ("Fleet 1").

The cruise reports from the IESNS (WD03) and spawning survey (WD04) in 2023 are available as working documents to this report. The spawning survey and IESNS in the Norwegian Sea were both carried out successfully in 2023, however like in 2022, the IESNS in the Barents Sea was not carried out by Russia this year.

The abundance estimates from "Fleet 1" are shown in Table 4.4.7.1 and Figure 4.4.7.2; from "Fleet 4" in Table 4.4.7.2 and Figure 4.4.7.1 and "Fleet 5" in Table 4.4.7.3 and Figure 4.4.7.1. In 2020 it
was decided to use the bootstrap mean values as point estimates of abundance instead of the baseline estimates. This applies to the years were the software Stox is used to estimate abundance. Variance estimates from the bootstrap runs were already being used in the assessment, thus it is more logical to also use point estimates from the bootstrap. A comparison using point estimates for both bootstrap and baseline was made, and the effect on the assessment was negligible.

Catch curves were made on the basis of the abundance estimates from the surveys "Fleet 1" (Figure 4.4.7.3) and "Fleet 5" (Figure 4.4.7.4). The same arguments are valid for the interpretation of the catch curves from the surveys as from the catches. In 2010, the numbers of all age groups decreased suddenly in "Fleet 5" and this is seen as a drop in the catch curves that year. This drop has continued for some of the year classes and the year classes 1998 and 1999 are disappearing faster from the stock than expected. This observed fast reduction in these age classes may also be influenced by the changes in "Fleet 5" catchability, with seemingly higher catchability in years 2006 - 2009. Like the catch curves from commercial landings, the corresponding curves from "Fleet 5" are also quite flat for year classes 2005 onwards. As "Fleet 1" was not conducted in the years 2009-2014, there is a gap in the catch curves, making it difficult to interpret them.

### 4.4.8 Sampling error in catches and surveys

Sampling errors for Norwegian catch-at-age for the years 2010-2022 is estimated using ECA (Salthaug and Aanes 2015, Hirst et al. 2012). Using the Taylor function (Aanes 2016a) to model the sampling variance of the catches yields a very good fit $\left(R_{a d j}^{2}=0.94\right)$ and using this function to impute missing sampling variances for catch-at-age yields relative standard errors shown in Table 4.4.8.1. It is assumed that the relative standard errors in the total catches are equal to the Norwegian catches (which comprise $\sim 60 \%$ of the total catches). Sampling errors for survey indices are estimated using StoX (http://www.imr.no/forskning/prosjekter/stox/nb-no) and Johnsen et al. (2019). For Fleet 1, estimates are available for the years 1988-1989, 1994-1996, 1998-2000, 2005-2008, and 2015-2023, for Fleet 4 estimates of sampling errors are available for 2009-2019 and 2021, and for Fleet 5 for 2008-2023. Missing values for sampling variances are imputed using the Taylor function which provides good fits ( $R_{a d j}^{2}$ 's are $0.95,0.98$ and 0.96 respectively). The resultant relative standard errors are given in Tables 4.4.8.2-4.4.8.4. Due to the very good fits of the Taylor functions, estimates of relative standard where empirical estimates are available, are also replaced by the model predicted values to reduce potential effects of imprecise estimates of errors.

### 4.4.9 Information from the fishing industry

No information was made available to the working group.

### 4.5 Stock assessment

The first benchmark of the NSSH assessment took place in 2008 with the assessment tool TASACS selected as the standard assessment tool for the stock. A second benchmark took place in 2016 (WKPELA - ICES, 2016a) where three assessment models were explored - TASACS, XSAM and one separable model. WKPELA accepted XSAM as the standard assessment tool for the NSSH. During the WGWIDE meeting in 2023, it was agreed to recommend using the SAM software with the same model and configuration as the XSAM software for WGWIDE 2024 and onwards. The reason was that the XSAM software is no longer maintained and will fail to compute if upgrading the R software or R packages.

### 4.5.1 XSAM final assessment 2023

The XSAM model is documented in Aanes 2016a and 2016b. XSAM includes the option to utilize the prediction of total catch in the assessment year (typically the sum of national quotas) along with the precision of the prediction. This approach was changed in 2017 when it was found that the model estimated a highly variable and significantly lower catch compared to the working group's prediction (sum of national quotas). In addition, this caused an abrupt change in the selection pattern from 2017 and onwards. The abrupt change in the selection pattern was not fully understood by the working group, but the effect was less pronounced if not using the catch prediction from the model for 2017. Therefore, it was decided to not utilize the prediction of total catches in 2017 when fitting the model to data (i.e. the assessment) and consequently in the shortterm forecast. The same approach is taken in the 2023 assessment, i.e. the catch prediction for 2023 is not included when fitting the model to data. The resulting estimated selection pattern is gradual (Figure 4.5.1.1) and in line with the current knowledge about the fishery. It is important to note that this has marginal effect on the assessment, but larger effects on the prediction and short-term forecast.

The 2023 XSAM assessment was performed with the same model options as in 2017. In summary, this means that the model was fit with time varying selectivity and effort according to $\operatorname{AR}(1)$ models in the model for fishing mortality; the recruitment was modelled as a process with constant mean and variance; the standard errors for all input data were predetermined using sample data (Tables 4.4.8.1-4.4.8.4), and a scaling constant common for all input data to allow additional variability in the input data that is not controlled by sampling is estimated. Additional details on the assessment settings are given in the Stock Annex.

The same input data over the same age ranges was used as in 2017. At the 2016 benchmark, data from 1988 and onwards was used from ages 3-12+ with input data catch-at-age, Fleet 1 and Fleet 5, At WGWIDE 2016, it was decided to start the model at age 2 to enable short-term predictions with reasonable levels of variability. To achieve this, age 2 from Fleet 4, and age 2 in catch-at-age was included in input data. Evaluation of diagnostics including lower ages than 2 and/or other fleets resulted in excluding lower ages than 2 and other fleets for the final assessment. It should be noted, that the recruitment index has not been updated since 2021.

The parameter estimates from the 2023 assessment are shown in Table 4.5.1.1 and in Figure 4.5.1.10. For a precise definition of the parameters, refer to Aanes 2016a in ICES (2016a). Note that the variance components $\sigma_{1}^{2}$ (variability in the separable model for $F$ ) and $\sigma_{R}^{2}$ (variability in recruitment) are rather imprecise. The estimate of the scaling constant $h$ is larger than 1, indicating that the model adds additional variability on the observation errors than explained by the sampling errors alone.

The catchabilities for all the fleets are on average positively correlated indicating some uncertainty due to a common scaling of all surveys to the total abundances although the correlations in general are small (Figure 4.5.1.2). There is a slight negative correlation between $\sigma_{1}^{2}$ (variability in the separable model for $\mathrm{F}, \operatorname{logs} 2 \_1$ in figure) and $\sigma_{2}^{2}$ (variability in the AR process for time varying selectivity, logs2_2 in figure) indicating little contrast in data for separating variability in the separable model from variability due to changes in selection pattern. The slopes in the multivariate AR model for time-varying selectivity gradually changes from negative to positive, but is expected as it is imposed due to the sum to zero constraint for the selection (see Aanes 2016a for details).

The weights each datum is given in the model fit (inverse of the sampling variance) is proportional to the empirical weights derived from sampling variances (Tables 4.4.8.1-4.4.8.4) which shows that the strong year classes in general are given larger weight to the model than weaker
year classes, and the ordering of the average weights (from high to low) is Catch-at-age, Fleet 5, Fleet 1 and Fleet 4 (Figure 4.5.1.3).

Two types of residuals are considered for this model. The first type is the model prediction (based on all data) vs. the data. In such time-series models, the residuals based on the prediction which uses all data points will be serially correlated although useful as they explain the unexplained part of the model (cf Harvey 1990 p 258). This means that patterns in residuals over time is to be expected and questions the use of $e . g$. qq-plots as an additional diagnostic tool to assess distributional assumptions. To obtain residuals which follow the assumptions about the data in the observation models (e.g. serially uncorrelated) single joint sample residuals are extracted (ICES, 2017). In short these are obtained by sampling predicted values from the conditional distribution of values given the observations. This sample corresponds to a sample from the joint distribution of latent variables and observations. A third approach could have been to extract the one step ahead observation residuals which are standard for diagnostics for regular state-space models (cf Harvey 1990). This is not done here.

The negative residuals tracing the 1983 year-class for catch-at-age represents low fishing mortalities examining the type 1 residuals (Figure 4.5.1.4). This effect is less pronounced considering the type 2 residuals. The type 2 residuals are qualitatively comparable with the type 1 residuals but generally display more mixed residuals as predicted by the theory. Otherwise the residuals for catch-at-age appears fairly mixed apart for some serial correlation for age 2 and 3 (which are very low), and some negative residuals for the plus group the most recent years. The residuals for Fleet 1 in year 1994, 1999, 2006 for young and old ages are all of the same signs and may appear as year effects. Also note that the residuals for Fleet 1 for ages 12+ from 2015 are all positive (Figure 4.5.1.4) which shows that the abundance indices from Fleet 1 displays a larger stock size over these ages and years compared to the assessment using all input data. Some serial correlation for residuals for ages 3 and 4 in Fleet 1 can also be detected, but is down weighted as these is found to be uncertain. Serial correlation in residuals for age 2 in Fleet 4 can also be detected indicating trends over time in mismatch between estimates and observations of abundance at age 2. Residuals for Fleet 5 appears adequate compared to previous years although some serial correlations can be detected also here.

The residuals for small values are bigger than residuals for the larger values since smaller values in general have higher variances than larger values (Tables 4.4.8.1-4.4.8.4) (Figure 4.5.1.5). The qq-plots for the standardized residuals show that the distributional assumptions on the observation errors are adequate, except for the smallest and largest values of catch-at-age and indices from Fleet 1 and the smallest indices for Fleet 4 . As qq-plots for residuals of type 1 may be questioned (see above) it is noted that qq-plots for residuals of type 2 is more relevant and generally shows a significantly better fit based on a visual inspection compared to using type 1.

The marginal likelihood and the components for each data source (see Aanes 2016b for details) are profiled over a range of the common scaling factor $h$ for all input data (Figure 4.5.1.6). It is apparent that the optimum of the marginal likelihood is clearly defined. The catch component is decreasing with decreasing values of $h$ indicating that the model puts more weight on the catch component than indicated by the comparison of sampling errors for all input data. This is in line with the findings in Aanes (2016a and 2016b) who showed that these types of models tend to put too much weight on the catch data if the weighting is not constrained. However, the likelihood component for the catch is overruled by the information in Fleets 1,4 and 5 such that the optimum for the marginal likelihood is clearly defined. The point estimates of SSB and F is insensitive to different values of $h$.

The retrospective runs for this model shows estimates within the estimated levels of precision (Figure 4.5.1.7), and has a reasonably low Mohn's rho value for SSB of -0.06 and -0.017 for F
(Mohn, 1999; Brooks and Legault, 2016). Note that the retrospective patterns are remarkably stable.

Figure 4.5.1.8 illustrates the conflict in data and increased uncertainty in estimates for the most recent years. The spawning-stock biomass shown for each survey index is calculated using the stock weights at age and proportion mature at age, with the abundance indices are scaled to the absolute abundance by the estimated catchabilities. A fairly good temporal match between the model estimate of SSB and the survey SSBs is seen, except for the years 2015 for Fleet 1, which displays a significantly faster reduction in the stock compared to Fleet 5 which shows a flatter trend in the same years. Both Fleet 1 and Fleet 5 indicate an increase in SSB from 2007 to 2009. It is worth noting that, although the point estimate of SSB based on Fleet 1 appears very much higher than Fleet 5 in 2015, the uncertainty in the estimates are very high, such that the respective estimates do not appear as significantly different. Since 2016 the conflict between Fleet 1 and Fleet 5 has become less.

The results of leave-one-out runs are presented in Figure 4.5.1.11 and can be used to assess the influence of individual data sources on the assessment. Removing Fleet 1 leads to a downward revision of SSB and an upward revision of F. The overall assessment uncertainty is similar to the base run which includes all data sources. Removing Fleet 5 results in an upward revision of SSB and a downward revision of F, with an increase in uncertainty. Removing Fleet 4 does not influence the SSB nor F.

The final 2023 assessment results are shown in Figure 4.5.1.9. The estimate of fishing mortality for 2019 to 2022 is rather high, as a response to the high catch in the last years with a point estimates from $\sim 0.17$ to $\sim 0.19$. In 2018 the fishing mortality is estimated to be lower than in 2017 and 2019. The spawning stock shows a declining trend since 2009 but an increase in 2021 and 2022 before continuing to decrease. The $95 \%$ confidence interval of the stock level in 2023 ranges from $\sim 2.974$ to $\sim 4.3$ million tonnes with a point estimate of 3.664 which is above $\mathrm{B}_{\mathrm{mp}}=3.184$ million tonnes, such that the probability of the stock being above $\mathrm{B}_{\mathrm{lim}}=2.5$ million tonnes is high.

The final results of the assessment are also presented in Tables 4.5.1.2 (stock in numbers), 4.5.1.3 (fishing mortality) and Table 4.5.1.4 is the summary table of the assessment.

### 4.6 NSSH reference points

ICES last reviewed the reference points of Norwegian spring spawning herring in April 2018 during WKNSSHREF (ICES, 2018a). ICES concluded that Blim should remain unchanged at 2.5 million tonnes and MSY $B_{\text {trigger }}=B_{\text {pa }}$ was estimated at 3.184 million tonnes. FMSY was estimated at the reference point workshop, but during the subsequent Management Strategy Evaluation WKNSSHMSE (ICES, 2018b) the fishing mortality reference points were revisited as issues were found with numerical instability and settings during the reference point workshop. FMSY was reestimated to be 0.157 .

### 4.6.1 PA reference points

The PA reference points for the stock were last estimated by WKNSSHREF and WKNSSHMSE in 2018. The WKNSSHREF group concluded that Blim should be kept at 2.5 million tonnes and $B_{p a}$ was estimated at 3.184 million tonnes. WKNSSHMSE estimated $F_{p a}=0.227$. However, following recent ICES guidelines $\mathrm{F}_{\mathrm{pa}}$ is now based on Fp 05 which was estimated at 0.157 by WKNSSHMSE in 2018.

### 4.6.2 MSY reference points

The MSY reference points were evaluated by WKNSSHREF and WKNSSHMSE in 2018. In the ICES MSY framework $\mathrm{B}_{\mathrm{pa}}$ is proposed/adopted as the default trigger biomass $\mathrm{B}_{\text {trigger }}$ and was estimated by WKNSSHREF at 3.184 million tonnes. Fmsy was estimated by WKNSSHMSE at 0.157.

### 4.6.3 Management reference points

In the current management strategy, which was agreed upon in October 2018, the Coastal States have agreed a target reference point defined at $F_{\text {target }}=0.14$ when the stock is above $B_{\text {pa }}$. If the SSB is below $\mathrm{B}_{\mathrm{pa}}$, a linear reduction in the fishing mortality rate will be applied from 0.14 at $\mathrm{B}_{\mathrm{pa}}$ to 0.05 at Blim.

### 4.7 State of the stock

The SSB on 1 January 2022 is estimated by XSAM to be 3.66 million tonnes which is above $B_{p a}$ ( 3.184 million t ). The spawning stock has been declining since 2009 but increased in 2021 and 2022 followed by a decrease again in 2023. The SSB time-series from the 2023 assessment is consistent with the SSB time-series from the 2022 assessment. In the last 20 years, several large year classes have been produced (1998, 1999, 2002, and 2004). The year classes 2005-2015 and 20172019 are estimated to be average or small, while the 2016 year-class is estimated to be above average in the 2023 assessment. Since there was no recruitment survey in 2023, the size of the 2021 year-class at age 2 was defined as the stochastic median recruitment in the time series. Fishing mortality in 2022 is estimated to be 0.182 which is above the management strategy $\mathrm{F}(0.140)$.

### 4.8 NSSH Catch predictions for 2024

### 4.8.1 Input data for the forecast

Forecasting was conducted using XSAM according to the method described in the Stock Annex and by Aanes (2016c). WGWIDE 2016 decided to use the point estimates from this forecast as basis for the advice. In short, the forecast is made by applying the point estimates of the stock status as input to set TAC, then based on the TAC a stochastic forecast was performed to determine levels of precision in the forecast. Table 4.8.1.1 lists the point estimates of the starting values for the forecast. The input stock numbers-at-age 2 and older were taken from the final assessment. The catch weight-at-age, used in the forecast, is the average of the observed catch weights over the last 3 years (2020-2022).

For the weight-at-age in the stock, the values for 2023 were obtained from the commercial fisheries in the wintering areas in January. For the years 2024 and 2025 the average of the last 3 years (2021-2023) was used.

Standard values for natural mortality were used. Maturity-at-age was based on the information presented in Section 4.4.5.

The exploitation pattern used in the forecast is taken from the predictions made by the model (see Aanes 2016c for details). The resultant mean annual exploitation pattern is shown in Figure 4.8.1.1 and displays a shift towards older fish in the recent years and further in the prediction. Prediction of recruitment at age 2 is obtained by the model with a mean that in practice represents the long term (1988-2022) estimated mean recruitment (back-transformed mean at log scale) and variance the corresponding recruitment variability over the period. Forecasted values of recruits are highly imprecise but have little influence on the short-term forecast of SSB as the
herring starts to mature at age 4 . Note that the 2016 year-class is regarded as large; hence, the maturity is set to be lower than for smaller year-classes. This results in the contribution of the 2016 year-class to the SSB being delayed.

The average fishing mortality is defined as the average over the ages 5 to $12+$, weighted over the population numbers in the relevant year

$$
\bar{F}_{y}=\sum_{a=5}^{12} N_{a, y} F_{a, y} / \sum_{a=5}^{12} N_{a, y}
$$

where $F_{a, y}$ and $N_{a, y}$ are fishing mortalities and numbers by age and year. This procedure is in accordance with that used in previous years for this stock although the age range was shifted from 5-11 to 5-12+ from 2018.

There was no agreement between the fishing parties on the sharing of the TAC for 2023. Therefore, to obtain an estimate of the total catch to be used as input for the catch-constraint projections for 2023, the sum of the unilateral quotas was used. In total, the expected outtake from the stock in 2023 amounts to 692942 tonnes. F in 2023 is estimated by XSAM based on this catch.

### 4.8.2 Results of the forecast

The Management Options Table with the results of the forecast is presented in Table 4.8.2.1. Assuming a total catch 692942 tonnes is taken in 2023, it is expected that the SSB will decrease from 3.664 million tonnes on 1 January in 2023 to 3.059 million tonnes in 2024, which is bellow Btrigger ( 3.184 million tonnes). The weighted F over ages $5-12+$ is 0.186 . The model predicts that the catch in 2023 to be dominated by three age groups, age $8(45 \%)$, age $11(8 \%)$, and age $12+(11 \%)$.

### 4.8.3 Comparison between forecast 2022 and 2023

The stock number at age used as the input for the forecast in 2022 and 2023 is seen in Fig. 4.8.3.1, along with the revised stock numbers in 2022 from the 2023 assessment and the predicted stock number in 2023 from the 2022 forecast. Small revisions were observed, i.e., for 2022 estimates, stock number at age 2 is negatively revised with the inclusion of new data and the stock number at age 6 is slightly positively revised. The predicted stock numbers in 2023 from the 2022 forecast similar to the stock number used in the 2023 forecast, with just smaller revisions.

Comparing the selection pattern used in the 2022 and the 2023 forecast, small differences were identified for the ages 9 to 12. The weight-in-stock between the two forecast years only has revisions. The difference between the basis for the short-term prediction in 2022 and the revised values from the 2023 assessment are summarised below.

|  | Forecast values 2022 | Assessment values 2023 | \%difference |
| :--- | :--- | :--- | :--- |
| Landings 2022 | 827963 | 813833 | -1.7 |
| Recruitment (age 2) <br> billions | 10.671 | 6.944 | -35 |
| SSB (2023) | 3.532 | 3.664 | 3.8 |
| F (2022) | 0.192 | 0.182 | -5 |

### 4.9 Comparison with previous assessment

A comparison between the assessments 2011-2023 is shown in Figure 4.9.1. In the years 20112015 the assessments were made with TASACS, but since 2016 XSAM has been applied, as accepted by WKPELA 2016. With the change of the assessment tool in 2016 the age of the recruitment changed from 0 to 2 and the age span in the reference $F$ changed from 5-14 to $5-11$. In WKNSSHREF (ICES, 2018a) this was further changed to 5-12+.

The table below shows the SSB (thousand tonnes) on 1 January in 2022 and weighted F in 2021 as estimated in 2022 and 2023.

|  | ICES 2022 | WG2023 | \%difference |
| :--- | :--- | :--- | :--- |
| SSB (2022) | 3867 | 4060 | $5.0 \%$ |
| Weighted F (2021) | 0.168 | 0.159 | $-5.4 \%$ |

### 4.10 Management plans and evaluations

The current management strategy for the Norwegian spring spawning herring fishery was agreed by the Coastal States in October 2018.

The implemented long-term management strategy of Norwegian spring spawning herring is consistent with the precautionary approach and the MSY approach (WKNSSHREF, ICES, 2018a; WKNSSHMSE, ICES, 2018b) and aims at ensuring harvest rates within safe biological limits.

Coastal States Agreed Record (4th November 2022)

1. The Parties agree to implement a long-term management strategy for the fisheries on the Norwegian Spring Spawning (Atlanto-Scandian) Herring stock, which is consistent with the precautionary approach and the MSY approach, aiming at ensuring harvest rates within safe biological limits.
2. For the purpose of this long-term management strategy, in the following text, "TAC" means the total allowable catch as agreed by Coastal States.
3. As a priority, the long-term management strategy shall ensure with high probability that the size of the stock is maintained above Blim.
4. In the case that the spawning biomass is forecast to be above or equal to $B_{\text {trigger }}\left(=B_{p a}\right)$ on 1 January of the year for which the TAC is to be set, the TAC shall be fixed to a fishing mortality of $\mathrm{F}_{\mathrm{mgt}}=0.14$.
5. Where the rules in paragraph 4 would lead to a TAC, that deviates by more than $20 \%$ below or $25 \%$ above the TAC of the preceding year, the Parties shall fix a TAC that is respectively no more than $20 \%$ less or $25 \%$ more than the TAC of the preceding year. The TAC constraint shall not apply if the spawning biomass at 1 January in the year for which the TAC is to be set is less than $B_{\text {trigger }}$
6. In the case that the spawning biomass (SSB) is forecast to be less than the precautionary biomass ( $\mathrm{B}_{\text {trigger }}$ ) but above or equal to Blim on 1 January of the year for which the TAC is to be set, the TAC shall be fixed at a level that is consistent with a fishing mortality given by:
7. $\quad$ Target $\mathrm{F}=0.05+\left[\left(\mathrm{SSB}-\mathrm{Blim}^{2}\right)^{*}\left(\mathrm{~F}_{\mathrm{mgt}}-0.05\right) \mathrm{I}\left(\mathrm{B}_{\text {trigger }}-\mathrm{Blim}_{\text {lim }}\right)\right]$
8. In the case that the spawning biomass is forecast to be less than Blim on 1 January of the year for which the TAC is to be set, the TAC will be fixed corresponding to a fishing mortality $\mathrm{F}=0.05$.
9. Each Party may transfer to the following year unutilised quantities of up to $10 \%$ of the quota allocated to it. The quantity transferred shall be in addition to the quota allocated to the Party concerned in the following year.
10. Each Party may authorize fishing by its vessels of up to $10 \%$ beyond the quota allocated. However, this shall not apply if the stock is forecast to be under Btrigger at the end of the TAC year. All quantities fished beyond the allocated quota for one year shall be deducted from the Party's quota allocated for the following year.
11. The Parties, on the basis of ICES advice, shall review this long-term management strategy at intervals not exceeding five years. The first such review shall take place no later than 2023, in time for ICES to issue advice for 2024.

Whilst the Coastal States stated a review of the long-term management strategy was due in 2023 this has not been undertaken as it was considered appropriate to do this after an impending benchmark.

A brief history of management strategies is in the stock annex. There has been no agreement on sharing of the TAC since 2013, resulting in the total catch being higher than the advised catch.

### 4.11 Management considerations

Perception of the stock has not changed since last year's assessment (estimated SSB in 2022 is 5\% higher in this year's assessment).

Historically, the size of the stock has shown large variations and dependency on the irregular occurrence of very strong year classes. Between 1998 and 2004 the stock produced several strong year classes which lead to an increase in SSB until 2009. Since then, SSB has declined due to absence of strong year classes in 2005-2015. The 2016 year-class was however, estimated to be well above average in the 2023 assessment and resulted in an increase in SSB from 2020 to 2021. SSB, however, declined in 2022-2023 and is predicted to be below Bmgt in both 2024 and 2025 even if the management strategy $\left(\mathrm{Fmgt}_{\mathrm{mg}}=0.123\right)$ is applied in 2024.

Between 1999 and 2013, catches were regulated through an agreed management. However, since 2013, a lack of agreement by the Coastal States on their share in the TAC has led to unilaterally set quotas which together are higher than the TAC indicated by the management strategy resulting in steeper reduction in the SSB than otherwise.

A new management strategy was implemented for the advisory year 2019.

### 4.12 Ecosystem considerations

NSS herring juveniles and adults are an important part of the ecosystems in the Barents Sea, along the Norwegian coast, in the Norwegian Sea and in adjacent waters. This refers both to predation on zooplankton by herring and herring being a food resource to higher trophic levels (e.g. cod, saithe, seabirds, and marine mammals). The predation intensity of and on herring have seasonal, spatial, and temporal variation because of variation in migration pattern, prey density, stock size, size of year classes and stock sizes of competing stocks for resources and predators. Recent features of some of these ecosystem factors of relevance for the stock are summarized below.

- Following a maximum in zooplankton biomass in May during the early 2000s the biomass declined with a minimum in 2006. From 2010, the trend turned to an increase and the last five years the zooplankton biomass has fluctuated around the long-term mean in the Lofoten and Norwegian Basins (IESNS survey report - ICES, 2023a), but is still low compared to the early years in East Iceland waters and the Jan Mayen front. Interestingly,
all the areas, excluding east of Iceland and on few occasions Jan Mayen, show co-varying changes in zooplankton biomass.
- The Atlantic water mass in the Norwegian Sea was warmer and saltier over the period 2000-2016 than the long-term mean (WGINOR - ICES, 2023b). However, since 2017, the extent of Artic water has increased resulting in a marked freshening of the water masses, but the temperature remain above long -term average. Two different mechanisms can explain this, increased fraction of subpolar water (fresh and cold) and low heat loss to the atmosphere in the Norwegian Atlantic flow. The recent trend of colder and fresher Atlantic Inflow into the Norwegian Sea has ceased (ICES, 2023b).
- The sea temperature in 2023 was generally below the long-term mean (1995-2021) in the Norwegian Sea, but the pattern was fragmented from $0 \mathrm{~m}-200 \mathrm{~m}$ depth, with the central part mostly above the mean (ICES, 2023a). The Arctic front in the southern Norwegian Sea was more southerly and easterly located in 2023 compared to the long-term mean.
- In general, the herring stock has had a more westerly feeding distribution (ICES, 2023a) in the recent years than what was previously observed. In May 2023, the herring in west was more northerly distributed than in recent years, with the southeastern areas virtually absent of herring. The large 2016-year class was by far the most abundant year class in areas where herring was found. The westerly distribution might be due to either better feeding opportunities there or a response to feeding competition with mackerel, but the consequence is a less spatial overlap of herring and mackerel in Norwegian Sea and adjoining waters since around 2014 (ICES, 2023c).
- Where herring and mackerel overlap spatially, they compete for food to some extent (Bachiller et al., 2016, 2018; Debes et al., 2012; Langøy et al., 2012; Óskarsson et al., 2016). There are studies showing mackerel being more effective feeder, which might indicate that the herring is forced to the southwestern and northeastern fringe of Norwegian Sea (ICES, 2023b). Alternatively, the higher zooplankton biomass in the southwest could also attract the herring into this location, since zooplankton biomass is much lower in the northeast (ICES, 2023b).
- $\quad$ Since 2005, herring have tended to stay longer in the western feeding areas into the autumn months (Homrum et al., 2022). During this same period, herring have had generally stable good somatic condition and started to build gonads in the autumn months. This trend is more pronounced for older herring. The most likely explanation is, that herring is able to utilize zooplankton resources in autumn - for example Calanus finmarchicus, which has been shown to be in the area in the autumn months (Strand et al., 2020).
- Results of stomach analyses of mackerel on the Norwegian coastal shelf (between about $66^{\circ} \mathrm{N}$ and $69^{\circ} \mathrm{N}$ ) suggest that mackerel fed opportunistically on herring larvae, and that predation pressure therefore largely depends on the degree of overlap in time and space (Skaret et al., 2015). Sampling in June 2017 and 2018, specifically studying mackerel predation on herring larvae, found significant numbers of herring larvae in mackerel stomachs in the area just south of Lofoten (Allan et al., 2021).
- The 2016 year-class of herring was the strongest since the 2004-year class in the Norwegian Sea as 4-year-olds, but as expected abundance is now beginning to visibly decrease (see the IESNS survey 2023 (Table 4.4.7.3).
- In the winter 2017/2018, the overwintering grounds shifted northward along the coast of Norway with older individuals occurring in oceanic areas. Such changes previously coincided with large year classes entering the spawning stock, however this recent change did not. Also, the onset of the overwintering period has been later in the year since the end of the 2000s.
Around spawning time of 2023 most of the spawning stock was found west of Lofoten and Vesterålen, further north and more concentrated than usual but similar to the 2022
survey. The observed maturity indicated a later spawning compared to the previous year (WGWIDE WD04).


### 4.13 Changes in fishing patterns

The fishery for Norwegian spring spawning herring has previously (before 2013) been described as progressing clockwise in the Nordic Seas during the year. However, the last 5-9 years the annual progression of the fishery has changed into a pendular behaviour, starting in the winter along the Norwegian coast, moving gradually to the west towards Iceland in the summer, and then east again into the central Norwegian Sea in the last quarter of the year.

The fishery reached its lowest catches since the mid-nineties in 2015, after which the catches increased again, reaching a maximum in 2021 of 850000 tonnes (Table 4.4.1.1). It is mainly the fishery in the fourth quarter that has increased since 2015, with up to $2 / 3$ of the catches taken in this quarter. The fishery in quarter four in the last few years has partly been north and west of Lofoten and partly in the central Norwegian Sea and east of Iceland, whereas before 2015 it used to be stretched out towards the coast of Norway and north towards the Bear Island.

The change in migration pattern since 2017/18, where the part of the stock (old fish) overwinters in the central Norwegian Sea, has caused the fishery in this area to be extended to later in the winter, and in 2022 there was fishery in the central Norwegian Sea in the first quarter as well as the fourth.

### 4.14 Recommendations

For some years there have been proposed issues with age reading of NSS herring. Following a recommendation from WGWIDE a scale/otolith exchange was organised in autumn 2022 and spring 2023 and a subsequent workshop was held in April 2023. The main results from the workshop were presented to the working group and final report will be published later this year. The main conclusions were that there was generally good agreement among age-readers up until age 8 (otoliths) - 9 (scales) when the outermost added winter-rings become difficult to distinguish. Agreement was particularly good within the groups that read scales and otoliths respectively, but also agreement was good between scale- and otolith readers. For older fish (~older than 8-9 years) agreement decreased and more significantly for otolith-readers than for scale readers. It was concluded that age-readers generally followed the same procedures (closely adhering to a proposed protocol formulated in 2016 (ICES, 2016b). The workshop could not come up with suggestions to improve agreement for age-reading of older fish, since all age-readers are adhering to the proposed protocol. Mixture with other herring stocks did not appear to largely hamper age-reading of NSS-herring, since there was good agreement among age-readers with regards to how to distinguish between NSS-herring and other herring stocks (i.e. otolith characteristics and/or maturity stage).

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### 4.16 Tables

Table 4.4.1.1 Total landings (ICES estimate) of Norwegian spring-spawning herring (tons) since 1972. Data provided by Working Group members.

| Year | Norway | USSR/ <br> Russia | Denmark | Faroes | Iceland | Ireland | Netherlands | Greenland | UK | Germany | France | Poland | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1972 | 13161 | - | - | - | - | - | - | - | - | - | - | - | - | 13161 |
| 1973 | 7017 | - | - | - | - | - | - | - | - | - | - | - | - | 7017 |
| 1974 | 7619 | - | - | - | - | - | - | - | - | - | - | - | - | 7619 |
| 1975 | 13713 | - | - | - | - | - | - | - | - | - | - | - | - | 13713 |
| 1976 | 10436 | - | - | - | - | - | - | - | - | - | - | - | - | 10436 |
| 1977 | 22706 | - | - | - | - | - | - | - | - | - | - | - | - | 22706 |
| 1978 | 19824 | - | - | - | - | - | - | - | - | - | - | - | - | 19824 |
| 1979 | 12864 | - | - | - | - | - | - | - | - | - | - | - | - | 12864 |
| 1980 | 18577 | - | - | - | - | - | - | - | - | - | - | - | - | 18577 |
| 1981 | 13736 | - | - | - | - | - | - | - | - | - | - | - | - | 13736 |
| 1982 | 16655 | - | - | - | - | - | - | - | - | - | - | - | - | 16655 |
| 1983 | 23054 | - | - | - | - | - | - | - | - | - | - | - | - | 23054 |
| 1984 | 53532 | - | - | - | - | - | - | - | - | - | - | - | - | 53532 |
| 1985 | 167272 | 2600 | - | - | - | - | - | - | - | - | - | - | - | 169872 |
| 1986 | 199256 | 26000 | - | - | - | - | - | - | - | - | - | - | - | 225256 |


| Year | Norway | $\begin{aligned} & \text { USSR/ } \\ & \text { Russia } \end{aligned}$ | Denmark | Faroes | Iceland | Ireland | Netherlands | Greenland | UK | Germany | France | Poland | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1987 | 108417 | 18889 | - | - | - | - | - | - | - | - | - | - | - | 127306 |
| 1988 | 115076 | 20225 | - | - | - | - | - | - | - | - | - | - | - | 135301 |
| 1989 | 88707 | 15123 | - | - | - | - | - | - | - | - | - | - | - | 103830 |
| 1990 | 74604 | 11807 | - | - | - | - | - | - | - | - | - | - | - | 86411 |
| 1991 | 73683 | 11000 | - | - | - | - | - | - | - | - | - | - | - | 84683 |
| 1992 | 91111 | 13337 | - | - | - | - | - | - | - | - | - | - | - | 104448 |
| 1993 | 199771 | 32645 | - | - | - | - | - | - | - | - | - | - | - | 232457 |
| 1994 | 380771 | 74400 | - | 2911 | 21146 | - | - | - | - | - | - | - | - | 479228 |
| 1995 | 529838 | 101987 | 30577 | 57084 | 174109 | - | 7969 | 2500 | 881 | 556 | - | - | - | 905501 |
| 1996 | 699161 | 119290 | 60681 | 52788 | 164957 | 19541 | 19664 | - | 46131 | 11978 | - | - | 22424 | 1220283 |
| 1997 | 860963 | 168900 | 44292 | 59987 | 220154 | 11179 | 8694 | - | 25149 | 6190 | 1500 | - | 19499 | 1426507 |
| 1998 | 743925 | 124049 | 35519 | 68136 | 197789 | 2437 | 12827 | - | 15971 | 7003 | 605 | - | 14863 | 1223131 |
| 1999 | 740640 | 157328 | 37010 | 55527 | 203381 | 2412 | 5871 | - | 19207 | - | - | - | 14057 | 1235433 |
| 2000 | 713500 | 163261 | 34968 | 68625 | 186035 | 8939 | - | - | 14096 | 3298 | - | - | 14749 | 1207201 |
| 2001 | 495036 | 109054 | 24038 | 34170 | 77693 | 6070 | 6439 | - | 12230 | 1588 | - | - | 9818 | 766136 |
| 2002 | 487233 | 113763 | 18998 | 32302 | 127197 | 1699 | 9392 | - | 3482 | 3017 | - | 1226 | 9486 | 807795 |
| 2003* | 477573 | 122846 | 14144 | 27943 | 117910 | 1400 | 8678 | - | 9214 | 3371 | - | - | 6431 | 789510 |
| 2004 | 477076 | 115876 | 23111 | 42771 | 102787 | 11 | 17369 | - | 1869 | 4810 | 400 | - | 7986 | 794066 |


| Year | Norway | USSR/ <br> Russia | Denmark | Faroes | Iceland | Ireland | Netherlands | Greenland | UK | Germany | France | Poland | Sweden | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 580804 | 132099 | 28368 | 65071 | 156467 | - | 21517 | - | - | 17676 | 0 | 561 | 680 | 1003243 |
| 2006 | 567237 | 120836 | 18449 | 63137 | 157474 | 4693 | 11625 | - | 12523 | 9958 | 80 | - | 2946 | 968958 |
| 2007 | 779089 | 162434 | 22911 | 64251 | 173621 | 6411 | 29764 | 4897 | 13244 | 6038 | 0 | 4333 | 0 | 1266993 |
| 2008 | 961603 | 193119 | 31128 | 74261 | 217602 | 7903 | 28155 | 3810 | 19737 | 8338 | 0 | 0 | 0 | 1545656 |
| 2009 | 1016675 | 210105 | 32320 | 85098 | 265479 | 10014 | 24021 | 3730 | 25477 | 14452 | 0 | 0 | 0 | 1687371 |
| 2010 | 871113 | 199472 | 26792 | 80281 | 205864 | 8061 | 26695 | 3453 | 24151 | 11133 | 0 | 0 | 0 | 1457015 |
| 2011 | 572641 | 144428 | 26740 | 53271 | 151074 | 5727 | 8348 | 3426 | 14045 | 13296 | 0 | 0 | 0 | 992997 |
| 2012 | 491005 | 118595 | 21754 | 36190 | 120956 | 4813 | 6237 | 1490 | 12310 | 11945 | 0 | 0 | 705 | 826000 |
| 2013 | 359458 | 78521 | 17160 | 105038 | 90729 | 3815 | 5626 | 11788 | 8342 | 4244 | 0 | 0 | 23 | 684743 |
| 2014 | 263253 | 60292 | 12513 | 38529 | 58828 | 706 | 9175 | 13108 | 4233 | 669 | 0 | 0 | 0 | 461306 |
| 2015 | 176321 | 45853 | 9105 | 33031 | 42625 | 1400 | 5255 | 12434 | 55 | 2660 | 0 | 0 | 0 | 328740 |
| 2016 | 197501 | 50455 | 10384 | 44727 | 50418 | 2048 | 3519 | 17508 | 4031 | 2582 | 0 | 0 | 0 | 383174 |
| 2017 | 389383 | 91118 | 19037 | 98170 | 90400 | 3495 | 6679 | 12569 | 4358 | 5201 | 0 | 1 | 1155 | 721566 |
| 2018 | 332028 | 64185 | 17052 | 82062 | 83393 | 2428 | 4290 | 2465 | 2582 | 1989 | 0 | 0 | 425 | 592899 |
| 2019 | 430507 | 84364 | 21207 | 113945 | 108045 | 2775 | 5111 | 3190 | 1801 | 4188 | 0 | 1327 | 705 | 777165 |
| 2020 | 409436 | 74936 | 16523 | 103029 | 98173 | 2704 | 5060 | 3546 | 143 | 2969 | 0 | 1352 | 3065 | 720937 |
| 2021** | 489632 | 92841 | 15854 | 114291 | 114299 | 1793 | 10939 | 6456 | 0 | 3365 | 0 | 1242 | 1101 | 851813 |
| 2022 | 445938 | 85870 | 15014 | 122083 | 112739 | 3209 | 3783 | 6818 | 9620 | 5600 | 0 | 0 | 3160 | 813834 |

## *In 2003 the Norwegian catches were raised of 39433 to account for changes in percentages of water content.

**The Russian catch for 2021 was taken from the ICES preliminary catches database

Table 4.4.1.2 Norwegian spring-spawning herring. Sampling coverage by year.

| Year | TOTAL CATCH | \% catch covered by sampling programme | No. samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 1207201 | 86 | 389 | 55956 | 10901 |
| 2001 | 766136 | 86 | 442 | 70005 | 11234 |
| 2002 | 807795 | 88 | 184 | 39332 | 5405 |
| 2003 | 789510 | 71 | 380 | 34711 | 11352 |
| 2004 | 794066 | 79 | 503 | 48784 | 13169 |
| 2005 | 1003243 | 86 | 459 | 49273 | 14112 |
| 2006 | 968958 | 93 | 631 | 94574 | 9862 |
| 2007 | 1266993 | 94 | 476 | 56383 | 14661 |
| 2008 | 1545656 | 94 | 722 | 81609 | 31438 |
| 2009 | 1686928 | 94 | 663 | 65536 | 12265 |
| 2010 | 1457015 | 91 | 1258 | 124071 | 12377 |
| 2011 | 992.997 | 95 | 766 | 79360 | 10744 |
| 2012 | 825.999 | 93 | 649 | 59327 | 14768 |
| 2013 | 684.743 | 91 | 402 | 33169 | 11431 |
| 2014 | 461.306 | 89 | 229 | 18370 | 5813 |
| 2015 | 328.739 | 92 | 177 | 25156 | 5039 |
| 2016 | 383.174 | 91 | 203 | 39120 | 5892 |
| 2017 | 721566 | 95 | 335 | 31755 | 7241 |
| 2018 | 592899 | 97 | 253 | 22106 | 6047 |
| 2019 | 777165 | 97 | 361 | 29856 | 7421 |
| 2020 | 720937 | 98 | 232 | 34232 | 6742 |
| 2021 | 851813 | 88 | 207 | 18830 | 5975 |
| 2022 | 813834 | 97 | 299 | 23100 | 5844 |

Table 4.4.1.3 Norwegian spring-spawning herring. Sampling coverage by country in 2022.

| COUNTRY | OFFICIAL CATCH | \% catch covered <br> by sampling pro- <br> gramme | NO. SAMPLES | NO. MEASURED | NO. AGED |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Denmark | 15014 | 100 | 11 | 850 | 256 |


| COUNTRY | OFFICIAL CATCH | \% catch covered by sampling programme | NO. SAMPLES | NO. MEASURED | NO. AGED |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Faroes | 122083 | 100 | 16 | 581 | 581 |
| Germany | 5600 | 0 | 0 | 0 | 0 |
| Greenland | 6818 | 0 | 1 | 200 | 0 |
| Iceland | 112739 | 100 | 65 | 2950 | 1498 |
| Ireland | 3209 | 83 | 3 | 262 | 159 |
| Netherlands | 3783 | 0 | 0 | 0 | 0 |
| Norway | 445938 | 99 | 85 | 2363 | 2363 |
| Russia | 85870 | 98 | 114 | 15536 | 757 |
| Scotland | 9620 | 100 | 4 | 358 | 230 |
| Sweden | 3160 | 0 | 0 | 0 | 0 |
| Total | 813834 | 97 | 299 | 23100 | 5844 |

Table 4.4.1.4 Norwegian spring-spawning herring. Sampling coverage by ICES Division in 2022.

| Area | Official Catch | No Sam- <br> ples | No Aged | No Meas- <br> ured | No Aged/ 1000 <br> tonnes | No Measured/ 1000 <br> tonnes |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2.a | 676478 | 244 | 4699 | 20701 | 7 | 31 |
| 2.b | 408 | 0 | 0 | 0 | 0 | 0 |
| 4.a | 185 | 0 | 0 | 0 | 0 | 0 |
| 5.a | 136618 | 55 | 1145 | 2399 | 8 | 0 |
| 5.b | 145 | 0 | 0 | 0 | 0 | 28 |
| Total | 813834 | 299 | 5844 | 23100 | 7 | 0 |

Table 4.4.1.5 Norwegian spring-spawning herring. Catch data provided by working group members and samples allocated to unsampled catches in SALLOC.

| Line | Country | Quarter | Div. | Catch (T) | Samples allocated (line) |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 1 | Norway | 1 | Ila | 147792 |  |
| 2 | Norway | 2 | Ila | 80 | $12,18,31$ |
| 3 | Norway | 3 | Ila | 4400 | $13,19,32$ |
| 4 | Norway | 4 | Ila | 293665 |  |
| 5 | Sweden | 1 | Ila | 3160 | $1,15,17,27,35$ |


| Line | Country | Quarter | Div. | Catch ( T ) | Samples allocated (line) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | Germany | 4 | $11 a$ | 5600 | 4,14,16,20,33 |
| 7 | Greenland | 3 | 11 a | 796 | 13,19,32 |
| 8 | Greenland | 4 | Ila | 5873 | 4,14,16,20,33 |
| 9 | Greenland | 3 | 11 b | 150 | 13,19,32 |
| 10 | Iceland | 3 | Va | 53762 |  |
| 11 | Iceland | 4 | Va | 47363 |  |
| 12 | Iceland | 2 | Ila | 278 |  |
| 13 | Iceland | 3 | 11 a | 10713 |  |
| 14 | Iceland | 4 | 11 a | 623 |  |
| 15 | Denmark | 1 | Ila | 7931 |  |
| 16 | Denmark | 4 | Ila | 7082 |  |
| 17 | Faroes | 1 | 11 a | 10537 |  |
| 18 | Faroes | 2 | Ila | 4370 |  |
| 19 | Faroes | 3 | Ila | 10230 |  |
| 20 | Faroes | 4 | 11 a | 60945 |  |
| 21 | Faroes | 3 | IIb | 258 | 13,19,32 |
| 22 | Faroes | 2 | IVa | 185 | 12,18,31 |
| 23 | Faroes | 4 | Va | 35493 |  |
| 24 | Faroes | 2 | Vb | 24 | 12,18,31 |
| 25 | Faroes | 3 | Vb | 1 | 10 |
| 26 | Faroes | 4 | Vb | 40 | 11,23 |
| 27 | Ireland | 1 | 1 la | 2676 |  |
| 28 | Ireland | 4 | lla | 533 | 4,14,16,20,33 |
| 29 | Netherlands | 4 | 11 a | 3783 | 4,14,16,20,33 |
| 30 | Russia | 1 | lla | 1640 | 1,15,17,27,35 |
| 31 | Russia | 2 | lla | 512 |  |
| 32 | Russia | 3 | lla | 6979 |  |
| 33 | Russia | 4 | 11 a | 76659 |  |
| 34 | Russia | 2 | Vb | 80 | 12,18,31 |


| Line | Country | Quarter | Div. | Catch (T) | Samples allocated (line) |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 35 | Scotland | 1 | Ila | 9620 |  |

## Table 4.4.3.1. Norwegian spring spawning herring. Catch in numbers (thousands).

|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1950 | 5112600 | 2000000 | 600000 | 276200 | 184800 | 185500 | 547000 | 628600 | 79500 | 88600 | 109500 | 86900 | 194500 | 368300 | 66400 | 344300 |
| 1951 | 1635500 | 7607700 | 400000 | 6600 | 383800 | 172400 | 164400 | 515600 | 602000 | 77100 | 82700 | 103100 | 107600 | 253500 | 348000 | 352500 |
| 1952 | 13721600 | 9149700 | 1232900 | 39300 | 60500 | 602300 | 136300 | 204500 | 380200 | 377900 | 79200 | 85700 | 107700 | 106800 | 186500 | 564400 |
| 1953 | 5697200 | 5055000 | 581300 | 740100 | 46600 | 100900 | 355600 | 81900 | 110900 | 314100 | 394900 | 61700 | 91200 | 94100 | 98800 | 730400 |
| 1954 | 10675990 | 7071090 | 855400 | 266300 | 1435500 | 142900 | 236000 | 490300 | 128100 | 199800 | 440400 | 460700 | 88400 | 100600 | 133000 | 803200 |
| 1955 | 5175600 | 2871100 | 510100 | 93000 | 276400 | 2045100 | 114300 | 189600 | 274700 | 85300 | 193400 | 295600 | 203200 | 58700 | 84600 | 580600 |
| 1956 | 5363900 | 2023700 | 627100 | 116500 | 251600 | 314200 | 2555100 | 110000 | 203900 | 264200 | 130700 | 198300 | 272800 | 163300 | 63000 | 565100 |
| 1957 | 5001900 | 3290800 | 219500 | 23300 | 373300 | 153800 | 228500 | 1985300 | 72000 | 127300 | 182500 | 88400 | 121200 | 149300 | 131600 | 281400 |
| 1958 | 9666990 | 2798100 | 666400 | 17500 | 17900 | 110900 | 89300 | 194400 | 973500 | 70700 | 123000 | 200900 | 98700 | 77400 | 70900 | 255600 |
| 1959 | 17896280 | 198530 | 325500 | 15100 | 26800 | 25900 | 146600 | 114800 | 240700 | 1103800 | 88600 | 124300 | 198000 | 88500 | 77400 | 235900 |
| 1960 | 12884310 | 13580790 | 392500 | 121700 | 18200 | 28100 | 24400 | 96200 | 73300 | 203900 | 1163000 | 85200 | 129700 | 153500 | 56700 | 168900 |
| 1961 | 6207500 | 16075600 | 2884800 | 31200 | 8100 | 4100 | 15000 | 19400 | 61600 | 49200 | 136100 | 728100 | 49700 | 45000 | 63000 | 60100 |
| 1962 | 3693200 | 4081100 | 1041300 | 1843800 | 8000 | 3100 | 7200 | 20200 | 11900 | 59100 | 52600 | 117000 | 813500 | 44200 | 54700 | 152300 |
| 1963 | 4807000 | 2119200 | 2045300 | 760400 | 835800 | 5300 | 1800 | 3600 | 18300 | 9300 | 107700 | 92500 | 174100 | 923700 | 79600 | 185300 |
| 1964 | 3613000 | 2728300 | 220300 | 114600 | 399000 | 2045800 | 13700 | 1500 | 3000 | 24900 | 29300 | 95600 | 82400 | 153000 | 772800 | 336800 |
| 1965 | 2303000 | 3780900 | 2853600 | 89900 | 256200 | 571100 | 2199700 | 19500 | 14900 | 7400 | 19100 | 40000 | 100500 | 107800 | 138700 | 883100 |
| 1966 | 3926500 | 662800 | 1678000 | 2048700 | 26900 | 466600 | 1306000 | 2884500 | 37900 | 14300 | 17400 | 26200 | 11000 | 69100 | 72100 | 556700 |


|  | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 |  | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1967 | 426800 | 9877100 |  | 70400 | 1392300 | 3254000 | 26600 | 421300 | 1132000 | 1720800 | 8900 | 5700 | 3500 | 8500 | 8900 | 17500 | 104400 |
| 1968 | 1783600 | 437000 |  | 388300 | 99100 | 1880500 | 1387400 | 14220 | 94000 | 134100 | 345100 | 2000 | 1100 | 830 | 2500 | 2600 | 17000 |
| 1969 | 561200 | 507100 |  | 141900 | 188200 | 800 | 8800 | 4700 | 700 | 11700 | 33600 | 36000 | 300 | 200 | 200 | 200 | 2400 |
| 1970 | 119300 | 529400 |  | 33200 | 6300 | 18600 | 600 | 3300 | 3300 | 1000 | 13400 | 26200 | 28100 | 300 | 100 | 200 | 2000 |
| 1971 | 30500 | 42900 |  | 85100 | 1820 | 1020 | 1240 | 360 | 1110 | 1130 | 360 | 4410 | 6910 | 5450 | 0 | 20 | 120 |
| 1972 | 347100 | 41000 |  | 20400 | 35376 | 3476 | 3583 | 2481 | 694 | 1486 | 198 | 0 | 494 | 593 | 593 | 0 | 0 |
| 1973 | 29300 | 3500 |  | 1700 | 2389 | 25200 | 651 | 1506 | 278 | 178 | 0 | 0 | 0 | 0 | 0 | 180 | 0 |
| 1974 | 65900 | 7800 |  | 3900 | 100 | 241 | 24505 | 257 | 196 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1975 | 30600 | 3600 |  | 1800 | 3268 | 132 | 910 | 30667 | 5 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1976 | . 20100 | 2400 |  | 1200 | 23248 | 5436 | 0 | 0 | 13086 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1977 | 43000 | 6200 |  | 3100 | 22103 | 23595 | 336 | 0 | 419 | 10766 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1978 | 20100 | 2400 |  | 1200 | 3019 | 12164 | 20315 | 870 | 0 | 620 | 5027 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1979 | 32600 | 3800 |  | 1900 | 6352 | 1866 | 6865 | 11216 | 326 | 0 | 0 | 2534 | 0 | 0 | 0 | 0 | 0 |
| 1980 | 6900 | 800 |  | 400 | 6407 | 5814 | 2278 | 8165 | 15838 | 441 | 8 | 0 | 2688 | 0 | 0 | 0 | 0 |
| 1981 | 8300 | 1100 |  | 11900 | 4166 | 4591 | 8596 | 2200 | 4512 | 8280 | 345 | 103 | 114 | 964 | 0 | 0 | 0 |
| 1982 | 22600 | 1100 |  | 200 | 13817 | 7892 | 4507 | 6258 | 1960 | 5075 | 6047 | 121 | 37 | 37 | 121 | 0 | 0 |
| 1983 | 127000 | 4680 | 1670 | 3183 | 21191 | 19521 | 6181 | 6823 | 1293 | 4598 | 7329 | 143 | 40 |  |  | 0 |  |


| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1984 | 33860 | 1700 | 2490 | 4483 | 5388 | 61543 | 18202 | 12638 | 15608 | 7215 | 16338 | 6478 | 0 | 0 | 0 | 1650 |
| 1985 | 28570 | 13150 | 207220 | 21500 | 15500 | 16500 | 130000 | 59000 | 55000 | 63000 | 10000 | 31000 | 50000 | 0 | 0 | 2640 |
| 1986 | 13810 | 1380 | 3090 | 539785 | 17594 | 14500 | 15500 | 105000 | 75000 | 42000 | 77000 | 19469 | 66000 | 80000 | 0 | 2470 |
| 1987 | 13850 | 6330 | 35770 | 19776 | 501393 | 18672 | 3502 | 7058 | 28000 | 12000 | 9500 | 4500 | 7834 | 6500 | 7000 | 450 |
| 1988 | 15490 | 2790 | 9110 | 62923 | 25059 | 550367 | 9452 | 3679 | 5964 | 14583 | 8872 | 2818 | 3356 | 2682 | 1560 | 540 |
| 1989 | 7120 | 1930 | 25200 | 2890 | 3623 | 5650 | 324290 | 3469 | 800 | 679 | 3297 | 1375 | 679 | 321 | 260 | 0 |
| 1990 | 1020 | 400 | 15540 | 18633 | 2658 | 11875 | 10854 | 226280 | 1289 | 1519 | 2036 | 2415 | 646 | 179 | 590 | 480 |
| 1991 | 100 | 3370 | 3330 | 8438 | 2780 | 1410 | 14698 | 8867 | 218851 | 2499 | 461 | 87 | 690 | 103 | 260 | 540 |
| 1992 | 1630 | 150 | 1340 | 12586 | 33100 | 4980 | 1193 | 11981 | 5748 | 225677 | 2483 | 639 | 247 | 1236 | 0 | 0 |
| 1993 | 6570 | 130 | 7240 | 28408 | 106866 | 87269 | 8625 | 3648 | 29603 | 18631 | 410110 | 0 | 0 | 0 | 0 | 0 |
| 1994 | 430 | 20 | 8100 | 32500 | 110090 | 363920 | 164800 | 15580 | 8140 | 37330 | 35660 | 645410 | 2830 | 460 | 100 | 2070 |
| 1995 | 0 | 0 | 1130 | 57590 | 346460 | 622810 | 637840 | 231090 | 15510 | 15850 | 69750 | 83740 | 911880 | 4070 | 250 | 450 |
| 1996 | 0 | 0 | 30140 | 34360 | 713620 | 1571000 | 940580 | 406280 | 103410 | 5680 | 7370 | 66090 | 17570 | 836550 | 0 | 0 |
| 1997 | 0 | 0 | 21820 | 130450 | 270950 | 1795780 | 1993620 | 761210 | 326490 | 60870 | 20020 | 32400 | 90520 | 19120 | 370330 | 300 |
| 1998 | 0 | 0 | 82891 | 70323 | 242365 | 368310 | 1760319 | 1263750 | 381482 | 129971 | 42502 | 25343 | 3478 | 112604 | 5633 | 108514 |
| 1999 | 0 | 0 | 5029 | 137626 | 35820 | 134813 | 429433 | 1604959 | 1164263 | 291394 | 106005 | 14524 | 40040 | 7202 | 88598 | 63983 |
| 2000 | 0 | 0 | 14395 | 84016 | 560379 | 34933 | 110719 | 404460 | 1299253 | 1045001 | 216980 | 71589 | 16260 | 22701 | 23321 | 71811 |


| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 2001 | 0 | 0 | 2076 | 102293 | 160678 | 426822 | 38749 | 95991 | 296460 | 839136 | 507106 | 73673 | 23722 | 3505 | 3356 | 22164 |
| 2002 | 0 | 0 | 62031 | 198360 | 643161 | 255516 | 326495 | 29843 | 93530 | 264675 | 663059 | 339326 | 52922 | 12437 | 7000 | 10087 |
| 2003 | 0 | 3461 | 4524 | 75243 | 323958 | 730468 | 175878 | 167776 | 22866 | 74494 | 217108 | 567253 | 219097 | 38555 | 8111 | 6192 |
| 2004 | 125 | 1846 | 43800 | 24299 | 92300 | 429510 | 714433 | 111022 | 137940 | 26656 | 52467 | 169196 | 401564 | 210547 | 28028 | 11883 |
| 2005 | 0 | 442 | 20411 | 447788 | 94206 | 170547 | 643600 | 930309 | 121856 | 123291 | 37967 | 65289 | 139331 | 344822 | 126879 | 15697 |
| 2006 | 0 | 1968 | 45438 | 75824 | 729898 | 82107 | 171370 | 726041 | 772217 | 88701 | 77115 | 30339 | 57882 | 133665 | 142240 | 49128 |
| 2007 | 0 | 4475 | 8450 | 224636 | 366983 | 1804495 | 152916 | 242923 | 728836 | 511664 | 47215 | 25384 | 15316 | 24488 | 64755 | 58465 |
| 2008 | 0 | 39898 | 123949 | 36630 | 550274 | 670681 | 2295912 | 199592 | 256132 | 586583 | 369620 | 29633 | 36025 | 23775 | 25195 | 63176 |
| 2009 | 0 | 3468 | 113424 | 192641 | 149075 | 1193781 | 914748 | 1929631 | 142931 | 262037 | 423972 | 238174 | 45519 | 9337 | 10153 | 70538 |
| 2010 | 0 | 75981 | 61673 | 101948 | 209295 | 189784 | 1064866 | 711951 | 1421939 | 175010 | 180164 | 340781 | 179039 | 12558 | 11602 | 49773 |
| 2011 | 0 | 126972 | 249809 | 61706 | 104634 | 234330 | 210165 | 755382 | 543212 | 642787 | 90515 | 117230 | 136509 | 45082 | 6628 | 11638 |
| 2012 | 0 | 2680 | 13083 | 211630 | 49999 | 119627 | 281908 | 263330 | 747839 | 314694 | 357902 | 53109 | 44982 | 64273 | 12420 | 3604 |
| 2013 | 0 | 1 | 20715 | 60364 | 276901 | 71287 | 112558 | 283658 | 242243 | 591912 | 169525 | 145318 | 24936 | 10614 | 9725 | 2299 |
| 2014 | 0 | 265 | 1441 | 28301 | 57838 | 257529 | 50424 | 71721 | 194814 | 147083 | 381317 | 83050 | 57315 | 12746 | 1809 | 7501 |
| 2015 | 0 | 647 | 3244 | 16139 | 55749 | 52369 | 152347 | 34046 | 65728 | 156075 | 103393 | 201141 | 24310 | 49373 | 3369 | 6397 |
| 2016 | 0 | 197 | 2351 | 45483 | 43416 | 112147 | 85937 | 164454 | 52267 | 73576 | 174655 | 96476 | 179051 | 38546 | 32880 | 8379 |
| 2017 | 0 | 618 | 16390 | 64275 | 305483 | 114976 | 248192 | 162566 | 289931 | 98836 | 133145 | 276874 | 107473 | 220368 | 22357 | 49442 |


| AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 2018 | 0 | 1261 | 22414 | 25638 | 59802 | 264182 | 150759 | 179628 | 109121 | 180968 | 85954 | 99061 | 212052 | 113841 | 136096 | 39249 |
| 2019 | 0 | 769 | 2205 | 148669 | 64237 | 185336 | 557804 | 146597 | 217346 | 119855 | 167569 | 133910 | 104730 | 220400 | 91773 | 121229 |
| 2020 | 0 | 1299 | 8252 | 49455 | 544337 | 70633 | 150932 | 412498 | 118081 | 156696 | 94975 | 188852 | 100408 | 96557 | 132619 | 103350 |
| 2021 | 204 | 3644 | 2368 | 25015 | 110359 | 1432164 | 162903 | 203923 | 345729 | 117846 | 127846 | 73558 | 68834 | 60477 | 40165 | 113929 |
| 2022 | 0 | 8136 | 26087 | 22384 | 91375 | 148378 | 1365208 | 97514 | 203324 | 248904 | 72439 | 79165 | 63398 | 73962 | 38289 | 63347 |

## Table 4.4.4.1. Norwegian spring spawning herring. Weight at age in the catch (kg).

| Year | age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1950 | 0.007 | 0.025 | 0.058 | 0.110 | 0.188 | 0.211 | 0.234 | 0.253 | 0.266 | 0.280 | 0.294 | 0.303 | 0.312 | 0.32 | 0.323 | 0.334 |
| 1951 | 0.009 | 0.029 | 0.068 | 0.130 | 0.222 | 0.249 | 0.276 | 0.298 | 0.314 | 0.330 | 0.346 | 0.357 | 0.368 | 0.377 | 0.381 | 0.394 |
| 1952 | 0.008 | 0.026 | 0.061 | 0.115 | 0.197 | 0.221 | 0.245 | 0.265 | 0.279 | 0.293 | 0.308 | 0.317 | 0.327 | 0.335 | 0.339 | 0.349 |
| 1953 | 0.008 | 0.027 | 0.063 | 0.120 | 0.205 | 0.230 | 0.255 | 0.275 | 0.290 | 0.305 | 0.320 | 0.330 | 0.34 | 0.347 | 0.351 | 0.363 |
| 1954 | 0.008 | 0.026 | 0.062 | 0.117 | 0.201 | 0.225 | 0.250 | 0.269 | 0.284 | 0.299 | 0.313 | 0.323 | 0.333 | 0.341 | 0.345 | 0.356 |
| 1955 | 0.008 | 0.027 | 0.063 | 0.119 | 0.204 | 0.229 | 0.254 | 0.274 | 0.289 | 0.304 | 0.318 | 0.328 | 0.338 | 0.346 | 0.350 | 0.362 |
| 1956 | 0.008 | 0.028 | 0.066 | 0.126 | 0.215 | 0.241 | 0.268 | 0.289 | 0.304 | 0.320 | 0.336 | 0.346 | 0.357 | 0.365 | 0.369 | 0.382 |
| 1957 | 0.008 | 0.028 | 0.066 | 0.127 | 0.216 | 0.243 | 0.269 | 0.290 | 0.306 | 0.322 | 0.338 | 0.348 | 0.359 | 0.367 | 0.371 | 0.384 |
| 1958 | 0.009 | 0.030 | 0.070 | 0.133 | 0.227 | 0.255 | 0.283 | 0.305 | 0.321 | 0.338 | 0.355 | 0.366 | 0.377 | 0.386 | 0.390 | 0.403 |


|  | age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1959 | 0.009 | 0.030 | 0.071 | 0.135 | 0.231 | 0.259 | 0.287 | 0.310 | 0.327 | 0.344 | 0.360 | 0.372 | 0.383 | 0.392 | 0.397 | 0.409 |
| 1960 | 0.006 | 0.011 | 0.074 | 0.119 | 0.188 | 0.277 | 0.337 | 0.318 | 0.363 | 0.379 | 0.360 | 0.420 | 0.411 | 0.439 | 0.450 | 0.447 |
| 1961 | 0.006 | 0.010 | 0.045 | 0.087 | 0.159 | 0.276 | 0.322 | 0.372 | 0.363 | 0.393 | 0.407 | 0.397 | 0.422 | 0.447 | 0.465 | 0.452 |
| 1962 | 0.009 | 0.023 | 0.055 | 0.085 | 0.148 | 0.288 | 0.333 | 0.360 | 0.352 | 0.350 | 0.374 | 0.384 | 0.374 | 0.394 | 0.399 | 0.414 |
| 1963 | 0.008 | 0.026 | 0.047 | 0.098 | 0.171 | 0.275 | 0.268 | 0.323 | 0.329 | 0.336 | 0.341 | 0.358 | 0.385 | 0.353 | 0.381 | 0.386 |
| 1964 | 0.009 | 0.024 | 0.059 | 0.139 | 0.219 | 0.239 | 0.298 | 0.295 | 0.339 | 0.350 | 0.358 | 0.351 | 0.367 | 0.375 | 0.372 | 0.433 |
| 1965 | 0.009 | 0.016 | 0.048 | 0.089 | 0.217 | 0.234 | 0.262 | 0.331 | 0.360 | 0.367 | 0.386 | 0.395 | 0.393 | 0.404 | 0.401 | 0.431 |
| 1966 | 0.008 | 0.017 | 0.040 | 0.063 | 0.246 | 0.260 | 0.265 | 0.301 | 0.410 | 0.425 | 0.456 | 0.460 | 0.467 | 0.446 | 0.459 | 0.472 |
| 1967 | 0.009 | 0.015 | 0.036 | 0.066 | 0.093 | 0.305 | 0.305 | 0.310 | 0.333 | 0.359 | 0.413 | 0.446 | 0.401 | 0.408 | 0.439 | 0.430 |
| 1968 | 0.010 | 0.027 | 0.049 | 0.075 | 0.108 | 0.158 | 0.375 | 0.383 | 0.364 | 0.382 | 0.441 | 0.410 |  | 0.517 | 0.491 | 0.485 |
| 1969 | 0.009 | 0.021 | 0.047 | 0.072 |  | 0.152 | 0.296 |  | 0.329 | 0.329 | 0.341 |  |  |  |  | 0.429 |
| 1970 | 0.008 | 0.058 | 0.085 | 0.105 | 0.171 |  | 0.216 | 0.277 | 0.298 | 0.304 | 0.305 | 0.309 |  |  |  | 0.376 |
| 1971 | 0.011 | 0.053 | 0.121 | 0.177 | 0.216 | 0.250 |  | 0.305 | 0.333 |  | 0.366 | 0.377 | 0.388 |  |  |  |
| 1972 | 0.011 | 0.029 | 0.062 | 0.103 | 0.154 | 0.215 | 0.258 |  | 0.322 |  |  |  |  |  |  |  |
| 1973 | 0.006 | 0.053 | 0.106 | 0.161 | 0.213 |  | 0.255 |  |  |  |  |  |  |  |  |  |
| 1974 | 0.006 | 0.055 | 0.117 |  |  | 0.249 |  |  |  |  |  |  |  |  |  |  |
| 1975 | 0.009 | 0.079 | 0.169 | 0.241 |  |  | 0.381 |  |  |  |  |  |  |  |  |  |


|  | age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1976 | 0.007 | 0.062 | 0.132 | 0.189 | 0.250 |  |  | 0.323 |  |  |  |  |  |  |  |  |
| 1977 | 0.011 | 0.091 | 0.193 | 0.316 | 0.350 |  |  |  | 0.511 |  |  |  |  |  |  |  |
| 1978 | 0.012 | 0.100 | 0.210 | 0.274 | 0.424 | 0.454 |  |  |  | 0.613 |  |  |  |  |  |  |
| 1979 | 0.010 | 0.088 | 0.181 | 0.293 | 0.359 | 0.416 | 0.436 |  |  |  | 0.553 |  |  |  |  |  |
| 1980 | 0.012 |  |  | 0.266 | 0.399 | 0.449 | 0.460 | 0.485 |  |  |  | 0.608 |  |  |  |  |
| 1981 | 0.010 | 0.082 | 0.163 | 0.196 | 0.291 | 0.341 | 0.368 | 0.380 | 0.397 |  |  |  |  |  |  |  |
| 1982 | 0.010 | 0.087 | 0.159 | 0.256 | 0.312 | 0.378 | 0.415 | 0.435 | 0.449 | 0.448 |  |  |  |  |  |  |
| 1983 | 0.011 | 0.090 | 0.165 | 0.217 | 0.265 | 0.337 | 0.378 | 0.410 | 0.426 | 0.435 | 0.444 |  |  |  |  |  |
| 1984 | 0.009 | 0.047 | 0.145 | 0.218 | 0.262 | 0.325 | 0.346 | 0.381 | 0.400 | 0.413 | 0.405 | 0.426 |  |  |  | 0.415 |
| 1985 | 0.009 | 0.022 | 0.022 | 0.214 | 0.277 | 0.295 | 0.338 | 0.360 | 0.381 | 0.397 | 0.409 | 0.417 | 0.435 |  |  | 0.435 |
| 1986 | 0.007 | 0.077 | 0.097 | 0.055 | 0.249 | 0.294 | 0.312 | 0.352 | 0.374 | 0.398 | 0.402 | 0.401 | 0.410 | 0.410 |  | 0.410 |
| 1987 | 0.010 | 0.075 | 0.091 | 0.124 | 0.173 | 0.253 | 0.232 | 0.312 | 0.328 | 0.349 | 0.353 | 0.370 | 0.385 | 0.385 | 0.385 |  |
| 1988 | 0.008 | 0.062 | 0.075 | 0.124 | 0.154 | 0.194 | 0.241 | 0.265 | 0.304 | 0.305 | 0.317 | 0.308 | 0.334 | 0.334 | 0.334 |  |
| 1989 | 0.010 | 0.060 | 0.204 | 0.188 | 0.264 | 0.260 | 0.282 | 0.306 |  |  | 0.422 | 0.364 |  |  |  |  |
| 1990 | 0.007 |  | 0.102 | 0.230 | 0.239 | 0.266 | 0.305 | 0.308 | 0.376 | 0.407 | 0.412 | 0.424 |  |  |  |  |
| 1991 |  | 0.015 | 0.104 | 0.208 | 0.250 | 0.288 | 0.312 | 0.316 | 0.330 | 0.344 |  |  |  |  |  |  |
| 1992 | 0.007 |  | 0.103 | 0.191 | 0.233 | 0.304 | 0.337 | 0.365 | 0.361 | 0.371 | 0.403 |  |  | 0.404 |  |  |


| Year | age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 1993 | 0.007 |  | 0.106 | 0.153 | 0.243 | 0.282 | 0.320 | 0.330 | 0.365 | 0.373 | 0.379 |  |  |  |  |  |
| 1994 |  |  | 0.102 | 0.194 | 0.239 | 0.280 | 0.317 | 0.328 | 0.356 | 0.372 | 0.390 | 0.379 | 0.399 | 0.403 |  |  |
| 1995 |  |  | 0.102 | 0.153 | 0.192 | 0.234 | 0.283 | 0.328 | 0.349 | 0.356 | 0.374 | 0.366 | 0.393 | 0.387 |  |  |
| 1996 |  |  | 0.136 | 0.136 | 0.168 | 0.206 | 0.262 | 0.309 | 0.337 | 0.366 | 0.360 | 0.361 | 0.367 | 0.379 |  |  |
| 1997 |  |  | 0.089 | 0.167 | 0.184 | 0.207 | 0.232 | 0.277 | 0.305 | 0.331 | 0.328 | 0.344 | 0.343 | 0.397 | 0.357 |  |
| 1998 |  |  | 0.111 | 0.150 | 0.216 | 0.221 | 0.249 | 0.277 | 0.316 | 0.338 | 0.374 | 0.372 | 0.366 | 0.396 | 0.377 | 0.406 |
| 1999 |  |  | 0.096 | 0.173 | 0.228 | 0.262 | 0.274 | 0.292 | 0.307 | 0.335 | 0.362 | 0.371 | 0.399 | 0.396 | 0.400 | 0.404 |
| 2000 |  |  | 0.124 | 0.175 | 0.222 | 0.242 | 0.289 | 0.303 | 0.310 | 0.328 | 0.349 | 0.383 | 0.411 | 0.410 | 0.419 | 0.409 |
| 2001 |  |  | 0.105 | 0.166 | 0.214 | 0.252 | 0.268 | 0.305 | 0.308 | 0.322 | 0.337 | 0.363 | 0.353 | 0.378 | 0.400 | 0.427 |
| 2002 |  |  | 0.056 | 0.128 | 0.198 | 0.255 | 0.281 | 0.303 | 0.322 | 0.323 | 0.334 | 0.345 | 0.369 | 0.407 | 0.410 | 0.435 |
| 2003 |  | 0.062 | 0.068 | 0.169 | 0.218 | 0.257 | 0.288 | 0.316 | 0.323 | 0.348 | 0.354 | 0.351 | 0.363 | 0.372 | 0.376 | 0.429 |
| 2004 | 0.022 | 0.066 | 0.143 | 0.18 | 0.227 | 0.26 | 0.29 | 0.323 | 0.355 | 0.375 | 0.383 | 0.399 | 0.395 | 0.405 | 0.429 | 0.439 |
| 2005 |  | 0.092 | 0.106 | 0.181 | 0.235 | 0.266 | 0.290 | 0.315 | 0.344 | 0.367 | 0.384 | 0.372 | 0.384 | 0.398 | 0.402 | 0.413 |
| 2006 |  | 0.055 | 0.102 | 0.171 | 0.238 | 0.268 | 0.292 | 0.311 | 0.330 | 0.365 | 0.374 | 0.376 | 0.388 | 0.396 | 0.398 | 0.407 |
| 2007 | 0.000 | 0.074 | 0.137 | 0.162 | 0.228 | 0.271 | 0.316 | 0.332 | 0.342 | 0.358 | 0.361 | 0.381 | 0.390 | 0.400 | 0.405 | 0.399 |
| 2008 | 0.000 | 0.026 | 0.106 | 0.145 | 0.209 | 0.254 | 0.296 | 0.318 | 0.341 | 0.353 | 0.363 | 0.367 | 0.395 | 0.396 | 0.386 | 0.413 |
| 2009 |  | 0.040 | 0.156 | 0.184 | 0.220 | 0.251 | 0.291 | 0.311 | 0.338 | 0.347 | 0.363 | 0.375 | 0.382 | 0.375 | 0.375 | 0.387 |


| Year | age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| 2010 |  | 0.059 | 0.107 | 0.177 | 0.218 | 0.261 | 0.279 | 0.311 | 0.325 | 0.343 | 0.362 | 0.370 | 0.388 | 0.391 | 0.376 | 0.441 |
| 2011 |  | 0.011 | 0.098 | 0.200 | 0.257 | 0.273 | 0.300 | 0.316 | 0.340 | 0.348 | 0.365 | 0.371 | 0.387 | 0.374 | 0.403 | 0.401 |
| 2012 |  | 0.034 | 0.126 | 0.211 | 0.272 | 0.301 | 0.308 | 0.331 | 0.335 | 0.351 | 0.354 | 0.370 | 0.389 | 0.389 | 0.382 | 0.388 |
| 2013 |  | 0.048 | 0.163 | 0.237 | 0.276 | 0.300 | 0.331 | 0.339 | 0.351 | 0.357 | 0.370 | 0.373 | 0.394 | 0.391 | 0.389 | 0.367 |
| 2014 |  | 0.057 | 0.179 | 0.233 | 0.271 | 0.293 | 0.322 | 0.342 | 0.353 | 0.367 | 0.365 | 0.374 | 0.375 | 0.378 | 0.418 | 0.371 |
| 2015 |  | 0.059 | 0.146 | 0.203 | 0.272 | 0.323 | 0.331 | 0.358 | 0.370 | 0.372 | 0.383 | 0.382 | 0.392 | 0.386 | 0.383 | 0.391 |
| 2016 |  | 0.048 | 0.111 | 0.212 | 0.255 | 0.290 | 0.333 | 0.339 | 0.361 | 0.367 | 0.370 | 0.381 | 0.378 | 0.388 | 0.383 | 0.395 |
| 2017 |  | 0.092 | 0.143 | 0.205 | 0.241 | 0.292 | 0.322 | 0.350 | 0.360 | 0.382 | 0.392 | 0.391 | 0.396 | 0.399 | 0.407 | 0.394 |
| 2018 |  | 0.068 | 0.127 | 0.207 | 0.240 | 0.276 | 0.321 | 0.348 | 0.371 | 0.380 | 0.399 | 0.404 | 0.400 | 0.407 | 0.408 | 0.418 |
| 2019 |  | 0.135 | 0.186 | 0.209 | 0.235 | 0.269 | 0.298 | 0.327 | 0.345 | 0.376 | 0.387 | 0.403 | 0.409 | 0.423 | 0.417 | 0.449 |
| 2020 |  | 0.131 | 0.170 | 0.204 | 0.236 | 0.274 | 0.306 | 0.317 | 0.342 | 0.358 | 0.374 | 0.395 | 0.402 | 0.408 | 0.415 | 0.444 |
| 2021 | 0.050 | 0.122 | 0.130 | 0.195 | 0.229 | 0.256 | 0.278 | 0.319 | 0.325 | 0.363 | 0.364 | 0.384 | 0.386 | 0.397 | 0.412 | 0.431 |
| 2022 |  | 0.086 | 0.161 | 0.224 | 0.231 | 0.268 | 0.287 | 0.316 | 0.345 | 0.356 | 0.382 | 0.397 | 0.417 | 0.423 | 0.431 | 0.468 |

Table 4.4.4.2. Norwegian spring spawning herring. Weight at age in the stock (kg).

| Year | AGE | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |



| Year | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |  | 15+ |
| 1984 |  | 0.001 | 0.010 | 0.085 | 0.140 | 0.204 | 0.295 | 0.338 | 0.376 | 0.395 | 0.407 | 0.413 | 0.422 | 0.437 | 0.437 | 0.437 | 0.437 |
| 1985 |  | 0.001 | 0.010 | 0.085 | 0.148 | 0.234 | 0.265 | 0.312 | 0.346 | 0.370 | 0.395 | 0.397 | 0.428 | 0.428 | 0.428 | 0.428 | 0.428 |
| 1986 |  | 0.001 | 0.010 | 0.085 | 0.054 | 0.206 | 0.265 | 0.289 | 0.339 | 0.368 | 0.391 | 0.382 | 0.388 | 0.395 | 0.395 | 0.395 | 0.395 |
| 1987 |  | 0.001 | 0.010 | 0.055 | 0.090 | 0.143 | 0.241 | 0.279 | 0.299 | 0.316 | 0.342 | 0.343 | 0.362 | 0.376 | 0.376 | 0.376 | 0.376 |
| 1988 |  | 0.001 | 0.015 | 0.050 | 0.098 | 0.135 | 0.197 | 0.277 | 0.315 | 0.339 | 0.343 | 0.359 | 0.365 | 0.376 | 0.376 | 0.376 | 0.376 |
| 1989 |  | 0.001 | 0.015 | 0.100 | 0.154 | 0.175 | 0.209 | 0.252 | 0.305 | 0.367 | 0.377 | 0.359 | 0.395 | 0.396 | 0.396 | 0.396 | 0.396 |
| 1990 |  | 0.001 | 0.008 | 0.048 | 0.219 | 0.198 | 0.258 | 0.288 | 0.309 | 0.428 | 0.370 | 0.403 | 0.387 | 0.440 | 0.440 | 0.440 | 0.44 |
| 1991 |  | 0.001 | 0.011 | 0.037 | 0.147 | 0.210 | 0.244 | 0.300 | 0.324 | 0.336 | 0.343 | 0.382 | 0.366 | 0.425 | 0.425 | 0.425 | 0.425 |
| 1992 |  | 0.001 | 0.007 | 0.030 | 0.128 | 0.224 | 0.296 | 0.327 | 0.355 | 0.345 | 0.367 | 0.341 | 0.361 | 0.430 | 0.470 | 0.470 | 0.46 |
| 1993 |  | 0.001 | 0.008 | 0.025 | 0.081 | 0.201 | 0.265 | 0.323 | 0.354 | 0.358 | 0.381 | 0.369 | 0.396 | 0.393 | 0.374 | 0.403 | 0.4 |
| 1994 |  | 0.001 | 0.010 | 0.025 | 0.075 | 0.151 | 0.254 | 0.318 | 0.371 | 0.347 | 0.412 | 0.382 | 0.407 | 0.410 | 0.410 | 0.410 | 0.41 |
| 1995 |  | 0.001 | 0.018 | 0.025 | 0.066 | 0.138 | 0.230 | 0.296 | 0.346 | 0.388 | 0.363 | 0.409 | 0.414 | 0.422 | 0.410 | 0.410 | 0.426 |
| 1996 |  | 0.001 | 0.018 | 0.025 | 0.076 | 0.118 | 0.188 | 0.261 | 0.316 | 0.346 | 0.374 | 0.390 | 0.390 | 0.384 | 0.398 | 0.398 | 0.398 |
| 1997 |  | 0.001 | 0.018 | 0.025 | 0.096 | 0.118 | 0.174 | 0.229 | 0.286 | 0.323 | 0.370 | 0.378 | 0.386 | 0.360 | 0.393 | 0.391 | 0.391 |
| 1998 |  | 0.001 | 0.018 | 0.025 | 0.074 | 0.147 | 0.174 | 0.217 | 0.242 | 0.278 | 0.304 | 0.310 | 0.359 | 0.340 | 0.344 | 0.385 | 0.369 |
| 1999 |  | 0.001 | 0.018 | 0.025 | 0.102 | 0.150 | 0.223 | 0.240 | 0.264 | 0.283 | 0.315 | 0.345 | 0.386 | 0.386 | 0.386 | 0.382 | 0.395 |
| 2000 |  | 0.001 | 0.018 | 0.025 | 0.119 | 0.178 | 0.225 | 0.271 | 0.285 | 0.298 | 0.311 | 0.339 | 0.390 | 0.398 | 0.406 | 0.414 | 0.427 |


| Year | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $0 \quad 1$ | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |  | 15+ |
| 2001 | 0.001 | 0.018 | 0.025 | 0.075 | 0.178 | 0.238 | 0.247 | 0.296 | 0.307 | 0.314 | 0.328 | 0.351 | 0.376 | 0.406 | 0.414 | 0.425 |
| 2002 | 0.001 | 0.010 | 0.023 | 0.057 | 0.177 | 0.241 | 0.275 | 0.302 | 0.311 | 0.314 | 0.328 | 0.341 | 0.372 | 0.405 | 0.415 | 0.438 |
| 2003 | 0.001 | 0.010 | 0.055 | 0.098 | 0.159 | 0.211 | 0.272 | 0.305 | 0.292 | 0.331 | 0.337 | 0.347 | 0.356 | 0.381 | 0.414 | 0.433 |
| 2004 | 0.001 | 0.010 | 0.055 | 0.106 | 0.149 | 0.212 | 0.241 | 0.279 | 0.302 | 0.337 | 0.354 | 0.355 | 0.360 | 0.371 | 0.400 | 0.429 |
| 2005 | 0.001 | 0.010 | 0.046 | 0.112 | 0.156 | 0.234 | 0.267 | 0.295 | 0.330 | 0.363 | 0.377 | 0.414 | 0.406 | 0.308 | 0.420 | 0.452 |
| 2006 | 0.001 | 0.010 | 0.042 | 0.107 | 0.179 | 0.232 | 0.272 | 0.297 | 0.318 | 0.371 | 0.365 | 0.393 | 0.395 | 0.399 | 0.415 | 0.428 |
| 2007 | 0.001 | 0.010 | 0.036 | 0.086 | 0.155 | 0.226 | 0.265 | 0.312 | 0.310 | 0.364 | 0.384 | 0.352 | 0.386 | 0.304 | 0.420 | 0.412 |
| 2008** | 0.001 | 0.010 | 0.044 | 0.077 | 0.146 | 0.212 | 0.269 | 0.289 | 0.327 | 0.351 | 0.358 | 0.372 | 0.411 | 0.353 | 0.389 | 0.393 |
| 2009*** | 0.001 | 0.010 | 0.044 | 0.077 | 0.141 | 0.215 | 0.270 | 0.306 | 0.336 | 0.346 | 0.364 | 0.369 | 0.411 | 0.353 | 0.389 | 0.393 |
| 2010**** | 0.001 | 0.01 | 0.044 | 0.077 | 0.188 | 0.22 | 0.251 | 0.286 | 0.308 | 0.333 | 0.344 | 0.354 | 0.373 | 0.353 | 0.389 | 0.393 |
| 2011 | 0.001 | 0.01 | 0.044 | 0.118 | 0.185 | 0.209 | 0.246 | 0.277 | 0.310 | 0.322 | 0.339 | 0.349 | 0.364 | 0.363 | 0.389 | 0.393 |
| 2012 | 0.001 | 0.01 | 0.044 | 0.138 | 0.185 | 0.256 | 0.273 | 0.290 | 0.305 | 0.330 | 0.342 | 0.361 | 0.390 | 0.377 | 0.389 | 0.393 |
| 2013 | 0.001 | 0.01 | 0.044 | 0.138 | 0.204 | 0.267 | 0.305 | 0.309 | 0.320 | 0.328 | 0.346 | 0.350 | 0.390 | 0.377 | 0.389 | 0.393 |
| 2014 | 0.001 | 0.01 | 0.044 | 0.138 | 0.198 | 0.274 | 0.301 | 0.326 | 0.333 | 0.339 | 0.347 | 0.344 | 0.362 | 0.362 | 0.389 | 0.393 |
| 2015 | 0.001 | 0.01 | 0.044 | 0.138 | 0.187 | 0.243 | 0.299 | 0.326 | 0.319 | 0.345 | 0.346 | 0.354 | 0.382 | 0.376 | 0.389 | 0.393 |
| 2016 | 0.001 | 0.01 | 0.054 | 0.115 | 0.186 | 0.247 | 0.293 | 0.320 | 0.334 | 0.353 | 0.354 | 0.352 | 0.361 | 0.370 | 0.380 | 0.388 |
| 2017 | 0.001 | 0.01 | 0.054 | 0.115 | 0.190 | 0.247 | 0.282 | 0.322 | 0.338 | 0.351 | 0.359 | 0.361 | 0.361 | 0.368 | 0.380 | 0.386 |


| Year | AGE |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |  | 15+ |
| 2018 |  | 0.001 | 0.01 | 0.054 | 0.115 | 0.149 | 0.225 | 0.260 | 0.289 | 0.312 | 0.343 | 0.359 | 0.361 | 0.369 | 0.368 | 0.377 | 0.386 |
| 2019 |  | 0.001 | 0.01 | 0.054 | 0.104 | 0.151 | 0.203 | 0.277 | 0.311 | 0.331 | 0.355 | 0.353 | 0.363 | 0.381 | 0.376 | 0.385 | 0.382 |
| 2020 |  | 0.001 | 0.01 | 0.054 | 0.104 | 0.150 | 0.203 | 0.266 | 0.301 | 0.328 | 0.343 | 0.358 | 0.366 | 0.374 | 0.367 | 0.384 | 0.391 |
| 2021 |  | 0.001 | 0.01 | 0.054 | 0.104 | 0.160 | 0.209 | 0.266 | 0.284 | 0.302 | 0.325 | 0.352 | 0.366 | 0.384 | 0.376 | 0.404 | 0.391 |
| 2022 |  | 0.001 | 0.01 | 0.054 | 0.104 | 0.125 | 0.168 | 0.243 | 0.287 | 0.303 | 0.323 | 0.352 | 0.366 | 0.384 | 0.376 | 0.404 | 0.391 |
| 2023 |  | 0.001 | 0.01 | 0.054 | 0.118 | 0.125 | 0.202 | 0.255 | 0.280 | 0.311 | 0.336 | 0.337 | 0.345 | 0.384 | 0.376 | 0.404 | 0.391 |

** mean weight at ages 11 and 13 are mean of 5 previous years at the same age. These age groups were not present in the catches of the wintering survey from which the stock weight are derived.
${ }^{* * *}$ derived from catch data from the wintering area north of $69^{\circ} \mathrm{N}$ during December 2008 - January 2009 for age groups 4-11.
${ }^{* * * *}$ derived from catch data from the wintering area north of $69^{\circ} \mathrm{N}$ during January 2010 for age groups 4-12.

Table 4.4.5.1. Norwegian Spring-spawning herring. Maturity at age.

| Year/Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1950 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1951 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1952 | 0 | 0 | 0 | 0 | 0.1 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1953 | 0 | 0 | 0 | 0 | 0.3 | 0.4 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1954 | 0 | 0 | 0 | 0 | 0.1 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1955 | 0 | 0 | 0 | 0.1 | 0.4 | 0.4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1956 | 0 | 0 | 0 | 0 | 0.5 | 0.7 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1957 | 0 | 0 | 0 | 0 | 0.3 | 0.8 | 0.8 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1958 | 0 | 0 | 0 | 0 | 0.3 | 0.5 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1959 | 0 | 0 | 0 | 0 | 0.7 | 0.8 | 1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1960 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1961 | 0 | 0 | 0 | 0 | 0.1 | 0.8 | 1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1962 | 0 | 0 | 0 | 0 | 0.1 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1963 | 0 | 0 | 0 | 0 | 0.1 | 0.4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1964 | 0 | 0 | 0 | 0 | 0.1 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1965 | 0 | 0 | 0 | 0 | 0.5 | 0.4 | 0.9 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1966 | 0 | 0 | 0 | 0 | 0.5 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1967 | 0 | 0 | 0 | 0 | 0.3 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1968 | 0 | 0 | 0 | 0 | 0 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1969 | 0 | 0 | 0 | 0.1 | 0.2 | 0.3 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1970 | 0 | 0 | 0 | 0 | 0.4 | 0.3 | 0.4 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1971 | 0 | 0 | 0 | 0 | 0.1 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1972 | 0 | 0 | 0 | 0 | 0.4 | 0.3 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1973 | 0 | 0 | 0 | 0.1 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1974 | 0 | 0 | 0 | 0 | 0.6 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1975 | 0 | 0 | 0 | 0.1 | 0.5 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1976 | 0 | 0 | 0 | 0.1 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1977 | 0 | 0 | 0 | 0.3 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |


| Year/Age | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 0 | 0 | 0 | 0.2 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1979 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1980 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1981 | 0 | 0 | 0 | 0.1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1982 | 0 | 0 | 0 | 0.1 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1983 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1984 | 0 | 0 | 0 | 0.1 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1985 | 0 | 0 | 0 | 0.1 | 0.8 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1986 | 0 | 0 | 0 | 0 | 0.5 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1987 | 0 | 0 | 0 | 0 | 0.1 | 0.8 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1988 | 0 | 0 | 0 | 0 | 0.2 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1989 | 0 | 0 | 0 | 0 | 0.4 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1990 | 0 | 0 | 0 | 0.2 | 0.5 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1991 | 0 | 0 | 0 | 0 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1992 | 0 | 0 | 0 | 0 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1993 | 0 | 0 | 0 | 0 | 0.5 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1994 | 0 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0.6 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0.5 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1997 | 0 | 0 | 0 | 0.1 | 0 | 0.4 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1998 | 0 | 0 | 0 | 0 | 0.6 | 0.4 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 1999 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2000 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2001 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2002 | 0 | 0 | 0 | 0 | 0.1 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2003 | 0 | 0 | 0 | 0 | 0.2 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2004 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2005 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2006 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |


| Year/Age | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | 13 | 14 | $15+$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2007 | 0 | 0 | 0 | 0 | 0.2 | 0.8 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2008 | 0 | 0 | 0 | 0 | 0.1 | 0.7 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2009 | 0 | 0 | 0 | 0 | 0.1 | 0.4 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2010 | 0 | 0 | 0 | 0 | 0.2 | 0.4 | 0.7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2013 | 0 | 0 | 0 | 0 | 0.4 | 0.7 | 0.8 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2014 | 0 | 0 | 0 | 0 | 0.5 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 20020 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2020 | 0 | 0 | 0 | 0 | 0.3 | 0.9 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| 2020 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 4.4.7.1. Norwegian Spring-spawning herring. Estimated indices (mean of bootstrap with 1000 iterations in StoX) from the acoustic surveys on the spawning grounds in February-March. Numbers in millions. Biomass in thousand tonnes. "Fleet 1".

| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | Total | Biomass |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0 | 392 | 307 | 8015 | 81 | 33 | 12 | 36 | 22 | 45 | 0 | 0 | 0 | 0 | 8943 | 1621 |
| 1989 | 161 | 16 | 338 | 91 | 3973 | 101 | 12 | 4 | 55 | 0 | 4 | 42 | 0 | 9 | 4813 | 1169 |
| 1990 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1991 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1992 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1993 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1994 | 37 | 100 | 48 | 848 | 483 | 62 | 13 | 144 | 49 | 1836 | 4 | 4 | 0 | 0 | 3665 | 1207 |
| 1995 | 4 | 450 | 4679 | 3211 | 1957 | 299 | 20 | 0 | 106 | 55 | 2327 | 0 | 0 | 0 | 13745 | 2860 |
| 1996 | 119 | 186 | 1976 | 7960 | 2326 | 875 | 301 | 0 | 0 | 136 | 0 | 1760 | 0 | 0 | 15645 | 3366 |
| 1997 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1998 | 51 | 308 | 978 | 2982 | 12859 | 8133 | 1851 | 592 | 163 | 43 | 0 | 329 | 0 | 1400 | 29705 | 6886 |
| 1999 | 114 | 1530 | 369 | 1351 | 2669 | 9334 | 7004 | 1666 | 511 | 130 | 0 | 0 | 353 | 373 | 25438 | 6262 |
| 2000 | 1394 | 691 | 2600 | 109 | 477 | 1144 | 4282 | 2838 | 493 | 50 | 2 | 0 | 7 | 228 | 14315 | 3285 |
| 2001 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2002 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2003 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2004 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Year | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | Total | Biomass |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 38 | 238 | 661 | 2128 | 5947 | 8328 | 613 | 503 | 156 | 92 | 576 | 1152 | 587 | 9 | 21026 | 5260 |
| 2006 | 26 | 90 | 6054 | 548 | 882 | 3362 | 3311 | 110 | 86 | 20 | 89 | 58 | 246 | 63 | 14951 | 3431 |
| 2007 | 33 | 367 | 1618 | 12397 | 815 | 655 | 2956 | 3205 | 141 | 228 | 40 | 204 | 284 | 470 | 23427 | 5350 |
| 2008 | 15 | 48 | 2564 | 2824 | 8882 | 522 | 471 | 1566 | 1567 | 161 | 102 | 46 | 128 | 136 | 19090 | 4553 |
| 2009 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2010 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2011 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2012 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2013 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2014 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 2015 | 204 | 533 | 2754 | 744 | 3267 | 388 | 692 | 2715 | 784 | 7222 | 367 | 1658 | 51 | 237 | 21662 | 6365 |
| 2016 | 18 | 197 | 237 | 594 | 365 | 2119 | 240 | 514 | 2930 | 652 | 3995 | 199 | 824 | 97 | 12982 | 4182 |
| 2017 | 19 | 110 | 1076 | 641 | 880 | 428 | 1326 | 181 | 206 | 2026 | 303 | 2542 | 80 | 729 | 10550 | 3314 |
| 2018 | 104 | 146 | 1720 | 2771 | 459 | 845 | 639 | 1095 | 444 | 370 | 1159 | 368 | 1538 | 354 | 12013 | 3262 |
| 2019 | 2 | 372 | 310 | 940 | 3778 | 754 | 879 | 660 | 1054 | 736 | 412 | 1807 | 182 | 2161 | 14166 | 4250 |
| 2020 | 6 | 44 | 3502 | 571 | 1212 | 3337 | 530 | 609 | 364 | 650 | 131 | 279 | 677 | 825 | 12750 | 3274 |
| 2021 | 21 | 112 | 293 | 10210 | 733 | 738 | 1932 | 427 | 451 | 312 | 219 | 395 | 208 | 1153 | 17250 | 4021 |
| 2022 | 27 | 72 | 162 | 760 | 6393 | 317 | 563 | 1515 | 301 | 486 | 301 | 255 | 385 | 630 | 12183 | 3302 |


| Year | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5 +}$ | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2023 | 0 | 205 | 189 | 568 | 594 | 6014 | 351 | 1057 | 1349 | 304 | 537 | 508 | 538 | 807 | $\mathbf{1 3 0}$ |

Table 4.4.7.2. Norwegian spring-spawning herring. Acoustic estimates (billion individuals) of immature herring in the Barents Sea in May/June from IESNS. Values in the years 2009-2022 are estimated with StoX (mean of bootstrap with 1000 iterations). "Fleet 4".

| age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 |
| 1991 | 24.3 | 5.2 |  |  |  |
| 1992 | 32.6 | 14 | 5.7 |  |  |
| 1993 | 102.7 | 25.8 | 1.5 |  |  |
| 1994 | 6.6 | 59.2 | 18 | 1.7 |  |
| 1995 | 0.5 | 7.7 | 8 | 1.1 |  |
| 1996* | 0.1 | 0.25 | 1.8 | 0.6 | 0.03 |
| 1997** | 2.6 | 0.04 | 0.4 | 0.35 | 0.05 |
| 1998 | 9.5 | 4.7 | 0.01 | 0.01 | 0 |
| 1999 | 49.5 | 4.9 | 0 | 0 | 0 |
| 2000 | 105.4 | 27.9 | 0 | 0 | 0 |
| 2001 | 0.3 | 7.6 | 8.8 | 0 | 0 |
| 2002 | 0.5 | 3.9 | 0 | 0 | 0 |
| 2003*** |  |  |  |  |  |
| 2004*** |  |  |  |  |  |
| 2005 | 23.3 | 4.5 | 2.5 | 0.4 | 0.3 |
| 2006 | 3.7 | 35.0 | 5.3 | 0.87 | 0 |
| 2007 | 2.1 | 3.7 | 12.5 | 1.9 | 0 |
| 2008^ |  |  |  |  |  |
| 2009 | 0.289 | 0.300 | 0.233 | 0.060 |  |
| 2010 | 5.196 | 1.380 | 0.000 | 0.000 |  |
| 2011 | 1.166 | 3.920 | 0.041 | 0.000 |  |
| 2012 | 0.787 | 0.030 | 0.000 | 0.000 |  |
| 2013 | 0.107 | 2.190 | 0.211 | 0.070 |  |
| 2014 | 4.239 | 3.110 | 1.728 | 0.127 | 0.043 |
| 2015 | 0.345 | 11.760 | 1.183 | 0.206 | 0.000 |
| 2016 | 1.826 | 5.620 | 1.568 | 0.101 | 0.038 |
| 2017 | 14.522 | 3.080 | 0.000 | 0.000 |  |


| age |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 |
| 2018 | 7.329 | 17.420 | 0.827 | 0.009 |  |
| 2019 | 0.113 | 2.370 | 17.481 | 0.044 |  |
| 2020*** |  |  |  |  |  |
| 2021 | 0.021 | 0.002 | 0.086 | 0.002 |  |
| 2022*** |  |  |  |  |  |
| 2023*** |  |  |  |  |  |

*Average of Norwegian and Russian estimates
**Combination of Norwegian and Russian estimates as described in 1998 WG report, since then only Russian estimates
***No surveys
${ }^{\wedge}$ Not a full survey

Table 4.4.7.3. Norwegian spring-spawning herring. Estimates from the international acoustic survey on the feeding areas in the Norwegian Sea in May (IESNS). Numbers in millions. Biomass in thousands. Values in the years 2008-2022 are estimated indices by StoX (mean of bootstrap with 1000 iterations). "Fleet 5".

|  | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | Total | Biomass |
| 1996 | 0 | 0 | 4114 | 22461 | 13244 | 4916 | 2045 | 424 | 14 | 7 | 155 | 0 | 3134 |  |  | 50514 | 8532 |
| 1997 | 0 | 0 | 1169 | 3599 | 18867 | 13546 | 2473 | 1771 | 178 | 77 | 288 | 190 | 60 | 2697 |  | 44915 | 9435 |
| 1998 | 24 | 1404 | 367 | 1099 | 4410 | 16378 | 10160 | 2059 | 804 | 183 | 0 | 0 | 35 | 0 | 492 | 37415 | 8004 |
| 1999 | 0 | 215 | 2191 | 322 | 965 | 3067 | 11763 | 6077 | 853 | 258 | 5 | 14 | 0 | 158 | 128 | 26016 | 6299 |
| 2000 | 0 | 157 | 1353 | 2783 | 92 | 384 | 1302 | 7194 | 5344 | 1689 | 271 | 0 | 114 | 0 | 75 | 20758 | 6001 |
| 2001 | 0 | 1540 | 8312 | 1430 | 1463 | 179 | 204 | 3215 | 5433 | 1220 | 94 | 178 | 0 | 0 | 6 | 23274 | 3937 |
| 2002 | 0 | 677 | 6343 | 9619 | 1418 | 779 | 375 | 847 | 1941 | 2500 | 1423 | 61 | 78 | 28 | 0 | 26089 | 4628 |
| 2003 | 32073 | 8115 | 6561 | 9985 | 9961 | 1499 | 732 | 146 | 228 | 1865 | 2359 | 1769 |  | 287 | 0 | 75580 | 6653 |
| 2004 | 0 | 13735 | 1543 | 5227 | 12571 | 10710 | 1075 | 580 | 76 | 313 | 362 | 1294 | 1120 | 10 | 88 | 48704 | 7687 |
| 2005 | 0 | 1293 | 19679 | 1353 | 1765 | 6205 | 5371 | 651 | 388 | 139 | 262 | 526 | 1003 | 364 | 115 | 39114 | 5109 |
| 2006 | 0 | 19 | 306 | 14560 | 1396 | 2011 | 6521 | 6978 | 679 | 713 | 173 | 407 | 921 | 618 | 243 | 35545 | 9100 |
| 2007 | 0 | 411 | 2889 | 5877 | 20292 | 1260 | 1992 | 6780 | 5582 | 647 | 488 | 372 | 403 | 1048 | 1010 | 49051 | 12161 |
| 2008 | 0 | 1213 | 655 | 10997 | 8406 | 14798 | 1543 | 2232 | 4890 | 2790 | 511 | 148 | 172 | 244 | 529 | 49187 | 10655 |
| 2009 | 0 | 137 | 1817 | 2280 | 12118 | 8599 | 9735 | 2054 | 1433 | 2608 | 1375 | 237 | 198 | 112 | 248 | 43057 | 9692 |
| 2010 | 231 | 119 | 572 | 2296 | 1828 | 8395 | 5918 | 5676 | 923 | 888 | 1002 | 550 | 89 | 42 | 62 | 28772 | 6649 |
| 2011 | 0 | 1110 | 921 | 1663 | 3592 | 2605 | 9303 | 4390 | 4257 | 771 | 956 | 732 | 269 | 29 | 33 | 30731 | 7336 |


| Year | Age |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Total <br> Biomass |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15+ | Total |  |
| 2012 | 0 | 396 | 2942 | 410 | 668 | 1736 | 2633 | 4328 | 1884 | 2148 | 297 | 604 | 303 | 139 | 41 | 18540 | 4476 |
| 2013 | 0 | 201 | 718 | 3555 | 425 | 1161 | 1859 | 2905 | 4449 | 2772 | 1865 | 678 | 790 | 222 | 102 | 21722 | 5653 |
| 2014 | 13 | 515 | 1258 | 784 | 2788 | 715 | 1118 | 2634 | 2268 | 2806 | 1118 | 703 | 337 | 72 | 212 | 17350 | 4504 |
| 2015 | 0 | 391 | 432 | 1316 | 1132 | 3535 | 1309 | 1191 | 3156 | 2526 | 4457 | 687 | 816 | 290 | 211 | 21450 | 5851 |
| 2016 | 0 | 75 | 3550 | 1538 | 2229 | 1749 | 2631 | 938 | 1092 | 1806 | 1882 | 2853 | 934 | 436 | 130 | 21851 | 5408 |
| 2017 | 10 | 131 | 948 | 4295 | 1198 | 1543 | 826 | 1414 | 317 | 738 | 1008 | 1741 | 2230 | 507 | 237 | 17159 | 4152 |
| 2018 | 0 | 496 | 1004 | 1968 | 5664 | 970 | 1409 | 569 | 1279 | 354 | 675 | 1564 | 1464 | 1498 | 500 | 19412 | 4987 |
| 2019 | 4 | 157 | 2625 | 680 | 2187 | 4656 | 1158 | 1223 | 952 | 1232 | 823 | 655 | 1406 | 917 | 803 | 19487 | 4805 |
| 2020 | 0 | 43 | 472 | 13065 | 513 | 1009 | 2492 | 786 | 629 | 434 | 694 | 324 | 505 | 726 | 902 | 22616 | 4210 |
| 2021 | 15 | 34 | 1109 | 1290 | 11906 | 698 | 1051 | 2039 | 501 | 551 | 476 | 462 | 442 | 615 | 1515 | 22984 | 5096 |
| 2022 | 0 | 507 | 383 | 1207 | 1286 | 9633 | 1151 | 1640 | 2064 | 577 | 339 | 325 | 293 | 115 | 288 | 19817 | 4427 |
| 2023 | 57 | 185 | 584 | 341 | 996 | 1218 | 9459 | 462 | 558 | 1100 | 304 | 214 | 218 | 256 | 514 | 16479 | 4055 |

Table 4.4.8.1 Norwegian spring-spawning herring. Relative standard error of estimated catch-at-age used by XSAM.

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0.368 | 0.193 | 0.263 | 0.094 | 0.364 | 0.498 | 0.424 | 0.315 | 0.371 | 0.544 | 0.382 |
| 1989 | 0.262 | 0.539 | 0.5 | 0.431 | 0.112 | 0.508 | 0.827 | 0.874 | 0.516 | 0.691 | 0.711 |
| 1990 | 0.308 | 0.29 | 0.555 | 0.337 | 0.347 | 0.126 | 0.706 | 0.668 | 0.606 | 0.573 | 0.621 |
| 1991 | 0.514 | 0.378 | 0.546 | 0.685 | 0.314 | 0.371 | 0.128 | 0.566 | 0.994 | 1.731 | 0.658 |
| 1992 | 0.697 | 0.33 | 0.24 | 0.45 | 0.724 | 0.336 | 0.429 | 0.126 | 0.567 | 0.891 | 0.673 |
| 1993 | 0.397 | 0.252 | 0.162 | 0.173 | 0.375 | 0.499 | 0.249 | 0.29 | 0.104 | NA | NA |
| 1994 | 0.383 | 0.241 | 0.161 | 0.108 | 0.14 | 0.308 | 0.382 | 0.23 | 0.234 | 0.089 | 0.436 |
| 1995 | 0.737 | 0.199 | 0.11 | 0.09 | 0.089 | 0.125 | 0.308 | 0.306 | 0.187 | 0.176 | 0.079 |
| 1996 | 0.247 | 0.237 | 0.086 | 0.066 | 0.079 | 0.104 | 0.164 | 0.431 | 0.395 | 0.19 | 0.081 |
| 1997 | 0.275 | 0.152 | 0.119 | 0.063 | 0.061 | 0.084 | 0.112 | 0.196 | 0.283 | 0.241 | 0.098 |
| 1998 | 0.176 | 0.186 | 0.123 | 0.107 | 0.064 | 0.071 | 0.106 | 0.152 | 0.22 | 0.262 | 0.126 |
| 1999 | 0.449 | 0.149 | 0.233 | 0.15 | 0.102 | 0.066 | 0.073 | 0.116 | 0.163 | 0.315 | 0.132 |
| 2000 | 0.316 | 0.176 | 0.093 | 0.235 | 0.16 | 0.104 | 0.071 | 0.076 | 0.128 | 0.185 | 0.15 |
| 2001 | 0.602 | 0.165 | 0.142 | 0.102 | 0.227 | 0.168 | 0.115 | 0.082 | 0.097 | 0.184 | 0.205 |
| 2002 | 0.194 | 0.132 | 0.089 | 0.121 | 0.112 | 0.248 | 0.169 | 0.12 | 0.088 | 0.11 | 0.177 |
| 2003 | 0.465 | 0.182 | 0.112 | 0.086 | 0.137 | 0.14 | 0.271 | 0.183 | 0.128 | 0.093 | 0.119 |
| 2004 | 0.218 | 0.265 | 0.17 | 0.102 | 0.086 | 0.16 | 0.149 | 0.257 | 0.205 | 0.139 | 0.089 |
| 2005 | 0.281 | 0.101 | 0.169 | 0.139 | 0.089 | 0.079 | 0.155 | 0.155 | 0.229 | 0.191 | 0.09 |
| 2006 | 0.216 | 0.182 | 0.086 | 0.177 | 0.139 | 0.086 | 0.084 | 0.173 | 0.181 | 0.247 | 0.106 |
| 2007 | 0.377 | 0.127 | 0.108 | 0.063 | 0.144 | 0.123 | 0.086 | 0.096 | 0.213 | 0.262 | 0.141 |
| 2008 | 0.154 | 0.232 | 0.094 | 0.088 | 0.058 | 0.132 | 0.121 | 0.092 | 0.107 | 0.249 | 0.145 |
| 2009 | 0.159 | 0.133 | 0.145 | 0.073 | 0.079 | 0.062 | 0.147 | 0.12 | 0.102 | 0.124 | 0.15 |
| 2010 | 0.195 | 0.165 | 0.13 | 0.134 | 0.075 | 0.086 | 0.069 | 0.138 | 0.136 | 0.11 | 0.122 |
| 2011 | 0.122 | 0.195 | 0.163 | 0.125 | 0.129 | 0.085 | 0.094 | 0.089 | 0.171 | 0.157 | 0.132 |
| 2012 | 0.326 | 0.129 | 0.209 | 0.156 | 0.117 | 0.12 | 0.085 | 0.113 | 0.108 | 0.205 | 0.154 |
| 2013 | 0.28 | 0.196 | 0.118 | 0.186 | 0.159 | 0.117 | 0.123 | 0.092 | 0.139 | 0.146 | 0.212 |
| 2014 | 0.68 | 0.252 | 0.199 | 0.121 | 0.208 | 0.185 | 0.133 | 0.146 | 0.106 | 0.176 | 0.179 |
| 2015 | 0.519 | 0.304 | 0.201 | 0.206 | 0.144 | 0.237 | 0.191 | 0.143 | 0.164 | 0.131 | 0.176 |
| 2016 | 0.578 | 0.215 | 0.219 | 0.16 | 0.174 | 0.14 | 0.206 | 0.184 | 0.138 | 0.168 | 0.121 |


| Year/Age | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2017 | 0.303 | 0.192 | 0.114 | 0.158 | 0.122 | 0.141 | 0.116 | 0.166 | 0.151 | 0.118 | 0.105 |
| 2018 | 0.273 | 0.261 | 0.197 | 0.12 | 0.145 | 0.136 | 0.161 | 0.136 | 0.174 | 0.166 | 0.097 |
| 2019 | 0.59 | 0.145 | 0.192 | 0.135 | 0.094 | 0.146 | 0.128 | 0.156 | 0.14 | 0.15 | 0.095 |
| 2020 | 0.38 | 0.21 | 0.094 | 0.186 | 0.145 | 0.103 | 0.157 | 0.143 | 0.169 | 0.134 | 0.102 |
| 2021 | 0.576 | 0.263 | 0.16 | 0.068 | 0.141 | 0.131 | 0.11 | 0.157 | 0.153 | 0.184 | 0.117 |
| 2022 | 0.259 | 0.273 | 0.171 | 0.145 | 0.069 | 0.167 | 0.131 | 0.122 | 0.185 | 0.179 | 0.124 |

Table 4.4.8.2 Norwegian spring-spawning herring. Relative standard error of Fleet 1 used by XSAM.

| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0.326 | 0.344 | 0.168 | 0.46 | 0.559 | 0.698 | 0.549 | 0.611 | 0.523 | NA |
| 1989 | 0.655 | 0.336 | 0.448 | 0.196 | 0.438 | 0.698 | 0.887 | 0.5 | NA | 0.5 |
| 1990 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1991 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1992 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1993 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1994 | 0.439 | 0.515 | 0.275 | 0.311 | 0.487 | 0.686 | 0.405 | 0.513 | 0.232 | 0.762 |
| 1995 | 0.316 | 0.189 | 0.206 | 0.229 | 0.346 | 0.624 | NA | 0.433 | 0.5 | 0.221 |
| 1996 | 0.383 | 0.229 | 0.169 | 0.221 | 0.273 | 0.345 | NA | NA | 0.41 | 0.235 |
| 1997 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 1998 | 0.343 | 0.267 | 0.209 | 0.152 | 0.168 | 0.232 | 0.298 | 0.395 | 0.528 | 0.236 |
| 1999 | 0.242 | 0.33 | 0.249 | 0.214 | 0.163 | 0.173 | 0.237 | 0.307 | 0.415 | 0.285 |
| 2000 | 0.288 | 0.215 | 0.431 | 0.312 | 0.258 | 0.193 | 0.211 | 0.31 | 0.511 | 0.364 |
| 2001 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2002 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2003 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2004 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2005 | 0.363 | 0.291 | 0.225 | 0.18 | 0.167 | 0.295 | 0.308 | 0.398 | 0.447 | 0.221 |
| 2006 | 0.449 | 0.179 | 0.303 | 0.273 | 0.204 | 0.204 | 0.43 | 0.454 | 0.624 | 0.315 |
| 2007 | 0.33 | 0.239 | 0.153 | 0.278 | 0.291 | 0.209 | 0.206 | 0.407 | 0.367 | 0.266 |
| 2008 | 0.515 | 0.216 | 0.212 | 0.165 | 0.306 | 0.313 | 0.241 | 0.241 | 0.396 | 0.322 |


| Year/Age | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2009 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2010 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2011 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2012 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2013 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2014 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2015 | 0.305 | 0.213 | 0.283 | 0.205 | 0.326 | 0.288 | 0.213 | 0.28 | 0.172 | 0.221 |
| 2016 | 0.379 | 0.364 | 0.297 | 0.331 | 0.225 | 0.363 | 0.307 | 0.21 | 0.291 | 0.186 |
| 2017 | 0.43 | 0.261 | 0.293 | 0.273 | 0.32 | 0.25 | 0.386 | 0.375 | 0.227 | 0.2 |
| 2018 | 0.404 | 0.236 | 0.212 | 0.315 | 0.275 | 0.293 | 0.26 | 0.317 | 0.33 | 0.203 |
| 2019 | 0.329 | 0.343 | 0.269 | 0.199 | 0.282 | 0.273 | 0.291 | 0.262 | 0.284 | 0.191 |
| 2020 | 0.525 | 0.202 | 0.3 | 0.255 | 0.204 | 0.305 | 0.296 | 0.331 | 0.292 | 0.23 |
| 2021 | 0.428 | 0.347 | 0.16 | 0.284 | 0.284 | 0.23 | 0.32 | 0.316 | 0.342 | 0.229 |
| 2022 | 0.472 | 0.395 | 0.282 | 0.177 | 0.341 | 0.301 | 0.242 | 0.345 | 0.311 | 0.241 |
| 2023 | 0.375 | 0.382 | 0.3 | 0.297 | 0.179 | 0.334 | 0.262 | 0.249 | 0.344 | 0.219 |

Table 4.4.8.3 Norwegian spring-spawning herring. Relative standard error of Fleet 4 used by XSAM.

| Year/Age | $\mathbf{2}$ |
| :--- | :--- |
| 1991 | 0.462 |
| 1992 | 0.419 |
| 1993 | 0.395 |
| 1994 | 0.364 |
| 1995 | 0.444 |
| 1996 | 0.620 |
| 1997 | 0.741 |
| 1998 | 0.466 |
| 1999 | 0.464 |
| 2000 | 0.392 |
| 2001 | 0.445 |
| 2002 | 0.475 |


| Year/Age | 2 |
| :---: | :---: |
| 2003 | NA |
| 2004 | NA |
| 2005 | 0.468 |
| 2006 | 0.383 |
| 2007 | 0.477 |
| 2008 | 0.595 |
| 2009 | 0.609 |
| 2010 | 0.525 |
| 2011 | 0.474 |
| 2012 | 0.763 |
| 2013 | 0.502 |
| 2014 | 0.485 |
| 2015 | 0.426 |
| 2016 | 0.458 |
| 2017 | 0.486 |
| 2018 | 0.410 |
| 2019 | 0.498 |
| 2020 | NA |
| 2021 | 1.006 |
| 2022 | NA |
| 2023 | NA |

Table 4.4.8.4 Norwegian spring-spawning herring. Relative standard error of Fleet 5 used by XSAM.

| Year/Age | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1996 | 0.203 | 0.136 | 0.154 | 0.195 | 0.239 | 0.346 | 0.772 | 0.909 | 0.439 | 0.216 |
| 1997 | 0.273 | 0.21 | 0.142 | 0.153 | 0.229 | 0.248 | 0.425 | 0.517 | 0.379 | 0.22 |
| 1998 | 0.358 | 0.277 | 0.2 | 0.147 | 0.164 | 0.239 | 0.298 | 0.422 | $N A$ | 0.329 |
| 1999 | 0.235 | 0.37 | 0.286 | 0.218 | 0.159 | 0.185 | 0.294 | 0.389 | 0.984 | 0.376 |
| 2000 | 0.264 | 0.223 | 0.496 | 0.355 | 0.266 | 0.178 | 0.191 | 0.25 | 0.385 | 0.419 |
| 2001 | 0.172 | 0.26 | 0.259 | 0.424 | 0.411 | 0.215 | 0.19 | 0.27 | 0.494 | 0.422 |


| Year/Age | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2002 | 0.183 | 0.166 | 0.261 | 0.3 | 0.357 | 0.294 | 0.242 | 0.228 | 0.261 | 0.431 |
| 2003 | 0.182 | 0.165 | 0.165 | 0.257 | 0.305 | 0.445 | 0.401 | 0.245 | 0.231 | 0.239 |
| 2004 | 0.256 | 0.192 | 0.156 | 0.162 | 0.278 | 0.322 | 0.519 | 0.372 | 0.36 | 0.228 |
| 2005 | 0.141 | 0.264 | 0.248 | 0.184 | 0.191 | 0.313 | 0.354 | 0.45 | 0.388 | 0.24 |
| 2006 | 0.374 | 0.151 | 0.262 | 0.24 | 0.182 | 0.179 | 0.31 | 0.307 | 0.428 | 0.236 |
| 2007 | 0.221 | 0.187 | 0.14 | 0.268 | 0.241 | 0.181 | 0.189 | 0.314 | 0.335 | 0.222 |
| 2008 | 0.313 | 0.161 | 0.172 | 0.15 | 0.256 | 0.234 | 0.195 | 0.222 | 0.332 | 0.277 |
| 2009 | 0.246 | 0.233 | 0.158 | 0.171 | 0.166 | 0.239 | 0.26 | 0.226 | 0.263 | 0.299 |
| 2010 | 0.323 | 0.233 | 0.246 | 0.172 | 0.186 | 0.188 | 0.289 | 0.291 | 0.283 | 0.304 |
| 2011 | 0.289 | 0.251 | 0.21 | 0.226 | 0.168 | 0.2 | 0.201 | 0.301 | 0.286 | 0.279 |
| 2012 | 0.22 | 0.349 | 0.311 | 0.249 | 0.226 | 0.201 | 0.244 | 0.237 | 0.377 | 0.278 |
| 2013 | 0.306 | 0.21 | 0.346 | 0.273 | 0.245 | 0.22 | 0.199 | 0.223 | 0.245 | 0.247 |
| 2014 | 0.268 | 0.3 | 0.222 | 0.306 | 0.276 | 0.225 | 0.234 | 0.222 | 0.276 | 0.265 |
| 2015 | 0.345 | 0.265 | 0.275 | 0.21 | 0.266 | 0.272 | 0.216 | 0.228 | 0.199 | 0.24 |
| 2016 | 0.21 | 0.256 | 0.235 | 0.248 | 0.226 | 0.287 | 0.277 | 0.246 | 0.244 | 0.2 |
| 2017 | 0.287 | 0.201 | 0.271 | 0.256 | 0.296 | 0.261 | 0.371 | 0.304 | 0.283 | 0.197 |
| 2018 | 0.283 | 0.241 | 0.188 | 0.285 | 0.261 | 0.323 | 0.267 | 0.361 | 0.311 | 0.194 |
| 2019 | 0.226 | 0.31 | 0.236 | 0.197 | 0.274 | 0.27 | 0.286 | 0.27 | 0.296 | 0.207 |
| 2020 | 0.338 | 0.155 | 0.331 | 0.283 | 0.228 | 0.3 | 0.316 | 0.345 | 0.309 | 0.229 |
| 2021 | 0.276 | 0.267 | 0.158 | 0.308 | 0.28 | 0.239 | 0.333 | 0.326 | 0.337 | 0.218 |
| 2022 | 0.355 | 0.271 | 0.267 | 0.166 | 0.274 | 0.252 | 0.239 | 0.322 | 0.365 | 0.282 |
| 2023 | 0.321 | 0.365 | 0.283 | 0.27 | 0.167 | 0.339 | 0.325 | 0.277 | 0.375 | 0.271 |

Table 4.5.1.1. Norwegian spring-spawning herring. Parameter estimates of the final XSAM model fit. The estimates from the final 2022 assessment are also shown.

| Parameter | Estimate | Std. Error | CV | Estimate 2022 | Std. Error 2022 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| $\boldsymbol{\operatorname { l o g } ( N _ { 3 , 1 9 8 8 } )}$ | 7.085 | 0.163 | 0.023 | 7.085 | 0.163 |
| $\boldsymbol{\operatorname { l o g } ( \boldsymbol { N } _ { 4 , 1 9 8 8 } )}$ | 6.631 | 0.201 | 0.03 | 6.631 | 0.201 |
| $\boldsymbol{\operatorname { l o g } ( \boldsymbol { N } _ { 5 , 1 9 8 8 } )}$ | 9.586 | 0.066 | 0.007 | 9.586 | 0.066 |
| $\boldsymbol{\operatorname { l o g } ( \boldsymbol { N } _ { 6 , 1 9 8 8 } )}$ | 4.834 | 0.375 | 0.078 | 4.837 | 0.380 |


| Parameter | Estimate | Std. Error | CV | Estimate 2022 | Std. Error 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\log \left(N_{7,1988}\right)$ | 3.521 | 0.522 | 0.148 | 3.527 | 0.532 |
| $\log \left(N_{8,1988}\right)$ | 3.078 | 0.583 | 0.189 | 3.079 | 0.594 |
| $\log \left(N_{9,1988}\right)$ | 4.066 | 0.449 | 0.11 | 4.073 | 0.455 |
| $\log \left(N_{10,1988}\right)$ | 3.277 | 0.657 | 0.201 | 3.282 | 0.669 |
| $\log \left(N_{11,1988}\right)$ | 3.172 | 0.684 | 0.215 | 3.191 | 0.691 |
| $\log \left(N_{12,1988}\right)$ | 3.574 | 0.741 | 0.207 | 3.585 | 0.753 |
| $\log \left(q_{3}^{F 1}\right)$ | -9.666 | 0.17 | 0.018 | -9.657 | 0.173 |
| $\log \left(q_{4}^{F 1}\right)$ | -8.158 | 0.123 | 0.015 | -8.143 | 0.124 |
| $\log \left(q_{5}^{F 1}\right)$ | -7.506 | 0.11 | 0.015 | -7.487 | 0.111 |
| $\log \left(q_{6}^{F 1}\right)$ | -7.302 | 0.11 | 0.015 | -7.283 | 0.110 |
| $\log \left(q_{7}^{F 1}\right)$ | -7.157 | 0.119 | 0.017 | -7.165 | 0.123 |
| $\log \left(q_{8}^{F 1}\right)$ | -6.917 | 0.085 | 0.012 | -6.926 | 0.086 |
| $\log \left(q_{2}^{F 4}\right)$ | -14.533 | 0.184 | 0.013 | -14.525 | 0.189 |
| $\log \left(q_{3}^{F 5}\right)$ | -7.678 | 0.102 | 0.013 | -7.654 | 0.105 |
| $\log \left(q_{4}^{F 5}\right)$ | -7.148 | 0.09 | 0.013 | -7.133 | 0.093 |
| $\log \left(q_{5}^{F 5}\right)$ | -6.924 | 0.088 | 0.013 | -6.913 | 0.091 |
| $\log \left(q_{6}^{F 5}\right)$ | -6.803 | 0.091 | 0.013 | -6.796 | 0.094 |
| $\log \left(q_{7}^{F 5}\right)$ | -6.708 | 0.096 | 0.014 | -6.721 | 0.101 |
| $\log \left(q_{8}^{F 5}\right)$ | -6.555 | 0.103 | 0.016 | -6.541 | 0.106 |
| $\log \left(q_{9}^{F 5}\right)$ | -6.555 | 0.114 | 0.017 | -6.537 | 0.118 |
| $\log \left(q_{10}^{F 5}\right)$ | -6.495 | 0.126 | 0.019 | -6.474 | 0.132 |
| $\log \left(q_{11}^{F 5}\right)$ | -6.457 | 0.121 | 0.019 | -6.433 | 0.126 |
| $\log \left(\sigma_{1}^{2}\right)$ | -5 | 1.419 | 0.284 | -5.000 | 1.409 |
| $\log \left(\sigma_{2}^{2}\right)$ | -2.827 | 0.247 | 0.087 | -2.777 | 0.243 |
| $\log \left(\sigma_{4}^{2}\right)$ | -2.31 | 0.295 | 0.128 | -2.281 | 0.299 |
| $\log \left(\sigma_{R}^{2}\right)$ | -0.047 | 0.249 | 5.278 | -0.022 | 0.255 |
| $\log (h)$ | 1.518 | 0.062 | 0.041 | 1.565 | 0.063 |
| $\mu_{R}$ | 9.27 | 0.171 | 9.27 | 9.275 | 0.176 |


| Parameter | Estimate | Std. Error | CV | Estimate 2022 | Std. Error 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| $\alpha_{Y}$ | -0.489 | 0.288 | -0.489 | -0.492 | 0.294 |
| $\beta_{Y}$ | 0.817 | 0.105 | 0.817 | 0.816 | 0.107 |
| $\alpha_{2 U}$ | -1.239 | 0.162 | -1.239 | -1.239 | 0.164 |
| $\alpha_{3 U}$ | -0.627 | 0.093 | -0.627 | -0.629 | 0.095 |
| $\alpha_{4 U}$ | -0.213 | 0.057 | -0.213 | -0.215 | 0.059 |
| $\alpha_{5 U}$ | 0.053 | 0.048 | 0.053 | 0.054 | 0.049 |
| $\alpha_{6 U}$ | 0.203 | 0.052 | 0.203 | 0.199 | 0.054 |
| $\alpha_{7 U}$ | 0.262 | 0.057 | 0.262 | 0.263 | 0.058 |
| $\alpha_{8 U}$ | 0.326 | 0.063 | 0.326 | 0.319 | 0.065 |
| $\alpha_{9 U}$ | 0.363 | 0.068 | 0.363 | 0.368 | 0.070 |
| $\alpha_{10 U}$ | 0.42 | 0.074 | 0.42 | 0.419 | 0.076 |
| $\boldsymbol{\beta}_{\boldsymbol{U}}$ | 0.601 | 0.051 | 0.601 | 0.603 | 0.052 |

Table 4.5.1.2 Norwegian spring-spawning herring. Point estimates of Stock in numbers (millions).

| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 673 | 1194 | 759 | 14554 | 126 | 34 | 22 | 58 | 27 | 24 | 36 |
| 1989 | 1172 | 260 | 964 | 635 | 12034 | 104 | 28 | 16 | 40 | 16 | 43 |
| 1990 | 4345 | 471 | 219 | 816 | 531 | 10030 | 86 | 22 | 13 | 30 | 47 |
| 1991 | 11486 | 1760 | 400 | 186 | 685 | 444 | 8384 | 70 | 18 | 10 | 61 |
| 1992 | 18699 | 4664 | 1507 | 341 | 157 | 575 | 372 | 6987 | 58 | 14 | 58 |
| 1993 | 50139 | 7596 | 3999 | 1280 | 286 | 131 | 480 | 309 | 5778 | 47 | 59 |
| 1994 | 60080 | 20363 | 6507 | 3373 | 1044 | 231 | 107 | 388 | 249 | 4579 | 82 |
| 1995 | 15817 | 24390 | 17438 | 5480 | 2640 | 780 | 178 | 83 | 300 | 187 | 3443 |
| 1996 | 5733 | 6412 | 20835 | 14604 | 4180 | 1759 | 510 | 128 | 60 | 207 | 2248 |
| 1997 | 2151 | 2319 | 5442 | 17238 | 11168 | 2805 | 1130 | 335 | 89 | 41 | 1357 |
| 1998 | 10937 | 868 | 1923 | 4384 | 13139 | 7774 | 1749 | 665 | 209 | 54 | 755 |
| 1999 | 6486 | 4415 | 715 | 1485 | 3384 | 9624 | 5449 | 1121 | 414 | 123 | 458 |
| 2000 | 32755 | 2625 | 3679 | 557 | 1134 | 2513 | 6830 | 3655 | 702 | 246 | 300 |
| 2001 | 29101 | 13269 | 2198 | 2743 | 417 | 833 | 1796 | 4670 | 2254 | 410 | 271 |
| 2002 | 11450 | 11797 | 11254 | 1750 | 2009 | 312 | 617 | 1292 | 3244 | 1489 | 452 |


| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 6733 | 4635 | 9968 | 9087 | 1290 | 1411 | 226 | 433 | 879 | 2162 | 1295 |
| 2004 | 57974 | 2730 | 3927 | 8241 | 7135 | 950 | 1033 | 164 | 305 | 592 | 2262 |
| 2005 | 24559 | 23525 | 2320 | 3272 | 6664 | 5494 | 708 | 749 | 119 | 214 | 1769 |
| 2006 | 43071 | 9960 | 19890 | 1912 | 2617 | 5105 | 3888 | 484 | 509 | 79 | 1137 |
| 2007 | 12179 | 17468 | 8470 | 16458 | 1537 | 2048 | 3745 | 2663 | 335 | 353 | 708 |
| 2008 | 17671 | 4932 | 14817 | 6974 | 12616 | 1165 | 1502 | 2552 | 1769 | 226 | 720 |
| 2009 | 7135 | 7128 | 4173 | 12213 | 5394 | 8791 | 824 | 1035 | 1635 | 1120 | 630 |
| 2010 | 5094 | 2862 | 5967 | 3425 | 9437 | 3837 | 5724 | 553 | 646 | 980 | 1083 |
| 2011 | 15441 | 2044 | 2386 | 4905 | 2730 | 7114 | 2674 | 3574 | 347 | 399 | 1130 |
| 2012 | 5712 | 6196 | 1708 | 1958 | 3954 | 2132 | 5364 | 1817 | 2393 | 226 | 974 |
| 2013 | 8495 | 2309 | 5188 | 1409 | 1576 | 3132 | 1630 | 3937 | 1281 | 1677 | 847 |
| 2014 | 5577 | 3441 | 1947 | 4253 | 1136 | 1249 | 2440 | 1217 | 2878 | 926 | 1973 |
| 2015 | 17619 | 2263 | 2928 | 1625 | 3454 | 920 | 1002 | 1919 | 933 | 2165 | 2318 |
| 2016 | 7702 | 7154 | 1932 | 2465 | 1344 | 2820 | 750 | 803 | 1518 | 722 | 3577 |
| 2017 | 4314 | 3126 | 6104 | 1621 | 2023 | 1080 | 2255 | 593 | 626 | 1156 | 3336 |
| 2018 | 42477 | 1749 | 2646 | 5004 | 1286 | 1521 | 801 | 1644 | 431 | 431 | 3205 |
| 2019 | 5170 | 17235 | 1486 | 2202 | 4042 | 984 | 1153 | 598 | 1224 | 312 | 2557 |
| 2020 | 4424 | 2096 | 14642 | 1225 | 1737 | 3001 | 722 | 819 | 423 | 865 | 1822 |
| 2021 | 2043 | 1793 | 1774 | 12075 | 967 | 1317 | 2201 | 517 | 572 | 292 | 1743 |
| 2022 | 6944 | 827 | 1513 | 1438 | 9159 | 697 | 944 | 1537 | 348 | 377 | 1402 |
| 2023 | 10619 | 2811 | 697 | 1224 | 1099 | 6644 | 502 | 645 | 1063 | 230 | 1235 |

Table 4.5.1.3 Norwegian spring-spawning herring. Point estimates of Fishing mortality.

| Year/Age | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1988 | 0.049 | 0.064 | 0.028 | 0.04 | 0.042 | 0.042 | 0.142 | 0.227 | 0.338 | 0.174 | 0.174 |
| 1989 | 0.011 | 0.021 | 0.017 | 0.027 | 0.032 | 0.035 | 0.077 | 0.11 | 0.152 | 0.093 | 0.093 |
| 1990 | 0.004 | 0.012 | 0.015 | 0.024 | 0.031 | 0.029 | 0.052 | 0.073 | 0.098 | 0.071 | 0.071 |
| 1991 | 0.001 | 0.005 | 0.011 | 0.019 | 0.025 | 0.025 | 0.032 | 0.043 | 0.057 | 0.049 | 0.049 |
| 1992 | 0.001 | 0.004 | 0.013 | 0.025 | 0.031 | 0.03 | 0.035 | 0.04 | 0.055 | 0.057 | 0.057 |
| 1993 | 0.001 | 0.005 | 0.02 | 0.054 | 0.063 | 0.058 | 0.064 | 0.068 | 0.083 | 0.104 | 0.104 |


| Year/Age | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1994 | 0.002 | 0.005 | 0.022 | 0.095 | 0.141 | 0.115 | 0.099 | 0.107 | 0.134 | 0.153 | 0.153 |
| 1995 | 0.003 | 0.008 | 0.027 | 0.121 | 0.256 | 0.275 | 0.176 | 0.17 | 0.222 | 0.329 | 0.329 |
| 1996 | 0.005 | 0.014 | 0.04 | 0.118 | 0.249 | 0.292 | 0.271 | 0.211 | 0.244 | 0.443 | 0.443 |
| 1997 | 0.008 | 0.038 | 0.066 | 0.122 | 0.212 | 0.323 | 0.381 | 0.323 | 0.351 | 0.465 | 0.465 |
| 1998 | 0.007 | 0.044 | 0.108 | 0.109 | 0.161 | 0.205 | 0.294 | 0.324 | 0.375 | 0.419 | 0.419 |
| 1999 | 0.004 | 0.032 | 0.099 | 0.12 | 0.147 | 0.193 | 0.249 | 0.318 | 0.369 | 0.511 | 0.511 |
| 2000 | 0.004 | 0.028 | 0.144 | 0.141 | 0.158 | 0.186 | 0.23 | 0.333 | 0.387 | 0.552 | 0.552 |
| 2001 | 0.003 | 0.015 | 0.078 | 0.161 | 0.141 | 0.149 | 0.179 | 0.214 | 0.264 | 0.259 | 0.259 |
| 2002 | 0.004 | 0.018 | 0.064 | 0.155 | 0.204 | 0.17 | 0.204 | 0.235 | 0.256 | 0.255 | 0.255 |
| 2003 | 0.003 | 0.016 | 0.04 | 0.092 | 0.155 | 0.162 | 0.17 | 0.201 | 0.246 | 0.274 | 0.274 |
| 2004 | 0.002 | 0.013 | 0.032 | 0.062 | 0.111 | 0.144 | 0.171 | 0.172 | 0.202 | 0.328 | 0.328 |
| 2005 | 0.003 | 0.018 | 0.044 | 0.074 | 0.117 | 0.196 | 0.232 | 0.236 | 0.263 | 0.406 | 0.406 |
| 2006 | 0.002 | 0.012 | 0.039 | 0.068 | 0.095 | 0.16 | 0.228 | 0.217 | 0.216 | 0.39 | 0.39 |
| 2007 | 0.004 | 0.015 | 0.044 | 0.116 | 0.127 | 0.16 | 0.234 | 0.259 | 0.244 | 0.238 | 0.238 |
| 2008 | 0.008 | 0.017 | 0.043 | 0.107 | 0.211 | 0.197 | 0.223 | 0.295 | 0.307 | 0.257 | 0.257 |
| 2009 | 0.014 | 0.028 | 0.047 | 0.108 | 0.191 | 0.279 | 0.249 | 0.322 | 0.362 | 0.33 | 0.33 |
| 2010 | 0.013 | 0.032 | 0.046 | 0.077 | 0.132 | 0.211 | 0.321 | 0.316 | 0.332 | 0.452 | 0.452 |
| 2011 | 0.013 | 0.03 | 0.048 | 0.066 | 0.097 | 0.132 | 0.236 | 0.251 | 0.277 | 0.3 | 0.3 |
| 2012 | 0.006 | 0.028 | 0.042 | 0.067 | 0.083 | 0.119 | 0.159 | 0.199 | 0.206 | 0.199 | 0.199 |
| 2013 | 0.004 | 0.021 | 0.049 | 0.066 | 0.082 | 0.1 | 0.142 | 0.164 | 0.175 | 0.096 | 0.096 |
| 2014 | 0.002 | 0.011 | 0.031 | 0.058 | 0.06 | 0.071 | 0.09 | 0.116 | 0.134 | 0.074 | 0.074 |
| 2015 | 0.001 | 0.008 | 0.022 | 0.04 | 0.053 | 0.055 | 0.072 | 0.085 | 0.106 | 0.076 | 0.076 |
| 2016 | 0.002 | 0.009 | 0.026 | 0.048 | 0.068 | 0.074 | 0.084 | 0.099 | 0.122 | 0.104 | 0.104 |
| 2017 | 0.003 | 0.017 | 0.049 | 0.081 | 0.136 | 0.149 | 0.166 | 0.171 | 0.222 | 0.188 | 0.188 |
| 2018 | 0.002 | 0.013 | 0.034 | 0.063 | 0.118 | 0.127 | 0.142 | 0.144 | 0.173 | 0.202 | 0.202 |
| 2019 | 0.003 | 0.013 | 0.044 | 0.087 | 0.148 | 0.159 | 0.192 | 0.196 | 0.197 | 0.304 | 0.304 |
| 2020 | 0.003 | 0.017 | 0.043 | 0.087 | 0.127 | 0.16 | 0.184 | 0.208 | 0.221 | 0.283 | 0.283 |
| 2021 | 0.004 | 0.02 | 0.06 | 0.126 | 0.177 | 0.183 | 0.209 | 0.246 | 0.267 | 0.222 | 0.222 |
| 2022 | 0.004 | 0.021 | 0.062 | 0.119 | 0.171 | 0.179 | 0.231 | 0.218 | 0.264 | 0.215 | 0.215 |


| Year/Age | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2 +}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2023 | 0.004 | 0.019 | 0.055 | 0.107 | 0.155 | 0.167 | 0.21 | 0.212 | 0.251 | 0.222 | 0.222 |

Table 4.5.1.4 Norwegian spring spawning herring. Final stock summary table. High and low represent approximate 95 \% confidence limits.

| Year | Recruitment (Age 2) | High | Low | Stock Size: SSB | High | Low | Catches | Fishing Pressure: F | High | Low |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | millions |  |  | thousand tonnes |  |  | thousand tonnes | Ages 5-12 |  |  |
| 1988 | 673 | 991 | 355 | 2128 | 2392 | 1865 | 135.301 | 0.042 | 0.058 | 0.026 |
| 1989 | 1172 | 1644 | 699 | 3292 | 3700 | 2884 | 103.83 | 0.033 | 0.046 | 0.019 |
| 1990 | 4345 | 5358 | 3332 | 3566 | 3998 | 3134 | 86.411 | 0.03 | 0.042 | 0.017 |
| 1991 | 11486 | 13358 | 9613 | 3343 | 3747 | 2940 | 84.683 | 0.031 | 0.044 | 0.018 |
| 1992 | 18699 | 21323 | 16075 | 3371 | 3757 | 2985 | 104.448 | 0.039 | 0.054 | 0.023 |
| 1993 | 50139 | 55387 | 44891 | 3344 | 3691 | 2997 | 232.457 | 0.076 | 0.099 | 0.053 |
| 1994 | 60080 | 65952 | 54208 | 3476 | 3822 | 3130 | 479.228 | 0.128 | 0.159 | 0.098 |
| 1995 | 15817 | 18145 | 13489 | 3543 | 3874 | 3212 | 905.501 | 0.218 | 0.258 | 0.179 |
| 1996 | 5733 | 6846 | 4620 | 4126 | 4456 | 3795 | 1220.283 | 0.192 | 0.222 | 0.162 |
| 1997 | 2151 | 2708 | 1594 | 5385 | 5775 | 4996 | 1426.507 | 0.193 | 0.22 | 0.166 |
| 1998 | 10937 | 12704 | 9171 | 5966 | 6394 | 5538 | 1223.131 | 0.186 | 0.214 | 0.158 |
| 1999 | 6486 | 7696 | 5277 | 5865 | 6315 | 5415 | 1235.433 | 0.213 | 0.246 | 0.18 |
| 2000 | 32755 | 36631 | 28879 | 4884 | 5298 | 4471 | 1207.201 | 0.257 | 0.299 | 0.215 |
| 2001 | 29101 | 32706 | 25496 | 4055 | 4427 | 3683 | 766.136 | 0.203 | 0.24 | 0.166 |
| 2002 | 11450 | 13326 | 9574 | 3579 | 3925 | 3234 | 807.795 | 0.223 | 0.264 | 0.182 |
| 2003 | 6733 | 8010 | 5456 | 4209 | 4590 | 3828 | 789.51 | 0.152 | 0.18 | 0.124 |
| 2004 | 57974 | 64050 | 51898 | 5300 | 5763 | 4837 | 794.066 | 0.128 | 0.152 | 0.105 |
| 2005 | 24559 | 27906 | 21212 | 5426 | 5915 | 4936 | 1003.243 | 0.172 | 0.203 | 0.142 |
| 2006 | 43071 | 48228 | 37913 | 5393 | 5873 | 4912 | 968.958 | 0.176 | 0.209 | 0.143 |
| 2007 | 12179 | 14310 | 10047 | 6937 | 7523 | 6351 | 1266.993 | 0.155 | 0.182 | 0.129 |
| 2008 | 17671 | 20469 | 14874 | 7025 | 7640 | 6410 | 1545.656 | 0.2 | 0.233 | 0.167 |
| 2009 | 7135 | 8546 | 5724 | 6997 | 7643 | 6351 | 1687.373 | 0.206 | 0.238 | 0.174 |
| 2010 | 5094 | 6178 | 4010 | 6205 | 6819 | 5592 | 1457.014 | 0.213 | 0.25 | 0.177 |
| 2011 | 15441 | 17926 | 12955 | 5867 | 6482 | 5253 | 992.998 | 0.158 | 0.187 | 0.129 |
| 2012 | 5712 | 6855 | 4569 | 5712 | 6335 | 5088 | 825.999 | 0.141 | 0.168 | 0.114 |
| 2013 | 8495 | 10026 | 6964 | 5344 | 5944 | 4744 | 684.743 | 0.121 | 0.146 | 0.096 |


| Year | Recruitment (Age 2) | High | Low | Stock Size: SSB | High | Low | Catches <br> thousand tonnes | Fishing Pressure: F <br> Ages 5-12 | High | Low |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | millions |  |  | thousand tonnes |  |  |  |  |  |  |
| 2014 | 5577 | 6700 | 4455 | 5164 | 5758 | 4571 | 461.306 | 0.085 | 0.103 | 0.066 |
| 2015 | 17619 | 20510 | 14728 | 4816 | 5380 | 4252 | 328.74 | 0.068 | 0.084 | 0.051 |
| 2016 | 7702 | 9333 | 6071 | 4263 | 4771 | 3755 | 383.174 | 0.085 | 0.105 | 0.065 |
| 2017 | 4314 | 5445 | 3182 | 4136 | 4619 | 3653 | 721.566 | 0.16 | 0.193 | 0.127 |
| 2018 | 42477 | 50403 | 34551 | 4046 | 4528 | 3563 | 592.899 | 0.127 | 0.154 | 0.1 |
| 2019 | 5170 | 6854 | 3486 | 3986 | 4479 | 3494 | 777.165 | 0.183 | 0.221 | 0.146 |
| 2020 | 4424 | 6247 | 2600 | 3461 | 3923 | 2998 | 720.937 | 0.185 | 0.227 | 0.143 |
| 2021 | 2043 | 3303 | 783 | 4077 | 4642 | 3512 | 851.813 | 0.159 | 0.194 | 0.124 |
| 2022 | 6944 | 12099 | 1790 | 4060 | 4714 | 3406 | 813.8337 | 0.182 | 0.226 | 0.137 |
| 2023 | 10619 | 31256 | 0 | 3664 | 4353 | 2974 |  |  |  |  |
| Average | 15888 | 18873 | 13182 | 4611 | 5091 | 4132 | 791 | 0.146 | 0.175 | 0.118 |

## Table 4.8.1.1 Norwegian Spring-spawning herring. Input to short-term prediction. Stock size is in millions and weight in kg.

| Input for | 2023 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Stockno. | Natural | Maturity | Proportion of M | Proportion of F | Weight | Exploitation | Weight |
| age | 1-Jan. | mortality | ogive | before spawning | before spawning | in stock | pattern | in catch |
| 2 | 10619 | 0.90 | 0.0 | 0 | 0 | 0.054 | 0.004 | 0.154 |
| 3 | 2811 | 0.15 | 0.0 | 0 | 0 | 0.118 | 0.02 | 0.207 |
| 4 | 697 | 0.15 | 0.4 | 0 | 0 | 0.125 | 0.058 | 0.232 |
| 5 | 1224 | 0.15 | 0.8 | 0 | 0 | 0.202 | 0.112 | 0.266 |
| 6 | 1099 | 0.15 | 0.9 | 0 | 0 | 0.255 | 0.162 | 0.29 |
| 7 | 6644 | 0.15 | 1.0 | 0 | 0 | 0.28 | 0.175 | 0.317 |
| 8 | 502 | 0.15 | 1.0 | 0 | 0 | 0.311 | 0.22 | 0.338 |
| 9 | 645 | 0.15 | 1.0 | 0 | 0 | 0.336 | 0.222 | 0.359 |
| 10 | 1063 | 0.15 | 1.0 | 0 | 0 | 0.337 | 0.263 | 0.373 |
| 11 | 230 | 0.15 | 1.0 | 0 | 0 | 0.345 | 0.233 | 0.392 |
| 12 | 1235 | 0.15 | 1.0 | 0 | 0 | 0.389 | 0.233 | 0.421 |


| Input for | 2024/2025 |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Stockno. | Natural | Maturity | Proportion of M | Proportion of F | Weight | Exploitation | Weight |
| age | 1-Jan. | mortality | ogive | before spawning | before spawning | in stock | pattern | in catch |
| 2 | 10619 | 0.90 | 0.0 | 0 | 0 | 0.054 | 0.004 | 0.154 |
| 3 |  | 0.15 | 0.0 | 0 | 0 | 0.118 | 0.02 | 0.207 |
| 4 |  | 0.15 | 0.4 | 0 | 0 | 0.125 | 0.058 | 0.232 |
| 5 |  | 0.15 | 0.8 | 0 | 0 | 0.202 | 0.112 | 0.266 |
| 6 |  | 0.15 | 0.9 | 0 | 0 | 0.255 | 0.162 | 0.29 |
| 7 |  | 0.15 | 1.0 | 0 | 0 | 0.28 | 0.175 | 0.317 |
| 8 |  | 0.15 | 1.0 | 0 | 0 | 0.311 | 0.22 | 0.338 |
| 9 |  | 0.15 | 1.0 | 0 | 0 | 0.336 | 0.222 | 0.359 |
| 10 |  | 0.15 | 1.0 | 0 | 0 | 0.337 | 0.263 | 0.373 |
| 11 |  | 0.15 | 1.0 | 0 | 0 | 0.345 | 0.233 | 0.392 |
| 12 |  | 0.15 | 1.0 | 0 | 0 | 0.389 | 0.233 | 0.421 |

Table 4.8.2.1 Norwegian spring spawning herring. Short-term prediction.

| Basis: |  |
| :--- | :--- |
| SSB (2023): | 3.664 million $t$ |
| Landings(2023): | 692294 (sum of national quotas) |
| SSB(2024): | 3.059 million $t$ |
| Fw5-12+(2023) | 0.186 |
| Recruitment(2023-2024): | $10.61916,10.61916,10.61916$ |

The catch options:

| Rationale | Catches (2023) | Basis | FW <br> (2023) | $\begin{aligned} & \text { SSB } \\ & (2024)^{*} \end{aligned}$ | $\begin{aligned} & \text { P(SSB2024 } \\ & \left.<B_{\text {lim }}\right) \end{aligned}$ | $\begin{aligned} & \text { \% SSB } \\ & \text { change* } \end{aligned}$ | \%TAC <br> change | \%CATCH <br> change |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Management strategy | 390010 | $\mathrm{F}=0.14$ | $\begin{aligned} & 0.124(0.098, \\ & 0.162) \end{aligned}$ | $\begin{aligned} & \text { 2913.275(2134.794, } \\ & 3758.246) \end{aligned}$ | 0.157 | $\begin{aligned} & -4.778(- \\ & 30,23) \end{aligned}$ | -23.7 | -44 |
| $\begin{aligned} & \mathrm{F}_{\mathrm{MSY}} \times \text { SSB } \\ & (2024) / \mathrm{MSY} \\ & \mathrm{~B}_{\text {trigger }} \end{aligned}$ | 469953 | $F=0.157$ | $\begin{aligned} & \text { 0.151(0.121, } \\ & 0.198) \end{aligned}$ | $\begin{aligned} & \text { 2844.892(2097.299, } \\ & 3829.202) \end{aligned}$ | 0.208 | $\begin{aligned} & -7.013(- \\ & 31,25) \end{aligned}$ | -8.1 | -32 |
| Zero Catch | 0 | $\mathrm{F}=0.0$ | $0(0,0)$ | $\begin{aligned} & 3248.398(2530.315, \\ & 4265.225) \end{aligned}$ | 0.019 | $\begin{aligned} & \text { 6.175(- } \\ & \text { 17,39) } \end{aligned}$ | -100 | -100 |
| $\mathrm{F}_{\mathrm{pa}}$ | 487686 | $\mathrm{F}=0.157$ | $\begin{aligned} & \text { 0.157(0.123, } \\ & 0.203) \end{aligned}$ | $\begin{aligned} & \text { 2829.739(2113.425, } \\ & 3746.169) \end{aligned}$ | 0.212 | $\begin{aligned} & -7.509(- \\ & 31,22) \end{aligned}$ | -4.6 | -30 |
| $\mathrm{Fl}_{\text {im }}$ | 850085 | $\mathrm{F}=0.291$ | $\begin{aligned} & 0.291(0.226, \\ & 0.397) \end{aligned}$ | $\begin{aligned} & \text { 2521.394(1779.128, } \\ & 3521.787) \end{aligned}$ | 0.514 | $\begin{aligned} & 17.587(- \\ & 42,15) \end{aligned}$ | 66.3 | 23 |
| $\mathrm{SSB}_{2025}=\mathrm{B}_{\text {lim }}$ | 875324 | $F=0.299$ | $\begin{aligned} & 0.301(0.234, \\ & 0.416) \end{aligned}$ | $\begin{aligned} & \text { 2500.022(1788.411, } \\ & 3430.535) \end{aligned}$ | 0.529 | $\begin{aligned} & \text { 18.286(- } \\ & 42,12) \end{aligned}$ | 71.2 | 26 |
| $S S S B_{2025}=\mathrm{B}_{\mathrm{pa}}$ | 74667 | $F=0.100$ | $\begin{aligned} & 0.023(0.018, \\ & 0.028) \end{aligned}$ | $\begin{aligned} & 3184.058(2471.529, \\ & 4076.192) \end{aligned}$ | 0.029 | $\begin{aligned} & 4.072(- \\ & 19,33) \end{aligned}$ | -85.4 | -89 |
| Status quo | 570511 | $\mathrm{F}=0.189$ | $\begin{aligned} & 0.186(0.149, \\ & 0.241) \end{aligned}$ | $\begin{aligned} & \text { 2759.039(2042.021, } \\ & 3728.537) \end{aligned}$ | 0.277 | $\begin{aligned} & -9.82(- \\ & 33,22) \end{aligned}$ | 11.6 | -18 |

*95\% confidence interval

### 4.17 Figures



| $\square$ |
| :--- |
| $\square 3000$ tonnes |
| $\square$ |
| $\square$ |
| $10-300-3000$ tonnes |
|  |

Figure 4.2.1.1. Total reported landings (ICES estimates) of Norwegian spring-spawning herring in 2022 by ICES rectangle. Landings below $\mathbf{1 0}$ tonnes per statistical rectangle are not included.


Figure 4.2.1.2. Total reported landings (ICES estimates) of Norwegian spring-spawning herring in 2022 by quarter and ICES rectangle. Landings below 10 tonnes per statistical rectangle are not included.


Figure 4.4.3.1. Norwegian spring spawning herring. Age disaggregated landings in numbers plotted on a log scale. Age is on x -axis. The labels indicate year classes and grey lines correspond to $\mathrm{Z}=\mathbf{0 . 3}$.


Figure 4.4.4.1. Norwegian spring spawning herring. Mean weight at age by age groups 3-14 in the years 1981-2022 in the landings.


Figure 4.4.4.2. Norwegian spring-spawning herring. Mean weight at age in the stock by age groups 3-14 for the years 1981-2023.


Figure 4.4.5.1. Old assumed (blue line) and new back-calculated (orange line) maturity-at-age for the year 2018.


Figure 4.4.7.1. Distribution of Norwegian spring-spawning herring as measured during the IESNS survey in April-June 2023 in terms of NASC values $\left(\mathrm{m}^{2} / \mathrm{nm}^{2}\right)$.


Figure 4.4.7.2. Norwegian acoustic survey on the NSSH spawning grounds. Distribution and acoustic density of herring recorded in 2023.


Figure 4.4.7.3. Norwegian spring spawning herring. Age disaggregated abundance indices (millions) from the acoustic survey on the spawning area in February-March (Fleet 1) plotted on a log scale. The labels indicate year classes and grey lines correspond to $\mathrm{Z}=\mathbf{0 . 3}$. Age is on x -axis.


Figure 4.4.7.4. Norwegian spring spawning herring. Age disaggregated abundance indices (millions) from the acoustic survey in the feeding area in the Norwegian Sea in May (Fleet 5) plotted on a log scale. The labels indicate year classes and grey lines correspond to $\mathrm{Z}=\mathbf{0 . 3}$.


Figure 4.5.1.1. Estimated exploitation pattern for the years 1988-2023 by the XSAM model fit. All panels show the same data, but depicted at different angles to improve visibility at different time periods


Figure 4.5.1.2. Norwegian spring spawning herring. Correlation between estimated parameters in the final XSAM model fit.

Catch at age



Fleet1



Figure 4.5.1.3. Norwegian spring spawning herring. Weights (inverse of variance) of data-input of the final XSAM model fit.









Figure 4.5.1.4. Norwegian spring spawning herring. Standardized residuals type 1 (left) and type 2 (right) (see text) of data-input of the final XSAM model fit. Red is negative and blue is positive residuals.


Figure 4.5.1.5. Norwegian spring spawning herring. Observed vs. predicted values (left column) and qq-plot based on type 1 (middle) and type $\mathbf{2}$ (right) residuals (see text) based on the final XSAM model fit.


Figure 4.5.1.6. Norwegian spring spawning herring. Profiles of marginal log-likelihood $l_{M}$, the catch component $l_{C}$, Fleet 1 component $l_{F 1}$, Fleet 4 component $l_{F 4}$, Fleet 5 component $l_{F 5}$, point estimate of SSB and average $F$ (ages 5-12+) in 2023 over the common scaling factor for variance in data $h$ for the final XSAM fit. The red dots indicate the value of the respective scaling factors for which the log-likelihood is maximized.


Figure 4.5.1.7. Norwegian spring spawning herring. Retrospective XSAM model fits of SSB and weighted average of fishing mortality ages 5-12 for the years 2017-2023. Mohn's rho computed to be -0.06 for SSB and -0.02 for F.


Figure 4.5.1.8. Norwegian spring spawning herring. Point estimates of Spawning-stock biomass by years 1988-2023 from model (black lines) and by survey indices from Fleet 1 (blue) and Fleet 5 (red). Shaded area is approximate to standard deviation.


Figure 4.5.1.9. Total reported landings 1988-2022, estimated recruitment, weighted average of fishing mortality (ages 512) and spawning-stock biomass for the years 1988-2023 based on the final XSAM model fit.


$\qquad$


Figure 4.5.1.10. Norwegian spring-spawning herring. A visual representation of parameter estimates of the final XSAM model fit (see table 4.5.1.1). The estimates from the 2022 assessment are also shown (red).


Figure 4.5.1.11. Norwegian spring-spawning herring. Alternative runs showing the effect of leaving one fleet out. The $F$ is shown to the left and SSB to the right. The base run is shown as purple, leaving out Fleet 1 is red, leaving out Fleet 4 is green and leaving out Fleet 5 is shown as blue. Shaded regions show the standard deviation.


Figure 4.8.1.1. XSAM estimated selection pattern; selected years (estimates for 2017-2022 and predictions for 20232024) are shown in colours as indicated in the legend.


Figure 4.8.3.1. In upper panel, the input of the stock number at age for the $\mathbf{2 0 2 2}$ and the $\mathbf{2 0 2 3}$ forecast is seen as green and blue lines respectively. The stock number at age in 2022 from the 2023 assessment is the red line, and purple line presents the predicted stock number in 2023 from the 2022 forecast. Middle panel, the selection pattern used in the 2022 (red) and 2023 (blue) forecast. Bottom panel, the weight in stock used in the 2022 (red) and 2023 (blue) forecast.
Spawning Stock Biomass

WG 2011
WG 2012
WG 2013
WG 2014
WVG 2015
WG 2016
WG 2017
WG 2018
WG 2019
WG 2020
WG 2021
WG 2022
WG 2023
Recruitment at age 2


Figure 4.9.1. Norwegian spring spawning herring. Comparisons of spawning stock; weighted fishing mortality F (5-11/512+); and recruitment at age 2 with previous assessments. In 2016 the proportion mature in the years 2006-2011 was changed; recruitment age changed from 0 to 2 and fishing mortality is calculated over ages 5 to 11. In 2018 (WKNSSHREF) the age range for the fishing mortality changed to ages 5 to $12+$. The vertical dotted lines indicate the benchmark years 2008 and 2016.

## 5 Horse mackerel in Northeast Atlantic

## Trachurus trachurus

### 5.1 Fisheries in 2022

The total international catches of horse mackerel in the North East Atlantic are shown in Table 5.1.1. Since 2011, the southern horse mackerel stock is assessed by ICES WGHANSA. The total catch from all areas in 2022 for the Western and North Sea stocks was $78,491 \mathrm{t}$ which is $14,148 \mathrm{t}$ less than in 2021 and the 2nd lowest in the time series.

Ireland, Germany and the Netherlands have a directed trawl fishery and Norway and France a directed purse-seine fishery for horse mackerel. Spain has directed as well as mixed trawl and purse-seine fisheries targeting horse mackerel. In earlier years, most of the catches were used for meal and oil while in later years most of the catches have been used for human consumption.

The quarterly catches of North Sea and Western horse mackerel by Division and Subdivision in 2022 are given in Table 5.1.2 and the distributions of the fisheries are given in Figures 5.1.1.a-d. Note that the figures also include catches of southern horse mackerel but without Portuguese catches in 9a. The maps are based on data provided by Belgium, Denmark, France, Germany, Ireland, Netherlands, Norway, Poland, Portugal, Spain, Sweden and the UK. They represent all catches of North Sea and Western Horse Mackerel. The distribution of the fishery is similar to recent years with the highest catches taken in the $1^{\text {st }}$ and $4^{\text {th }}$ quarter. The historic catch by rectangle and year is also shown in section 1.10 of this report.
The Dutch, Danish, Irish and German fleets operated mainly in the North and West of Ireland and the Western waters off Scotland. The French fleet were in the Bay of Biscay and West Scotland whereas the Norwegian fleet fished in the North-eastern part of the North Sea. The Spanish fleet operated mainly in waters of Cantabrian Sea and Bay of Biscay.

First quarter: The fishing season with most of the catches $37,568 \mathrm{t}(48 \%$ of the total catch of the combined Western and North Sea horse mackerel catch). The fishery was mainly carried out west of Scotland and West and North of Ireland and along the Spanish coast (Figure 5.1.1.a).
Second quarter: $7,854 \mathrm{t}$. As usual, catches were significantly lower than in the first quarter as the second quarter is the main spawning period. Most of the catch were taken along the Spanish coast and Bay of Biscay. (Figure 5.1.1.b)
Third quarter: 11,574 t. Most of the catch were taken in Spanish waters, West of Ireland, in the Channel area and at the Norwegian coast (Figure 5.1.1.c).

Fourth quarter: Catches were 19,972 t ( $25 \%$ of the total catch). The catches were distributed in five main areas (Figure 5.1.1.d):

- Spanish waters,
- Western and Northern Irish waters and West of Scotland
- Norwegian coast
- Eastern part of the Channel
- Along the shelf edge of the Celtic Sea


### 5.2 Stock units

For many years the Working Group has considered the horse mackerel in the Northeast Atlantic as consisting of three separate stocks: the North Sea, the Southern and the Western stocks (ICES 1990, ICES 1991). For further information, see the Western Horse Mackerel Stock Annex and the WD document on horse mackerel stock structure (WD Brunel et al., 2016). The boundaries for the different stocks are given in Figure 5.2.1.

### 5.3 WG catch estimates

In 2017, a review of catch statistics for North Sea and Western horse mackerel stocks was carried out. The results of this report have been reported in previous Working Groups (Costas, 2017a).

As a result of this review, catches and catch-at-age of reported historical data of both North Sea and Western stocks of horse mackerel were updated (Figures 5.3.1 and 5.3.2). Catch statistics were reviewed since 1990 onward for the Western stock and since 2000 onward for the North Sea stock. The main mismatches between the catch statistics in working group reports and these reviewed data were due to several reasons such as late availability of some data for the report or the availability of official catch data only.
It is expected that the upcoming benchmark will also result in a review of the catch statistics for the North Sea and Western horse mackerel.

### 5.4 Allocation of catches to stocks

The distribution areas for the three stocks are given in the Stock Annex for the Western Horse Mackerel. The catches in 2021 were allocated to the three stocks as follows:

Western stock: $3^{\text {rd }}$ and $4^{\text {th }}$ quarters: Divisions 3.a and 4.a. Quarters 1-4: 2.a, 5.b, 6.a, 7.a-c, e-k and 8.a-e.

North Sea stock: $1^{\text {st }}$ and $2^{\text {nd }}$ quarters: Divisions 3.a and 4.a Quarters 1-4: Divisions 4.b, 4.c and 7.d.

Southern stock: Division 9.a. All catches from these areas were allocated to the southern stock. This stock is now dealt with by another working group (ICES WGHANSA).
The catches by stock are given in Table 5.4.1 and Figure 5.4.1. The catches by ICES sub-area and division for the Western and North Sea stocks for period 1982-2021 are shown in Figures 5.4.2-3. The catches by stock and countries for the period 1997-2021 are given in Table 5.4.2-5.4.3.
Recent genetic investigations show that the current boundaries might need to be newly evaluated in future (see also section 1.4.8.3 and section 5.10).

### 5.5 Estimates of discards

Only the Netherlands have provided data on discards over an extended period with occasional estimates from Germany and Spain. Since 2017 however, additional countries have provided estimates of discards with 8 countries reporting in 2022. However, following the introduction of the European landing obligation for the pelagic fisheries targeting horse mackerel in large areas of the overall fishing area and for Norwegian waters there is a general discard ban in place and discards in recent years have decreased substantially. The discard rate is estimated to be $4.8 \%$ in weight for the combined horse mackerel stocks. The discard rate for the North Sea stock is estimated to be $0.5 \%$ and for the Western stock $4.8 \%$ in 2022.

### 5.6 Trachurus species mixing

Three species of genus Trachurus: T. trachurus, T. mediterraneus and T. picturatus are found together and are commercially exploited in NE Atlantic waters. Following the Working Group recommendation (ICES 2002/ACFM: 06) special care was taken to ensure that catch and length distributions and numbers-at-age of T. trachurus supplied to the Working Group did not include T. mediterraneus and/or T. picturatus.

In general, T. mediterraneus, fishery takes place mainly in the eastern part of the ICES division. 8.c. There is no clear trend in the catches of T. mediterraneus in this area and they have been fairly stable over the last few years. (Table 5.6.1). Information on the T. picturatus fishery is available in the WGHANSA Report (Working Group on Horse Mackerel, Anchovy and Sardine).

Taking into account that the WGWIDE horse mackerel assessments are only made for T. trachurus, the Working Group recommends that the TACs and any other management regulations which might be established in the future should be related only to T. trachurus and not to Trachurus spp. More information is needed about the Trachurus spp. before the fishery and the stock can be evaluated.

### 5.7 Length distribution by fleet and by country:

France, Germany, Ireland, the Netherlands, UK (England, Northern Ireland and Scotland), Norway and Spain provided length distributions for their catches in 2022. The length distributions cover approximately $75 \%$ of the total landings of the Western and North Sea horse mackerel catches and are shown in Figure 5.7.1.

### 5.8 Comparing trends between areas and stocks

Horse mackerel (Trachurus trachurus) in the northeast Atlantic is assumed to consist of three separate stocks:

- $\quad$ North Sea (4a part of the year, $4 b, 4 c$ and 7 d )
- Western (4a part of the year, 5b, 6a, 7a-c,e-k, 8a-d)
- $\quad$ Southern (9a)

Catches between 2000 and 2022 are shown in figure 5.4 .1 and indicate an overall decline in the catches of horse mackerel since 2009.

A detailed analysis on the development of the catch by age data was presented to the 2017 working group (Pastoors, 2017). In this analysis it was indicated that there is an increase in the catches of juveniles in the Western and North Sea stocks in recent years. This could be an indication of a stronger recruitment of horse mackerel which has been reported by surveys and fishermen. However, it is also an alarming signal if a larger proportion of the catch consists of juveniles. These catches could be seen mostly in division 7.d and to a lesser extent, 7.e.

### 5.9 Quality and adequacy of fishery and sampling data

Table 5.9 .1 shows a summary of the overall sampling intensity on horse mackerel catches in recent years based on the InterCatch input. Since 2011 the Southern horse mackerel is dealt with by ICES WGHANSA.

Countries that historically sample are Ireland, the Netherlands, Germany, Norway and Spain, covering $42-100 \%$ of their respective catches. In 2020, due to the Covid pandemic, sampling
activities in some countries were hampered which lead to an overall lower sampling coverage for 2020. However, due to the fact that for the recent years it was possible to upload age samples taken from English vessels in the Netherlands for North Sea horse mackerel the proportion of sampling increased in comparison to the years before.

Table 5.9.2 shows the sampling intensity for the Western stock in 2022 and table 5.9.3 shows the sampling intensity for the North Sea stock in 2022 by country.
In 2022, France, Germany, Ireland, the Netherlands, Norway, Spain and UK (England), UK Scotland and UK (Northern Ireland) provided samples and length distributions and Germany, Ireland, the Netherlands, Norway, Spain and UK (England), and UK (Northern Ireland) provided also age distributions. However, the lack of age and length distribution data for relatively large portions of the horse mackerel catches continues to have a serious effect on the accuracy and reliability of the assessment and the Working Group remain especially concerned about the low number of fish which are aged.

An analysis of the sampling intensity was carried out for the period 2000-2019 for both the North Sea and the Western stock. Sampling intensity in fisheries can be defined as the ratio of sampled catch to the total catch. The precision and accuracy of sampled catch are of considerable importance to obtain a reliable estimate of the commercial catch. Sampled catch is used to extrapolate to total catch in order to obtain a catch-at-age (or at-length) and weight-at-age which are often used as inputs for the stock assessment models. In addition, in the case of horse mackerel the impact of temporal (quarter) and spatial (area by ICES division) factors have to be taken in account in order to obtain a reliable estimate of the commercial catches.

Figure 5.9.1 shows the proportion of sampled catches by division for the North Sea stock. In general, all ICES divisions show low levels of sampling, especially in recent years. In relation to age composition sampling level are dramatically lower in recent years (Figure 5.9.2) but due to the inclusion of samples of English vessels sampled in the Netherlands this situation improved substantially in the last two years. However, divisions that are usually not sampled can affect the precision and accuracy of total catch-at-age and weight-at-age. For the North Sea stock, samples were only available for area 4.a (Quarter 1,2), 4.c for quarters 1,3 and 4 and from $7 . \mathrm{d}$ from also the $1^{\text {st, }} 3^{\text {rd }}$ and $4^{\text {th }}$ quarter. Therefore, these estimates can be biased, especially, since samples are sometimes less than the recommended 100 fish per sample. (Table 5.9.1)

The proportion of the sampled catches by region for the Western stock are shown in figure 5.9.3. The general index of sampling intensity dropped in comparison to last year but remains on a good level (74\%).Divisions (regions) that are not sampled can affect the precision and accuracy of total catch-at-age and weight at age (Figure 9.5.4).

Length distributions were supplied by a number of countries. However, as some countries only deliver catch-at-age distributions and others only length distributions of the catch, the obtained catch-at-age and length distributions do not reflect the total catch especially in case of North Sea horse mackerel. Furthermore, some of the length distributions are only taken from discards of non-horse mackerel targeting fleets and omit the horse mackerel target fleet. This lack of coverage may also affect the accuracy and reliability of the assessment and is a matter of concern for the Working Group.

### 5.10 Recommendation on genetic stock identification

ICES has long considered horse mackerel in the northeast Atlantic to consist of three stocks, the separation of which was based on a variety of factors including the temporal and spatial distribution of the fisheries, the observed egg and larval distributions, information from acoustic and
trawl surveys and from parasite infestation rates. Further refinements of the definitions of stock units were based on the results from the EU-funded HOMSIR project (2000-2003).

However, initial results of a recent comprehensive genetic research project on stock identification of horse mackerel suggest that the boundaries of the stocks might require revision. This relates to the Northern boundary in 4a where the western horse mackerel stock might be present all year round, and not only in quarters 3 and 4 as samples analysed from Q1 and Q2 in 4a were determined to be western horse mackerel. Furthermore, analysed samples indicate that the southern boundary of the western stock might within division 9.a rather than between 8.c and 9.a. The western stock has also been shown to mix with the North Sea stock in division 7.d., in the commercial catches that have been analysed. Ongoing and new studies are being undertaken to investigate these issues. Potential changes in the perception of the stock distributions could impact the reliability of the assessments for the three current stocks of horse mackerel in the Northeast Atlantic (see also section 1.4.8.3 for details).

Based on this genetic investigations WGWIDE recommends that the stock identification methods working group (SIMWG) conduct a review of ongoing genetic research into the stock structure of horse mackerel.

### 5.11 References

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### 5.12 Tables

Table 5.1.1 HORSE MACKEREL general. Catches ( t ) by Sub-area. Data as submitted by Working Group members. Data of limited discard information are only available for some years.

| Sub-area | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 2 | - | + | - | 412 | 23 | 79 | 214 |
| $4+3 . a$ | 1,412 | 2,151 | 7,245 | 2,788 | 4,420 | 25,987 | 24,238 | 20,746 |
| 6 | 7,791 | 8,724 | 11,134 | 6,283 | 24,881 | 31,716 | 33,025 | 20,455 |
| 7 | 43,525 | 45,697 | 34,749 | 33,478 | 40,526 | 42,952 | 39,034 | 77,628 |
| 8 | 47,155 | 37,495 | 40,073 | 22,683 | 28,223 | 25,629 | 27,740 | 43,405 |
| 9 | 37,619 | 36,903 | 35,873 | 39,726 | 48,733 | 23,178 | 20,237 | 31,159 |
| Total | 137,504 | 130,970 | 129,074 | 104,958 | 147,195 | 149,485 | 144,353 | 193,607 |


| Sub-area | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 3,311 | 6,818 | 4,809 | 11,414 | 3200 | 13457 | 0 | 759 |
| $4+3 . a$ | 20,895 | 62,892 | 112,047 | 145,062 | 71,195 | 120,054 | 145,965 | 111,899 |
| 6 | 35,157 | 45,842 | 34,870 | 20,904 | 29,726 | 39,061 | 65,397 | 69,616 |
| 7 | 100,734 | 90,253 | 138,890 | 192,196 | 150,575 | 183,458 | 202,083 | 196,192 |
| 8 | 37,703 | 34,177 | 38,686 | 46,302 | 42,840 | 54,172 | 44,726 | 35,501 |
| 9 | 24,540 | 29,763 | 29,231 | 24,023 | 34,992 | 27,858 | 31,521 | 28,442 |
| Disc | 222,340 | 269,745 | 358,533 | 439,901 | 337,968 | 440,280 | 499,222 | 446,974 |
| Total |  |  |  |  | 5,440 | 2,220 | 9,530 | 4,565 |


| Sub-area | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 13151 | 3366 | 2601 | 2544 | 2557 | 919 | 310 | 1324 |
| $4+3 . a$ | 100,916 | 25,998 | 79,761 | 34,917 | 58,745 | 31,435 | 18,513 | 52,337 |
| 6 | 83,568 | 81,311 | 40,145 | 35,073 | 40,381 | 20,735 | 24,839 | 14,843 |
| 7 | 328,995 | 263,465 | 326,469 | 300,723 | 186,622 | 140,190 | 138,428 | 98,677 |
| 8 | 28,707 | 48,360 | 40,806 | 38,571 | 48,350 | 54,197 | 75,067 | 55,897 |
| 9 | 25,147 | 20,400 | 29,491 | 41,574 | 27,733 | 26,160 | 24,912 | 23,665 |
| Disc | 2,076 | 17,082 | 168 | 996 | 0 | 385 | 254 | 307 |
| Total | 582,560 | 459,982 | 519,441 | 454,398 | 364,388 | 274,022 | 282,323 | 247,049 |


| Sub-area | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ | $\mathbf{2 0 0 7}$ | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 0 9}$ | $\mathbf{2 0 1 0}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2 | 36 | 42 | 176 | 27 | 366.34 | 572 | 1847 | 1667 |
| $4+3 . a$ | 34,095 | 30,736 | 40,594 | 37,583 | 16,226 | 15,628 | 78,064 | 13,600 |
| 6 | 23,772 | 22,177 | 22,053 | 15,722 | 25,949 | 25,867 | 17,775 | 23,199 |
| 7 | 123,428 | 115,739 | 106,671 | 101,183 | 93,013 | 102,755 | 96,915 | 148,701 |
| 8 | 41,711 | 24,126 | 41,491 | 34,121 | 28,396 | 33,756 | 33,580 | 39,659 |
| 9 | 19,570 | 23,581 | 23,111 | 24,557 | 23,423 | 23,596 | 26,496 | 27,217 |
| Disc | 842 | 2,356 | 1,864 | 1,431 | 509 | 474 | 1,483 | 434 |
| Total | 243,455 | 218,758 | 235,961 | 214,624 | 187,882 | 202,649 | 256,161 | 254,478 |


| Sub-area | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 647.588 | 66.02912 | 30 | 424.291 | 10 | 45.276 | 5 | 718 |
| $4+3 . a$ | 25,158 | 5,234 | 8,183 | 17,270 | 10,560 | 11,565 | 12,609 | 11,758 |
| 6 | 39,496 | 44,971 | 43,266 | 32,444 | 24,153 | 32,186 | 28,170 | 38,896 |
| 7 | 120,340 | 120,476 | 100,859 | 66,853 | 49,644 | 46,901 | 33,297 | 38,816 |
| 8 | 35,245 | 17,209 | 26,983 | 30,844 | 19,822 | 17,511 | 18,307 | 23,393 |
| $9^{1}$ | 22,575 | 25,316 | 29,382 | 29,205 | 33,179 | 41,081 | 37,080 | 31,920 |
| Disc | 430 | 3,279 | 4,582 | 1,904 | 6,232 | 5,944 | 5,488 | 2,873 |
| Total | 243,892 | 216,552 | 213,285 | 178,945 | 143,600 | 155,232 | 134,956 | 148,374 |
| Sub-area |  | 2019 |  |  | 2021 |  | 2022 |  |
| 2 |  | 867 | 29 |  | 12 |  | 157* |  |
| $4+3 . a$ |  | 12,593 |  |  | 7,672 |  | 5,102 |  |
| 6 |  | 47,351 |  |  | 13,727 |  | 12,016 |  |
| 7 |  | 42,973 |  |  | 49,934 |  | 38,285 |  |
| 8 |  | 29,640 |  |  | 19,602 |  | 17125 |  |
| $9^{1}$ |  | 34,080 |  |  | 26,745 |  | 24,997 |  |
| Disc |  | 3,326 |  |  | 1,692 |  | 3,786 |  |
| Total |  | 170,829 |  |  | 119,384 |  | 103,488 |  |

${ }^{1}$ - Southern Horse Mackerel (ICES Division 9) is assessed by ICES WGHANSA since 2011,

*     - 10t of landings in 5 b included

Table 5.1.2 HORSE MACKEREL Western and North Sea Stock combined.
Quarterly catches ( t ) by Division and Subdivision in 2022.

| Division | 1Q | 2Q | 3Q | 4Q | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2.a+5.b | 51 | 69 | 16 | 19 | 157 |
| 3 | 0 | 0 | 29 | 190 | 220 |
| 4.a | 1357 | 276 | 1638 | 1610 | $5596{ }^{1}$ |
| 4.bc | 10 | 270 | 1244 | 521 | $2065{ }^{2}$ |
| 7.d | 1452 | 188 | 974 | 2043 | $4678{ }^{3}$ |
| 6.a,b | 8925 | 90 | 5 | 2545 | $12329{ }^{4}$ |
| 7.a-c,e-k | 21701 | 331 | 2550 | 9885 | 34467 |
| 8.a-e | 4072 | 6633 | 5117 | 3158 | 18979 |
| Sum | 37568 | 7858 | 11574 | 19972 | $78491{ }^{5}$ |

${ }^{1}$ for the total 714 t were added which were only declared as yearly catch
${ }^{2}$ for the total $20 t$ were added which were only declared as yearly catch
${ }^{3}$ for the total $22 t$ were added which were only declared as yearly catch
${ }^{4}$ for the total 764t were added which were only declared as yearly catch
${ }^{5}$ for the total 1519 t were added which were only declared as yearly catch

Table 5.4.1 HORSE MACKEREL General. Landings and discards ( $t$ ) by year and ICES Division, for the North Sea, Western, and Southern horse mackerel stocks. (Data submitted by Working Group members.)

| Year | 3.a | 4.a | 4.b,c | 7.d | Disc | NS <br> Stock | 2.a 5.b | 3.a | 4.a | 6.a,b | $\begin{aligned} & \text { 7.a-c, e- } \\ & \text { k } \end{aligned}$ | 8.a-e | Disc | Western Stock | W + NS <br> Stock | Southern <br> Stock(9.a) ${ }^{\text {x }}$ | All Stocks |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 2,788* |  | - | 1,247 |  | 4,035 | - |  | - | 6,283 | 32,231 | 3,073 | - | 61,197 | 65,232 | 39,726 | 104,958 |
| 1983 | 4,420* |  | - | 3,600 |  | 8,020 | 412 |  | - | 24,881 | 36,926 | 28,223 | - | 90,442 | 98,462 | 48,733 | 147,195 |
| 1984 | 25,893* |  | - | 3,585 |  | 29,478 | 23 |  | 94 | 31,716 | 38,782 | 25,629 | 500 | 96,744 | 126,222 | 23,178 | 149,400 |
| 1985 | - |  | 22,897 | 2,715 |  | 26,750 | 79 |  | 203 | 33,025 | 35,296 | 27,740 | 7,500 | 103,843 | 129,455 | 20,237 | 150,830 |
| 1986 | - |  | 19,496 | 4,756 |  | 24,648 | 214 |  | 776 | 20,343 | 72,761 | 43,405 | 8,500 | 145,999 | 170,251 | 31,159 | 201,806 |
| 1987 | 1,138 |  | 9,477 | 1,721 |  | 11,634 | 3,311 |  | 11,185 | 35,197 | 99,942 | 37,703 | - | 187,338 | 199,674 | 24,540 | 223,512 |
| 1988 | 396 |  | 18,290 | 3,120 |  | 23,671 | 6,818 |  | 42,174 | 45,842 | 81,978 | 34,177 | 3,740 | 214,729 | 236,535 | 29,763 | 268,163 |
| 1989 | 436 |  | 25,830 | 6,522 |  | 33,265 | 4,809 |  | 85304** | 34,870 | 131,218 | 38,686 | 1,150 | 296,037 | 328,825 | 29,231 | 358,533 |
| 1990 | 2,261 |  | 17,437 | 1,325 |  | 18,762 | 11,414 | 14,878 | 112753** | 20,794 | 182,580 | 46,302 | 9,930 | 398,645 | 419,668 | 24,023 | 441,430 |
| 1991 | 913 | 0 | 11,400 | 600 | 0 | 12,913 | 3,200 | 2,725 | 56,157 | 29,726 | 149,975 | 42,840 | 5,440 | 290,063 | 302,976 | 34,992 | 337,968 |
| 1992 | 0 | 0 | 13,955 | 688 | 400 | 15,043 | 13,457 | 2,374 | 103,725 | 39,061 | 182,770 | 54,172 | 1,820 | 397,379 | 412,422 | 27,858 | 440,280 |


| Year | 3.a | 4.a | 4.b,c | 7.d | Disc | NS <br> Stock | 2.a 5.b | 3.a | 4.a | 6.a,b | $\begin{aligned} & \text { 7.a-c, e- } \\ & \text { k } \end{aligned}$ | 8.a-e | Disc | Western Stock | W + NS <br> Stock | Southern <br> Stock(9.a) ${ }^{\mathrm{x}}$ | All Stocks |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0 | 0 | 3,895 | 8,792 | 930 | 13,617 | 0 | 850 | 141,220 | 65,397 | 193,291 | 44,726 | 8,600 | 454,084 | 467,701 | 31,521 | 499,222 |
| 1994 | 0 | 0 | 2,496 | 2,503 | 630 | 5,629 | 759 | 2,492 | 106,911 | 69,616 | 193,689 | 35,501 | 3,935 | 412,903 | 418,532 | 28,442 | 446,974 |
| 1995 | 112 | 0 | 7,948 | 8,666 | 30 | 16,756 | 13,151 | 128 | 92,728 | 83,568 | 320,329 | 28,707 | 2,046 | 540,657 | 557,413 | 25,147 | 582,560 |
| 1996 | 1,657 | 0 | 7,558 | 9,416 | 212 | 18,843 | 3,366 | 0 | 16,783 | 81,311 | 254,049 | 48,360 | 16,870 | 420,739 | 439,582 | 20,400 | 459,982 |
| 1997 | 0 | 0 | 14,078 | 5,452 | 10 | 19,540 | 2,601 | 2,037 | 63,646 | 40,145 | 321,017 | 40,806 | 158 | 470,410 | 489,950 | 29,491 | 519,441 |
| 1998 | 3,693 | 0 | 10,530 | 16,194 | 83 | 30,500 | 2,544 | 3,693 | 17,001 | 35,073 | 284,529 | 38,571 | 913 | 382,324 | 412,824 | 41,574 | 454,398 |
| 1999 | 0 | 0 | 9,335 | 27,889 | 0 | 37,224 | 2,557 | 2,095 | 47,315 | 40,381 | 158,733 | 48,350 | 0 | 299,431 | 336,655 | 27,733 | 364,388 |
| 2000 | 0 | 176 | 25,931 | 19,019 | 4 | 45,130 | 919 | 1,014 | 4,314 | 20,735 | 121,171 | 54,197 | 382 | 202,732 | 247,862 | 26,160 | 274,022 |
| 2001 | 43 | 212 | 6,686 | 21,390 | 0 | 28,331 | 310 | 134 | 11,438 | 24,839 | 117,038 | 75,067 | 254 | 229,081 | 257,411 | 24,912 | 282,323 |
| 2002 | 0 | 639 | 15,303 | 11,323 | 0 | 27,264 | 1,324 | 174 | 36,221 | 14,843 | 87,354 | 55,897 | 307 | 196,120 | 223,384 | 23,665 | 247,049 |
| 2003 | 49 | 622 | 10,309 | 21,049 | 0 | 32,028 | 36 | 1,843 | 21,272 | 23,772 | 102,379 | 41,711 | 842 | 191,856 | 223,885 | 19,570 | 243,455 |
| 2004 | 303 | 133 | 18,544 | 16,455 | 0 | 35,435 | 42 | 48 | 11,708 | 22,177 | 99,284 | 24,126 | 2,356 | 159,742 | 195,177 | 23,581 | 218,758 |


| Year | $3 . \mathrm{a}$ | 4.a | 4.b,c | 7.d | Disc | NS Stock | 2.a 5.b | 3.a | 4.a | 6.a,b | $\begin{aligned} & \text { 7.a-c, e- } \\ & \text { k } \end{aligned}$ | 8.a-e | Disc | Western Stock | W + NS <br> Stock | Southern Stock(9.a) ${ }^{\mathrm{x}}$ | All Stocks |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 0 | 1,331 | 13,995 | 15,460 | 62 | 30,848 | 176 | 284 | 24,983 | 22,053 | 91,211 | 41,491 | 1,802 | 182,001 | 212,850 | 23,111 | 235,961 |
| 2006 | 185 | 2,192 | 7,996 | 23,789 | 78 | 34,240 | 27 | 58 | 27,152 | 15,722 | 77,394 | 34,121 | 1,353 | 155,827 | 190,067 | 24,557 | 214,624 |
| 2007 | 11 | 2,051 | 9,114 | 29,789 | 139 | 41,103 | 366 | 110 | 4,940 | 25,949 | 63,224 | 28,396 | 370 | 123,356 | 164,459 | 23,423 | 187,882 |
| 2008 | 27 | 910 | 2,582 | 32,185 | 0 | 35,704 | 572 | 3 | 12,107 | 25,867 | 70,570 | 33,756 | 474 | 143,349 | 179,053 | 23,596 | 202,649 |
| 2009 | 21 | 314 | 18,975 | 25,537 | 1,036 | 45,883 | 1,847 | 17 | 58,738 | 17,775 | 71,378 | 33,580 | 447 | 183,782 | 229,665 | 26,496 | 256,161 |
| 2010 | 0 | 100 | 1,969 | 22,077 | 2 | 24,149 | 1,667 | 88 | 11,442 | 23,199 | 126,624 | 39,659 | 432 | 203,112 | 227,261 | 27,217 | 254,478 |
| 2011 | 0 | 0 | 10,435 | 17,184 | 0 | 27,619 | 648 | 0 | 14,723 | 39,496 | 103,156 | 35,245 | 430 | 193,698 | 221,317 | 22,575 | 243,892 |
| 2012 | 0 | 355 | 1,559 | 19,464 | 0 | 21,378 | 66 | 9 | 3,311 | 44,971 | 101,012 | 17,209 | 3,279 | 169,858 | 191,236 | 25,316 | 216,552 |
| 2013 | 0 | 17 | 1,453 | 17,175 | 0 | 18,645 | 30 | 10 | 6,702 | 43,266 | 83,684 | 26,983 | 4,582 | 165,258 | 183,903 | 29,382 | 213,285 |
| 2014 | 1 | 2 | 2,597 | 10,772 | 7 | 13,380 | 424 | 4,096 | 10,573 | 32,444 | 56,081 | 30,844 | 1,896 | 136,360 | 149,740 | 29,205 | 178,945 |
| 2015 | 3 | 644 | 770 | 8,581 | 2,004 | 12,002 | 10 | 65 | 9,078 | 24,153 | 41,063 | 19,822 | 4,228 | 98,419 | 110,421 | 33,179 | 143,600 |
| 2016 | 2 | 1,628 | 975 | 11,209 | 1,527 | 15,341 | 45 | 0 | 8,960 | 32,186 | 35,692 | 17,511 | 4,417 | 98,811 | 114,151 | 41,081 | 155,232 |


| Year | $3 . \mathrm{a}$ | 4.a | 4.b,c | 7.d | Disc | NS Stock | 2.a 5.b | 3.a | 4.a | 6.a,b | $\begin{aligned} & \text { 7.a-c, e- } \\ & \text { k } \end{aligned}$ | 8.a-e | Disc | Western Stock | w + NS <br> Stock | Southern <br> Stock(9.a) ${ }^{\mathrm{x}}$ | All Stocks |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2017 | 0 | 22 | 2,557 | 10,787 | 1,213 | 14,579 | 5 | 697 | 9,332 | 28,170 | 22,510 | 18,307 | 3,939 | 82,961 | 97,540 | 37,088 | 134,956 |
| 2018 | 0 | 1,418 | 1,413 | 11,677 | 265 | 14,773 | 718 | 380 | 8,547 | 38,896 | 27,140 | 23,393 | 2,609 | 101,683 | 116,456 | 31,920 | 148,376 |
| 2019 | 0.5 | 2,571 | 1,217 | 7,829 | 185 | 11,803 | 867 | 490 | 8,314 | 47,351 | 35,144 | 29,640 | 3,141 | 124,947 | 136,750 | 34,080 | 170,830 |
| 2020 | 0 | 2,211 | 1,099 | 9,077 | 201 | 12,587 | 290 | 96 | 10,387 | 19,037 | 24,232 | 19,359 | 2,741 | 76,422 | 89,009 | 31,344 | 120,347 |
| 2021 | 1 | 2,270 | 1,639 | 7,120 | 52 | 11,082 | 12 | 12 | 3,751 | 13,727 | 42,813 | 19,602 | 1,641 | 81,557 | 92,639 | 26,745 | 119,384 |
| 2022 | 0 | 1,632 | 2,020 | 4,678 | 435 | 8,377 | 157 | 191 | 3,280 | 12,016 | 33,994 | 17,125 | 3,352 | 70,114 | 78,491 | 24,997 | 103,488 |

*Divisions 3.a and 4.b,c combined•
**Norwegian catches in 4.b included in Western horse mackerel-
${ }^{\mathrm{x}}$ Southern Horse Mackerel is assessed by ICES WGHANSA since 2011

Table 5.4.2 National catches of the Western Horse mackerel stock.

| Country | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 18 | 19 | 21 | 0 | - | - | - | - | - |
| Denmark | 62,897 | 31,023 | 26,040 | 16,385 | 21,254 | 10,147 | 11340 | 11,667 | 10,155 |
| Estonia | 78 | 22 | - | 0 | - | - | - | 3,826 | 3,695 |
| Faroe Islands | 1,095 | 216 | 1,040 | 24 | 800 | 671 | 4 | 8,056 | 10,690 |
| France | 39,188 | 26,667 | 25,141 | 20,457 | 15,145 | 18,951 | 10,381 | 17,744 | 16,364 |
| Germany | 28,533 | 33,716 | 23,549 | 13,014 | 11,491 | 12,658 | 15,696 | 26,432 | 34,607 |
| Ireland | 74,250 | 73,672 | 57,983 | 55,229 | 51,874 | 36,422 | 35,857 | - | - |
| Lithuania | - | - | - | - | - | - | - | 40986 | 41,057 |
| Netherlands | 82,885 | 103,246 | 83,450 | 57,261 | 73,440 | 44,997 | 48,924 | 10729 | 24,909 |
| Norway | 45,058 | 13,363 | 46,648 | 1,982 | 7,956 | 36,164 | 20,371 | 16,272 | 16,636 |
| Russia | 554 | 345 | 121 | 80 | 16 | 3 | 2 | 567 | 216 |
| Spain | 31,087 | 43,829 | 39,831 | 24,204 | 23,537 | 24,763 | 24,599 | 4,617 | 3,560 |
| Sweden | 1,761 | 3411 | 1,957 | 1009 | 68 | 561 | 1,002 | 458 | 210 |
| UK (Engl. + Wales) | 19,778 | 13,068 | 9,268 | 4,554 | 7,096 | 5,970 | 4,438 | 1,522 | 143 |
| UK (N. Ireland) | - | 1,158 | - | 625 | 1140 | 1129 | 914 | 14,506 | 17,962 |
| UK (Scotland) | 32,865 | 18,283 | 11,197 | 10,283 | 8,026 | 2,905 | 721 | 2356 | 1802 |
| Unallocated | 17,158 | 15,262 | 23,763 | -2757 | 6,978 | 472 | 16,765 | 159,737 | 182,006 |
| Discard | 158 | 913 | - | 382 | 254 | 307 | 842 | - | - |
| Total | 437,363 | 378,213 | 350,009 | 202,732 | 229,075 | 196,120 | 191,856 | 11,667 | 10,155 |


| Country | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | - | - | - | - | 19 | 2 | 0.2 | 14 |
| Denmark | 8,411 | 7,617 | 5,261 | 6,027 | 5,940 | 6,108 | 4,002 | 6,820 |
| Faroe Islands | - | 478 | 841 | - | 377 | 349 | - |  |
| France | 11,031 | 12,748 | 12,626 | - | 260 | 8,271 | 1,797 | 3,595 |
| Germany | 10,862 | 5,784 | 11,801 | 15,122 | 17,688 | 21,114 | 17,063 | 24,835 |
| Ireland | 26,779 | 29,759 | 35,332 | 40,754 | 44,488 | 38,466 | 45,239 | 35,791 |
| Lithuania | 6,828 | 5,467 | 5,548 | - | - | - | - |  |
| Netherlands | 37,130 | 29,462 | 43,648 | 39,453 | 61,504 | 55,690 | 66,396 | 53,697 |
| Norway | 27,114 | 4,182 | 12,223 | 59,764 | 11,978 | 13,755 | 3,251 | 6,596 |
| Spain | 13,877 | 14,277 | 19,851 | 21,077 | 38,745 | 34,581 | 13560 | 22,541 |
| Sweden | - | 76 | 8 | 258 | 2 | 90 | - | 1 |
| UK (Engl. + Wales) | 3,574 | 5,482 | 3,365 | 6,482 | 12,714 | 11,716 | 12,122 | 3,959 |
| UK (N. Ireland) | 103 | - | - | - | 59 | 198 | - | 2,325 |
| UK (Scotland) | 468 | 776 | 1,077 | 1,412 | 2,349 | 2,928 | 1,335 | 504 |
| Unallocated | 8,292 | 6,878 | $-8,703$ | -7,014 | 6,556 | - | 1815 | - |
| Discard | 1353 | 370 | 474 | 447 | 432 | 430 | 3,280 | 4,582 |
| Total | 155,822 | 123,356 | 143,352 | 183,782 | 203,111 | 193,698 | 169,860 | 165,260 |


| Country | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium |  |  |  |  |  |  |  |  |  |
| Denmark | 5,945 | 4,556 | 321 | 4,541 | 6,302 | 7,764 | 5,487 | 6,042 | 5,265 |
| Faroe Islands | 68 | - | - | 180 | - | 26 | - |  | 106 |
| France | 3,428 | 3,247 | 2,797 | 3,923 | 3,443 | 4,382 | 2,217 | 2,710 | 2,613 |
| Germany | 17,161 | 9,417 | 11,414 | 7,172 | 4,734 | 9,211 | 954 | 5,530 | 4,263 |
| Ireland | 32,667 | 21,654 | 27,605 | 23,560 | 25,347 | 28,899 | 17,390 | 18,770 | 15,799 |
| Lithuania | - | - | 2,596 | - | - | - | 0 |  | - |
| Netherlands | 25,053 | 24,958 | 23,792 | 14,269 | 25,942 | 29,656 | 14,240 | 20,786 | 18,474 |
| Norway | 14,353 | 8,897 | 9,438 | 9,885 | 9,319 | 9,021 | 10,666 | 3,663 | 3,131 |
| Poland | - | - | -- | - | - | 127 | 1,002 | 1,605 | - |
| Spain | 19,442 | 13,071 | 14,235 | 14,901 | 20,362 | 25,776 | 18,582 | 16,191 | 13,424 |
| Sweden | 0 | 10 | - | 41 | 23 | 323 | 83 | 4 | 3 |
| UK (Engl. + Wales) | 4,832 | 2,063 | 842 | 549 | 2,443 | 4,036 | 1,496 | 2,651 | 2,316 |
| UK (N. Ireland) | 1,579 | 1,204 | - |  | 1,080 | 1,907 | 1,231 | 1,350 | 811 |
| UK (Scotland) | 1,389 | 738 | 970 | - | - | 678 | 333 | 615 | 558 |
| Unallocated | 8,545 | 4,377 | 1,010 | 3,994 | 74 | - | - | - | - |
| Discard | 1,896 | 4,228 | 4,417 | 3,928 | 2,609 | 3,141 | 2,741 | 1,641 | 3,352 |
| Total | 136,360 | 98,419 | 98,810 | 82,950 | 101,682 | 124,947 | 76,422 | 81,557 | 70,114 |

Table 5.4.3. National catches of the North Sea Horse mackerel stock.

| Country | $\mathbf{1 9 9 7}$ | $\mathbf{1 9 9 8}$ | $\mathbf{1 9 9 9}$ | $\mathbf{2 0 0 0}$ | $\mathbf{2 0 0 1}$ | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 3}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 5}$ | $\mathbf{2 0 0 6}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium | - | 19 | 21 |  | 30 | 5 | 4 | 4 | - |  |
| Denmark | 180 | 1,481 | 3,377 | 4,403 | 885 | 2,315 | 3,301 | 8,690 | 3,987 | 8,353 |
| Faroe Islands | - | - | 135 | - | - | 28 | 804 | 21 | - | - |
| France | 3,246 | 2,399 | - | - |  | 1,246 | 2,326 | 231 | 5,236 | 1,205 |
| Germany | 7,847 | 5,844 | 5,920 | 3,728 | 974 | 6,532 | 2,936 | 5,194 | 2,725 | 11,034 |
| Ireland | - | 2,861 | 27 | 201 | 338 | 61 | - | 1 | 753 | 10,863 |
| Lithuania | - | 10,711 | - | - | - | - | - | - | - | 26,779 |
| Netherlands | 36,855 | - | 8,117 | 8,697 | 13,867 | 12,209 | 24,119 | 26,303 | 27,730 | 6,829 |
| Norway | - | - | 238 | 105 | 36 | 525 | 144 | 22 | 204 | 37,130 |


| Country | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sweden | - | 3,401 | 5 | 40 | 46 | 16 | 72 | 98 | 4 | 27,114 |
| UK (Engl. + Wales) | 269 | 907 | 11 | 1,585 | 3,425 | 2,322 | 1,966 | 5,633 | 3,859 | - |
| UK (Scotland) | 29 | - - | - 4 | 421 | - | 2 | 1 | 2 | - | 13,878 |
| Unallocated | -28,896 | 2,794 | 19,373 | 25,944 | 48,805 | 1,981 | -3,645 | -13,064 | -13,719 | - |
| Discard | 10 | 83 | - 4 | 4 | - |  | - | - | 62 | 3,583 |
| Total | 19,540 | 30,500 | 37,224 | 45,128 | 8 28,376 | 27,267 | 32,029 | 33,135 | 30,845 | 155,094 |
| Country | 2006 | 2007 | 2008 |  | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| Belgium |  |  |  |  | 4 | 16 |  | 46 | 51.077 | 74 |
| Denmark | 1,283 | 252 | 57 |  | 72 | 15 | 142 | 1514 | 1,020 | 552 |
| Faroe Islands | - | - | - | - | - - | - | - | 0 |  |  |
| France | 4,380 | 5,349 | 2,247 | , | - 8 | 813 | 273 | 1,047 | 1,010 | 1,742 |
| Germany | 1,125 | 65 | 1,081 |  | 1,539 | 3,794 | 3,461 | 5,356 | 2,941 | 1,619 |
| Ireland | 2,077 |  | 887 |  | 25 | - | - | 0 |  | 0 |
| Lithuania | 1,999 | 297 | - | - | - - | - | - | 0 |  | 0 |
| Netherlands | 27,285 | 31,153 | 319,439 |  | 22,546 | 17,093 | 16,289 | 12,157 | 8,725 | 4,925 |
| Norway | 113 | 1,243 | 21 |  | 12,855 | 526 | 7,359 | 129 | 377 | 0 |
| Sweden | 9 | 21 | 36 |  | 401 | - | - | 0 |  | 1 |
| UK (Engl. + Wales) | 595 | 6921 | 1,061 |  | 1,435 | 1,890 |  | 935 | 4,401 | 4,198 |
| UK (Scotland) | 300 | 625 | 7 |  | 4 | 111 | 93 | 240 | 172 | 262 |
| Unallocated | -5,004 | -4,960 | 10,869 |  | 5,964 | -116 | 0 | 0 | 0 |  |
| Discard | 78 | 139 | - |  | 1,036 | 2 | 0 | 0 | 0 | 7 |
| Total | 34,240 | 41,105 | 535,705 |  | 45,881 | 24,144 | 27,617 | 21,424 | 18,696 | 13,380 |
| Country | 2015 | 2016 | 2017 |  | 2018 | 2019 | 2020 |  |  | 2022 |
| Belgium | 63 | 51 | 67 |  | 44 | 18 | 39 | 38 |  | 30 |
| Denmark | 800 | 268 | 294 |  | 397 | 100 | 177 | 72 |  | 37 |
| Faroe Islands | 0 | 0 | 4 |  | 0 | 10 | 109 |  |  |  |
| France | 934 | 1,322 | 1,863 |  | 1,443 | 935 | 758 | 50 |  | 433 |


| Germany | 644 | 1,879 | 949 | 2,766 | 946 | 3 | 87 | 67 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ireland | 0 | 0 | 0 | 0 | 0 | 0 | 174 | 0 |
| Lithuania | 0 | 0 | 0 | 0 | 1,254 | 0 | 0 | 0 |
| Netherlands | 3,305 | 3,892 | 5,638 | 5,184 | 2,089 | 4,803 | 3,377 | 2,490 |
| Norway | 662 | 1,701 | 5 | 1,423 | 2,543 | 2,090 | 2,091 | 1,640 |
| Sweden | 9 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| UK (Engl. + Wales) | 3,581 | 4,697 | 4,546 | 3,250 | 3,632 | 4,381 | 4,669 | 3,203 |
| UK (Northern Ireland) | 0 | 0 | 0 | 0 | 53 | 0 | 0 | 0 |
| UK (Scotland) | 0 | 0 | 0 | 0 | 38 | 24 | 19 | 42 |
| Unallocated | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Discard | 2,004 | 1,527 | 1,213 | 265 | 185 | 201 | 52 | 435 |
| Total | 12,002 | 15,337 | 14,579 | 14,773 | 11,802 | 12,587 | 11,082 | 8,377 |

Table 5.6.1. Catches ( t ) of Trachurus mediterraneus in Divisions 8.ab, 8.c and Sub-Area 7

|  | 7 | 8.ab | 8.c East | 8.c West | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 0 | 23 | 3903 |  | 3926 |
| 1990 | 0 | 298 | 2943 |  | 3241 |
| 1991 | 0 | 2122 | 5020 |  | 7142 |
| 1992 | 0 | 1123 | 4804 |  | 5927 |
| 1993 | 0 | 649 | 5576 |  | 6225 |
| 1994 | 0 | 1573 | 3344 |  | 4917 |
| 1995 | 0 | 2271 | 4585 |  | 6856 |
| 1996 | 0 | 1175 | 3443 |  | 4618 |
| 1997 | 0 | 557 | 3264 |  | 3821 |
| 1998 | 0 | 740 | 3755 |  | 4495 |
| 1999 | 0 | 1100 | 1592 |  | 2692 |
| 2000 | 59 | 988 | 808 |  | 1854 |
| 2001 | 1 | 525 | 1293 |  | 1820 |
| 2002 | 1 | 525 | 1198 |  | 1724 |
| 2003 | 0 | 340 | 1699 |  | 2039 |


|  | 7 | 8.ab | 8.c East | 8.c West | TOTAL |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 0 | 53 | 841 |  | 894 |
| 2005 | 1 | 155 | 1005 |  | 1162 |
| 2006 | 1 | 168 | 794 |  | 963 |
| 2007 | 0 | 126 | 326 |  | 452 |
| 2008 | 0 | 82 | 405 |  | 487 |
| 2009 | 0 | 42 | 1082 |  | 1124 |
| 2010 | 0 | 97 | 370 |  | 467 |
| 2011 | 0 | 119 | 1096 |  | 1225 |
| 2012 | 0 | 186 | 667 | 116 | 969 |
| 2013 | 0 | 52 | 238 | 0 | 290 |
| 2014 | 0 | 130 | 1160 | 0 | 1290 |
| 2015 | 0 | 8 | 890 | 0 | 899 |
| 2016 | 0 | 5 | 471 | 0 | 476 |
| 2017 | 0 | 18 | 684 | 0 | 702 |
| 2018 | 0.4 | 38 | 640 | 0 | 678 |
| 2019 | 0.02 | 81 | 384 | 1 | 466 |
| 2020 | 0 | 0 | 558 | 2 | 560 |
| 2021 | 0.9 | 265 | 390 | 0 | 656 |
| 2022 | 29 | 306 | 267 | 2 | 605 |

Table 5.9.1. Summary of the overall sampling intensity on horse mackerel catches in recent years in all areas 1992-2022

| Year | Total Catch (ICES estimate) | \% catch covered by sampling programme* | No. samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 436500 | 45 | 1803 | 158447 | 5797 |
| 1993 | 504190 | 75 | 1178 | 158954 | 7476 |
| 1994 | 447153 | 61 | 1453 | 134269 | 6571 |
| 1995 | 580000 | 48 | 2041 | 177803 | 5885 |
| 1996 | 460200 | 63 | 2498 | 208416 | 4719 |
| 1997 | 518900 | 75 | 2572 | 247207 | 6391 |
| 1998 | 399700 | 62 | 2539 | 245220 | 6416 |
| 1999 | 363033 | 51 | 2158 | 208387 | 7954 |


| Year | Total Catch (ICES estimate) | \% catch covered by sampling programme* | No. samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 247862 | 50 | 378 | 33317 | 4126 |
| 2001 | 257411 | 61 | 467 | 46885 | 7141 |
| 2002 | 223384 | 68 | 540 | 79103 | 6831 |
| 2003 | 223885 | 77 | 434 | 59241 | 8044 |
| 2004 | 195177 | 62 | 518 | 62720 | 9273 |
| 2005 | 212850 | 76 | 573 | 67898 | 8840 |
| 2006 | 190067 | 75 | 602 | 57701 | 9905 |
| 2007 | 164459 | 58 | 397 | 41046 | 8061 |
| 2008 | 179053 | 72 | 488 | 46768 | 8870 |
| 2009 | 229665 | 84 | 902 | 57505 | 10575 |
| 2010 | 227261 | 82 | 710 | 49307 | 14159 |
| 2011 | 221317 | 71 | 502 | 40492 | 7484 |
| 2012 | 191236 | 69 | 501 | 41148 | 8220 |
| 2013 | 183903 | 75 | 686 | 87300 | 9776 |
| 2014 | 149740 | 83 | 650 | 53945 | 8085 |
| 2015 | 110421 | 68 | 825 | 39415 | 7034 |
| 2016 | 114151 | 76 | 1033 | 93853 | 6675 |
| 2017 | 97539 | 63 | 1113 | 116722 | 8221 |
| 2018 | 116455 | 74 | 1584 | 117768 | 6965 |
| 2019 | 136750 | 64 | 1014 | 77211 | 7476 |
| 2020 | 89,009 | 52 | 516 | 41811 | 5662 |
| 2021 | 92,639 | 77 | 977 | 59222 | 8080 |
| 2023 | 78,491 | 78 | 1162 | 65991 | 5814 |

*Percentage related to catch (catch at age) according to ICES estimation

Table 5.9.2. Horse mackerel sampling intensity for the Western stock in 2022.

| Country | Catch | \% Catch Sampled* | No. Samples | No. Measured | No. Aged |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Denmark | 5265 | 0 | 0 | 0 | 0 |
| Faroe Islands | 106 | 0 | 0 | 0 | 0 |
| France** | 3288 | 12 | 426 | 2769 | 0 |


| Country | Catch | \% Catch Sampled* | No. Samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Germany | 4263 | 76 | 56 | 9767 | 780 |
| Ireland | 15799 | 94 | 78 | 9065 | 2012 |
| Netherlands | 18474 | 81 | 48 | 5663 | 1145 |
| Norway | 3131 | 97 | 7 | 285 | 193 |
| Spain | 13424 | 97 | 740 | 28502 | 47 |
| Sweden | 3 | 0 | 0 | 0 | 0 |
| UK (England) | 2316 | 87 | 77 | 4187 | 401 |
| UK(Northern Ireland) | 811 | 58 | 184 | 184 | 46 |
| UK(Scotland) | 558 | 81 | 89 | 3473 | 0 |
| Total | 70114 | 74 | 1112 | 56891 | 4624 |

*Percentage based on ICES estimate with regards to age samples
**provided only length distributions
*** age samples processed by the Netherlands

Table 5.9.3. Horse mackerel sampling intensity for the North Sea stock in 2022.

| Country | Catch | \% Catch Sampled* | No. Samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 30 | 0 | 0 | 0 | 0 |
| Denmark | 37 | 0 | 0 | 0 | 0 |
| Faroe Islands | 0 | 0 | 0 | 0 | 0 |
| France** | 433 | 25 | 59 | 464 | 0 |
| Germany | 67 | 79 | 11 | 307 | 208 |
| Netherlands | 2490 | 76 | 13 | 2556 | 322 |
| Norway | 1640 | 99 | 4 | 193 | 117 |
| Sweden | 0 | 0 | 0 | 0 | 0 |
| UK (England)**** | 3202 | 94 | 22 | 6046 | 543 |
| UK(Scotland) ${ }^{* * *}$ | 42 | 0 | 0 | 0 | 0 |
| Total | 8377 | 78 | 109 | 9566 | 1190 |

[^4]
### 5.13 Figures



Figure 5.1.1a. Horse mackerel catches (prov. without Portuguese catch) 1st quarter 2022


Figure 5.1.1b. Horse mackerel catches (prov. without Portuguese catch) $\mathbf{2}^{\text {nd }}$ quarter 2022


Figure 5.1.1c. Horse mackerel catches (prov. without Portuguese catch) $\mathbf{3}^{\text {rd }}$ quarter 2022


Figure 5.1.1d. Horse mackerel catches (prov. without Portuguese catch) 4th quarter 2022.


Figure 5.2.1: Distribution of Horse Mackerel in the Northeast-Atlantic: Stock definitions as used by the 2004 WGMHSA. Note that the "Juvenile Area" is currently only defined for the Western Stock distribution area - juveniles do also occur in other areas (like in Div. 7.d). Map source: GEBCO, polar projection, $\mathbf{2 0 0} \mathbf{m}$ depth contour drawn.

Western HOM Catches


Figure 5.3.1. Total catch for Western Horse Mackerel stock, 1982-2022.
North Sea HOM Catches


Figure 5.3.4. Total catch for North Sea Horse Mackerel stock, 1982-2022


Figure 5.4.1 Horse mackerel general overview. Total catches in the northeast Atlantic during the period 1982-2022. The catches taken from the southern, western and North Sea horse mackerel stocks are shown in relation to the total catches in the northeast Atlantic. Catches from Div. 8.c were transferred from southern stock to western stock from 1982 onwards. Southern horse mackerel is assessed by ICES WGHANSA since 2011.

## North Sea Stock <br> Catches by division



Figure 5.4.2. North Sea horse mackerel stock. Total catches by Division during the period 1982-2022.

Western Stock 1982-2022
Catches by division


Figure 5.7.1. Length distributions contributed by country and area of the Western and North Sea horse mackerel 2022.

NSHOM \% observed vs. estimated 2000-22


Figure 5.9.1 North Sea horse mackerel stock. Percentage sampled catch (blue) vs. unsampled catch (red) by division and year, 2000-2022.


Figure 5.9.2 North Sea horse mackerel stock. Sampling intensity index as percentage sampled catch in total catch by year, 2000-2022.

WHOM \% observed vs. estimated 2000-22


Figure 5.9.5. Western horse mackerel stock. Percentage sampled catch (blue) vs. unsampled catch (red) by division and year. Period 2000-2022. Area of distribution of Western stock was divided into different regions. Chan: (7.e,f,h); WSCO+IRL (7.a-c, 7.j-k and 6.a); BoB (8.a,b,d); CantSea(8.c); N-Nsea (3.a and 4.a); NOR (2.a and 5.a).


Figure 5.9.6. Western horse mackerel stock. Sampling intensity index as percentage sampled catch in total catch by year, 2000-2022.

## 6 Horse mackerel in Skagerrak, Kattegat, southern and central North Sea, and eastern English Channel

Trachurus trachurus in divisions 3.a, 4.b-c, and 7.d (hom.27.3a4bc7d)

### 6.1 ICES advice 2022

In 2012, the North Sea horse mackerel (NSHOM) was classified as a category 5 stock, based on the ICES approach to data-limited stocks (DLS). Since then, a progressive reduction in TAC was advised by ICES, from 25500 tonnes in 2013-2014 to 15200 tonnes in 2015-2016. This reduction in the advised catch was supported by the analysis of information from the North Sea International Bottom Trawl Survey (NS-IBTS) traditionally used in the assessment, but also new information from the French Channel Ground Fish Survey (CGFS) since 2014. Additionally, in 2015, information on discards in non-directed fisheries became available that has been taken into account in the advice since 2017.
In 2017, this stock was benchmarked and the NS-IBTS and CGFS survey indices where modelled together. The resulting joint index was considered an appropriate indicator of trend in abundance over time and the NSHOM stock assessment was upgraded to category 3 . The joint index showed an increasing trend in 2014 to 2016, but was followed by a decrease in 2017. In 2018, the index remained at a similar level as in 2017, while the index slightly increased in 2019. In 2020 no index value was calculated due to the absence of UK-stations in the French Ground fish survey. In 2021 the survey index decreased to values similar to 2018, but in 2022 the index is at the highest estimate since the last 5 years at a value of 265 fish/hour. Length-based DLS methods have been applied to data from 2016 onwards. The length-based $\mathrm{F} / \mathrm{F}$ MSY ratio indicated that F is still slightly above Fmsy in till 2022. Stock size relative to reference points is unknown.
Biannual advice for 2024 and 2025 was provided in 2023, based on the data up to 2022 (ICES, 2023b). The advice follows the rfb rule (ICES 2022, ICES 2023a). The stock biomass trend indicated a ratio of 1.23 for last two years compared to the three years prior. The fishing pressure proxy Lr=m/Lmean $n 2022$ was 1.02 , indicating that the fishing mortality was above Fmsy. The biomass safeguard multiplier was above 1 , so it did not have to be applied. The stability clause also did not have to be applied, since the advice did only change $8.5 \%$. The approach resulted in a catch advice for 2024 and 2025 of 9730 tonnes.

### 6.2 Fishery of North Sea horse mackerel stock

Based on historical catches taken by the Danish industrial fleet for reduction to fishmeal and fish oil in the 1970s and 1980s, approximately $48 \%$ of the EU North Sea horse mackerel TAC was taken by Denmark. Catches were taken in the fourth quarter mainly in divisions 4 b and 7 d . The 1990s saw a drop in the value of industrial fish, limited fishing opportunities and steep increases in fuel costs that affected the Danish quota uptake. In 2001, an individual quota scheme for a number of species was introduced in Denmark, but not for North Sea horse mackerel. This lead to a rapid restructuring and lower capacity of the Danish fleet, which in combination with the above mentioned factors led to a decrease of the Danish North Sea horse mackerel catches.

Since the 1990's, a larger proportion of the catches have been taken in a directed horse mackerel fishery for human consumption by the Dutch-owned freezer-trawler fleet. This is possible because Denmark has traded parts of its quota with the Netherlands for other species. However,
due to the structure of the Danish quota management setup only a limited amount of quota can be made available for swaps with other countries. These practical implications of the management scheme largely explain the consistent underutilisation of the TAC over the period 20102014 (approximately $50 \%$; Figure 6.2.1)). However, following the sharp reduction in TAC in 2015, uptake increased significantly in the years thereafter. In 2020, $91 \%$ of the TAC was used, with the highest catches taken by the Netherlands, followed by UK, Norway and France. For 2021 the TAC utilization is $79 \%$ with the highest catches taken by the UK, followed by Netherlands, Norway and France, and in 2022 the uptake was $94 \%$ with the highest catches taken again by the same countries in that consecutive order (Figure 6.2.2)
Catches taken in Divisions 3a and 4a during the two first quarters and all year round in Divisions $4 b, 4 c$ and $7 d$ are currently regarded as North Sea horse mackerel (Section 5, Table 5.4.1). The catches were relatively low during the period $1982-1997$ with an average of 18000 t , but increased between 1998 ( 30500 t) and 2000 ( 45130 t). From 2000 to 2010, the catches varied between 24149 and 45883 t . Since 2014, a steep decline in catches is observed, both due to the reduction in the TAC since 2014 but also due to the underutilization of the quota. In 2021 the catch was 11 082 t similar to 2019, but in 2022 catch has decreased to 8377 and catches from 7 d are on a decline. In 2020 catches from that area were $72 \%$. In 2021 catches from that area were $65 \%$ and in 2022 catches from that area amounted to $56 \%$ of the total catch (Figure 6.2.4). This is because catches in area 4c have increased to $24 \%$. Relative catches in other areas were comparable to 2021.

Over the period 1985-2001 most catches were taken in the 4 b (Figure 6.2.3). However, since the early 2000s the proportion of catches from 7d increased steadily until 2013, when the $92 \%$ of total catches were fished in this area (Figure 6.2.4). In 2022, the UK accounted for most of the landings, followed by Netherlands, Norway and France (Figure 6.2.5). The majority was still caught in quarter 4 in 7 d , whereas the Norwegian catches were taken during quarters 1 and 2 in 4 a . The relative contribution of the catches in 4 a has been increasing since 2015, with the exception of 2017 when catches in 4a were negligible. In 2014 the contribution from 4a was only $0.02 \%$ but in 2022 the contribution is $20 \%$, partially because absolute catches in other areas have gone down.

Most of the discards reported were from 7d by the French bottom-trawl fleet. Discarding in the target pelagic fisheries is considered negligible. New information in 2015 from bottom-trawl fisheries (not directed at horse mackerel) indicated an overall discard rate of $16.7 \%$ for the stock as a whole, while in 2016 this rate was $10 \%$. Complete discard information for earlier years has not been submitted to ICES. Information from national discard reports for the non-directed bottomtrawl fisheries indicates a similar level of discarding in earlier years. In 2017 and 2018 the discard rate was $8.3 \%$ and $1.8 \%$, respectively, while it decreased to $1.6 \%$ in 2019 and 2020. In 2021 the discard rate dropped to $0.5 \%$, but in 2022 the rate has increased to $5.2 \%$.

### 6.3 Biological data

### 6.3.1 Catch in Numbers at Age

For the third year in a row in 2022, it was possible to include samples taken from English vessels in the Netherlands, increasing further the biological sampling coverage. Also Norway had supplied samples for 2022. For 2022, the proportion of sampling increased to $78 \%$ in comparison to $67 \%$ in $2021,56 \%$ in 2020 and to 2019 where only $1 / 3$ of the landings were sampled. Age samples in 2022 were available from four countries (the Netherlands and UK/England, Norway and Germany). Coverage was decent, as the temporal and spatial overlap between the catch and sampling was high. In prior years there had been a biased perception of the age distribution of catches over the year and areas as sampling effort was partially and unevenly distributed.

Annual catch numbers at age are shown in Table 6.3.1. Catch-at-age for the whole period 19952022 are given in Table 6.3.2 and in Figures 6.3.1 and 6.3.2. These data show that since 2005 the age distribution of catches has experienced a reduction, with a decrease in the range of ages of importance in total catches. However, this decrease could be due to the low age sampling, in particular in 2018 (maximum age observed 7 years). In parallel to the rejuvenation of catches, the comparison of catch-at-age data after 1998 by area (Figure 6.3.2) shows that since 2010 commercial catches have stayed relatively stable in 7 d in comparison to 3 a and $4 \mathrm{a}, \mathrm{b}$ and c where a decreasing pattern is seen. Due to the low level of sampling effort in 2018, data for this year are only based on a single sample from 7d in Q4. Following 2018, the older ages are represented again, but the representation of older ages in the catch is still marginal, especially in 7 d .

Although the 2015 cohort seems to be clear in the catch-at-age distribution, in general, cohort structure is not clearly detectable in the data. In addition to the low sampling levels, this may partly be due to the shifts in the distribution of the fishery. It may also be due to age reading difficulties, which are a known to be encountered (e.g. Bolle et al., 2011). Most clearly detectable is the relatively large 2001 year-class, although it is not clearly present in the catch data in all years. There are indications that environmental conditions may be an important factor (possibly stronger than stock size) contributing to spawning success of horse mackerel. This is, for example, illustrated by the largest year-classes (1982 and 2001) observed in the Western stock which were produced at the lowest observed stock sizes. Since 2001 is considered to have been a relatively strong year class in the Western stock, it is plausible that circumstances in the North Sea were similar to those in Western areas and also allowed for relatively high spawning success in the North Sea.

The potential for mixing of fish from the Western and North Sea stock in 7d and 7e in winter may also confuse the cohort signals. For example, the large recruitment in the Western stock may have led to more of these fish being located in the North Sea stock area as age 1 fish in 2002. On behalf of the Pelagic Advisory Council and the EAPO Northern Pelagic Working Group, a research project on genetic composition of horse mackerel stocks was initiated in 2015 with University College Dublin (Ireland) with the intention of clarifying the mixing among the North Sea and the Western horse mackerel stocks. Genetic samples have been taken over the entire distribution area of horse mackerel during the years 2015, 2016, and 2017, with a specific focus on the separation between horse mackerel in the western waters and horse mackerel in the North Sea. The results of the whole-genome sequencing indicated that the North Sea horse mackerel stock is clearly genetically different from the Western stock (Farrell and Carlsson, 2019; Fuentes-Pardo et al., 2020). Markers were identified that could distinguish with up to $95 \%$ accuracy between individuals collected in the North Sea and Western stocks. Follow-up work on this project is described in Section 6.7.

### 6.3.2 Mean weight at age and mean length at age

The mean annual weight and length over the period 2000-2022 are presented in Table 6.3.2 and Figures 6.3.3 and 6.3.4, respectively.

There do not seem to be strong differences over this period, although for the last two years a decline had been observed. In 2022 however, the annual means for weight and length have experienced an increase again. Zooming into the values per area however, mean length and weight of the stock for age 7-9 for area 4a are substantially higher than for the other areas.

Catches defined as NSHOM catches in this area are taken in Q1 and Q2. A closer look at the distribution of those catches revealed them to originate from within the Norwegian fjords, where the size at age was sigificantly higher than in landings farther west in the North Sea during Q3 and Q4. That, and the ongoing work on stock delineation supports the notion that these fish are
not North Sea horse mackerel. Additionally, the fish caught in 2022 were only age 7 and older, no fish younger than age 7 were caught in the fjords. Further work on the biology of the horse mackerel need to be done to explain this anomaly.

### 6.3.3 Maturity-at-age

Peak spawning in the North Sea occurs in May and June (Macer, 1974), and spawning occurs in the coastal regions of the southern North Sea along the coasts of Belgium, the Netherlands, Germany, and Denmark.

There is no information available about the maturity-at-age of the North Sea Horse mackerel stock.

### 6.3.4 Natural mortality

There is no specific information available about natural mortality of this stock.

### 6.4 Data exploration

### 6.4.1 Catch curves

The log-catch numbers were plotted by cohort to calculate the negative slope to get an estimate of total mortality (Z). Fully selected ages 3 to $15+$ from the 1992-2015 period provide complete data for the 1992 to 2007 cohorts and partial data for younger cohorts (Figure 6.4.1). The estimated negative slopes by cohort (Figure 6.4.2) indicate an increasing trend in total mortality up to the late 1990s, after which Z fluctuates from year to year up till late 00 's after which a steeper increase followed by a steep decrease are observed. However, due to the low quality of the signals for some cohorts these Z estimates should be considered with caution.

An analysis of the catch number at age data carried out in 2011 showed that only the 1 vs.2, 2 vs.3, 7 vs .8 and 9 vs .10 age groups were positively and significantly correlated in the catch. This analysis has not been updated since, but these results suggest limitations in the catch-at-age data.

### 6.4.2 Assessment models and alternative methods to estimate the biomass

In 2002 Rückert et al. estimated the North Sea horse mackerel biomass based on a ratio estimate that related CPUE data from the IBTS to CPUE data of whiting (Merlangius merlangus). The applied method assumes that length specific catchability of whiting and horse mackerel are the same for the IBTS gear. Subsequently, they use the total biomass of whiting derived from an analytical stock assessment (MSVPA) to estimate the relationship between CPUE and biomass.

At the 2014 WGWIDE meeting exploratory model fits were attempted with the JAXass model, a simple statistical catch-at-age model fitted to an age-aggregated index of ( $2+$ ) biomass, total catch data and proportions at age from the catch. JAXass is based on Per Sparre's "separable VPA" model, an ad hoc method tested for the first time at WGWIDE in 2003, and later 2004. A new analysis using this model was also carried out in 2007 using an IBTS index. In 2014 the model has been coded in ADMB (Fournier et al., 2012) and updated with an improved objective function (dnorm), additional years of data and new methods for calculating the index (see above).

Difficulties in fitting an assessment model for this stock include:

- Unclear stock boundaries
- Difficulty aging horse mackerel
- Lack of strong cohort signals in catch-at-age data
- $\quad$ Scientific index derived from a survey not specifically designed for horse mackerel and limited coverage of one of the main fishing grounds for the stock (7d)

Catches taken in 7d are close to the management boundary between the (larger) Western horse mackerel stock and the NS horse mackerel stock. It is quite possible that given changes in oceanographic conditions, or changes in abundance of either of the two stocks, that some proportion of the catches taken in 7d actually originated from the Western horse mackerel stock. Nevertheless, all assessment models used assume that $100 \%$ of fish caught in area 7 d belong to the North Sea horse mackerel stock. This is in agreement with current stock and management definitions.
In 2018, the working group explored the Surplus Production model in Continuous Time (SPiCT) model for North Sea horse mackerel. SPiCT is one of the methods in the ICES guidelines to estimate MSY reference points for category 3 and 4 stocks (ICES, 2018). The model was run using the joint survey index as input or with separate survey indices (NS-IBTS and CGFS). The model with the joint survey index led to conflicting results with the perception of the stock, as biomass was estimated to be above $B_{\text {msy }}$ and fishing mortality below $\mathrm{F}_{\text {msy. }}$. The model with two separate indices resulted in stock biomass and fishing mortality that were more in line with the perception of the stock. However, there were strong retrospective patterns and wide confidence intervals in recent years. Furthermore, additional work is necessary on the setting of the priors, and on ensuring that model assumptions are not violated.

### 6.4.3 Survey data

### 6.4.3.1 Egg Surveys

No egg surveys for horse mackerel have been carried out in the North Sea since 1991. Such surveys were carried out during the period 1988-1991. SSB estimates are available historically. However, they were calculated assuming horse mackerel to be a determinate spawner. Horse mackerel is now considered an indeterminate spawner (Gordo et al. 2008). Therefore, egg abundance could only be considered a relative index of SSB. The Mackerel and Horse Mackerel Egg Surveys in the North Sea do not cover the spawning area of the North Sea horse mackerel stock.

### 6.4.3.2 North Sea International Bottom Trawl Survey

Many pelagic species are frequently found close to the bottom during daytime (which is when the North Sea IBTS survey operates) and migrate upwards predominantly during the night when they are susceptible to semi-pelagic fishing gear and to bottom trawls (Barange et al. 1998). Macer (1977) observed that dense shoals are formed close to the bottom during daytime, but the top of the shoals may extend into midwater. Eaton et al. (1983) argued that horse mackerel of 2 years and older are predominantly demersal in habit. Therefore, in the absence of a targeted survey for this stock, the NS-IBTS is considered a reasonable alternative.

NS-IBTS data from quarter 3 were obtained from DATRAS and analysed. Based on a comparison of NS-IBTS data from all 4 quarters in the period 1991-1996, Rückert et al. (2002) showed that horse mackerel catches in the NS-IBTS were most abundant in the third quarter of the year. In 2013 WGWIDE considered that using an 'exploitable biomass index' estimated with the abundance by haul of individuals of 20 cm and larger is the most appropriate for the purpose of interpreting trend in the stock.

To create indices, a subset of ICES statistical rectangles was identified. Rectangles that were not covered by the survey more than once during the period 1991-2012 were excluded from the index area. In 2012, WGWIDE expressed concern that the previously selected index area did not sufficiently cover the distribution area of the stock, especially in years that the stock would be
relatively more abundant and spread out more. Rückert et al. (2002) also identified a larger distribution area of the North Sea stock. Based on the above, WGWIDE 2013 identified 61 rectangles to be included in the index area as shown in Figure 6.4.3.

### 6.4.3.3 French Channel Groundfish Survey

In order to improve data basis for the North Sea horse mackerel assessment, alternative survey indices have been explored. Previous indices only covered the North Sea distribution of the stock, while the majority of catches in recent years come from the eastern English Channel (7d). We evaluated the potential contribution of the French Channel Groundfish Survey (FR-CGFS) in 7d in quarter 4. The FR-CGFS has been carried out since 1990 and has frequent captures of horse mackerel. Although this survey is conducted in a different quarter to the NS-IBTS, the observed seasonal migration patterns of horse mackerel indicate that fish move into the Channel following quarter 3, so the timing is considered appropriate.
In 2015, the RV Gwen Drez was replaced by the RV Thalassa to carry out the FR-CGFS. In 2014 an inter-calibration process was conducted to quantify the differences in catchability for a large number of species. ICES reviewed this inter-calibration exercise and found a number of drawbacks that may undermine the reliability of the estimated conversion factors. The main concerns were:

- The analyses were limited in the number of tows. Considering that a number of these tows could be zeros for one of the two vessels and possibly resulting in highly uncertain estimates.
- Lack of length-specific correction factor.
- At a standardized depth of 50 m and above, wing spread estimates for the R/V Thalassa as measured by the MARPORT sensor were deemed erroneous, which may question the validity of estimated area swept by the net on the R/V Thalassa and the effect it may have on correction factors for species caught at depth at 50 m and greater.
- A number of tow locations including areas outside 7d were excluded. Changing the depth range of a survey can add serious bias in the calibration and the current approach seems to be ignoring this issue.
- Correction coefficients were not measured without error.

However, these limitations were considered by WGWIDE to be of minor importance for the North Sea horse mackerel since:

- Despite being still a low sample size the North Sea horse mackerel was present in all the 32 paired hauls.
- $\quad$ There are no important differences in size distribution (Figure 6.4.4).
- The analysis with and without the areas excluded in the new sampling design did not show important differences (ICES, 2017).
- CPUE of North Sea horse mackerel for hauls deeper than 50 m was relatively low (Figure 6.4.5), and it is expected than the potential problems in determining the conversion factor below that depth range would have a relatively minor impact in the estimated abundance.

For these reasons it was considered appropriate to continue using the FR-CGFS, standardizing the time-series of abundance for the period 1990-2015 with the estimated conversion factor 10.363 .

### 6.4.3.4 Impact of Covid-19

Due to the Covid-19 pandemic and the lockdown in place in France at that time there was a delay in submitting the cruise application form for the FR-CGFS in 2020 to the French Foreign Ministry. The result was that no authorisation was provided in time to allow the survey to trawl within

UK waters in 2020. Therefore, only French waters were sampled, meaning that only $70 \%$ of the core survey stations were completed (ICES, 2021b).

To assess the potential impact of missing UK stations in the FR-CGFS on the resulting abundance index for the exploitable stock, we tested the impact of
i) removing all UK sampling stations from the 1992-2019 time series,
ii) removing UK sampling stations from 2016-2019, one year at the time, and
iii) removing the FR-CGFS in 2016-2019, one year at the time, when modelling the abundance and calculating the index.
Removing all UK sampling stations from all years did not change the overall trend of the abundance index, but there were quite some deviations for individual years (Figure 6.4.6). Removing UK stations from one year at a time for 2016-2019 resulted in virtually no change for 2017 and 2018, but more apparent changes for 2016 and 2019 (Figure 6.4.7). Both these exercises suggest that basing the abundance index on NS-IBTS and French stations from FR-CGFS only may lead to different index values compared to when UK stations are included. The French sampling stations in the FR-CGFS only are thus not representative for the abundance of adult horse mackerel in the entire eastern Channel. As a further exploration, the abundance index was modelled by leaving out the FR-CGFS entirely for 2019. However, the hurdle model was not able to run, and therefore a zero-inflated model was run instead. This model was considered to be the secondbest model during the benchmark process in 2017 and performed almost equally well as the hurdle model (ICES, 2017). Removing the FR-CGFS one year at a time for 2016-2019 resulted in minimal change for 2017 and 2018, but more apparent changes for 2016 and 2019 (Figure 6.4.8). Similar to (i) and (ii), leaving out the FR-CGFS may lead to different index values compared to when FR-CGFS is included.

As the investigations suggest that the missing UK stations from the FR-CGFS or leaving out the FR-CGFS entirely may lead to changes in the abundance index, it was decided that no reliable index value for 2020 could be produced. For the 2021 assessment, the approach of previous year was continued and thus no index value for 2020 was modelled. UK stations were visited in 2021 so the index value for 2021 was modelled according to the method with the hurdle model.

### 6.4.4 Length distributions from the surveys

The highest proportion of fish caught in 2022 were around 6-8 cm in the NS-IBTS, this was similar to 2021 (Figure 6.4.9). No group of strong year classes from previous years could be observed. In the FR-CGFS, the highest proportion of fish were between $18-30 \mathrm{~cm}$, similar to 2019, when also the larger fish were dominating the catches (Figure 6.4.10). Despite that in 2020 the length frequencies are only based on French sampling stations, the length frequencies from 2021 show a similar pattern.

### 6.4.5 Length distributions from commercial catches

Currently, length distributions from catch data are available from 2016 to 2021. Future work is needed to retrieve historic length data in order to present a longer time series. The data used for the analysis come from the commercial catch sampling by national sampling programmes. For comparison, the analysis has also been run in the past with length data from the self-sampling programme of the Pelagic-Freezer-trawler Association (PFA), see for instance ICES $(2019,2020)$.

The length distributions based on the commercial catch data from 7d show a consistent distribution in time with a mean length between 22.15 and 25.5 cm each year, although with the exception of 24.7 cm in 2019 and the value for 2022 at 25.2 (Figure 6.4.11). Lengths in 4 c were on average 21.7 cm in 2019, 22.0 cm in 2020, 23.8 in 2021, and 22.4 in 2022 (Figure 6.4.12).

An error was found in the calculation of the length frequency distributions in the previous 2016 to 2020 assessments. Furthermore, the length frequency distribution calculated in 2019 included French data from only quarters 3 and 4 , whereas data is also available for quarters 1 and 2 . The length frequency distributions for 2018-2020 were re-calculated using all available data, and upgraded code. No recalculation could be done for 2016-2017 since the original files are not available.

### 6.5 Stock assessment

### 6.5.1 Modelling the survey data

In January 2017, a benchmark of the North Sea horse mackerel assessment was conducted (ICES, 2017). Based on a capacity to model the over-dispersion and the high proportion of zero values in the survey catch data, a hurdle model was considered the best option of all model alternatives tested. The log-likelihood ratio test, AIC and the evidence ratio statistic supported that the model that best represented the data was a hurdle model with Year and Survey as explanatory factors (including the interaction term) in the count model (GLM-negative binomial), and Year and Survey (without the interaction) in the zero model (GLM-binomial).
The probability of having a CPUE of zero was modelled by a logistic regression with a GLMbinomial distribution model:

$$
\operatorname{logit}\left(\pi_{i}\right)=\text { Intercept }_{z e r o}+\text { Year }_{i, z e r o}+\text { Survey }_{i, z e r o}
$$

where $\pi_{i}$ is the mean probability of having a CPUE of zero in haul $i$ as a function Year and Survey.
The expected CPUE of North Sea horse mackerel per haul $i$, conditional to not having a zero in hurdle models (not having a false zero in zero-inflated models), was modelled with a GLM-negative binomial distribution model:

$$
\log \left(C P U E_{i}\right)=\text { Intercept }_{\text {count }}+\text { Year }_{i, \text { count }} x \text { Survey }_{i, \text { count }}
$$

This model was used to synthesise the information from both the FR-CGFS and NS-IBTS and predict the average annual CPUE index as an indicator of trends in stock abundance. Separate models were fitted to the juvenile $(<20 \mathrm{~cm})$ and adult exploitable $(\geq 20 \mathrm{~cm})$ sub-stocks. The contribution of the two surveys to the combined index is weighted taken into consideration their respective area coverage as well as the mean wing spread. This index model allowed upgrading of the NSHOM to a category 3 stock within the ICES classification.

Similar to the 2019 assessment (ICES, 2019) ,2020 assessment (ICES, 2020), and 2021 assessment (ICES, 2021a), the model for the adult sub-stock that was run this year returned a warning despite the fact that the model converged. All parameter coefficients were estimated, but not the standard error for the intercept and the parameter $\theta$ of the count model. To check the robustness of the hurdle model with the warning, a zero-inflated model was run with the same set-up as the hurdle model. This zero-inflated model was considered to be the second-best model during the benchmark process in 2017 and performed almost equally well as the hurdle model (ICES, 2017). The fitted values of the zero-inflated model were very similar to that of the hurdle model with warning (Figure 6.5.1). The hurdle model from this year and its resulting index values where thus considered robust. Should the warning continue to occur in future assessments, additional testing and investigation should be conducted.

Due to the exclusion of the 2020 survey for modelling the abundance index, during the 2021 assessment the same time period (1992-2019) was used as during the 2020 assessment (ICES, 2020). In the current 2022 assessment the 2020 survey is still excluded. Since the last the 2021 assessment the updated abundance index resulted in a higher value for 2016 for the exploitable
stock compared to last year (ICES 2021; Figure. 6.5.2). For each assessment, survey data from all years are extracted so that any underlying changes in the raw data stored in DATRAS are taken account of. Changes in reported raw HOM catches in 2016 in the NS-IBTS led to a higher mean catch rate of HOM (Figure 6.5.3), resulting in a higher abundance index value for 2016.

### 6.5.2 Summary of index trends and survey length distributions

The survey index for both the juvenile and exploitable sub-stock experienced a marked decline in the early 1990s and fluctuated at relatively low levels thereafter (Figures 6.5.4; Table 6.5.1). This reduction was partly due to the decline of the average abundance per haul over time, but also due to the increase of hauls with zero catch of the adult sub-stock (Figure 6.5.5). The survey index was at its third and second lowest in 2017 and 2018 (lowest in 2009), shows a slight increase again in 2019, but shows an average decline again in 2021, because of the low index of the NSIBTS. In 2022 the combined index has increased again since both indices show an increase (Figure 6.5.4).

The index trend for the juvenile sub-stock shows large fluctuations since 2015 (Figure 6.5.4). These are mainly attributed to the fluctuating trend of juveniles in the NS-IBTS (Figure 6.5.6), caused by some hauls with high catches of small horse mackerel in 2016 and 2018 (Figure 6.4.9). Fitted values for juveniles in the FR-CGFS show decreasing trend since 2014, but a slight increase again in 2019 and in 2021. The 2022 value is slightly lower again but similar to the year before (Figure 6.5.6). The index of abundance of individuals $<20 \mathrm{~cm}$ could be considered a recruitment index, but future analyses should be carried out to study the correlation between the abundances and survey indices of year classes over time in more detail.

### 6.5.3 Length-based indicator and MSY proxy reference points

As part of the ICES approach to provide advice within the MSY framework for stocks of category 3 and 4, different Data Limited Stock (DLS) methods to estimate MSY proxy reference points (ICES, 2012, 2018) for the North Sea horse mackerel were previously explored (Pérez-Rodríguez, 2017). The Length Based Indicators analysis is the DLS method used in previous years assessments.

As most length samples and catches originate from 7d, length distributions from this area were used to calculate the MSY proxy. In 2021, the F/FMSY proxy based on the commercial catch samples indicated that fishing mortality was still slightly above $\mathrm{F}_{\mathrm{MSY}}$, with $\mathrm{Lmean} / \mathrm{Lf}_{\mathrm{F}=\mathrm{M}}=22.86 \mathrm{~cm}$ / $24.33 \mathrm{~cm}=0.94$ (Figure 6.5.7).

Due to the recalculation of the 2018-2020 length distributions the F/Fmsy ratios in those years changed from 0.927 to 0.932 for 2018, from 0.978 to 0.972 for 2019, and from 0.927 to 0.912 from 2020. These revisions have no effect on the advice given in the 2021 assessment.

For WGWIDE 2023 the new rfb rules for category 3 stocks are implemented (ICES 2022, ICES 2023a). This rule still uses the 2-over-3 rule for the the calculation of the trend in the abundance index. The approach with a cap and buffer as was previously done, is dropped. A multiplier and a biomass safeguard is now calculated instead. Additionally, there is a slight change in the calculation of the Fmsy proxy value. Whereas previously the Lfm was calculated based on the yearly Lc, currently this value is based on the Lc derived from pooled length data from 20192022. This is because for these years enough detailed information is available to provide Lc. As a result, Lmean and Lfm for 2019-2022 can be recalculated using this new pooled Lc value. For 2016-2018, this was not possible and the values provided in the advice for those years are still derived with the yearly Lc.

The implication of this new approach is that the values for $L_{\text {mean }} / L_{F=m}$ from 2019-2021 have changed again; for 2019 the value changed from 0.9721 .046 , for 2020 it changed from 0.912 to 0.929 , for 2021 from 0.926 to 0.939 . These revisions have no implications on the advice given in the 2021 assessment year. For the current assessment year the value is 0.980

### 6.6 Basis for 2024 and 2025 advice

Stock advice for North Sea horse mackerel is biennial. The NS-IBTS and FR-CGFS were modelled together to produce a joint abundance index for the exploitable part of the stock ( $\geq 20 \mathrm{~cm}$ ). No index value for 2020 could be produced. For this reason, the 2 -over- 3 rule applied to the index could only make use of index values from 2018 to 2019 and from 2021-2022. The resulting index ratio indicated that the adult sub-stock declined by $23 \%$. The Lmean/Lf=m ratio in 2022 was 0.98 indicating that the fishing mortality is above $\mathrm{F}_{\mathrm{MSY}}$. The biomass safeguard estimated by dividing the last index value by the Itrigger value, which is the lowest value observed in the index multiplied by 1.4 , was higher than 1 , and thus the biomass safeguard was set at 1 . The multiplier based on life history characteristics was 0.9 under a $k$ value of 0.205 . Under these circumstances, and based on the previous catch advice of 8969 t , ICES advised that catches of North Sea horse mackerel in 2024 and 2025 should be no more than 9730 t .

There are some signs of improved recruitment in some years (e.g. 2016, 2018 and to a lesser extend in 2021), but the trend of the abundance index for the juvenile sub-stock is fluctuating and, when separated, the two surveys, NS-IBTS and FR-CGFS, do not show the same trend. It remains to be seen if the weak signs of improved recruitment result in higher adult abundance. In 2019 there was a slight increase in the index of the exploitable sub-stock but that trend has not continued in 2021. In 2022 a slight increase can be recognized again.

### 6.7 Ongoing work

On behalf of the Pelagic Advisory Council and the EAPO Northern Pelagic Working Group, a research project on genetic composition of horse mackerel stocks was initiated in 2015 with University College Dublin (Ireland). Genetic samples have been taken over the whole distribution area of horse mackerel during the years 2015, 2016, and 2017, with a specific focus on the separation between horse mackerel in the western waters and horse mackerel in the North Sea. The result of the research indicated that the western horse mackerel stock is clearly genetically different from the North Sea stock (Farrell and Carlsson, 2019; Fuentes-Pardo et al., 2020). Markers were identified that are able to reveal the stock identity of individual horse mackerel from potential mixing areas, namely Division 7.d, 7.e and 4.a. Following this, the Institute of Marine Research in Norway sampled horse mackerel in coastal waters within 4.a during all quarters in 2019. Preliminary results presented at WGWIDE 2021 showed that the genetic profile of individuals caught in all quarters matched well with the genetic profile of the Western HOM stock, with just one or two individuals matching better with North Sea HOM profile (Florian Berg, pers. comm.). Results presented at WGWIDE 2023 endorsed the view of the preliminary results presented during WGWIDE 2021. In another research project, horse mackerel from 7d and 7e have been collected by the PFA on board of commercial vessels in the Autumn of 2020, while during the same period horse mackerel from 4a have been collected during the NS-IBTS in Q3. The stock identity of the sampled fish will be investigated. The Norwegian research as well as the ongoing research described here may have large implications for stock delineation as the North Sea horse mackerel stock could have a more restricted distribution than currently presumed. During WGWIDE 2023 it was presented that samples analysed from Q1 and Q2 in 4a were determined to be Western Horse mackerel. In addition there is also mixing occurring with Western horse mackerel in area 7.d. Potential changes in the perception of the stock distributions could impact
the reliability of the assessments for the three current stocks of horse mackerel in the Northeast Atlantic.

### 6.8 Management considerations

In the past, Division 7d was included in the management area for Western horse mackerel together with Divisions $2 \mathrm{a}, 7 \mathrm{a}-\mathrm{c}, 7 \mathrm{e}-\mathrm{k}, 8 \mathrm{a}, 8 \mathrm{~b}, 8 \mathrm{~d}, 8 \mathrm{e}$, Subarea 6, EU and international waters of Division 5b, and international waters of Subareas 12 and 14. ICES considers Division 7d now to be part of the North Sea horse mackerel distribution area. Since 2010, the TAC for the North Sea area has included Divisions 4.b, c and 7d. Considering that a majority of the catches are taken in Division 7d, the total North Sea horse mackerel catches are effectively constrained by the TAC since the realignment of the management areas in 2010.

Catches in Divisions 3a (Western Skagerrak) and 4a in quarters 3 and 4 are considered to be from the Western horse mackerel stock, while catches in quarters 1 and 2 are considered to be from the North Sea horse mackerel stock. Catches in area 4a and 3a are variable. In recent years only Norway has had significant catches in this area, but these are only taken in some years. Recent work suggest that all horse mackerel caught in 4 a belong to the Western stock, and ongoing genetic research on samples from 4 a and 7 d will shed more light on the proportions of the two stocks in catches from these areas.

### 6.9 Deviations from stock annex caused by missing information from COVID-19 disruption

1. Stock: hom.27.3a4bc7d
2. Missing or deteriorated survey data:

The assessment is based on two surveys, NS-IBTS and FR-CGFS. Due to the pandemic, trawling authorization in UK EEZ was not delivered in time, consequently FR-CGFS survey was not allowed to sample stations within UK waters in 2020.

## 3. Missing or deteriorated catch data:

Related to age sampling coverage was $56 \%$ and was covering only Q3, Q4 in areas 27.4.c and 27.7.d. Although most landed catch is taken from 27.7.d in Q 4 , other areas and quarters remain uncovered. Length sampling were impacted by the pandemic as samples were only available by two countries.
4. Missing or deteriorated commercial LPUE/CPUE data:

Not applicable
5. Missing or deteriorated biological data:

Not applicable
6. Brief description of methods explored to remedy the challenge:

Effects of having only UK stations in FR-CGFS in all years or a single year, and excluding FRCGFS entirely for a single year on the combined survey index were investigated.
7. Suggested solution to the challenge, including reason for this selecting this solution:

Exploration methods suggested that leaving out UK stations or FR-CGFS entirely may affect the survey index and would lead to a survey index value not representative of stock abundance. It was therefore decided to produce no survey index value for 2020.
8. Was there an evaluation of the loss of certainty caused by the solution that was carried out?

The chosen solution affects the 2-over-3 rule by that only four instead of five index values can be used to assess the change in stock abundance. Like this year's assessment for 2022 and 2023, this will also affect the advice given in 2023 for 2024 and 2025.

### 6.10 References

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### 6.11 Tables

Table 6.3.1. North Sea Horse Mackerel stock. Catch in numbers (1000) by quarter and area in 2022

| Number/1000 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | 0.02 | 0.04 | 0.03 | 1.84 | 507.5 | 509.43 |
| 4 | 0.07 | 0.21 | 0.13 | 8.68 | 2388.27 | 2397.36 |
| 5 | 0.07 | 0.21 | 0.14 | 8.75 | 2410.71 | 2419.88 |
| 6 | 0.03 | 0.1 | 0.06 | 3.99 | 1099.54 | 1103.72 |
| 7 | 0.01 | 107.83 | 0.02 | 0.99 | 163.85 | 272.69 |
| 8 | 0.05 | 215.65 | 0.08 | 5.35 | 1258.52 | 1479.65 |
| 9 | 0.01 | 100.26 | 0.02 | 1.3 | 258.11 | 359.69 |
| 10 | 0.02 | 508.54 | 0.04 | 2.26 | 114.2 | 625.05 |
| 11 | 0 | 0 | 0 | 0 | 0 | 0 |


| 12 | 0 | 90.96 | 0.01 | 0.4 | 20.43 | 111.8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 13 | 0.01 | 312.54 | 0.02 | 1.39 | 70.18 | 384.15 |
| 14 | 0.03 | 833.2 | 0.06 | 3.71 | 187.11 | 1024.11 |
| 15 | 0.03 | 761.58 | 0.05 | 3.39 | 171.02 | 936.07 |
| Sum | 0.35 | 2931.12 | 0.66 | 42.05 | 8649.44 | 11623.6 |
| 2Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1 | 0 | 4.08 | 11.07 | 188.2 | 138.82 | 342.18 |
| 2 | 0.01 | 6.28 | 17.03 | 289.4 | 213.47 | 526.18 |
| 3 | 0 | 3.05 | 8.27 | 140.64 | 103.74 | 255.71 |
| 4 | 0.01 | 6.74 | 18.27 | 310.48 | 229.02 | 564.51 |
| 5 | 0 | 3.48 | 9.44 | 160.41 | 118.32 | 291.66 |
| 6 | 0 | 1.73 | 4.7 | 79.97 | 58.99 | 145.4 |
| 7 | 0 | 4.3 | 1.26 | 21.4 | 15.79 | 42.75 |
| 8 | 0 | 10.19 | 5.21 | 88.48 | 65.26 | 169.14 |
| 9 | 0 | 22.26 | 1.2 | 20.39 | 15.04 | 58.88 |
| 10 | 0 | 45.55 | 1.96 | 33.31 | 24.57 | 105.39 |
| 11 | 0 | 20.61 | 0.07 | 1.24 | 0.91 | 22.84 |
| 12 | 0 | 46.43 | 0.49 | 8.26 | 6.09 | 61.27 |
| 13 | 0 | 73.62 | 1.37 | 23.21 | 17.12 | 115.32 |
| 14 | 0 | 58.37 | 3.15 | 53.59 | 39.53 | 154.65 |
| 15 | 0 | 221.82 | 3.48 | 59.12 | 43.6 | 328.02 |
| Sum | 0.02 | 528.51 | 86.97 | 1478.1 | 1090.27 | 3183.9 |
| 30 |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 65.55 | 2564.29 | 1509.6 | 4139.44 |
| 2 | 0 | 0 | 100.8 | 3699.21 | 2565.34 | 6365.36 |
| 3 | 0 | 0 | 40.3 | 1250.35 | 1254.01 | 2544.66 |


| 4 | 0 | 0 | 67.27 | 2101.47 | 2079.18 | 4247.92 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 0 | 0 | 14.64 | 564.76 | 345.33 | 924.73 |
| 6 | 0 | 0 | 9.07 | 283.96 | 279.58 | 572.61 |
| 7 | 0 | 0 | 2.73 | 103.79 | 65.82 | 172.33 |
| 8 | 0 | 0 | 5.43 | 179.35 | 158.09 | 342.87 |
| 9 | 0 | 0 | 0.52 | 0.47 | 31.79 | 32.78 |
| Sum | 0 | 0 | 306.31 | 10747.65 | 8288.74 | 19342.7 |
| 4Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 12.17 | 731.98 | 2957.93 | 3702.08 |
| 2 | 0 | 0 | 17.69 | 1379.58 | 3968.73 | 5366.01 |
| 3 | 0 | 0 | 14.56 | 757.47 | 3636.92 | 4408.95 |
| 4 | 0 | 0 | 22.63 | 1382.77 | 5429.11 | 6834.51 |
| 5 | 0 | 0 | 4.82 | 144.77 | 1306.41 | 1456 |
| 6 | 0 | 0 | 2.01 | 166.57 | 436.26 | 604.84 |
| 7 | 0 | 0 | 0.65 | 10.39 | 182.99 | 194.03 |
| 8 | 0 | 0 | 0.33 | 52.94 | 46.31 | 99.59 |
| Sum | 0 | 0 | 74.86 | 4626.47 | 17964.66 | 22666.01 |
| 1-4Q |  |  |  |  |  |  |
| Ages | 27.3.a | 27.4.a | 27.4.b | 27.4.c | 27.7.d | Total |
| 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 1 | 0 | 4.08 | 90.77 | 3508.5 | 4634.71 | 8238.07 |
| 2 | 0.01 | 6.28 | 138.48 | 5404.12 | 6789.93 | 12338.82 |
| 3 | 0.02 | 3.09 | 65.1 | 2173.87 | 5529.98 | 7772.06 |
| 4 | 0.08 | 6.94 | 111.78 | 3845.67 | 10175.45 | 14139.93 |
| 5 | 0.08 | 3.69 | 30.25 | 893.49 | 4198.24 | 5125.75 |
| 6 | 0.04 | 1.83 | 16.42 | 541.46 | 1882.6 | 2442.34 |
| 7 | 0.01 | 112.13 | 4.78 | 138.13 | 430.3 | 685.35 |
| 8 | 0.05 | 225.83 | 11.46 | 331.07 | 1534.03 | 2102.44 |


| 9 | 0.01 | 122.51 | 1.8 | 22.96 | 305.88 | 453.17 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 10 | 0.02 | 554.09 | 1.99 | 35.57 | 138.77 | 730.44 |
| 11 | 0 | 20.61 | 0.07 | 1.24 | 0.91 | 22.84 |
| 12 | 0 | 137.39 | 0.49 | 8.66 | 26.52 | 173.06 |
| 13 | 0.01 | 386.16 | 1.39 | 24.6 | 87.31 | 499.47 |
| 14 | 0.03 | 891.58 | 3.21 | 57.3 | 226.64 | 1178.76 |
| 15 | 0.03 | 983.4 | 3.53 | 62.5 | 214.63 | 1264.1 |
| Sum | 0.39 | 3459.61 | 481.52 | 17049.14 | 36175.9 | 57166.6 |

Table 6.3.2. Numbers at age (millions). weight at age (kg) and length at age (cm) for the North Sea horse mackerel 1995-2022 in the commercial fleet catches (2018 distribution based on one sample only due to low sampling level).


|  | 윽 |  | 축 | $\stackrel{\infty}{\circ}$ | O | Oi | Ö | Oì | $\underset{\sim}{\mathbf{O}}$ | O | No | O O | 숭 | Oi | Oio | Oì | ت국 | Nì | $\underset{\sim}{\underset{\sim}{2}}$ | $\underset{\sim}{\underset{N}{N}}$ | $\underset{\sim}{n}$ | $\begin{aligned} & 0 \\ & \underset{N}{7} \end{aligned}$ | $\stackrel{N}{\mathrm{~N}}$ | $\underset{\sim}{\infty}$ | $\underset{\sim}{\underset{\sim}{0}}$ | N్ద్ | N్ | N |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 12 | 8.8 | 1.3 | 0.3 | 0.7 | 0.4 | 1.7 | 5.4 | 3.4 | 3.1 | 5 | 1.3 | 1.5 | 0.2 | 2.1 | 3.6 | 0.3 | 3.5 | 0.1 | 1 | 0.4 | 0.9 | 0.4 | 0 |  | 0.7 | 0.13 | 0.19 | 0.17 |
| 13 | 0.2 | 8.9 |  | 1.8 | 1.4 | 1.2 | 3.7 | 2.4 | 2.4 | 6.8 | 2.7 | 0.4 | 0.8 | 0.4 | 0.3 | 0.3 | 0.4 | 0 | 0.6 | 0 | 0.2 | 1.4 | 0 |  | 0.7 | 0.03 | 0.12 | 0.5 |
| 14 | 4.4 | 8 | 1.4 | 0.3 | 3.8 | 0 | 2 | 1.4 | 3.4 | 2.5 | 2.1 | 0.8 | 0.8 | 2.4 | 1 | 0.2 | 0.3 | 0.2 | 0 | 0.2 | 0.2 | 0.5 | 0.3 |  | 0.3 | 0.03 |  | 1.18 |
| 15 |  |  |  | 5.1 | 4 | 6.7 | 5.8 | 3 | 3.7 | 8.8 | 5.5 | 0.7 | 0 | 1.1 | 6.1 | 1.1 | 0.5 | 0 | 0.1 | 0.1 | 0 | 3.1 | 0.3 |  | 7.7 | 0.07 |  | 1.26 |
| ＋ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| $\stackrel{\text { 品 }}{\substack{0}}$ | 슥 | இЮ | 숙 | が | ন্ন | O O | O-O | No | ÒN | O | NiN | O O | Nò | O~N | Oio | Oì | $\underset{\sim}{\underset{N}{7}}$ | Nì | $\underset{\sim}{\underset{N}{2}}$ | $\underset{\sim}{\text { I }}$ | $\stackrel{\sim}{N}$ | $\begin{aligned} & 0 \\ & \underset{N}{1} \end{aligned}$ | ̇ㅜN | $\underset{\sim}{\infty}$ | $\underset{\sim}{7}$ | Oi | Nָ | N |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA | $\begin{aligned} & 0.0 \\ & 47 \end{aligned}$ | NA | NA |
|  | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.06 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |  | 0.0 | 0.0 | 0.0 |
| 1 | 76 | 07 | 63 | 63 | 63 | 75 | 67 | 66 | 74 | 76 | 7 | 74 | 15 | 63 | 74 | 77 | 6 | 69 | 77 | 78 | 62 | 7 | 6 | 61 | NA | 57 | 68 | 61 |
|  | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |  | 0.0 | 0.0 | 0.1 | 0.1 | 0.0 | 0.0 | 0.08 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 |
| 2 | 26 | 23 | 02 | 02 | 02 | 0.1 | 9 | 96 | 05 | 05 | 87 | 98 | 1 | 96 | 87 | 01 | 92 | 9 | 99 | 1 | 99 | 93 | 86 | 93 | 11 | 86 | 79 | 95 |
|  | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.0 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.10 | 0.1 | 0.1 | 0.1 | 0.0 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.0 | 0.1 |
| 3 | 25 | 43 | 26 | 26 | 26 | 37 | 94 | 29 | 22 | 22 | 04 | 16 | 4 | 09 | 13 | 18 | 96 | 18 | 12 | 13 | 3 | 15 | 13 | 31 | 25 | 19 | 99 | 2 |
|  | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.11 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 4 | 33 | 56 | 42 | 42 | 42 | 52 | 17 | 55 | 36 | 46 | 33 | 24 | 5 | 25 | 34 | 37 | 15 | 42 | 38 | 35 | 5 | 26 | 31 | 47 | 55 | 32 |  | 34 |
|  | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |  | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |
| 5 | 46 | 77 | 6 | 6 | 6 | 65 | 59 | 71 | 64 | 74 | 59 | 41 | 0.13 | 45 | 51 | 55 | 45 | 52 | 66 | 44 | 69 | 58 | 73 | 7 | 65 | 91 | 43 | 58 |
|  | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.16 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.2 | 0.1 | 0.1 | 0.1 |
| 6 | 64 | 87 | 75 | 75 | 75 | 92 | 83 | 95 | 8 | 98 | 97 | 78 | 3 | 6 | 82 | 83 | 66 | 72 | 8 | 77 | 96 | 55 | 89 | 89 | 02 | 74 | 38 | 88 |


| 品 | 슥 | ஜঃ | 숙 | ®® | 욱 | O O | O- | No | ÒN | O | Li in | OO | Ò | ষ্N | OiO | Oì | ت묵 | Ṅ্N | $\underset{\sim}{\underset{\sim}{N}}$ | $\underset{\sim}{\text { I }}$ | $\stackrel{n}{2}$ | $\begin{aligned} & 0 \\ & \underset{N}{1} \\ & \hline \end{aligned}$ | 수N | $\stackrel{\infty}{\underset{\sim}{N}}$ | $\underset{\sim}{\sim}$ | N్N | N | $\underset{\sim}{N}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 7 | 0.1 | 0.2 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.2 | 0.1 | 0.2 | 0.2 | 0.2 | 0.19 | 0.1 | 0.1 | 0.2 | 0.1 | 0.1 |  | 0.1 | 0.2 | 0.1 | 0.1 | 0.2 | 0.2 | 0.2 | 0.1 | 0.2 |
|  | 61 | 03 | 99 | 99 | 99 | 94 | 98 | 16 | 93 | 24 | 38 | 12 | 2 | 93 | 95 | 06 | 93 | 83 | 0.2 | 84 | 6 | 62 | 77 | 01 | 61 | 47 | 98 | 71 |
| 8 | 0.1 | 0.1 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.19 | 0.2 | 0.2 | 0.1 | 0.1 | 0.1 | 0.2 | 0.2 | 0.2 | 0.2 | 0.1 |  | 0.2 | 0.3 | 0.1 | 0.2 |
|  | 78 | 95 | 31 | 31 | 31 | 16 | 01 | 27 | 12 | 29 | 48 | 47 | 7 | 21 | 58 | 99 | 93 | 88 | 16 | 01 | 9 | 35 | 88 | NA | 48 | 06 | 85 | 43 |
| 9 | 0.1 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.25 | 0.2 | 0.2 | 0.2 | 0.3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 |  | 0.2 | 0.3 | 0.2 | 0.3 |
|  | 65 | 18 | 5 | 5 | 5 | 44 | 37 | 28 | 4 | 56 | 59 | 36 | 7 | 86 | 53 | 41 | 05 | 12 | 23 | 22 | 65 | 46 | 22 | NA | 61 | 08 | 58 | 16 |
| 10 | 0.1 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.2 | 0.25 | 0.2 | 0.3 | 0.2 | 0.3 | 0.2 | 0.2 | 0.2 | 0.3 | 0.3 | 0.2 |  | 0.3 | 0.3 |  | 0.4 |
|  | 73 | 41 | 59 | 59 | 59 | 83 | 46 | 53 | 69 | 9 | 87 | 86 | 5 | 96 | 22 | 27 | 34 | 04 | 26 | 2 | 12 | 59 | 33 | NA | 04 | 13 | 0 | 53 |
| 11 | 0.3 | 0.3 |  |  |  | 0.2 | 0.2 | 0.3 | 0.2 |  | 0.3 | 0.2 | 0.51 | 0.2 | 0.4 | 0.2 | 0.3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.3 | 0.2 |  | 0.3 | 0.3 | 0.2 | 0.5 |
|  | 17 | 07 | 0.3 | 0.3 | 0.3 | 86 | 6 | 03 | 4 | 0.3 | 35 | 37 | 7 | 73 | 22 | 84 | 45 | 75 | 42 | 64 | 62 | 69 | 57 | NA | 01 | 39 | 9 | 26 |
| 12 | 0.2 | 0.2 | 0.3 | 0.3 | 0.3 | 0.3 | 0.2 | 0.2 | 0.2 | 0.2 | 0.3 | 0.2 | 0.27 | 0.3 | 0.4 | 0.2 | 0.4 | 0.1 | 0.2 | 0.2 | 0.3 | 0.3 |  |  | 0.4 | 0.3 | 0.4 | 0.4 |
|  | 33 | 11 | 29 | 29 | 29 | 54 | 86 | 93 | 98 | 97 | 49 | 61 | 9 | 09 | 47 | 34 | 08 | 95 | 63 | 87 | 18 | 79 | NA | NA | 11 | 78 | 31 | 8 |
| 13 | 0.2 | 0.2 | 0.3 | 0.3 | 0.3 | 0.3 | 0.2 | 0.3 | 0.3 | 0.3 | 0.3 | 0.2 | 0.33 | 0.3 | 0.3 | 0.2 | 0.4 |  | 0.2 | 0.2 | 0.3 | 0.2 |  |  | 0.4 | 0.3 | 0.2 | 0.4 |
|  | 41 | 58 | 67 | 67 | 67 | 16 | 87 | 17 | 56 | 01 | 38 | 67 | 9 | 75 | 83 | 88 | 74 | NA | 62 | 52 | 51 | 42 | NA | NA | 2 | 25 | 62 | 78 |
| 14 | 0.3 | 0.2 | 0.2 | 0.2 | 0.2 |  | 0.2 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.41 | 0.2 | 0.3 | 0.3 | 0.4 | 0.1 | 0.5 | 0.4 | 0.2 | 0.3 | 0.2 |  | 0.4 | 0.3 |  | 0.4 |
|  | 48 | 77 | 99 | 99 | 99 | NA | 95 | 2 | 16 | 38 | 73 | 02 | 4 | 77 | 62 | 15 | 15 | 87 | 59 | 08 | 35 | 9 | 14 | NA | 29 | 89 | NA | 75 |
| 15+ | 0.3 | 0.2 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.3 | 0.4 | 0.3 | 0.4 |  | 0.3 | 0.4 | 0.3 | 0.4 |  | 0.3 | 0.2 |  | 0.3 | 0.2 |  | 0.4 | 0.3 |  | 0.5 |
|  | 48 | 77 | 6 | 6 | 6 | 5 | 36 | 9 | 53 | 02 | 75 | 04 | 0 | 89 | 6 | 51 | 75 | 0 | 39 | 73 | 0 | 78 | 6 | NA | 31 | 7 | NA | 08 |
| cm | length |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | - | \% | 숙 | $\stackrel{\infty}{\circ}$ | 육 | 잉 | O- | No | ÒN | ষ্ণ | 잉 | OO O | 잉 | O~N | Oio | Oì | ت국 | Nì | $\underset{\sim}{\underset{\sim}{n}}$ | $\stackrel{\rightharpoonup}{\mathrm{N}}$ | $\stackrel{n}{N}$ | $\begin{aligned} & 0 \\ & \underset{N}{1} \end{aligned}$ | Nì | $\stackrel{\sim}{\sim}$ | $\underset{\sim}{7}$ | O్N | ㄷN | $\underset{\sim}{N}$ |
| 0 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 18 |  | NA |
| 1 | $\begin{aligned} & 19 . \\ & 2 . \end{aligned}$ | $\begin{aligned} & 19 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 19 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 19 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 19 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 19 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 19 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 19 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 20 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 19 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 18 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 20 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 19 . \\ & 9 \end{aligned}$ | 20 | $\begin{aligned} & 20 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 20 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 19 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 19 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 20 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 20 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 19 . \\ & 8 \end{aligned}$ | 20 | $\begin{aligned} & 19 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 19.5 \\ & 1 \end{aligned}$ |  | $\begin{aligned} & 19 . \\ & 3 \end{aligned}$ | 20 | $\begin{aligned} & 19 . \\ & 4 \end{aligned}$ |


| $\stackrel{\text { 曻 }}{\substack{4}}$ | 늑 | இু | 숙 | © | 욱 | O O | O- | Nì | 으N | O | 잇 | O | Nò | Oi | 웃 | Oi | $\stackrel{\underset{\sim}{7}}{ }$ | N | $\underset{\sim}{\underset{\sim}{N}}$ | $\stackrel{\underset{N}{A}}{\substack{2}}$ | $\stackrel{\sim}{N}$ | $\begin{aligned} & 0 \\ & \underset{N}{N} \end{aligned}$ | $\stackrel{N}{N}$ | $\stackrel{\infty}{\underset{\sim}{N}}$ | 국 | N్N | N | N N |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | 22 | 22 | 22 | 22 | 22 | $\begin{aligned} & 21 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 21 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 21 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 22 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 22 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 21 . \\ & 5 \end{aligned}$ | 22 | $\begin{aligned} & 20 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 21 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 21 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 22 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 21 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 21 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 22 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 22 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 22 . \\ & 9 \end{aligned}$ | 22 | $\begin{aligned} & 21 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 22.1 \\ & 9 \end{aligned}$ | $\begin{aligned} & 23 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 21 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 21 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 22 . \\ & 4 \end{aligned}$ |
| 3 | 23. 5 | $\begin{aligned} & 23 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 23 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 23 . \\ & 5 \end{aligned}$ | 23. 5 | 23. 9 | 21. 9 | 23. 8 | 23. 7 | 23. 6 | 22. 9 | 23. 4 | 22. 6 | 23. 2 | 23. 2 | 23. 9 | 23 | 23. 5 | 23. 5 | $\begin{aligned} & 23 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 23 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 23 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 24.6 \\ & 7 \end{aligned}$ | $24 .$ $4$ | $\begin{aligned} & 24 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 22 . \\ & 9 \end{aligned}$ | $24 .$ $1$ |
| 4 | $\begin{aligned} & 24 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 23 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 7 \end{aligned}$ | $24 .$ $1$ | $\begin{aligned} & 23 . \\ & 6 \end{aligned}$ | $24 .$ $1$ | $\begin{aligned} & 24 . \\ & 6 \end{aligned}$ | 25 | $\begin{aligned} & 24 . \\ & 5 \end{aligned}$ | 25 | $\begin{aligned} & 25 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 25 \\ & 8 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 8 \end{aligned}$ | $24 .$ $1$ | $\begin{aligned} & 25.5 \\ & 8 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 9 \end{aligned}$ |
| 5 | $\begin{aligned} & 25 \\ & 5 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 5 \end{aligned}$ | 26 | $\begin{aligned} & 26 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 24 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 7 \end{aligned}$ | 27 | $\begin{aligned} & 25 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 26.7 \\ & 8 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 25 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 6 \end{aligned}$ |
| 6 | $26 .$ $4$ | $\begin{aligned} & 26 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 6 \end{aligned}$ | $27 .$ $5$ | $\begin{aligned} & 27 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 7 \end{aligned}$ | 27 | $\begin{aligned} & 26 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 2 \end{aligned}$ | $27 .$ $1$ | $\begin{aligned} & 27 . \\ & 6 \end{aligned}$ | 27 | $27 .$ $1$ | $\begin{aligned} & 27 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 26 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 5 \end{aligned}$ | 27.5 | $\begin{aligned} & 28 . \\ & 1 \end{aligned}$ | 27 | $\begin{aligned} & 25 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 2 \end{aligned}$ |
| 7 | $\begin{aligned} & 27 . \\ & 2 \end{aligned}$ | $27 .$ $2$ | $\begin{aligned} & 27 . \\ & 2 \end{aligned}$ | $27 .$ $2$ | $\begin{aligned} & 27 . \\ & 2 \end{aligned}$ | $28 .$ $1$ | $28 .$ $1$ | $\begin{aligned} & 28 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 9 \end{aligned}$ | $28 .$ $1$ | $\begin{aligned} & 28 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 7 \end{aligned}$ | $27 .$ $1$ | $\begin{aligned} & 28 . \\ & 3 \end{aligned}$ | $27 .$ $5$ | $\begin{aligned} & 30 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 28.0 \\ & 4 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | 28 | $\begin{aligned} & 30 . \\ & 5 \end{aligned}$ |
| 8 | $\begin{aligned} & 29 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 3 \end{aligned}$ | 29 | $\begin{aligned} & 29 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 9 \end{aligned}$ | 28 | $\begin{aligned} & 31 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 2 \end{aligned}$ | 28 |  | 30 | $\begin{aligned} & 31 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 27 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 3 \end{aligned}$ |
| 9 | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 28 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 1 \end{aligned}$ |  | $\begin{aligned} & 30 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 9 \end{aligned}$ |
| 10 | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 5 \end{aligned}$ | 31 | $\begin{aligned} & 31 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 5 \end{aligned}$ | 30 | $\begin{aligned} & 32 . \\ & 5 \end{aligned}$ | 28 | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 34 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 29 . \\ & 5 \end{aligned}$ |  | $\begin{aligned} & 32 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 1 \end{aligned}$ |  | $\begin{aligned} & 36 . \\ & 3 \end{aligned}$ |
| 11 | $\begin{aligned} & 30 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 33 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 39 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 6 \end{aligned}$ | 35 | $\begin{aligned} & 32 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 33 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 1 \end{aligned}$ | 30 | $\begin{aligned} & 30 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 35 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 1 \end{aligned}$ |  | $\begin{aligned} & 32 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 37 . \\ & 7 \end{aligned}$ |
| 12 | $\begin{aligned} & 32 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 1 \end{aligned}$ | $33 .$ $6$ | 32 | $\begin{aligned} & 31 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 9 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 35 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 35 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 34 . \\ & 6 \end{aligned}$ | $27 .$ $5$ | $\begin{aligned} & 30 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 35 . \\ & 5 \end{aligned}$ |  |  | 36 | $34 .$ $4$ | $\begin{aligned} & 34 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 36 . \\ & 8 \end{aligned}$ |
| 13 | 33. 3 | $\begin{aligned} & 33 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 33 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 33 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 33 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 33 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 7 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 9 \end{aligned}$ | 34 | $\begin{aligned} & 32 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 33 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 33 . \\ & 9 \end{aligned}$ | 34 | $\begin{aligned} & 31 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 36 . \\ & 4 \end{aligned}$ |  | $\begin{aligned} & 32 . \\ & 1 \end{aligned}$ | 30 | $\begin{aligned} & 32 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 5 \end{aligned}$ |  |  | $\begin{aligned} & 36 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 5 \end{aligned}$ | 37 |


| $\underset{\substack{\infty \\ \multirow{2}{*}{\hline}\\ \hline}}{ }$ | 늑 | ু | 욱 | $\stackrel{\infty}{\circ}$ | 욱 | 웃 | O-̇ | Nì | Nò | O | Nì | OO | Nò | Oi ì | Oio | Oì | $\underset{\sim}{7}$ | $\underset{\sim}{\sim}$ | $\underset{\sim}{n}$ | $\underset{\sim}{\text { N }}$ | $\stackrel{n}{\sim}$ | $\begin{aligned} & 0 \\ & \underset{N}{1} \end{aligned}$ | Nì | $\stackrel{\infty}{\sim}$ | $\stackrel{\sim}{\mathrm{N}}$ | N్N | N | $\underset{\sim}{\sim}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 14 | $\begin{aligned} & 31 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 1 \end{aligned}$ |  | $\begin{aligned} & 32 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 6 \end{aligned}$ | 33 | $\begin{aligned} & 34 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 34 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 34 . \\ & 2 \end{aligned}$ | 33 | 36 | $\begin{aligned} & 27 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 38 . \\ & 5 \end{aligned}$ | 36 | $\begin{aligned} & 30 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 36 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 30 . \\ & 5 \end{aligned}$ |  | $\begin{aligned} & 36 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 35 . \\ & 5 \end{aligned}$ |  | 37 |
| 15+ | $\begin{aligned} & 32 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 33 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 33 . \\ & 4 \end{aligned}$ | $\begin{aligned} & 34 . \\ & 3 \end{aligned}$ | $\begin{aligned} & 33 . \\ & 6 \end{aligned}$ | $\begin{aligned} & 34 . \\ & 8 \end{aligned}$ | $\begin{aligned} & 35 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 35 . \\ & 3 \end{aligned}$ |  | $\begin{aligned} & 35 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 36 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 34 . \\ & 5 \end{aligned}$ | $\begin{aligned} & 36 . \\ & 9 \end{aligned}$ |  | $\begin{aligned} & 34 . \\ & 2 \end{aligned}$ | $\begin{aligned} & 32 . \\ & 5 \end{aligned}$ |  | $\begin{aligned} & 36 . \\ & 1 \end{aligned}$ | $\begin{aligned} & 31 . \\ & 5 \end{aligned}$ |  | $\begin{aligned} & 36 . \\ & 5 \end{aligned}$ | 33 |  | $37 .$ $8$ |

Table 6.3.3. North Sea Horse Mackerel stock. Mean weight at age (grams) in the catch by area for all quarters in 2022

Q1-Q4

| Ages | 27.3.a (Q1.2) | 27.4.a(Q1.2) | 27.4.b | 27.4.c | 27.7.d | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 60.5 | 60.5 | 60.7 | 60.4 | 61.6 | 61.1 |
| 2 | 97.4 | 97.4 | 96.6 | 96.5 | 93.4 | 94.8 |
| 3 | 92.1 | 120.7 | 125.5 | 127.7 | 117.6 | 120.5 |
| 4 | 113.1 | 128.2 | 137.5 | 136.1 | 133.8 | 134.5 |
| 5 | 151.4 | 161.7 | 175.5 | 186.3 | 152.4 | 158.5 |
| 6 | 189.5 | 194.4 | 197.5 | 198.4 | 185.5 | 188.4 |
| 7 | 304 | 425.4 | 267.6 | 285.6 | 226 | 270.9 |
| 8 | 245.9 | 434.8 | 232.3 | 228.8 | 217.7 | 242.8 |
| 9 | 312.5 | 452.2 | 290.9 | 314.3 | 261.9 | 316.1 |
| 10 | 448.2 | 453.8 | 453.7 | 453.4 | 449 | 452.9 |
| 11 | 526 | 526 | 526 | 526 | 526 | 526 |
| 12 | 459.4 | 483.3 | 483 | 482.1 | 463.8 | 480.3 |
| 13 | 466.6 | 479.8 | 479.6 | 479.1 | 468.7 | 477.8 |
| 14 | 471.2 | 475.8 | 475.7 | 475.4 | 471.8 | 475 |
| 15+ | 494.8 | 510.4 | 510.2 | 509.5 | 497.3 | 508.1 |

Table 6.3.4. North Sea Horse Mackerel stock. Mean length (mm) at age in the catch by area for all quarters in 2022

1-4Q

| Ages | 27.3.a (Q1.2) | 27.4.a(Q1.2) | 27.4.b | 27.4.c | 27.7.d | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 191.6 | 191.6 | 192.3 | 192.1 | 195.1 | 191.6 |
| 2 | 222.9 | 222.9 | 223.1 | 222.8 | 224.3 | 222.9 |
| 3 | 241.5 | 240.6 | 243.2 | 243.8 | 240.4 | 227.5 |
| 4 | 266.7 | 286.3 | 248.8 | 248.1 | 249.7 | 241.5 |
| 5 | 321.6 | 254 | 278.7 | 277.2 | 283.2 | 291 |
| 7 |  |  | 300.4 | 304.7 | 292.9 | 321.6 |


| Ages | 27.3.a (Q1.2) | 27.4.a(Q1.2) | 27.4.b | 27.4.c | 27.7.d | Total |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 8 | 306.2 | 358 | 295.1 | 292.8 | 296.7 | 306.2 |
| 9 | 330.1 | 361.4 | 316.1 | 328.1 | 315.8 | 330.1 |
| 10 | 362 | 363.1 | 363 | 363 | 362.2 | 362 |
| 11 | 364.3 | 369.1 | 377 | 377 | 377 | 377 |
| 12 | 368.1 | 369.8 | 370.4 | 370.3 | 368.5 | 368.1 |
| 13 | 369 | 378.7 | 369.8 | 369.8 | 369.1 | 369 |
| 14 | 376.1 |  | 378.7 | 378.5 | 376.5 | 376.1 |
| $15+$ | 376.5 |  |  | 364.3 |  |  |

Table 6.5.1. North Sea Horse Mackerel. CPUE Indices of abundance (number/hour) for the juvenile (<20cm) and exploitable ( $\geq 20 \mathrm{~cm}$ ) sub-stock. estimated as a combined index for the NS-IBTS Q3 and the FR-CGFS in Q4. The survey indices are derived from the prediction of a hurdle model fit to data over the period 1992-2022 and include a 95\% confidence interval based on a bootstrapping procedure (CI_low = lower bound. CI_high = upper bound). Survey data from 2020 were not included in the modelling procedure as not all sampling stations of FR-CGFS could be visited in 2020, and therefore no reliable index value for $\mathbf{2 0 2 0}$ could be calculated.

|  | Exploitable sub-stock (>20 cm) |  |  | Juvenile sub-stock (<20 cm) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Index | Cl_low | Cl_high | Index | CI_low | Cl_high |
| 1992 | 1376 | 572 | 2889 | 4272 | 1928 | 9541 |
| 1993 | 555 | 274 | 986 | 1857 | 912 | 3738 |
| 1994 | 1168 | 519 | 2156 | 2591 | 1269 | 5252 |
| 1995 | 1347 | 535 | 2790 | 2020 | 1072 | 4349 |
| 1996 | 1054 | 460 | 1998 | 732 | 312 | 1505 |
| 1997 | 626 | 274 | 1156 | 2166 | 972 | 4857 |
| 1998 | 407 | 191 | 769 | 650 | 318 | 1247 |
| 1999 | 447 | 210 | 808 | 1439 | 750 | 2602 |
| 2000 | 422 | 212 | 762 | 1567 | 815 | 3050 |
| 2001 | 517 | 240 | 956 | 2181 | 1196 | 4981 |
| 2002 | 425 | 204 | 829 | 2394 | 1205 | 4780 |
| 2003 | 288 | 134 | 574 | 1787 | 927 | 3350 |
| 2004 | 351 | 156 | 644 | 1003 | 525 | 1805 |
| 2005 | 658 | 304 | 1252 | 803 | 419 | 1524 |
| 2006 | 697 | 312 | 1330 | 531 | 282 | 969 |
| 2007 | 344 | 154 | 708 | 601 | 320 | 1099 |


| 2008 | Exploitable sub-stock ( $\mathbf{2 0} \mathbf{~ c m}$ ) |  |  | Juvenile sub-stock (<20 cm) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 163 | 79 | 366 | 532 | 287 | 941 |
| 2009 | 98 | 44 | 206 | 691 | 364 | 1253 |
| 2010 | 195 | 81 | 402 | 2259 | 1067 | 4443 |
| 2011 | 234 | 108 | 474 | 504 | 257 | 936 |
| 2012 | 151 | 80 | 410 | 316 | 167 | 649 |
| 2013 | 188 | 75 | 413 | 1089 | 543 | 2060 |
| 2014 | 329 | 154 | 735 | 1547 | 795 | 2992 |
| 2015 | 447 | 186 | 923 | 1471 | 701 | 2907 |
| 2016 | 519 | 227 | 996 | 3056 | 1478 | 6178 |
| 2017 | 134 | 55 | 298 | 940 | 428 | 1925 |
| 2018 | 114 | 48 | 218 | 3163 | 1609 | 7417 |
| 2019 | 201 | 92 | 434 | 812 | 372 | 1677 |
| 2020 |  |  |  |  |  |  |
| 2021 | 122 | 49 | 245 | 1663 | 788 | 3226 |
| 2022 | 265 | 107 | 578 | 1447 | 698 | 2912 |

### 6.12 Figures

Catch and TAC


Figure 6.2.1. North Sea horse mackerel. Utilisation of quota from 2000 to 2022.


Figure 6.2.2. North Sea horse mackerel. Utilisation of quota by country in 2022.

North Sea Stock: Catch by division


Figure 6.2.3. North Sea horse mackerel. Catch in (1000 t) by division and year from 1982 to 2022.

Proportional catch by area


Figure 6.2.4. North Sea horse mackerel. Proportion of catches by ICES division from 2000 to 2022.


Figure 6.2.5. North Sea Horse Mackerel. Total catch (in tonnes) by ICES division, quarter, catch category and country in 2022.

NSHM: catch at age ( N ; observed) all areas


Figure 6.3.1. North Sea horse mackerel age distribution in the catch for 1995-2022. The size of bubbles is proportional to the catch number. Note that age 15 is a plus $g$

NSHM: catch at age ( N ; observed) 27.7.d


NSHM: catch at age ( N ; observed) miunus 27.7.d


Figure 6.3.2. North Sea horse mackerel. Bubble plots of age distribution in the catch by area for 1998-2022 for area 7.d (upper panel) and without 7.d (bottom panel). The size of bubbles is proportional to the catch numbers. Note that age 15 is a plus group.


Figure 6.3.3. North Sea horse mackerel. Mean weight at age in commercial catches over the period 2000-2022. Note that only age 1-10 are presented and that 10 is not a plus group.


Figure 6.3.4. North Sea horse mackerel. Mean length at age in commercial catches over the period 2000-2022. Note that only age 1-10 are presented and that 10 is not a plus group.


Figure 6.4.1. North Sea Horse Mackerel. Catch curves for the 1992 to 2015 cohorts, ages from 3 to $15+$. Values plotted on the vertical axis are the log(catch) values for each cohort in each year. The negative slope of these curves estimates total mortality $(Z)$ in the cohort.

## Total mortality by cohort



Figure 6.4.2. North Sea Horse Mackerel. Total mortality by cohort (Z) estimated from the negative gradients of the 19922015 cohort catch curves (Figure 6.4.1).


Figure 6.4.3. North Sea horse mackerel. ICES rectangles selected by WGWIDE in 2013 and currently used by the working group.


Figure 6.4.4. North Sea horse mackerel. Size distribution of North Sea horse mackerel catches during the inter-calibration exercise conducted in 2014 between the RV Gwen Drez (red bars) and Thalassa (blue bars).


Figure 6.4.5. North Sea horse mackerel. CPUE by depth for the CGFS survey from 1992 to 2017.


Figure 6.4.6. North Sea horse mackerel. Modelled abundance index from 1992-2019 including both UK and French stations in the FR-CGFS (blue) and excluding UK stations in the FR-CGFS (red) for the exploitable sub-stock ( $\geq 20 \mathrm{~cm}$ ).


Figure 6.4.7. North Sea horse mackerel. Modelled abundance index from 1992-2019 for the exploitable sub-stock ( $\geq 20$ cm ) for when UK sampling stations from FR-CGFS have been excluded for 2016 (top left), 2017 (top right), 2018 (bottom left) and 2019 (bottom right).


Figure 6.4.8. North Sea horse mackerel. Modelled abundance index from 1992-2019 for the exploitable sub-stock ( $\geq 20$ cm ) for when the FR-CGFS has been excluded for 2016 (top left), 2017 (top right), 2018 (bottom left) and 2019 (bottom right).


Figure 6.4.9. North Sea horse mackerel. Relative occurrence by length for the period 2017-2022 in the NS-IBTS.


Figure 6.4.10. North Sea horse mackerel. Relative occurrence by length for the period 2017-2022 in the FR-CGFS. Note that stations in UK waters could not be visited in 2020.

NSHM length frequency catches 27.7.d


Figure 6.4.11. North Sea horse mackerel. Length distributions in proportion to catch numbers from commercial catches in 27.7.d for the period 2016-2022.

NSHM length frequency catches 27.4.c


Figure 6.4.12. North Sea horse mackerel. Length distributions in proportion to catch numbers from commercial catches in 27.4.c in 2019-2022.


Figure 6.5.1. North Sea horse mackerel. CPUE per year of the exploitable sub-stock ( $\geq 20 \mathrm{~cm}$ ) from 1992 to 2021 as modelled by the hurdle model (red) that returned a warning when ran, and the zero-inflated model (grey).


Figure 6.5.2. North Sea horse mackerel. CPUE per year of the exploitable sub-stock ( $\geq 20 \mathrm{~cm}$ ) from 1992 to 2022 as modelled by the hurdle model at WGWIDE 2022 (grey) and WGWIDE 2023 (red). In the model the 2020 index value for the exploitable sub-stock ( $\mathbf{2 0} \mathbf{~ c m}$ ) is left out and the index is modelled using a Hurdle model. Complete overlap between last and current year is observed.


Figure 6.5.3. North Sea horse mackerel. Mean CPUE across hauls of the exploitable sub-stock ( $\geq 20 \mathrm{~cm}$ ) from 1992 to 2023 for the FR-CGFS (blue WGWIDE 2020 (not visible), grey WGWIDE 2021) and the NS-IBTS (black WGWIDE 2020, red WGWIDE 2023). Small changes in reported catches of NS-IBTS resulted in a higher index value for 2016.


Abundance index juvenile substock


Figure 6.5.4. North Sea Horse Mackerel. Joint CPUE survey index (number/hour) derived from the hurdle model fit to the NS-IBTS survey in the North Sea and the FR-CGFS survey in the Eastern English channel for the period 1991-2022. No index value for 2020 could be produced due to sampling issues in the FR-CGFS. Top: exploitable sub-stock ( $\geq 20 \mathrm{~cm}$ ), bottom: juvenile sub-stock ( $<20 \mathrm{~cm}$ ). Red shaded area represent the $95 \%$ confidence interval, which is determined by bootstrap resampling of Pearson residuals with 999 iterations.


Figure 6.5.5. North Sea horse mackerel. Proportion of hauls with zero catch for the exploitable ( $\mathbf{2 0} \mathbf{c m}$ ) and juvenile (<20 cm) sub-stocks in the NS-IBTS (blue) and the FR-CGFS (red) from 1992 to 2022. Note that the FR-CGFS 2020 values are based on French stations only, as UK stations could not be sampled.

## $>20 \mathrm{~cm}$ substock


<20cm substock


Figure 6.5.6. North Sea Horse Mackerel. Mean CPUE survey index (number/hour) obtained from the hurdle model fit to the NS-IBTS survey in the North Sea (in red), the FR-CGFS survey in the English channel (in grey) and the joint survey index (in blue). Top: exploitable sub-stock ( $\geq 20 \mathrm{~cm}$ ), bottom: juvenile sub-stock ( $<20 \mathrm{~cm}$ ). No index values for 2020 could be produced due to COVID-19 pandemic impacting the FR-CGFS.


Figure 6.5.7. Length distribution (cm), estimated parameters $L_{c}, L_{\text {mean, }} L_{F=M}(c m)$ and $F / F_{M S Y}$ as Fmsy Proxy ratio for 20162022. Length samples from commercial catches in ICES division 27.7.d. Recalculations for 2018 till 2020 have been performed for the constructions of these plots. Values for plots of 2019 till 2022 are based on the latest WKLife method for cat3 stocks.

## 7 Western horse mackerel

Trachurus trachurus in Subarea 8 and divisions 2.a, 4.a, 5.b, 6.a, 7.a-c, e-k (hom.27.2a4a5b6a7a-ce-k8)

### 7.1 TAC and ICES advice applicable to 2022 and 2023

Since 2011, the TACs cover areas in line with the distribution areas of the stock.
For 2022 the TAC was the following (EU 2022/109):

| Areas | TAC 2022 | Stocks fished in this area |
| :--- | :--- | :--- |
| 2.a, 4.a, 5.b, 6, 7.a-c, 7.e-k, 8.abde, 12, 14 | 49178 t | Western stock \& North Sea stock in 4.a 1-2 <br> quarters |
| 4.b,c, 7.d | 3461 t | North Sea stocks |
| Division 8.c | 2780 t | Western stock |

For 2023 the TAC was the following (EU 2023/194):

| Areas | TAC 2023 | Stocks fished in this area |
| :--- | :--- | :--- |
| 2.a, 4.a, 5.b, 6, 7.a-c, 7.e-k, 8.abde, 12, 14 | 13157 t | Western stock \& North Sea stock in 4.a 1-2 <br> quarters |
| 4.b,c, 7.d | 8969 t | North Sea stocks |
| Division 8.c | 2120 t | Western stock |

The TAC for the western stock should apply to the distribution area of western horse mackerel as follows:

All Quarters: 2.a, 5.b, 6.a, 7.a-c, 7.e-k, 8.a-e
Quarters 3\&4: 3.a (west), 4.a
The TAC for the North Sea stock should apply to the distribution area of North Sea horse mackerel as follows:

All Quarters: 3.a (east), 4.b-c, 7.d
Quarters 1\&2: 3.a (west), 4.a
In 2022, ICES advised on the basis of MSY approach that Western horse mackerel catches in 2023 should be zero. The Western horse mackerel TAC for 2023 is 15277 tonnes. The TAC should apply to the total distribution area of this stock. The horse mackerel catches in Division 3.a are taken outside the horse mackerel TACs.

### 7.1.1 The fishery in 2022

Information on the development of the fisheries by quarter and division is shown in Tables 5.1.1 and 5.1.2 and in Figures 5.1.1.a-d. The total catch allocated to Western horse mackerel in 2022 was 70114 tonnes, which is 11443 tonnes less than in 2021 and exceeds the ICES advice by 18156 tonnes. The catches of horse mackerel by country and area are shown in Tables 7.1.1.1-5 while the catches by quarter since 2000 are shown in Figure 7.1.1.1.

### 7.1.2 Estimates of discards

Discard data are available since 2000 for some countries. Prior to 2013, the estimates available are considered to be an underestimate (Figure 7.1.2.1).
In 2022, most countries have submitted discard information. Countries that reported discard estimates for horse mackerel were Denmark, France, Germany, Ireland, Sweden, Spain, UK (England and Wales, and Scotland). 2022 discard estimates for the Netherlands and Norway are considered to be equal to zero. Total discards for Western horse mackerel were 3352 tonnes, equal to $4.8 \%$ in weight of the total catches, a increase in comparison to last year.

Discard data are included in the assessment as part of the total catches.
Length frequency distributions of discards were provided by Spain, France, Germany, Ireland, and UK but are not included in the assessment length-frequency input data.

### 7.1.3 Stock description and management units

The Western horse mackerel stock is distributed in divisions 2.a, 5.b, 3.a, 4.a, 6.a, 7.a-c, 7.e-k and 8.a-e (for more details see Section 5.3 and Figure 7.1.3.1) and spawns in the Bay of Biscay, and in UK and Irish waters before parts of the stock migrate northwards into the Norwegian Sea and the North Sea where they are fished in the third and fourth quarter (note for area 4.a, only catches taken in quarters 3 and 4 are considered to be from the western stock). The western stock is considered a management unit and advised accordingly with TAC set in accordance with the distribution of the stock (note that catches in division 3.a are taken outside the TAC).

### 7.2 Scientific data

### 7.2.1 Egg survey estimates

In 2022 a new egg survey was carried out in the western and southern spawning areas with details given in section 8.6 .1 of this report. Data from the 1992-2022 surveys are used in the assessment. The time series of TAEP estimates used in the assessment is shown in table 7.2.1.1.

## 2022 egg survey results

The time series of egg production estimates for western horse mackerel is presented in Table 7.2.1.1 and Figure 7.2.1.1. In 2022, the total annual egg production (TAEP) of Western horse mackerel is $5.51 \cdot 10^{14}$, a $310 \%$ increase in comparison to the 2019 TAEP (ICES, 2023). The estimated TAEP in 2019 was the lowest production in the historical time series.
Figure 7.2.1.3 shows the horse mackerel production by period since 2007. In 2022, egg production was very low throughout all survey periods, with a clear peak in period 6 (June) (Figure 7.2.1.2).

In addition, egg production has been estimated using the Daily Egg Production Method (DEPM). Horse mackerel eggs at stage 1a were used to estimate the daily egg production. Peak spawning for western horse mackerel was expected to occur in period 6 . The average daily egg production and total spawning area were estimated for calculating egg production (Table 7.2.1.2) (ICES, 2023).

## Fecundity parameters

The sampling of horse mackerel was based on the DEPM method and took place during survey periods 6 and 7 (June and July). The sampling was carried out as described in the survey protocols (ICES, 2019), with emphasis on the need to collect sufficient samples for the fecundity analysis. During the 2022 MEGS surveys, 1648 female horse mackerel were sampled during the peak spawning periods (periods 6 and 7). A significant contribution to the number of females sampled came from the commercial fleet (ICES, 2023). Preliminary results of the adult DEPM parameters are summarized in Table 7.2.1.3.

SSB is then calculated as a ratio of total daily egg production to daily fecundity (DF, eggs day/g female) during period 6. Preliminary results are shown in Table 7.2.1.4.

### 7.2.2 Other surveys for Western horse mackerel

## Bottom-trawl surveys

A bottom-trawl survey index for recruitment was available for 2022. The recruitment index is based on IBTS surveys conducted by Ireland, France and Scotland, covering the main distribution of the stock (Bay of Biscay, Celtic Sea, West of Ireland and West of Scotland) from 2003 to 2022. A Bayesian Delta-GLMM is used to calculate an index of juvenile abundance based on catch rates, and the index is updated every year when new data become available (ICES, 2017a). The updated values are shown in Figure 7.2.2.1 (middle panel) and the indices estimated in 20202022 are given in Table 7.2.2.1. Annual revisions of the index are minor. The 2017 data point was highly uncertain due to very limited coverage of the French survey: the French research vessel had technical issues and could therefore only cover less than $1 / 3$ of the stations usually sampled. Despite this high uncertainty, the 2017 data point suggested a very strong recruitment to be expected the following year. This perception was confirmed by the presence of numerous small fish in the 2017 and 2018 catch data. The overall trend suggests an increase in recruitment from 2013 to 2017 and a decrease back down to 2016 levels in 2018. Recruitment in 2019 and 2020 decreased further and was close to the lowest values of the time series, followed by a small increase again in 2021 and 2022.

## Acoustic surveys

In the Bay of Biscay two coordinated acoustic surveys take place in spring, PELGAS (IfremerFrance) and PELACUS (IEO-Spain). Only the PELACUS survey, which cover the ICES division 8c, is used in the assessment. There is no biomass estimate for 2020 because the survey was cancelled due to the Covid-19 pandemic. The estimate for 2023 is shown in this report (Figure 7.2.2.1, Table 7.2.2.2.), but it is not part of the assessment this year (no catches available yet for 2023).

The biomass estimated by the PELACUS survey was high in the 1990s, reaching the maximum value in 1998 ( 139395 t ). Biomass values are lower in the $21^{\text {st }}$ century, peaking in $2010(53417 \mathrm{t}$ ) and $2015(67068 \mathrm{t})$. Biomass has fluctuated around 10000 t over the most recent 4 surveys.

### 7.2.3 Effort and catch per unit effort

No new information was presented on effort and catch per unit effort.

### 7.2.4 Catch in numbers

In 2022, the Netherlands (4a, 6a, 7bfghj, 8a), Ireland (6a, 7bgj), Norway (2a,4a), Spain (8bc), UK (England \& Northern Ireland) (6a, 7beffghj) and Germany (7ghj) provided catch in numbers-atage (Figure 7.2.4.1). The catch sampled for age readings in 2022 covered $74 \%$ of the total reported catch, which is a $5 \%$ less in comparison to last year. Catch in number-at-length were available from the Netherlands (4a, 6a, 7bcefghj, 8a), Germany (4.a, 7ghj), Ireland (6a, 7bghj), Norway (2a, 4a), Spain (8abcd) and UK (England, Northern Ireland \& Scotland) (4a, 6a, 7befghj) as well as from France ( $7 \mathrm{e}, 8 \mathrm{ab}$ ).

The total annual and quarterly catches in number for western horse mackerel in 2022 are shown in Table 7.2.4.1. The sampling intensity is discussed in Section 5.9.

The catch-at-age matrix is given in Table 7.2.4.2 and illustrated in Figures 7.2.4.2 and 7.2.4.3. The latter shows the dominance of the 1982-year class in the catches since 1984 until it entered the plus group in 1997. Since 2002, the 2001-year class, which entered the plus group in 2016, has been caught in considerable numbers. The 2008-year class can be followed in the catch data suggesting it was stronger than other year classes subsequent to the 2001.

Spain, Ireland, the Netherlands and UK (England) also provided the age length keys (ALK) for 2022. The combined ALK is showed in Table 7.2.4.4.

### 7.2.5 Length and age data

Mean length-at-age and mean weight-at-age in the catches
The mean weight and mean length-at-age in the catches by area, and by quarter in 2022 are shown in Tables 7.2.5.1 and 7.2.5.2. Weight-at-age time-series is shown in Figure 7.2.5.1.

Mean weight at age in the stock
Prior to 2017, estimates of mean weight-at-age in the stock for the assessment were based on catch weight-at-age from Q1 and Q2, (Table 7.2.5.3). At present, the stock weight-at-age used in the forecast is an output of the assessment (presented in Table 7.4.1). Further information can be found in the stock annex.

### 7.2.6 Maturity ogive

Maturity-at-age is presented in Table 7.2.6.1. In the assessment model a constant logistic function was used (Figure 7.2.6.1). Further information can be found in the stock annex.

### 7.2.7 Natural mortality

A fixed natural mortality of 0.15 year $^{-1}$ is assumed for all ages and years in the assessment. Further information can be found in the stock annex.

### 7.2.8 Fecundity data

Potential fecundity data ( $10^{6} \mathrm{eggs}$ ) per kg spawning females are available for the years 1987, 1992, 1995, 1998, 2000, 2001: the data are presented in Table 7.2.8.1 but were not used in the assessment model. In the assessment the fecundity is modelled as linear eggs/kg on body weight. Further information can be found in the stock annex.

### 7.2.9 Information from stakeholders

The EU fishing industry, partly in conjunction with the Pelagic Advisory Council (PELAC), has been working on a number of research projects relevant to Western horse mackerel that are briefly reported here. More details can be found in section 1.4.8 of this report and (WD02: Hintzen 2023). The Pelagic Freezer-trawler Association (PFA) provided an annual report on the self-sampling programme that started in 2015. The Western horse mackerel fishery takes place from October through to March of the subsequent year. Overall, the self-sampling activities for the Western horse mackerel fisheries during the years 2017-2023 (up to 02/06/2023) covered 244 fishing trips with 3096 hauls, a total catch of 116548 tonnes and 121815 individual length measurements. The main fishing areas are ICES division 27.6.a, division 27.7.b and division 27.7.j. Western horse mackerel have a wide range in the length distributions in the catch. Median lengths have fluctuated between 22.2 and 30.0 cm (for more details see WD02: Hintzen 2023).

There is also an industry-science collaboration aimed at improving the knowledge on gonad development of mackerel and horse mackerel. Samples were taken by the fishing industry (PFA vessels) on both targeted and by-catches of mackerel and/or horse mackerel. During 2023, a dedicated PFA industry researcher carried out one sampling trip on-board of a commercial trawler with the aim to collect frozen gonad samples, all frozen in a different way to assess if the way of freezing has an impact on the fecundity estimates. Samples were delivered to Wageningen Marine Research and results are still pending.

Recently genetic samples have been also collected on board of commercial vessels in the Autumn of 2020, as well as from 4.a during the NS-IBTS in Q3. Preliminary results suggest that the western horse mackerel stock could have a wider distribution than currently presumed, as the samples analysed from Q1 and Q2 in 4a were determined to be western horse mackerel. There is also mixing occurring with North Sea horse mackerel in area 7.d. See section 1.4.8.3 for further details.

### 7.2.10 Data exploration

The length frequency distributions of the landings for the entire fleet included in the model are shown in Figures 7.2.10.1-2. The length distributions 2015-2020 show a considerable amount of very small fish, mostly from Spanish catches. In 2021 the recent trend of large catches of small fish changed and the length distribution was a more normal distribution with the most common landed lengths around 30 cm .2022 trends are similar to 2021 but there are once again large amounts of fish caught below 20 cm . The main mode of the distribution continuously increased since 2004 to 2017 due to the growth of the small individuals observed in recent years. The length distribution of discards has been provided by some countries since 2018. However, this information was not available at the last benchmark (2017) and therefore they are not included in the current assessment.

Within-cohort consistency of the catch-at-age matrix is investigated in Figure 7.2.10.3: this shows that the catch-at-age data contains information on year-class strength that could form the basis
for an age-structured model. The numbers at age in the catch by decade show a slight trend towards younger individuals when moving from the beginning of the time-series towards the end (Figure 7.2.10.4).

The indices of abundance used in the assessment cover different areas and therefore represent different parts of the stock. Negative correlations between indices that should represent the same portion of the population may lead to problems in the fitting of the model. The correlation between time-series was therefore estimated and is presented in Figure 7.2.10.5. There was no strong correlation between the IBTS recruitment index and the other two surveys. The egg survey index, which aims to represent the adult portion of the stock was strongly positively correlated with the PELACUS acoustic survey biomass estimate.

### 7.2.11 Assessment model, diagnostics

A one fleet, one sex, one area stock synthesis model (SS; Stock Synthesis v3.30) is used for the assessment of Western horse mackerel stock in the Northeast Atlantic. A description of the model can be found in the stock annex. The assessment presented is an update of the 2022 assessment, with the inclusion of the 2022 biomass estimates from the three fishery independent surveys (IBTS, PELACUS, and the egg survey), the 2022 length frequency distribution of the landings and the survey PELACUS, and the 2022 total catch and conditional ALKs. As in last year 's assessment, the length and age distributions were tuned using the Francis reweighting approach within the model.

Fits to the available data are given in Figure 7.2.11.1, and model estimates with associated precision in Figure 7.2.11.2. The model does not fit well to the biomass estimates and length composition provided by the PELACUS survey. The fitting to the most recent length frequency distributions and the conditional ALKs remains sub-optimal. The model does not capture the small fish observed in recent years and substantially over predicts the proportion of large $>20 \mathrm{~cm}$ fish in 2022.

The assessment shows a strong retrospective pattern, with peels falling outside the confidence intervals of SSB (1 peel) and recruitment (3 peels) estimates (Figure 7.2.11.3). However, the ouptuts of the 2022 and 2023 assesements are similar, and the consistent downward rescaling of the SSB and upward scaling of F perceived in previous years has not been observed. The Mohn's rho values improved significantly compared last year, being 0.179 for SSB and -0.163 for F .

### 7.3 State of the stock

### 7.3.1 Stock assessment

The SS model with new biomass, length and age data is presented as the final assessment model. Stock numbers-at-age and fishing mortality-at-age are given in Tables 7.3.1.1 and 7.3.1.2, and a stock summary is provided in Table 7.3.1.3, and illustrated in Figure 7.2.11.2. SSB peaked in 1988 following the recruitment of the exceptionally strong 1982 year-class. Subsequently, SSB slowly declined until 2003 and then recovered again following the moderate-to-strong year-class of 2001 (a third of the size of the 1982 year-class). SSB reached the minimum values of the time series in 2017 (834 480 t), increasing slightly in recent years. In 2023, SSB is estimated to be below $\mathrm{Blim}_{\text {lim }}$.

The recruitment has been weak since 2001, reaching the lowest value in 2011. Recruitment estimates for 2014-2018 are the highest observed since 2008 and are higher than the geometric mean estimated over the years 1983-2022 (2 088520 thousands of individuals). Recruitment has been very low again since 2019.

Fishing mortality (ages 1-10) has oscillated over the time series. It increased after 2007 as a result of increasing catches and decreasing biomass as the 2001 year-class was reduced. The fishing mortality decreased between 2013 and 2017 due to a decrease in catches and a reduced proportion of the adult population in the exploited stock. The fishing mortality in $2022(0.075)$ is just above $\mathrm{F}_{\text {MSY }}(0.074)$ and lower than the previous year, making it the lowest value in the time series since 2008.

### 7.4 Short-term forecast

A deterministic short-term forecast was conducted using the ' fwd()$^{\prime}$ ' method in FLR (Flash R addon package).

## Input

Table 7.4.1. lists the input data for the short-term predictions. Weight at age in the stock and weight at age in the catch are equal to the year-invariant weight at age function used in the stock synthesis model. Exploitation pattern is based on estimated fishing mortality in 2022 and is the average of ages 1 to 10 . Natural mortality is assumed to be 0.15 across all ages. The proportion mature for this stock has a logistic form with fully mature individuals at age 4 as used in the assessment model.

The expected catch for the intermediate year was set at $100 \%$ of the TAC (15 277 t). Note that although the plus group in the catch was set at $15+$, the true population in SS model is set to arrive up to age 20 (as from literature) and is therefore estimated accordingly.

## Output

A range of predicted catch and SSB options from the short-term forecast are presented in Table 7.4.2.

### 7.5 Uncertainties in the assessment and forecast

Recent assessments (2019-2022) show a strong retrospective pattern with a consistent downward revision in absolute level of SSB and an upward revision in $\mathrm{F}_{1-10}$, although this year assessment is in line with last year.

The fitting to the fishery independent indices remains good for two of the three surveys used: IBTS and MEGS. A degradation of the fitting to the IBTS recruitment index was observed 20152018 but more recent estimates now once again fall within the confidence intervals. The fit to the PELACUS acoustic index remains poor (Figure 7.2.11.1).

The change in selectivity, which is detected from both the length and the age composition of the catch data, is not entirely picked up from the model. In general, the model tends to overestimate the mean age of the last decade. The selectivity issue should be further investigated and addressed: for example, it is not clear whether the high presence of small specimens in the landings data is due to the inclusion of BMS individuals in the overall catch instead of having it as discard (the discard ban was implemented in 2015 for pelagic species) or if this is due to an effective change in selectivity (i.e. catchability of the gear and availability of the stock).

The model fixes the realised fecundity with a constant number of eggs $/ \mathrm{kg}$ independently of the individual weight. However, Western horse mackerel is known to be an indeterminate spawner, which implies this relationship may not be appropriate when it comes to the use of an egg survey as index of spawning biomass. During the benchmark in 2017 an attempt was made to estimate the parameters relative to fecundity, however, the information provided to the model was not
sufficient. The inclusion of this feature, whenever appropriate data become available, would help to improve the reliability of the assessment.

The assumed value for natural mortality should be investigated. However, there is no data available (such as tagging) that could assist in estimating natural mortality more accurately. Nevertheless, total mortality appears to be low, given the persistence of the 1982-year class in the catch data for 14 years.

The assessment, as was developed at the benchmark, has an increased amount of information for providing more robust estimates of recruitment, also informed when occasional strong year classes are observed in the catch. On the contrary, the SSB is informed only by the triennial egg survey and by the acoustic survey (which only covers a small part of the stock distribution and size ranges, has a very low weight in the model and is very noisy): a new index for the spawning biomass would therefore be beneficial for the future stability of this assessment. The development of a combined SSB index estimated from appropriate surveys in the area (e.g., PELACUS, PELGAS, WESPAS) should be pursued.

### 7.6 Comparison with previous assessment and forecast

A comparison of the update assessment with the historic ones (previous 4 years) is shown in Figure 7.2.11.4: the new information broke the pattern of recent years asessements and did not create a downward rescaling of the assessment biomass and upward revision of F. Recruitment, on the other hand, remains fairly stable until 2015 but a downward revision is estimated from then on.

### 7.7 Management options

### 7.7.1 MSY approach

In 2017 stochastic equilibrium analyses were carried out using the EqSim software (WKWIDE 2017) to provide an estimate for $\mathrm{F}_{\text {MSY }}$ and other biological reference points. During WGWIDE 2017 further investigations were carried out and summarised in a Working Document attached to WGWIDE 2017 report (ICES, 2017b).

Reference points were subsequently revised during an inter-benchmark workshop carried out in July-August 2019 as those derived during the 2017 benchmark were deemed no longer appropriate in light of the retrospective pattern observed in the model. More robust reference points were therefore put forward after a number of alternatives were examined, following ICES guidelines, and based on the 2018 assessment. The detailed rationale can be found in the inter-benchmark report (ICES, 2019).

SSB in 2003 was adopted as a proxy for $\mathrm{B}_{\mathrm{pa}}$ on the basis that fishing mortality had been relatively low for the data period ( $\mathrm{Fbar}^{\text {mean }} \sim 0.11$, natural mortality $=0.15$ ), and there was no indication of impaired recruitment below the associated Blim, despite a continuing decline in SSB. Fmsy was derived from stochastic simulations as before and evaluated at 0.074 . In 2021, $\mathrm{F}_{\mathrm{pa}}$ was re-defined as $\mathrm{F}_{\mathrm{p} 05}$ (ICES, 2021a). These updated reference points were used in determining the MSY based 2024 catch advice.

### 7.7.2 Management plans and evaluations

An overview of earlier management plans and management plan evaluations was presented at WGWIDE 2017. To date, no agreed management plan is available for this stock despite several
attempts to develop such management plans. The Pelagic Advisory Council (PELAC), together with several researchers have carried out an evaluation of potential harvest control rules for Western horse mackerel (ICES, 2021b). This rebuilding plan has not been approved by the European Commission and the UK, and the working group no longer considers it appropriate as it is outdated by 3 years.

### 7.8 Management considerations

The 2001 year-class has now entered the plus group but no other detectable very strong yearclasses entering the fishery, even though a higher amount of age 1-2 years old fish have been observed in the catches in the past 4-5 years.

Due to the SSB remaining below $\mathrm{B}_{\text {lim, }}$ the advice for 2023 is catches in 2024 should be zero. It is expected that even with 0 catch there will be some discard landings in 2024 available as with previous years. Nevertheless, a scientific monitoring quota should be considered to support the stock assessment and the advice in future years. Mixed fisheries considerations might be relevant due to horse mackerel being taken as by-catch in some areas.

There are potential issues with stock boundaries for horse mackerel (see section 1.4.8.3), which are subject to ongoing investigation. Mismatches between the stock and management unit could undermine the effectiveness of any management action. Note that subarea 8.c is included in the ICES advice for Western horse mackerel but currently an independent TAC is applied to horse mackerel in this Subarea.

The stock is distributed within the NEAFC regulatory areas, but the fishery rarely takes place in there. In the last three years only 10 tonnes were estimated to belong the NEAFC regulatory area (Table 7.4.3).

### 7.9 Ecosystem considerations

Knowledge about the distribution of the Western horse mackerel stock is mostly gained from the egg surveys and the seasonal changes in the fishery. Based on these observations it is not possible to infer a trend in the distribution of Western horse mackerel. However, from catch data it appears that the stock is concentrated in the southern areas, and it is mostly characterized by small individuals.

### 7.10 Regulations and their effects

There are horse mackerel management agreements between EU and the UK, but not with Norway. The TAC set by EU and the UK therefore only applies to EU and UK waters and the EU and UK fleet in international waters. The minimum landing size of horse mackerel by the EU and UK fleet is 15 cm ( $10 \%$ undersized allowed in the catches). In Norwegian waters there is no quota for horse mackerel but existing regulations on bycatch proportions as well as a general discard prohibition (for all species) apply to horse mackerel.

An overview of the scientific advice, the TACs (or sum of unilateral quota) and the catches is shown in Figure 7.10.1. From 2001 onwards, TACs and catches have fluctuated around the scientific advice, where in some years the TACs were set higher and in other years lower than the scientific advice.

### 7.11 Changes in fishing technology and fishing patterns

The description of the fishery is given in Section 5.1 and no large changes in fishing areas or patterns have taken place.

### 7.12 Changes in the environment

Migrations are closely associated with the slope current, and horse mackerel migrations are known to be modulated by temperature. Continued warming of the slope current is likely to affect the timing and spatial extent of this migration.

It has been reported a good correspondence between the modelled influx of Atlantic water to the North Sea in the first quarter and the horse mackerel catches taken by Norwegian purse-seiners in the Norwegian EEZ later in the year (October-November) since 1987 (Iversen et al. 2002).

### 7.13 References

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### 7.14 Tables

Table 7.1.1.1. Western horse mackerel. Catches ( $t$ ) in Subarea $\mathbf{2}$ by country (Data as submitted by Working Group members).

| Country | 1980 | 1981 | 1982 | 21983 |  | 1984 |  | 1985 | 1986 | 1987 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | - | - | - | - |  | - |  | - | - | 39 |
| France | - | - | - | - |  | 1 |  | 1 | - ${ }^{2}$ | - 2 |
| Germany, Fed.Rep | - | + | - | - |  | - |  | - | - | - |
| Norway | - | - | - | 412 |  | 22 |  | 78 | 214 | 3,272 |
| USSR | - | - | - | - |  | - |  | - | - | - |
| Total | - | + | - | 412 |  | 23 |  | 79 | 214 | 3,311 |
|  | 1988 | 1989 | 1990 | 1991 |  | 1992 |  | 1993 | 1994 | 1995 |
| Faroe Islands | - | - | 9643 | 1,115 |  | 9,157 ${ }^{3}$ |  | 1,068 | - | 950 |
| Denmark | - | - | - | - |  | - |  | - | - | 200 |
| France | -2 | - | - | - |  | - |  | - | 55 | - |
| Germany, Fed. Rep. | 64 | 12 | + | - |  | - |  | - | - | - |
| Norway | 6,285 | 4,770 | 9,135 | 3,200 |  | 4,300 |  | 2,100 | 4 | 11,300 |
| USSR / Russia (1992-) | 469 | 27 | 1,298 | 172 |  | - |  | - | 700 | 1,633 |
| UK (England + Wales) | - | - | 17 |  |  | - |  | - | - | - |
| Total | 6,818 | 4,809 | 11,414 | 4,487 |  | 13,457 |  | 3,168 | 759 | 14,083 |
|  | 1996 | 1997 |  | 1998 1 | 1999 | 2 | 2000 | 2001 | 2002 | 2003 |
| Faroe Islands | 1,598 | 7993 |  | $188^{3} \quad 13$ | $132^{3}$ |  |  | - | - | - |
| Denmark | - | - |  | 1,755 ${ }^{3}$ | - |  |  | - | - | - |
| France | - | - |  | - | - |  |  | - | - | - |
| Germany | - | - |  | - - | - |  |  | - | - | - |
| Norway | 887 | 1,170 |  | 234 2, | 2,304 |  | 841 | 44 | 1,321 | 22 |
| Russia | 881 | 554 |  | 345 | 121 |  | 78 | 16 | 3 | 2 |
| UK (England + Wales) | - | - |  | - - | - |  | - | - | - | - |
| Estonia | - | 78 |  | 22 - | - |  | - | - | - | - |
| Total | 3,366 | 2,601 |  | 2,544 | 2557 | 9 | 919 | 60 | 1,324 | 24 |


|  |  | 2004 | 2005 | 2006 | 2007 |  | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands |  | - | - | 3 | - |  | - | - | 222 | 224 | - | - |
| Denmark |  | - | - | - | - |  | - | - | - | - | - | - |
| France |  | - | - | - | - |  | - | - | - | - | + | - |
| Germany |  | - | - | - | $\cdot$ |  | - | - | - | - | - | - |
| Ireland |  | - | - | - | - |  | - | - | - | - | - | - |
| Netherlands |  | - | - | - | - |  | - | - | - | 1 | - | - |
| Norway |  | 42 | 176 | 27 | - |  | 572 | 1,847 | 1,364 | 298 | 66 | 30 |
| Russia |  | - | - | - | - |  | - | - | - | - | - | - |
| UK (England + Wales) |  | - | - | - | - |  | - | - | - | - | - | - |
| Estonia |  | - | - | - | - |  | - | - | - | - | - | - |
| Total |  | 42 | 176 | 27 | 0 |  | 572 | 1,847 | 1,586 | - | 66 | 30 |
|  | 2014 | 2015 |  | 2016 | 2017 |  | 2018 | 2019 | 2020 | 2021 | 2022 |  |
| Faroe Islands | - | - |  | - | - |  | - | - | - | - | 10* | - |
| Denmark | - | - |  | - | - |  | - | - | - | - | - | - |
| France | - | - |  | - | - |  | - | - | - | - | - | - |
| Germany | - | - |  | - | - |  | - | - | - | - | - | - |
| Ireland | - | - |  | - | - | - | - | - | - | - | - | - |
| Netherlands | 107 | - |  | - | - |  | - | - | - | - | - | - |
| Norway | 302 | 10 |  | 45 | 5 |  | 718 | 867 | 290 | 12 | 146 |  |
| Russia | - |  |  | - | - |  | - | - | - | - | - | - |
| UK (England + Wales) | - |  |  | - | - |  | - | - | - | - | - | - |
| Estonia | - |  |  | - | - |  | - | - | - | - | - | - |
| Total | 409 | 10 |  | 45 | 5 |  | 718 | 867 | 290 | 12 | 156 |  |
| ${ }^{2}$ Included in 4. <br> ${ }^{3}$ Includes catches in Div. ${ }^{4}$ Taken in Div. 5.b. |  |  |  |  |  |  |  |  |  |  |  |  |

Table 7.1.1.2. Western horse mackerel. Catches ( t ) in North Sea Subarea 4 and Skagerrak Division 3.a by country (Data submitted by Working Group members). Catches partly concern the North Sea horse mackerel.

| Country | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium | 8 | 34 | 7 | 55 | 20 | 13 | 13 | 9 | 10 |
| Denmark | 199 | 3,576 | 1,612 | 1,590 | 23,730 | 22,495 | 18,652 | 7,290 | 20,323 |
| Faroe Islands | 260 | - | - | - | - | - | - | - | - |
| France | 292 | 421 | 567 | 366 | 827 | 298 | 2312 | 1891 | 7841 |
| Germany, | + | 139 | 30 | 52 | + | + | - | 3 | 153 |
| Fed.Rep. | 1,161 | 412 | - | - | - | - | - | - | - |
| Ireland | 101 | 355 | 559 | 2,0292 | 824 | 1602 | 6002 | 8503 | 1,0603 |
| Netherlands | 119 | 2,292 | 7 | 322 | 2 | 203 | 776 | 11,7283 | 34,4253 |
| Norway2 | - | - | - | 2 | 94 | - | - | - | - |
| Poland | - | - | - | - | - | - | 2 | - | - |
| Sweden | 11 | 15 | 6 | 4 | - | 71 | 3 | 339 | 373 |
| UK (Engl. + | - | - | - | - | 3 | 998 | 531 | 487 | 5,749 |
| Wales) | - | - | - | 489 | - | - | - | - |  |
| UK (Scotland) |  |  |  |  |  |  |  |  |  |
| USSR |  |  |  |  |  |  |  |  |  |
| Total | 2,151 | 7,253 | 2,788 | $4,420,887$ | 24,238 | 20,808 | 20,895 | 62,877 |  |


| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Belgium | 10 | 13 | - | + | 74 | 57 | 51 | 28 | - |
| Denmark | 23,329 | 20,605 | 6,982 | 7,755 | 6,120 | 3,921 | 2,432 | 1,433 | 976 |
| Estonia | - | - | - | 293 | - | - | 17 | - | - |
| Faroe Islands | - | 942 | 340 | - | 360 | 275 | - | - | 296 |
| France | 248 | 220 | 174 | 162 | 302 | - | - | - | - |
| Germany, Fed.Rep. | 506 | $2,469^{4}$ | 5,995 | 2,801 | 1,570 | 1,014 | 1,600 | 7 | 37 |
| Ireland | - | 687 | 2,657 | 2,600 | 4,086 | 415 | 220 | 1,100 | 8,152 |
| Netherlands | 14,172 | 1,970 | 3,852 | 3,000 | 2,470 | 1,329 | 5,285 | 6,205 | 52 |
| Norway | 84,161 | 117,903 | 50,000 | 96,000 | 126,800 | 94,000 | 84,747 | 14,639 | 43,888 |
| Poland | - | - | - | - | - | - | - | - | - |
| Sweden | - | 102 | 953 | 800 | 697 | 2,087 | - | 95 | 1761 |
| UK (Engl. + Wales) | 10 | 10 | 132 | 4 | 115 | 389 | 478 | 40 | 10 |
| UK (N. Ireland) | - | - | 350 | - | - | - | - | - | - |
| UK (Scotland) | 2,093 | 458 | 7,309 | 996 | 1,059 | 7,582 | 3,650 | 2,442 | 10,511 |
| USSR / Russia (1992 -) | - | - | - | - | - | - | - | - | - |
| Unallocated+discards | $12,482^{3}$ | $-317^{3}$ | $-750^{3}$ | -2785 | $-3,270$ | 1,511 | -28 | 136 | $-31,615^{6}$ |
| Total | 112,047 | 145,062 | 77,904 | 114,133 | 140,383 | 112,580 | 98,452 | 26,125 | 34,068 |


| Country | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 19 | 21 | - | - | - | - | - | - | - |
| Denmark | 2,048 | 2,026 | 7 | 98 | 53 | 841 | 48 | 216 | 60 |
| Estonia | - | - | - | - | - | - | - | - | - |
| Faroe Islands | 28 | 908 | 24 | 0 | 671 | 5 | 76 | 35 | 0 |
| France | 379 | 60 | 49 | - | - | 255 | - | 1 | - |
| Germany | 4,620 | 4,072 | 0 | 0 | 4 | 534 | 0 | 44 | 1 |
| Ireland | - | 404 | 32 | 332 | 11 | 93 | 378 | - | - |
| Lithuania | - | - | - | - | - | - | - | - | - |
| Netherlands | 4,548 | 3,285 | 10 | 1 | 0 | 36 | 0 | 0 | 0 |
| Norway | 13,129 | 44,344 | 1,141 | 7,912 | 34,843 | 20,349 | 10,687 | 24,733 | 27,087 |
| Russia | - | - | 2 | - | - | - | - | - | - |
| Sweden | 1,761 | 1,957 | 1,009 | 68 | 561 | 1,002 | 567 | 216 | 0 |
| UK (Engl. + Wales) | 1 | 12 | - | - | - | - | 0 | - | - |
| UK (Scotland) | 3,041 | 1,658 | 3,054 | 3,161 | 252 | 0 | 0 | 22 | 61 |
| Unallocated+discards | 737 | -325 | 10 | 0 | 0 | -36 | 0 | 0 | 0 |
| Total | 30,311 | 58,422 | 5,338 | 11,572 | 36,395 | 23,079 | 11,756 | 25,267 | 27,210 |

${ }^{1}$ Includes Division 2.a. ${ }^{2}$ Estimated from biological sampling. ${ }^{3}$ Assumed to be misreported. ${ }^{4}$ Includes 13 t from the German Democratic Republic. ${ }^{5}$ Includes a negative unallocated catch of $-4,000 \mathrm{t} .{ }^{6}$ Negative values when there were overestimations of catch when comparing scientific with official data

| Country | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 74 | 2 | 207 | 61 | 19 | 9 | 0 | 23 |
| Faroe Islands | 3 | 55 | 0 | 8 | 0 | 0 | 0 | 53 |
| France | - | 1 | - | - | 268 | - | - | 17 |
| Germany, Fed.Rep. | 6 | 93 | 0 | 4 | 0 | 0 | 20 | 0 |
| Ireland | 651 | 298 | 342 | 14 | 755 | 25 | 7 | - |
| Netherlands | - | - | - | - | - | - | - | - |
| Lithuania | 22 | 0 | 7 | 339 | 81 | 92 | 0 | 310 |
| Norway | 4180 | 11631 | 57890 | 10556 | 13409 | 3183 | 6566 | 14051 |
| Sweden | 76 | 9 | 258 | 2 | 90 | 0 | 1 | 0 |
| UK (Engl. + Wales) | 31 | - | - | - | - | - | 16 | 203 |
| UK (Scotland) | 7 | 20 | 51 | 546 | 101 | 12 | 102 | 11 |
| Unallocated +discards | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 30 |
| Total | 5050 | 12110 | 58755 | 11531 | 14723 | 3320 | 6712 | 14699 |
| Country | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| Denmark | 37 | 7 | 21 | 289 | 183 | 22 | 11 | 211 |
| Faroe Islands | 0 | 0 | 67 | 0 | 6 | - | - | 96 |
| France | 12 | 4 | 1 | 2 | 98 | 0 | 2 | 8 |
| Germany, Fed.Rep. | 6 | 28 | 1 | 1 | 5 | 0.5 | 3 | 11 |
| Ireland | 8 | - | - | - | - | - | - |  |
| Netherlands | - | 0 | 14 | 7 | 72 | 1 | 27 | 53 |
| Lithuania | 12 | 130 | - | - | - | 0 | - |  |
| Norway | 8,887 | 8,765 | 9,880 | 8,601 | 8,154 | 10,376 | 3,651 | 2984 |
| Sweden | 10 | 0 | 41 | 23 | 323 | 83 | 4 | 3 |
| UK (Engl. + Wales) | 134 | 13 | 4 | 0 | - | 0 | 0.5 | 3 |
| UK (Scotland) | 36 | 14 | - | - | 50 | - | 63 | 102 |
| Unallocated +discards | 32 | 97 | 87 | 162** | 339 | 1239 | 160 | 711 |
| Total | 9,175 | 9,057 | 10,117 | 9,085 | 9144 | 11,700 | 3,923 | 4182 |

Table 7.1.1.3 Western horse mackerel. Catches ( $\mathbf{t}$ ) in Subarea 6 by country (Data submitted by Working Group members).

| Country | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 734 | 341 | 2,785 | 7 | - | - | - | 769 | 1,655 |
| Faroe Islands | - | - | 1,248 | - | - | 4,014 | 1,992 | 4,450 ${ }^{2}$ | 4,000 ${ }^{2}$ |
| France | 45 | 454 | 4 | 10 | 14 | 13 | 12 | 20 | 10 |
| Germany, Fed. Rep. | 5,550 | 10,212 | 2,113 | 4,146 | 130 | 191 | 354 | 174 | 615 |
| Ireland | - | - | - | 15,086 | 13,858 | 27,102 | 28,125 | 29,743 | 27,872 |
| Netherlands | 2,385 | 100 | 50 | 94 | 17,500 | 18,450 | 3,450 | 5,750 | 3,340 |
| Norway | - | 5 | - | - | - |  | 83 | 75 | 41 |
| Spain | - | - | - | - | - |  | -1 | _1 | -1 |
| UK (Engl. + Wales) | 9 | 5 | + | 38 | + | 996 | 198 | 404 | 475 |
| UK (N. Ireland) |  |  |  |  |  | - | - | - | - |
| UK (Scotland) | 1 | 17 | 83 | - | 214 | 1,427 | 138 | 1,027 | 7,834 |
| USSR. | - | - | - | - | - | - | - | - | - |
| Unallocated + disc |  |  |  |  |  | -19,168 | $-13,897$ | -7,255 | - |
| Total | 8,724 | 11,134 | 6,283 | 19,381 | 31,716 | 33,025 | 20,455 | 35,157 | 45,842 |
| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| Denmark | 973 | 615 | - | 42 | - | 294 | 106 | 114 | 780 |
| Faroe Islands | 3,059 | 628 | 255 | - | 820 | 80 | - | - | - |
| France | 2 | 17 | 4 | 3 | + | - | - | - | 53 |
| Germany, Fed. Rep. | 1,162 | 2,474 | 2,500 | 6,281 | 10,023 | 1,430 | 1,368 | 943 | 229 |
| Ireland | 19,493 | 15,911 | 24,766 | 32,994 | 44,802 | 65,564 | 120,124 | 87,872 | 22,474 |
| Netherlands | 1,907 | 660 | 3,369 | 2,150 | 590 | 341 | 2,326 | 572 | 1335 |
| Norway | - | - | - | - | - | - | - | - | - |
| Spain | -1 | -1 | 1 | 3 | - | - | - | - | - |
| UK (Engl. + Wales) | 44 | 145 | 1,229 | 577 | 144 | 109 | 208 | 612 | 56 |
| UK (N.Ireland) | - | - | 1,970 | 273 | - | - | - | - | 767 |
| UK (Scotland) | 1,737 | 267 | 1,640 | 86 | 4,523 | 1,760 | 789 | 2,669 | 14,452 |
| USSR/Russia (1992-) | - | 44 | - | - | - | - | - | - | - |
| Unallocated + disc. | 6,493 | 143 | -1,278 | -1,940 | $-6,960{ }^{3}$ | -51 | -41,326 | -11,523 | 837 |


| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total | 34,870 | 20,904 | 34,456 | 40,469 | 53,942 | 69,527 | 83,595 | 81,259 | 40,983 |
| Country | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| Denmark |  | 79 |  |  |  |  |  |  |  |
| Faroe Islands | - | - |  |  |  |  |  |  |  |
| France | 221 |  |  | 428 | 55 | 209 | 172 | 41 | 411 |
| Germany | 414 | 1031 | 209 | 265 | 149 | 1337 | 1413 | 1958 | 1025 |
| Ireland | 21951 | 31736 | 15843 | 20162 | 12341 | 20903 | 15702 | 12395 | 9780 |
| Lithuania |  |  |  |  |  |  |  |  | 2822 |
| Netherlands | 983 | 2646 | 686 | 600 | 450 | 847 | 3702 | 6039 | 1892 |
| Spain | - | - |  |  |  |  |  | 0 | 0 |
| UK (Engl.+Wales) | 227 | 344 | 41 | 91 |  | 46 | 5 | 52 |  |
| UK (N.Ireland) | 1132 | - | 79 | 272 | 654 | 530 | 249 | 210 | 82 |
| UK (Scotland) | 10147 | 4544 | 1839 | 3111 | 1192 | 453 | 377 | 62 | 43 |
| Unallocated+disc. | 98 | 1507 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total | 34815 | 41887 | 18697 | 24929 | 14840 | 24325 | 21619 | 20757 | 16055 |

${ }^{1}$ Included in Subarea 7. ${ }^{2}$ Includes Divisions 3.a, 4.a, band 6.b. ${ }^{3}$ Includes a negative unallocated catch of $\mathbf{- 7 0 0 0} \mathbf{t}$.

| Country | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  |  |  |  | 58 | 1,131 | 433 | 856 | 3,045 |
| Faroe Islands |  | 573 |  | 66 |  |  |  |  |  |
| France |  | 73 |  |  | 246 |  |  | 195 | 65 |
| Germany | 1,835 | 5,097 | 635 | 773 | 6,508 | 671 | 8,616 | 4,194 | 1,980 |
| Ireland | 20,010 | 18,751 | 16,596 | 19,985 | 23,556 | 29,282 | 19,979 | 15,745 | 10,894 |
| Lithuania | 80 | 641 |  |  |  |  |  |  |  |
| Netherlands | 2,177 | 3,904 | 2,332 | 1,684 | 6,353 | 12,653 | 11,078 | 8,580 | 6,211 |
| Norway | 2 | 20 | 27 | 18 | 48 | 2 |  |  |  |
| Spain | 0 |  |  |  |  |  |  |  |  |
| UK (Engl. + Wales) | 332 |  |  | 463 |  |  | 451 | 18 | 58 |
| UK (N.Ireland) |  |  |  | 59 | 198 |  | 2,325 | 1,579 | 1,204 |
| UK (Scotland) | 38 | 588 | 243 | 89 | 2,528 | 1,231 | 385 | 1,277 | 696 |
| Unallocated+disc. | 0 | 0 | 0 | 0 | 230 | 2 | - | 123 |  |
| Total | 24,474 | 29,648 | 19,833 | 23,136 | 39,726 | 44,973 | 43,266 | 32,567 | 24,153 |


| Country | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark |  | 3,462 | 4,982 | 6,467 | 2,267 | 1,853 | 1,043 |
| Faroe Islands |  | 113 |  | 20 |  | - |  |
| France | 23 | 1,025 | 197 | 550 | 3 | 908 | 74 |
| Germany | 4,069 | 2,884 | 2,779 | 1,418 | 0 | - |  |
| Ireland | 15,381 | 15,123 | 17,959 | 21,109 | 9,187 | 8,530 | 9462 |
| Lithuania | 2,510 |  |  |  |  | - |  |
| Netherlands | 9,246 | 5,497 | 11,921 | 14,421 | 5,202 | 1,309 | 500 |
| Norway |  |  |  |  |  |  |  |
| Spain |  |  |  |  |  |  |  |
| UK (Engl. + Wales) |  | 66 | 32 | 830 | 817 | 249 | 8 |
| UK (N.Ireland) | 0 |  | 1,026 | 1,907 | 1,229 | 417 | 471 |
| UK (Scotland) | 956 |  |  | 627 | $331^{* *}$ | 459 | 456 |


| Country | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Unallocated+disc. | 116 | 55 | 129 | 108 | 91 | 313 |  |
| Total | 32,186 | 28,286 | 38,950 | 47,480 | 19,146 | 13,818 | 12,329 |

** 1.4t BMS included

Table 7.1.1.4. Western horse mackerel. Catches ( $t$ ) in Subarea 7 by country (Data submitted by the Working Group members).

| Country | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | - | 1 | 1 | - | - | + | + | 2 | - |
| Denmark | 5,045 | 3,099 | 877 | 993 | 732 | 1477 | 30408 | 27,368 | 33,202 |
| France | 1,983 | 2,800 | 2,314 | 1,834 | 2,387 | 1,881 | 3,801 | 2,197 | 1,523 |
| Germany, Fed.Rep. | 2,289 | 1,079 | 12 | 1,977 | 228 | - | 5 | 374 | 4,705 |
| Ireland | - | 16 | - | - | 65 | 100 | 703 | 15 | 481 |
| Netherlands | 23,002 | 25,000 | 27500 | 34,350 | 38,700 | 33,550 | 40,750 | 69,400 | 43,560 |
| Norway | 394 | - | - | - | - | - | - | - | - |
| Spain | 50 | 234 | 104 | 142 | 560 | 275 | 137 | 148 | 150 |
| UK (Engl. + Wales) | 12,933 | 2,520 | 2,670 | 1,230 | 279 | 1,630 | 1,824 | 1,228 | 3,759 |
| UK (Scotland) | 1 | - | - | - | 1 | 1 | + | 2 | 2,873 |
| USSR | - | - | - | - | - | 120 | - | - | - |
| Total | 45,697 | 34,749 | 33,478 | 40,526 | 42,952 | 39,034 | 77,628 | 100,734 | 90,253 |
| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| Faroe Islands | - | 28 | - | - | - | - | - | - | - |
| Belgium | - | + | - | - | - | 1 | - | - | 18 |
| Denmark | 34,474 | 30,594 | 28,888 | 18,984 | 16,978 | 41,605 | 28,300 | 43,330 | 60,412 |
| France | 4,576 | 2,538 | 1,230 | 1,198 | 1,001 | - | - | - | 30,571 |
| Germany, Fed.Rep. | 7,743 | 8,109 | 12,919 | 12,951 | 15,684 | 14,828 | 17,436 | 15,949 | 28,267 |
| Ireland | 12,645 | 17,887 | 19,074 | 15,568 | 16,363 | 15,281 | 58,011 | 38,455 | 43,624 |
| Netherlands | 43,582 | 111,900 | 104,107 | 109,197 | 157,110 | 92,903 | 116,126 | 114,692 | 131,701 |
| Norway | - | - | - | - | - | - | - | - | - |
| Spain | 14 | 16 | 113 | 106 | 54 | 29 | 25 | 33 | 6 |
| UK (Engl. + Wales) | 4,488 | 13,371 | 6,436 | 7,870 | 6,090 | 12,418 | 31,641 | 28,605 | 17,464 |


| Country | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| UK (N.Ireland) | - | - | 2,026 | 1,690 | 587 | 119 | - | - | 1,093 |
| UK (Scotland) | + | 139 | 1,992 | 5,008 | 3,123 | 9,015 | 10,522 | 11,241 | 7,902 |
| Unallocated +discards | 28,368 | 7,614 | 24,541 | 15,563 | 4,010 | 14,057 | 68,644 | 26,795 | 58,718 |
| Total | 135,890 | 192,196 | 201,326 | 188,135 | 221,000 | 200,256 | 330,705 | 279,100 | 379,776 |
| Country |  | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| Faroe Islands | - | - |  |  |  |  |  |  |  |


| Country | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Faroe Islands | 475 | 212 |  | - | - | 0 |  |  |  |
| Belgium |  |  | 19 | 2 |  | 14 |  |  |  |
| Denmark | 4856 | 1970 | 2710 | 5247 | 5831 | 2281 | 6373 | 5066 | 1474 |
| France | 2007 | 9703 |  | 260 | 7431 | 579 | 744 | 940 | 1552 |
| Germany | 3943 | 5693 | 14205 | 16847 | 14545 | 16391 | 15781 | 12948 | 7382 |
| Ireland | 8039 | 16282 | 23816 | 24491 | 14154 | 15893 | 15805 | 16922 | 10751 |
| Lithuania | 5387 | 4907 |  |  |  | - | 0 |  |  |


| Country | 2007 | 2008 | 2009 | 2010 | 2011 |  | 2012 | 2013 | 2014 | 2015 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Netherlands | 32654 | 28077 | 23263 | 65865 | 4920 |  | 53644 | 41562 | 15529 | 18100 |
| Norway | - | - | - | 40 |  |  | - | 0 |  |  |
| Spain | 11 | 11 | 6 | 3 |  |  | 10 | 0 |  |  |
| UK (Engl. + Wales) | 5119 | 3245 | 6257 | 12139 | 1168 |  | 12122 | 3388 | 4576 | 1798 |
| UK (Scotland) |  | 469 | 1119 | 1713 | 299 |  | 91 | 17 | 101 | 6 |
| Unallocated+discards | 6012 | -4624 | -10891 | 6511 | 1 |  | 3038 | 4399 | 974 | 1929 |
| Total | 68504 | 65946 | 60487 | 133136 | 61031 |  | 104049 | 88083 | 57055 | 42992 |
| Country | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 20 |  |  |  |
| Denmark | 314 | 1057 | 1,031 | 690 | 3,198 | 3,540 | $2,2$ |  |  |  |
| France | 551 | 595 | 1,067 | 907 | 1,486 | 990 | 1,8 |  |  |  |
| Germany | 7313 | 4077 | 1,401 | 7,673 | 952 | 5,525 | 3,9 |  |  |  |
| Ireland | 12193 | 7857 | 7,169 | 7,753 | 7,870 | 10,24 | 0 6,3 |  |  |  |
| Lithuania | 86 |  |  |  |  |  |  |  |  |  |
| Netherlands | 14415 | 8445 | 14,009 | 15,159 | 9,036 | 17,47 | 316 |  |  |  |
| Poland |  |  |  | 127 | 1,000 | 1,605 |  |  |  |  |
| Spain | 0 |  | 0 | 1 | 6 | 14 | 16 |  |  |  |
| UK (Engl. + Wales) | 820 | 478 | 2,410 | 2,862 | 679** | 2,401 | *** 2,20 |  |  |  |
| UK (Scotland) |  |  |  |  | 3 | 92 |  |  |  |  |
| UK (Northern Ireland) |  |  | 52 | 0 | 2 | 933 | 34 |  |  |  |
| Unallocated+discards | 1692 | 830 | 548 | 918 | 311 | 677 | 47 |  |  |  |
| Total | 37384 | 23340 | 27,687 | 36,062 | 24,544 | 43,49 | 0 34 |  |  |  |

${ }^{2}$ French catches landed in the Netherlands **21t BMS landings included

Table 7.1.1.5. Western horse mackerel. Catches $(t)$ in Subarea 8 by country (Data submitted by Working Group members).

| Country | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Denmark | - | - | - | - | - | - | 446 | 3,283 | 2,793 |  |
| France | 3,361 | 3,711 | 3,073 | 2,643 | 2,489 | 4,305 | 3,534 | 3,983 | 4,502 |  |
| Netherlands | - | - | - | - | -2 | -2 | -2 | -2 | - |  |
| Spain | - |  |  |  |  |  |  |  |  |  |


| Country | 1998 |  | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total | 40,455 |  | 37,692 | 54,222 | 75,120 | 57,246 | 41,711 | 24,125 | 41,260 | 34,122 |
| Country | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 |
| Denmark | 2,687 | 3,289 | 3,109 | 632 | 200 | 581 | 14 |  |  |  |
| France | 10,741 | 2,848 |  |  | 326 | 1,218 | 2,849 | 2,277 | 1,618 | 2,219 |
| Germany |  | 918 | 281 | 64 | 61 |  | 417 | 19 | 49 | 4 |
| Ireland | 694 |  |  |  |  | 39 |  |  | 0 | 32 |
| Netherlands | 211 | 6,269 | 1,848 | 98 | 49 | 7 | 1,057 | 526 | 635 | 1 |
| Spain | 14,265 | 19,840 | 0 21,07 | 138,742 | 34,581 | 133,502 | 222,542 | 19,443 | 13,072 | 14,235 |
| UK (Engl. + Wales) |  | 120 | 224 | 112 | 28 |  | 104 | 35 | 72 | 9 |
| Unallocated+discards |  | 67 | 913 | 7,412 | 417 | 431 | 2,055 | 182 | 9,314 | 6,643 |
| Total | 28,598 | 33,352 | 2 27,44 | 77,060 | 35,662 | 2 15,777 | 7 29,039 | 22,483 | 24,760 | 23,143 |
| Country |  | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |  |  |  |
| Denmark |  | 1 |  | 422 |  | 638 | 1,582 |  |  |  |
| France |  | 2,303 | 2,176 | 2,914 | 728 | 808 | 728 |  |  |  |
| Germany |  | 210 | 554 | 144 | 2 | 2 | 321 |  |  |  |
| Ireland |  | 580 | 219 | 36 | 332 | - |  |  |  |  |
| Netherlands |  | 313 | 6 | 3 | 0.5 | 1,976 | 1,085 |  |  |  |
| Spain |  | 14,901 | 20,362 | 25,775 | 19,163 | 16,177 | 13,380 |  |  |  |
| UK (Engl. + Wales) |  |  | 2 | 344 |  |  | 1 |  |  |  |
| Unallocated+discards |  | 2,907 | 1,921 | 1,755 | 1,104 | 713 | 1,853 |  |  |  |
| Total |  | 21,213 | 25,240 | 31,396 | 20,742 | 20,314 | 18,951 |  |  |  |

${ }^{2}$ Included in Subarea 7. ${ }^{3}$ French catches landed in the Netherlands

Table 7.2.1.1. Western horse mackerel. The time series of Total Annual Egg Production (TAEP) estimates (10 ${ }^{12}$ eggs).

| Year | TAEP (*e12) |
| :--- | :--- |
| 1992 | 2094 |
| 1995 | 1344 |
| 1998 | 1242 |
| 2001 | 864 |


| 2004 | 884 |
| :--- | :--- |
| 2007 | 1486 |
| 2010 | 1033 |
| 2013 | 366 |
| 2016 | 331 |
| 2022 | 178 |

Table 7.2.1.2. 1 Daily egg production estimate for the 2022 using DEPM.

| Period | Po |  |  |
| :--- | :--- | :--- | :--- |
| (eggs/m2/day) | Spawning area <br> $(\mathbf{m 2})(\mathbf{x 1 0 1 1 )}$ | Ptot <br> (eggs/day)(x1013) |  |
| 6 | 133.857 | 1.390603962 | 0.186 |

Table 7.2.1.3. Preliminary mean parameter estimates for 2022 adult horse mackerel period 6 (spawning peak)

| Period | Parameter | mean | var | cv |
| :--- | :--- | :--- | :--- | :--- |
| 6 | Female Mean weight (g) | 275.4 | 124.9 | 0.673 |
| 6 | Mean Sex Ratio | 0.508 | 0 | 0.006 |
| 6 | Mean Spawning Fraction <br> 6 | Mean batch fecundity <br> (N.eggs) <br> Relative mean batch fecundity (eggs/g fe- <br> male) <br> Daily fecundity <br> (eggs/g female day) | 219.5 | 0.187 |
| 6 |  | 20.9 | 2.56 | 0.049 |

Table 7.2.1.4. Preliminary SSB estimate for the 2022 Western horse mackerel DEPM

| Period | Daily fecundity <br> (eggs/g female day) | Ptot <br> (eggs/day)( $\mathbf{x 1 0 1 3 )}$ | SSB (kt) |
| :--- | :--- | :--- | :--- |
| 6 | 20.865 | 0.186 | 891.445 |

Table 7.2.2.1. Western horse mackerel. Time series of recruitment index estimated from the IBTS Surveys (2003-2022) in 2021-2023

| Year | Index 2023 |  | Index 2022 | Index 2021 |
| :--- | :--- | :--- | :--- | :--- |
|  | Mean | CV |  |  |
| 2003 | 805259 | 0.30 | 728100 | 732297 |
| 2004 | 2705361 | 0.29 | 2516442 | 2453310 |
| 2005 | 2337356 | 0.34 | 2199332 | 2151351 |
| 2006 | 1616567 | 0.34 | 1501474 | 1499811 |
| 2007 | 3447979 | 0.31 | 3125619 | 3121579 |
| 2008 | 8041550 | 0.32 | 7824230 | 7481365 |
| 2009 | 1240984 | 0.26 | 1127972 | 1148964 |
| 2010 | 976048 | 0.34 | 872244 | 864772 |
| 2011 | 189761 | 0.35 | 175162 | 178188 |
| 2012 | 4674962 | 0.31 | 4435133 | 4339882 |
| 2013 | 1182836 | 0.25 | 1099932 | 1111210 |
| 2014 | 3108626 | 0.29 | 2905589 | 2931963 |
| 2015 | 4269281 | 0.25 | 0.26 | 4123241 |

Table 7.2.2.2. Western horse mackerel. Time series of biomass from the PELACUS acoustic survey (in tonnes).

| Year | Biomass | CV |
| :--- | :--- | :--- |
| 1992 | 57188 | 0.32 |


| Year | Biomass | cV |
| :---: | :---: | :---: |
| 1993 | 25028 | 0.32 |
| 1995 | 93825 | 0.32 |
| 1997 | 74364 | 0.32 |
| 1998 | 139395 | 0.32 |
| 1999 | 71744 | 0.32 |
| 2000 | 26192 | 0.32 |
| 2001 | 40864 | 0.32 |
| 2002 | 41788 | 0.32 |
| 2003 | 26647 | 0.32 |
| 2004 | 23992 | 0.32 |
| 2005 | 40082 | 0.32 |
| 2006 | 13934 | 0.32 |
| 2007 | 28173 | 0.32 |
| 2008 | 33614 | 0.32 |
| 2009 | 24020 | 0.32 |
| 2010 | 53417 | 0.32 |
| 2011 | 7687 | 0.32 |
| 2012 | 15479 | 0.32 |
| 2013 | 5532 | 0.32 |
| 2014 | 30454 | 0.32 |
| 2015 | 67068 | 0.32 |
| 2016 | 32581 | 0.32 |
| 2017 | 13845 | 0.32 |
| 2018 | 9270 | 0.32 |
| 2019 | 13075 | 0.32 |
| 2020 | NA | NA |
| 2021 | 10233 | 0.32 |
| 2022 | 18584 | 0.32 |
| 2023 | 10336 | 0.32 |

Table 7.2.4.1. Western Horse Mackerel stock. Catch in numbers (thousands) at age by quarter and area in 2022 ( $15=15+$ group)

| Q1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | 27.2.a | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.k | 27.8.a | 27.8.b | 27.8.c | 27.8.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 7487.76 | 16020.66 | 0.57 | 1140.21 | 84.20 | 1.88 | 0.50 | 24735.77 |
| 2 | 0 | 0 | 0.00 | 2.92 | 0.03 | 0.03 | 54.37 | 16.00 | 6.24 | 66.09 | 13 | 0.21 | 0.00 | 632.56 | 583.58 | 0.09 | 302.15 | 575.32 | 0.16 | 0.05 | 2253.12 |
| 3 | 0.00 | 0.00 | 0.00 | 7.83 | 0.09 | 0.09 | 66.23 | 154.08 | 16.71 | 121.87 | 35.68 | 29.91 | 0.00 | 810.24 | 51.66 | 0.76 | 98.18 | 1711.23 | 0.20 | 0.04 | 3104.81 |
| 4 | 2.52 | 439.71 | 0.06 | 238.77 | 1.77 | 1.82 | 226.71 | 2019.51 | 331.46 | 272.13 | 707.77 | 5860.61 | 0.00 | 518 | 13.11 | 0.80 | 53.20 | 726.58 | 0.10 | 0.01 | 11414.41 |
| 5 | 26.43 | 4611.19 | 0.20 | 2363.11 | 6.17 | 6.35 | 529.79 | 3388.76 | 1158.80 | 398.63 | 2474.39 | 24294.22 | 0.01 | 1005 | 15.62 | 0.23 | 92.73 | 617.38 | 0.11 | 0.02 | 40989.60 |
| 6 | 10.77 | 1878.40 | 0.06 | 807.09 | 1.96 | 2.02 | 168.24 | 1161.41 | 367.63 | 92.75 | 784.99 | 7572.33 | 0.00 | 1569 | 36.28 | 0.10 | 215.48 | 977.44 | 0.19 | 0.03 | 15646.36 |
| 7 | 16.12 | 2812.72 | 0.05 | 660.40 | 1.42 | 1.46 | 119.43 | 683.27 | 266.35 | 67.20 | 568.73 | 5594.57 | 0.00 | 1363 | 36.64 | 0.04 | 195.56 | 522.94 | 0.14 | 0.02 | 12909.86 |
| 8 | 92.80 | 16190.64 | 0.20 | 4487.59 | 6.20 | 6.38 | 528.90 | 3102.41 | 1164.39 | 293.76 | 2486 | 22384.05 | 0.01 | 1275 | 49.53 | 0.03 | 170.52 | 333.30 | 0.11 | 0.02 | 52571.95 |
| 9 | 5.14 | 896.48 | 0.02 | 329.91 | 0.65 | 0.67 | 54.53 | 311.95 | 121.60 | 30.68 | 260 | 2519.92 | 0.00 | 289 | 36.46 | 0.01 | 121.15 | 111.78 | 0.04 | 0.01 | 5089.85 |
| 10 | 13.63 | 2378.02 | 0.01 | 243.33 | 0.19 | 0.20 | 16.23 | 92.85 | 36.20 | 9.13 | 77 | 575.70 | 0.00 | 379 | 71.12 | 0.02 | 242.35 | 162.59 | 0.06 | 0.01 | 4297.57 |
| 11 | 3.48 | 607.47 | 0.01 | 123.94 | 0.17 | 0.18 | 14.62 | 83.66 | 32.61 | 8.23 | 70 | 630.29 | 0.00 | 141 | 41.04 | 0.01 | 203.89 | 80.78 | 0.04 | 0.00 | 2041.29 |
| 12 | 1.55 | 269.67 | 0.00 | 59.63 | 0.12 | 0.12 | 9.98 | 57.08 | 22.25 | 5.61 | 48 | 461.16 | 0.00 | 122 | 31.40 | 0.01 | 194.94 | 55.53 | 0.03 | 0.00 | 1338.86 |
| 13 | 5.44 | 948.87 | 0.01 | 107.40 | 0.21 | 0.22 | 17.95 | 102.72 | 40.04 | 10.10 | 85 | 829.80 | 0.00 | 43 | 9.15 | 0.00 | 74.63 | 14.62 | 0.01 | 0.00 | 2289.39 |
| 14 | 12.46 | 2173.35 | 0.02 | 290.77 | 0.54 | 0.56 | 45.41 | 259.80 | 101.27 | 25.55 | 216 | 2080.05 | 0.00 | 110 | 19.27 | 0.01 | 207.78 | 25.14 | 0.03 | 0.00 | 5567.77 |
| 15+ | 11.20 | 1954.82 | 0.00 | 206.05 | 0.07 | 0.07 | 5.83 | 33.35 | 13.00 | 3.28 | 28 | 64.03 | 0.00 | 131 | 16.50 | 0.01 | 243.94 | 41.47 | 0.03 | 0.00 | 2752.62 |


| Q2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.5.b | 27.6.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 2214.06 | 1989.25 | 0.27 | 3387.27 | 1449.47 | 0.63 | 3.18 | 9044.13 |
| 2 | 0 | 0.00 | 0.00 | 0.05 | 0.00 | 0.09 | 0.24 | 0.01 | 0.06 | 0.12 | 0.97 | 0.91 | 0.01 | 3387.29 | 194.28 | 4.25 | 2326.96 | 7936.53 | 0.96 | 2.92 | 13855.64 |
| 3 | 0.00 | 0 | 0.00 | 0.14 | 0.00 | 0.27 | 0.72 | 0.03 | 0.17 | 0.35 | 2.90 | 2.71 | 0.02 | 4033.99 | 90.45 | 13.36 | 2624.04 | 9757.00 | 1.15 | 0.28 | 16527.58 |
| 4 | 0.00 | 0 | 0.00 | 2.70 | 0.02 | 5.24 | 14.03 | 0.64 | 3.32 | 6.75 | 56.54 | 52.88 | 0.44 | 1087.61 | 31.73 | 5.94 | 2080.83 | 1244.54 | 0.31 | 0.39 | 4593.90 |
| 5 | 0.00 | 0 | 0.00 | 8.90 | 0.08 | 17.27 | 46.28 | 2.11 | 10.95 | 22.28 | 186.55 | 174.48 | 1.44 | 691.34 | 30.37 | 4.53 | 1770.98 | 328.71 | 0.20 | 0.53 | 3296.98 |
| 6 | 0.00 | 0 | 0.00 | 2.99 | 0.03 | 5.80 | 15.55 | 0.71 | 3.68 | 7.49 | 62.67 | 58.61 | 0.48 | 666.70 | 40.66 | 7.63 | 1798.68 | 205.31 | 0.19 | 1.16 | 2878.33 |
| 7 | 0.85 | 0 | 1.27 | 2.45 | 02 | 4.75 | 12.73 | 0.58 | 3.01 | 6.13 | 51.30 | 47.98 | 0.40 | 282.61 | 38.03 | 8.22 | 715.64 | 102.23 | 0.08 | 0.40 | 1278.81 |
| 8 | 1.82 | 0 | 2.75 | 10.27 | 0.09 | 19.93 | 53.39 | 2.43 | 12.63 | 25.71 | 215.22 | 201.29 | 1.66 | 187.80 | 83.73 | 7.21 | 394.58 | 88.83 | 0.05 | 0.39 | 1310.07 |
| 9 | 4.80 | 1 | 7.24 | 1.04 | 0.01 | 2.02 | 5.41 | 0.25 | 1.28 | 2.61 | 21.81 | 20.40 | 0.17 | 94.71 | 99.42 | 2.01 | 121.89 | 65.93 | 0.03 | 0.48 | 452.32 |
| 10 | 9.87 | 2 | 14.89 | 0.37 | 0.00 | 0.72 | 1.94 | 0.09 | 0.46 | 0.93 | 7.81 | 7.30 | 0.06 | 221.22 | 301.58 | 1.27 | 233.29 | 132.49 | 0.06 | 6.79 | 942.80 |
| 11 | 4.53 | 1 | 6.84 | 0.25 | 0.00 | 0.49 | 1.33 | 0.06 | 0.31 | 0.64 | 5.34 | 5.00 | 0.04 | 199.10 | 254.10 | 0.20 | 263.10 | 92.98 | 0.06 | 5.89 | 841.04 |
| 12 | 10.18 | 2 | 15.36 | 0.17 | 0.00 | 0.33 | 0.88 | 0.04 | 0.21 | 0.42 | 3.53 | 3.31 | 0.03 | 213.51 | 228.16 | 0.01 | 339.21 | 87.26 | 0.06 | 7.18 | 911.56 |
| 13 | 16.09 | 3 | 24.28 | 0.30 | 0.00 | 0.58 | 1.55 | 0.07 | 0.37 | 0.75 | 6.25 | 5.85 | 0.05 | 79.72 | 76.99 | 0.00 | 132.50 | 34.98 | 0.02 | 2.73 | 385.78 |
| 14 | 12.59 |  | 19 |  | 0 | 1 | 4 | - | 1 |  | 16 | 15.00 | 0.12 | 201.30 | 206.60 | 0.02 | 318.90 | 94.18 | 0.06 | 6.76 | 901.97 |
| 15+ | 48.55 | 8 | 73 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |  | 2.12 | 0.02 | 224.29 | 184.46 | 0.17 | 436.01 | 72.29 | 0.06 | 4.48 | 1057.39 |

Table 7.2.4.1 cont. Western Horse Mackerel stock. Catch in numbers (thousands) at age by quarter and area in 2022 (15 = 15+ group)

| Q3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.3.a | 27.3.a. 20 | 27.4.a | 27.6.a | 27.7.b | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | Total |
| 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 55.12 | 39.26 | 47 | 6.80 | 0 | 3 | 156.37 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 146 | 46.80 | 4.68 | 3785.04 | 683.67 | 0.01 | 7.27 | 4673.23 |
| 2 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 44 | 11.36 | 4.37 | 407.15 | 930.65 | 0.00 | 2.05 | 1399.23 |
| 3 | 0 | 0 | 0 | 1 | 0.00 | 0.06 | 0.06 | 2.27 | 0.07 | 0.14 | 0.97 | 18.83 | 19.01 | 34.11 | 8.88 | 2.00 | 214 | 832.55 | 0.00 | 1.60 | 1135.99 |
| 4 | 0 | 0.68 | 0.00 | 38.19 | 0.13 | 0.95 | 1.00 | 37.65 | 1.11 | 2.37 | 12.63 | 53.99 | 618.84 | 25.43 | 6.62 | 1.47 | 194 | 586.06 | 0.00 | 1.19 | 1582.99 |
| 5 | 0.49 | 0.86 | 0.00 | 48.06 | 0.16 | 1.16 | 1.23 | 46.18 | 1.36 | 2.91 | 29.76 | 27.75 | 787.38 | 27.64 | 7.40 | 0.83 | 224.29 | 624.72 | 0.00 | 1.31 | 1833.51 |
| 6 | 2.24 | 3.94 | 0.01 | 219.92 | 0.73 | 1.04 | 1.10 | 41.41 | 1.22 | 2.61 | 6.73 | 22.61 | 731.85 | 32.69 | 9.34 | 0.02 | 276.43 | 728.01 | 0.00 | 1.57 | 2083.49 |
| 7 | 11.55 | 20.35 | 0.04 | 1134.99 | 3.75 | 1.90 | 2.01 | 75.32 | 2.22 | 4.74 | 121.46 | 30.40 | 1216.51 | 15.71 | 4.67 | 0.01 | 87.90 | 394.71 | 0.00 | 0.74 | 3128.97 |
| 8 | 6.31 | 11.11 | 0.02 | 619.48 | 2.05 | 6.54 | 6.92 | 259.71 | 7.66 | 16.35 | 568.94 | 36.00 | 4099.76 | 41.54 | 19.00 | 0.05 | 157.93 | 1111.80 | 0.00 | 1.95 | 6973.11 |
| 9 | 3.90 | 6.86 | 0.01 | 382.76 | 1.26 | 0.55 | 0.59 | 22.00 | 0.65 | 1.39 | 60.52 | 0.89 | 335.48 | 35.60 | 33.33 | 0.10 | 149.23 | 923.74 | 0.00 | 1.70 | 1960.57 |
| 10 | 3.45 | 6.08 | 0.01 | 338.91 | 1.12 | 0.42 | 0.44 | 16.64 | 0.49 | 1.05 | 8.76 | 0.67 | 296.85 | 52.74 | 65.31 | 0.06 | 361.84 | 1213.18 | 0.00 | 2.53 | 2370.55 |
| 11 | 1.15 | 2.02 | 0.00 | 112.66 | 0.37 | 0.05 | 0.05 | 2.02 | 0.06 | 0.13 | 16.10 | 0.08 | 18.46 | 37 | 37.85 | 0.07 | 337.73 | 780.79 | 0.00 | 1.78 | 1348.57 |
| 12 | 2.14 | 3.77 | 0.01 | 210.30 | 0.69 | 0.01 | 0.01 | 0.52 | 0.02 | 0.03 | 4.01 | 0.02 | 4.92 | 23 | 22.04 | 0.06 | 236.78 | 450.26 | 0.00 | 1.08 | 959.50 |
| 13 | 3.65 | 6.43 | 0.01 | 358.94 | 1.19 | 0.00 | 0.00 | 0.12 | 0.00 | 0.01 | 0.01 | 0.00 | 2.15 | 26 | 22.88 | 0.03 | 255.85 | 519.66 | 0.00 | 1.21 | 1197.85 |
| 14 | 3.16 | 5.57 | 0.01 | 310.43 | 1.02 | 0.05 | 0.05 | 2.01 | 0.06 | 0.13 | 0.67 | 0.08 | 36.31 | 34 | 21.17 | 0.02 | 354.17 | 687.54 | 0.00 | 1.61 | 1458.29 |
| 15+ | 10.56 | 18.60 | 0.04 | 1037.65 | 3.43 | 0.05 | 0.05 | 1.82 | 0.05 | 0.11 | 0.40 | 0.07 | 33.03 | 134 | 24.40 | 0.04 | 517 | 1368.10 | 0.00 | 2.98 | 3151.88 |


| Q4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.3.a | 27.4.a | 27.6.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 377 | 110 | 111 | 0 | 13 | 0 | 4 | 615 |
| 1 | 0 | 0 | 0 | 0 | 2 | 0 | 1 | 98 | 1952 | 405 | 258 | 0 | 9 | 1544 | 125 | 27 | 54 | 3278 | 0 | 5 | 7759 |
| 2 | 0 | 0 | 0 | 0 | 12 | 1 | 1 | 134 | 1592 | 1630 | 354 | 1 | 12 | 1034 | 83 | 16 | 84 | 2217 | 0 | 1 | 7172 |
| 3 | 0.23 | 2.28 | 1.51 | 48.34 | 42.74 | 0.83 | 1.02 | 210.57 | 1764.30 | 2873.16 | 559.77 | 0.77 | 66.76 | 435.40 | 18.84 | 2.62 | 55.25 | 933.76 | 0.00 | 0.25 | 7018.40 |
| 4 | 6.13 | 60.88 | 40.39 | 1289.48 | 658.77 | 1.59 | 1.95 | 417.58 | 3526.79 | 4120.76 | 1039.39 | 1.47 | 809.90 | 270.50 | 7.63 | 0.08 | 25.68 | 368.60 | 0.00 | 0.26 | 12647.84 |
| 5 | 33.13 | 329.04 | 1013.72 | 6174.35 | 775.86 | 1.57 | 1.93 | 611.64 | 1264.10 | 1081.02 | 1581.56 | 1.46 | 5254.76 | 356.88 | 11.93 | 0.09 | 40.55 | 402.01 | 0.00 | 0.34 | 18935.95 |
| 6 | 5.46 | 54.27 | 238.18 | 947.30 | 237.80 | 0.44 | 0.54 | 298.18 | 69.29 | 100.25 | 685.53 | 0.41 | 1571.16 | 548.61 | 17.05 | 0.13 | 62.25 | 543.18 | 0.00 | 0.22 | 5380.24 |
| 7 | 7.00 | 69.56 | 1035.19 | 484.28 | 76.53 | 0.56 | 0.68 | 203.67 | 0.00 | 66.12 | 1407.00 | 0.52 | 1996.06 | 840.80 | 9.26 | 0.12 | 29.68 | 152.63 | 0.00 | 0.03 | 6379.68 |
| 8 | 23.09 | 229.34 | 2657.39 | 2352.60 | 281.17 | 1.74 | 2.13 | 479.56 | 0.00 | 266.19 | 3491.52 | 1.61 | 7192.55 | 914.63 | 37.52 | 0.14 | 84.47 | 219.96 | 0.00 | 0.35 | 18235.96 |
| 9 | 0.41 | 4.03 | 2.67 | 85.32 | 4.14 | 0.20 | 0.24 | 65.38 | 0.00 | 31.46 | 502.94 | 0.18 | 720.00 | 412.50 | 88.22 | 0.07 | 109.65 | 51.42 | 0.00 | 1.70 | 2080.53 |
| 10 | 2.68 | 26.60 | 437.92 | 143.09 | 9.11 | 0.14 | 0.17 | 32.38 | 0.00 | 29.66 | 110.05 | 0.13 | 738.81 | 243.66 | 206.97 | 0.08 | 190.31 | 49.41 | 0.00 | 4.52 | 2225.67 |
| 11 | 1.05 | 10.39 | 195.89 | 31.16 | 0.18 | 0.05 | 0.06 | 10.99 | 0.00 | 27.95 | 42.00 | 0.04 | 230.66 | 153.53 | 133.42 | 0.05 | 147.54 | 32.98 | 0.00 | 3.20 | 1021.12 |
| 12 | 0.05 | 0.50 | 0.33 | 10.61 | 0.03 | 0.01 | 0.01 | 2.15 | 0.00 | 14.93 | 40.22 | 0.01 | 2.38 | 88.34 | 74.05 | 0.03 | 94.47 | 11.33 | 0.00 | 1.97 | 341.44 |
| 13 | 0.25 | 2.46 | 1.63 | 52.05 | 0.03 | 0.01 | 0.01 | 1.96 | 33.70 | 2.52 | 8.33 | 0.01 | 8.09 | 82.49 | 68.26 | 0.03 | 92.19 | 9.49 | 0.00 | 1.79 | 365.28 |
| 14 | 0.14 | 1.39 | 0.92 | 29.35 | 0.10 | 0.03 | 0.03 | 6.02 | 0.00 | 65.00 | 20.33 | 0.02 | 78.00 | 78.44 | 51.67 | 0.03 | 107.33 | 10.59 | 0.00 | 1.53 | 450.92 |
| 15+ | 0.29 | 2.87 | 44.12 | 18.47 | 3.42 | 0.01 | 0.01 | 1.67 | 0.00 | 34.04 | 4.40 | 0.01 | 2.99 | 86.32 | 43.02 | 0.03 | 129.69 | 19.83 | 0.00 | 1.40 | 392.58 |

## Table 7.2.4.1 cont. Western Horse Mackerel stock. Catch in numbers (thousands) at age by quarter and area in 2022 (15 = 15+ group)

| $\begin{gathered} \text { all }, ~ \\ \text { Anges } \end{gathered}$ | 27.2.a | 27.3.a | 27.3.20 | 27.4.a | 27.5.b | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.e. 2 | 27.7.e | 27.7.f | 27.7.8 | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k | 27.7.1. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.e.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| , | O | , | , | 1 | 0 | , | - | , | 0 | 0 | , | , | 0 | , | , | , | , | , | 381 | 165 | 150 | 47 | 20 | . | , | ${ }_{7}^{\text {Total }}$ |
| 1 | 0 | 0 | 0 |  | 0 | 0 | 0 | 2 | 0 | 1 | 98 | 1952 | 405 | 258 | 0 | 9 | 0 | 0 | 11391 | 18181 | 32 | 8367 | 5495 | 3 | 16 | 46212 |
|  | 0 | 0 | 0 | 0 | 0 | , | 0 | 15 | 1 | 1 | 189 | 1608 | 1637 | 420 | 15 | 13 | 0 | 0 | 5097 | 872 | 25 | 3121 | 11660 | 1 |  | 24680 |
| 3 | 0.24 | 2.30 | 0.00 | 5.00 | 0.00 | 50.82 | 0.00 | 50.76 | 0.92 | 1.44 | 279.79 | 1918.48 | 2890.18 | 68.96 | 58.18 | 118.40 | 0.00 | 0.02 | 5313.73 | 169.82 | 18.74 | 2991.71 | 13234.53 | 1.36 | 2.17 | 27991.57 |
| 4 | 9.04 | ${ }^{61.56}$ | 0.00 | ${ }_{102.47}$ | 0.00 | ${ }_{1819.13}$ | 0.06 | 901.19 | 3.38 | 10.01 | 695.96 | 5548.05 | 4457.91 | 1330.90 | 819.78 | 7342.23 | 0.00 | 0.44 | 1901.33 | 59.08 | 8.29 | 2334.00 | 2925.78 | 0.41 | 1.85 | 30412.85 |
| 5 | 60.05 | 329.91 | 0.00 | 1618.08 | 0.00 | 11381.33 | 0.20 | 3149.03 | 7.82 | 26.79 | 1233.89 | 4656.33 | 2253.67 | 2032.24 | 2690.15 | 30510.84 | 0.01 | 1.44 | 2081.30 | 65.34 | 5.68 | 2128.55 | 1972.82 | 0.32 | 2.19 | 66207.97 |
| 6 | 18.47 | 58.21 | 0.01 | 616.02 | 0.00 | 2995.51 | 0.06 | 1048.92 | 2.43 | 9.46 | 523.37 | ${ }^{1232.63}$ | 474.16 | 792.49 | 87.68 | 9933.96 | 0.00 | 0.48 | 2817.21 | 103.33 | 7.88 | 2352.84 | 2453.95 | 0.38 | 2.98 | 26315.45 |
| 7 | 35.53 | 89.90 | 0.04 | 2435.80 | 0.14 | 3586.43 | 0.05 | 741.28 | 2.00 | 8.90 | 411.15 | 686.07 | 340.21 | 1601.78 | 650.94 | 8855.13 | 0.00 | 0.40 | 2501.94 | 88.59 | 8.39 | 1028.79 | 1172.51 | 0.22 | 1.18 | 24247.36 |
| 8 | 124.02 | 240.45 | 0.02 | 4338.07 | 0.31 | 19684,27 | 0.20 | 4785.57 | 8.03 | 35.37 | 1321.56 | 3112.50 | 1459.56 | 4379.92 | 2739.15 | 38877.65 | 0.01 | 1.66 | 2418.70 | 189.78 | 7.44 | 807.50 | 1753.89 | 0.17 | 2.71 | 81288.52 |
| 9 | 14.24 | 10.89 | 0.01 | 453.23 | 0.81 | 1062.90 | 0.02 | 335.64 | 0.85 | 3.51 | 147.31 | 312.85 | 155.72 | 59.75 | 282.54 | 3595.80 | 0.00 | 0.17 | 83.01 | 257.43 | 2.20 | 501.93 | 1152.88 | 0.07 | 3.88 | 9723.66 |
| 10 | 29.63 | 32.67 | 0.01 | 938.23 | 1.66 | 2709.92 | 0.01 | 253.23 | 0.33 | 1.53 | 67.19 | 93.43 | 67.36 | 128.87 | 85.89 | 1618.66 | 0.00 | 0.06 | 896.27 | 644.99 | 1.43 | 1027.79 | 1557.67 | 0.13 | 13.85 | 10170.79 |
| 11 | 10.21 | 12.41 | 0.00 | ${ }^{354.64}$ | 0.76 | ${ }^{695} 52$ | 0.01 | 124.42 | 0.22 | 0.78 | 28.96 | 83.78 | ${ }^{61.00}$ | ${ }_{6.97}$ | 75.10 | 884.41 | 0.00 | 0.04 | 531.06 | 466.41 | 0.34 | 952.25 | 987.54 | 0.10 | 10.87 | 5347.48 |
| 12 | 13.92 | 4.27 | 0.01 | 235.13 | 1.71 | 332.57 | 0.00 | 59.84 | 0.13 | 0.47 | 13.52 | 57.14 | 37.42 | 50.27 | 51.08 | 471.78 | 0.00 | 0.03 | 446.94 | 355.66 | 0.11 | 865.39 | 604.38 | ${ }_{0} 0.09$ | 10.24 | 3602.10 |
| 13 | 25.43 | 8.89 | 0.01 | 428.22 | 2.70 | 1098.81 | 0.01 | 107.74 | 0.22 | 0.81 | 21.58 | 136.49 | 42.94 | 19.18 | 91.76 | 845.90 | 0.00 | 0.05 | 230.61 | 177.28 | 0.07 | 555.16 | 578.74 | 0.04 |  |  |
| 14 | 28.35 | 6.95 | 0.01 | 436.24 | 2.11 | 2356.45 | 0.02 | 291.68 | 0.57 | 2.13 | 57.42 | 260.04 | 167.34 | 48.46 | 232.39 | 2209.36 | 0.00 | 0.12 | 423.50 | 298.72 | 0.08 | 988.18 | 817.45 | 0.09 | 9.90 | 863.57 |
| $15+$ | 70.60 | 21.47 | 0.04 | 1234.02 | 8.14 | 2212.98 | 0.00 | 209.62 | 0.08 | ${ }_{0} 0.34$ | 9.88 | 33.43 | 47.28 | 8.35 | 30.10 | 102.17 | 0.00 | 0.02 | 575.38 | 268.37 | 0.25 | 1326.58 | 1501.69 | 0.10 | 8.85 | 7669.74 |

Table 7.2.4.2. Western horse mackerel. Catch-at-age (thousands).

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0 | 3713 | 21072 | 134743 | 11515 | 13197 | 11741 | 8848 | 1651 | 414 | 1651 | 6582 | 18483 | 28679 | 19432 | 8210 |
| 1983 | 0 | 7903 | 2269 | 32900 | 53508 | 15345 | 44539 | 52673 | 17923 | 3291 | 5505 | 3386 | 17017 | 23902 | 38352 | 46482 |
| 1984 | 0 | 0 | 241360 | 4439 | 36294 | 149798 | 22350 | 38244 | 34020 | 14756 | 4101 | 0 | 639 | 1757 | 5080 | 50895 |
| 1985 | 0 | 1633 | 4901 | 602992 | 4463 | 41822 | 100376 | 12644 | 16172 | 6200 | 9224 | 339 | 850 | 3723 | 1250 | 34814 |
| 1986 | 0 | 0 | 0 | 1548 | 676208 | 8727 | 65147 | 109747 | 25712 | 21179 | 15271 | 3116 | 1031 | 855 | 292 | 51531 |
| 1987 | 0 | 99 | 493 | 0 | 2950 | 891660 | 2061 | 41564 | 90814 | 11740 | 9549 | 19363 | 8917 | 1398 | 200 | 32899 |
| 1988 | 876 | 27369 | 6112 | 2099 | 4402 | 18968 | 941725 | 12115 | 39913 | 67869 | 9739 | 16326 | 17304 | 5179 | 4892 | 32396 |
| 1989 | 0 | 0 | 0 | 20766 | 18282 | 5308 | 14500 | 1276730 | 12046 | 59357 | 83125 | 13905 | 24196 | 13731 | 8987 | 18132 |
| 1990 | 0 | 20406 | 45036 | 138929 | 61442 | 33298 | 10549 | 20607 | 1384850 | 37011 | 70512 | 101945 | 14987 | 34687 | 18077 | 56598 |
| 1991 | 20176 | 24021 | 56066 | 17977 | 159643 | 97147 | 49515 | 21713 | 17148 | 1028420 | 20309 | 12161 | 43665 | 8141 | 7053 | 25553 |
| 1992 | 14888 | 229694 | 36332 | 80550 | 56280 | 255874 | 126816 | 48711 | 18992 | 23447 | 1099780 | 13409 | 23002 | 65250 | 11967 | 33246 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 46 | 131108 | 109807 | 16738 | 62342 | 105760 | 325674 | 141148 | 68418 | 55289 | 30689 | 1075610 | 11373 | 24018 | 68137 | 32140 |
| 1994 | 3686 | 60759 | 911713 | 115729 | 53056 | 44520 | 38769 | 221863 | 106390 | 40988 | 43083 | 22380 | 918512 | 10143 | 14599 | 36635 |
| 1995 | 2702 | 233030 | 646753 | 526053 | 269658 | 74592 | 114649 | 36076 | 228687 | 113304 | 96624 | 59874 | 63187 | 951901 | 39278 | 148243 |
| 1996 | 10729 | 19774 | 659641 | 864188 | 189273 | 87562 | 52050 | 55914 | 53835 | 57361 | 56962 | 91690 | 67114 | 56012 | 349086 | 165611 |
| 1997 | 4860 | 110451 | 471611 | 732959 | 408648 | 256563 | 141168 | 143166 | 143769 | 123044 | 133166 | 96058 | 176730 | 98196 | 51674 | 283110 |
| 1998 | 744 | 91505 | 184443 | 488661 | 359590 | 217571 | 153136 | 119309 | 77494 | 67072 | 50108 | 58791 | 30535 | 65839 | 57583 | 141362 |
| 1999 | 14822 | 97561 | 83715 | 176919 | 265820 | 254516 | 212217 | 187196 | 147271 | 77622 | 35582 | 22909 | 34440 | 29743 | 41830 | 122176 |
| 2000 | 565 | 66210 | 130897 | 64801 | 119297 | 232346 | 202175 | 165745 | 109218 | 54365 | 14594 | 17509 | 18642 | 18585 | 10031 | 73174 |
| 2001 | 60561 | 93125 | 204360 | 166641 | 113659 | 120410 | 141419 | 259974 | 218002 | 110319 | 38576 | 22749 | 17102 | 14092 | 18857 | 64868 |
| 2002 | 14044 | 505717 | 122603 | 158114 | 123258 | 66640 | 68890 | 95052 | 132743 | 87285 | 46167 | 29692 | 25333 | 11305 | 12753 | 72682 |
| 2003 | 1913 | 323194 | 509889 | 141442 | 148989 | 89122 | 59047 | 48582 | 52305 | 102089 | 57089 | 31748 | 27158 | 8832 | 7683 | 40641 |
| 2004 | 22237 | 159011 | 116055 | 486195 | 81099 | 98855 | 69441 | 48969 | 32589 | 51953 | 54542 | 33298 | 12581 | 13407 | 4305 | 21278 |
| 2005 | 1305 | 74538 | 171420 | 310767 | 540649 | 69957 | 74746 | 61889 | 44443 | 22726 | 27019 | 42746 | 23677 | 6849 | 7491 | 18626 |
| 2006 | 1905 | 53322 | 58091 | 75505 | 91274 | 482229 | 57377 | 37222 | 41970 | 16865 | 11828 | 17073 | 32025 | 12877 | 7464 | 24645 |
| 2007 | 5121 | 32399 | 38598 | 40530 | 61938 | 112724 | 347284 | 48160 | 29112 | 21504 | 8728 | 7015 | 8462 | 14021 | 7618 | 18335 |
| 2008 | 30155 | 78121 | 24456 | 53525 | 57125 | 84358 | 54701 | 297879 | 49889 | 36692 | 25172 | 14466 | 12787 | 9269 | 13194 | 24124 |
| 2009 | 47421 | 86053 | 31431 | 56816 | 40104 | 36174 | 62700 | 57683 | 273217 | 68318 | 42063 | 30583 | 21230 | 8266 | 6811 | 39752 |
| 2010 | 4331 | 68198 | 122386 | 69381 | 29371 | 30496 | 51312 | 110033 | 73973 | 285281 | 70041 | 34486 | 24421 | 14887 | 14942 | 44201 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 1136 | 17035 | 61864 | 106032 | 51259 | 35380 | 38626 | 59428 | 59031 | 61017 | 239472 | 88764 | 29187 | 17731 | 9783 | 35379 |
| 2012 | 5350 | 48100 | 42653 | 64221 | 171284 | 56012 | 37917 | 28132 | 25608 | 45490 | 41255 | 162118 | 50523 | 24043 | 11621 | 30567 |
| 2013 | 94165 | 138663 | 34651 | 34171 | 76847 | 248958 | 67370 | 25070 | 18447 | 20746 | 31217 | 20836 | 106242 | 21316 | 16279 | 24536 |
| 2014 | 19215 | 26080 | 83034 | 34591 | 28200 | 62102 | 152650 | 56679 | 21786 | 16441 | 23876 | 23654 | 24509 | 57284 | 25197 | 23878 |
| 2015 | 85629 | 108174 | 25416 | 51631 | 31604 | 24613 | 46201 | 118679 | 27331 | 12698 | 10883 | 12584 | 11794 | 7272 | 48586 | 15935 |
| 2016 | 133936 | 168323 | 97368 | 18662 | 31033 | 18762 | 14519 | 22754 | 80818 | 19004 | 10531 | 10298 | 14703 | 16212 | 18451 | 62769 |
| 2017 | 104771 | 135690 | 26426 | 132175 | 34464 | 49849 | 23046 | 14115 | 22170 | 52786 | 12603 | 6491 | 6110 | 6919 | 7284 | 33718 |
| 2018 | 25736 | 107004 | 42957 | 54376 | 257565 | 43887 | 39837 | 14438 | 8809 | 19014 | 44833 | 10875 | 8065 | 4589 | 3645 | 35529 |
| 2019 | 7643 | 53043 | 59271 | 50945 | 52717 | 280292 | 42996 | 38021 | 16292 | 12752 | 19572 | 33296 | 10418 | 4690 | 3940 | 30219 |
| 2020 | 22256 | 57801 | 40360 | 50895 | 17318 | 32781 | 162029 | 19134 | 13415 | 4799 | 4292 | 5888 | 14437 | 5012 | 2647 | 11550 |
| 2021 | 2921 | 22236 | 20310 | 55002 | 91506 | 39834 | 50698 | 107895 | 14412 | 13362 | 9923 | 6880 | 5403 | 7861 | 3217 | 9047 |
| 2022 | 772 | 46212 | 24680 | 27792 | 30413 | 66208 | 26315 | 24247 | 81289 | 9724 | 10171 | 5347 | 3602 | 4378 | 8638 | 7670 |

Table 7.2.4.3. Western horse mackerel. Marginal age-distribution (Timing = month of year, Fleet =1[commercial], sex = mixed, and sample size = no. samples/100).

| year | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 7 | 7 | 7 | 7 | 7 | 7 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Timing | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |


| year | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sample size | 6.8 | 6.8 | 6.8 | 6.8 | 6.8 | 6.8 | 6.8 | 6.8 | 6.8 | 6.8 | 4.5 | 7.5 | 6.1 | 4.8 | 6.3 | 7.5 | 6.2 | 5.1 | 2.8 | 3.2 | 3.6 |
| 0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.013 | 0.007 | 0.000 | 0.001 | 0.001 | 0.004 | 0.001 | 0.000 | 0.008 | 0.000 | 0.036 | 0.009 |
| 1 | 0.013 | 0.022 | 0.000 | 0.002 | 0.000 | 0.000 | 0.023 | 0.000 | 0.010 | 0.015 | 0.107 | 0.058 | 0.023 | 0.065 | 0.007 | 0.033 | 0.042 | 0.054 | 0.051 | 0.056 | 0.322 |
| 2 | 0.073 | 0.006 | 0.400 | 0.006 | 0.000 | 0.000 | 0.005 | 0.000 | 0.022 | 0.035 | 0.017 | 0.049 | 0.345 | 0.179 | 0.233 | 0.140 | 0.085 | 0.046 | 0.101 | 0.123 | 0.078 |
| 3 | 0.465 | 0.090 | 0.007 | 0.717 | 0.002 | 0.000 | 0.002 | 0.013 | 0.068 | 0.011 | 0.038 | 0.007 | 0.044 | 0.146 | 0.305 | 0.217 | 0.226 | 0.098 | 0.050 | 0.100 | 0.101 |
| 4 | 0.040 | 0.147 | 0.060 | 0.005 | 0.690 | 0.003 | 0.004 | 0.012 | 0.030 | 0.099 | 0.026 | 0.028 | 0.020 | 0.075 | 0.067 | 0.121 | 0.166 | 0.147 | 0.092 | 0.068 | 0.078 |
| 5 | 0.046 | 0.042 | 0.248 | 0.050 | 0.009 | 0.801 | 0.016 | 0.003 | 0.016 | 0.060 | 0.120 | 0.047 | 0.017 | 0.021 | 0.031 | 0.076 | 0.101 | 0.141 | 0.179 | 0.072 | 0.042 |
| 6 | 0.040 | 0.122 | 0.037 | 0.119 | 0.066 | 0.002 | 0.780 | 0.009 | 0.005 | 0.031 | 0.059 | 0.144 | 0.015 | 0.032 | 0.018 | 0.042 | 0.071 | 0.118 | 0.156 | 0.085 | 0.044 |
| 7 | 0.031 | 0.144 | 0.063 | 0.015 | 0.112 | 0.037 | 0.010 | 0.814 | 0.010 | 0.013 | 0.023 | 0.063 | 0.084 | 0.010 | 0.020 | 0.042 | 0.055 | 0.104 | 0.128 | 0.156 | 0.060 |
| 8 | 0.006 | 0.049 | 0.056 | 0.019 | 0.026 | 0.082 | 0.033 | 0.008 | 0.676 | 0.011 | 0.009 | 0.030 | 0.040 | 0.063 | 0.019 | 0.043 | 0.036 | 0.082 | 0.084 | 0.131 | 0.084 |
| 9 | 0.001 | 0.009 | 0.024 | 0.007 | 0.022 | 0.011 | 0.056 | 0.038 | 0.018 | 0.639 | 0.011 | 0.024 | 0.016 | 0.031 | 0.020 | 0.036 | 0.031 | 0.043 | 0.042 | 0.066 | 0.056 |
| 10 | 0.006 | 0.015 | 0.007 | 0.011 | 0.016 | 0.009 | 0.008 | 0.053 | 0.034 | 0.013 | 0.514 | 0.014 | 0.016 | 0.027 | 0.020 | 0.039 | 0.023 | 0.020 | 0.011 | 0.023 | 0.029 |
| 11 | 0.023 | 0.009 | 0.000 | 0.000 | 0.003 | 0.017 | 0.014 | 0.009 | 0.050 | 0.008 | 0.006 | 0.476 | 0.008 | 0.017 | 0.032 | 0.028 | 0.027 | 0.013 | 0.013 | 0.014 | 0.019 |
| 12 | 0.064 | 0.047 | 0.001 | 0.001 | 0.001 | 0.008 | 0.014 | 0.015 | 0.007 | 0.027 | 0.011 | 0.005 | 0.348 | 0.018 | 0.024 | 0.052 | 0.014 | 0.019 | 0.014 | 0.010 | 0.016 |
| 13 | 0.099 | 0.065 | 0.003 | 0.004 | 0.001 | 0.001 | 0.004 | 0.009 | 0.017 | 0.005 | 0.031 | 0.011 | 0.004 | 0.264 | 0.020 | 0.029 | 0.030 | 0.016 | 0.014 | 0.008 | 0.007 |
| 14 | 0.067 | 0.105 | 0.008 | 0.001 | 0.000 | 0.000 | 0.004 | 0.006 | 0.009 | 0.004 | 0.006 | 0.030 | 0.006 | 0.011 | 0.123 | 0.015 | 0.027 | 0.023 | 0.008 | 0.011 | 0.008 |
| 15 | 0.028 | 0.127 | 0.084 | 0.041 | 0.053 | 0.030 | 0.027 | 0.012 | 0.028 | 0.016 | 0.016 | 0.014 | 0.014 | 0.041 | 0.058 | 0.084 | 0.065 | 0.068 | 0.056 | 0.039 | 0.046 |


| year | 2003* | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Timing | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 |
| Fleet | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| Sex | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sample size | 7.9 | 6.8 | 7.8 | 7.2 | 6.2 | 7.7 | 8.7 | 7.8 | 6.2 | 6.8 | 7.7 | 8.1 | 6.4 | 8.2 | 6.8 | 6.9 | 6.9 | 5.1 | 7.9 | 7.4 |
| 0 | 0.001 | 0.017 | 0.001 | 0.002 | 0.006 | 0.035 | 0.052 | 0.004 | 0.001 | 0.006 | 0.096 | 0.028 | 0.134 | 0.181 | 0.157 | 0.036 | 0.011 | 0.048 | 0.006 | 0.002 |
| 1 | 0.196 | 0.122 | 0.050 | 0.052 | 0.040 | 0.090 | 0.095 | 0.065 | 0.019 | 0.057 | 0.142 | 0.038 | 0.169 | 0.228 | 0.203 | 0.148 | 0.074 | 0.124 | 0.048 | 0.122 |
| 2 | 0.309 | 0.089 | 0.114 | 0.057 | 0.048 | 0.028 | 0.035 | 0.117 | 0.068 | 0.050 | 0.035 | 0.122 | 0.040 | 0.132 | 0.040 | 0.060 | 0.083 | 0.087 | 0.044 | 0.065 |
| 3 | 0.086 | 0.372 | 0.207 | 0.074 | 0.051 | 0.062 | 0.063 | 0.066 | 0.116 | 0.076 | 0.035 | 0.051 | 0.081 | 0.025 | 0.198 | 0.075 | 0.071 | 0.110 | 0.118 | 0.074 |
| 4 | 0.090 | 0.062 | 0.361 | 0.089 | 0.077 | 0.066 | 0.044 | 0.028 | 0.056 | 0.203 | 0.078 | 0.042 | 0.049 | 0.042 | 0.052 | 0.357 | 0.074 | 0.037 | 0.197 | 0.081 |
| 5 | 0.054 | 0.076 | 0.047 | 0.472 | 0.141 | 0.097 | 0.040 | 0.029 | 0.039 | 0.066 | 0.254 | 0.091 | 0.039 | 0.025 | 0.075 | 0.061 | 0.391 | 0.071 | 0.086 | 0.175 |
| 6 | 0.036 | 0.053 | 0.050 | 0.056 | 0.433 | 0.063 | 0.069 | 0.049 | 0.042 | 0.045 | 0.069 | 0.225 | 0.072 | 0.020 | 0.034 | 0.055 | 0.060 | 0.349 | 0.109 | 0.070 |
| 7 | 0.029 | 0.038 | 0.041 | 0.036 | 0.060 | 0.344 | 0.063 | 0.105 | 0.065 | 0.033 | 0.026 | 0.083 | 0.186 | 0.031 | 0.021 | 0.020 | 0.053 | 0.041 | 0.232 | 0.064 |
| 8 | 0.032 | 0.025 | 0.030 | 0.041 | 0.036 | 0.058 | 0.301 | 0.071 | 0.065 | 0.030 | 0.019 | 0.032 | 0.043 | 0.109 | 0.033 | 0.012 | 0.023 | 0.029 | 0.031 | 0.215 |
| 9 | 0.062 | 0.040 | 0.015 | 0.017 | 0.027 | 0.042 | 0.075 | 0.272 | 0.067 | 0.054 | 0.021 | 0.024 | 0.020 | 0.026 | 0.079 | 0.026 | 0.018 | 0.010 | 0.029 | 0.026 |
| 10 | 0.035 | 0.042 | 0.018 | 0.012 | 0.011 | 0.029 | 0.046 | 0.067 | 0.263 | 0.049 | 0.032 | 0.035 | 0.017 | 0.014 | 0.019 | 0.062 | 0.027 | 0.009 | 0.021 | 0.027 |
| 11 | 0.019 | 0.025 | 0.029 | 0.017 | 0.009 | 0.017 | 0.034 | 0.033 | 0.097 | 0.192 | 0.021 | 0.035 | 0.020 | 0.014 | 0.010 | 0.015 | 0.046 | 0.013 | 0.015 | 0.014 |
| 12 | 0.016 | 0.010 | 0.016 | 0.031 | 0.011 | 0.015 | 0.023 | 0.023 | 0.032 | 0.060 | 0.108 | 0.036 | 0.018 | 0.020 | 0.009 | 0.011 | 0.015 | 0.031 | 0.012 | 0.010 |


| year | $2003^{*}$ | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | $\mathbf{2 0 1 0}$ | $\mathbf{2 0 1 1}$ | $\mathbf{2 0 1 2}$ | $\mathbf{2 0 1 3}$ | $\mathbf{2 0 1 4}$ | $\mathbf{2 0 1 5}$ | $\mathbf{2 0 1 6}$ | $\mathbf{2 0 1 7}$ | $\mathbf{2 0 1 8}$ | $\mathbf{2 0 1 9}$ | $\mathbf{2 0 2 0}$ | $\mathbf{2 0 2 1}$ | $\mathbf{2 0 2 2}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 13 | 0.005 | 0.010 | 0.005 | 0.013 | 0.017 | 0.011 | 0.009 | 0.014 | 0.019 | 0.028 | 0.022 | 0.084 | 0.011 | 0.022 | 0.010 | 0.006 | 0.007 | 0.011 | 0.017 | 0.012 |
| 14 | 0.005 | 0.003 | 0.005 | 0.007 | 0.010 | 0.015 | 0.007 | 0.014 | 0.011 | 0.014 | 0.017 | 0.037 | 0.076 | 0.025 | 0.011 | 0.005 | 0.006 | 0.006 | 0.007 | 0.023 |
| 15 | 0.025 | 0.016 | 0.012 | 0.024 | 0.023 | 0.028 | 0.044 | 0.042 | 0.039 | 0.036 | 0.025 | 0.035 | 0.025 | 0.085 | 0.050 | 0.049 | 0.042 | 0.025 | 0.019 | 0.020 |

*From 2003 the marginal age composition is replaced by the age-length key in the assessment.
Table 7.2.4.4. Western horse mackerel. Conditional age-length key.

|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 18 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 19 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 20 | 0 | 2 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 21 | 0 | 2 | 11 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 22 | 0 | 3 | 18 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 23 | 0 | 0 | 13 | 15 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 24 | 0 | 1 | 24 | 63 | 32 | 7 | 2 | 2 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 25 | 0 | 0 | 8 | 72 | 88 | 22 | 8 | 2 | 1 | 4 | 5 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 26 | 0 | 0 | 2 | 41 | 111 | 57 | 11 | 14 | 18 | 12 | 1 | 0 | 0 | 0 | 1 | 0 |
| 2003 | 27 | 0 | 0 | 0 | 9 | 72 | 81 | 33 | 29 | 29 | 32 | 5 | 1 | 1 | 0 | 0 | 0 |
| 2003 | 28 | 0 | 0 | 0 | 1 | 34 | 54 | 43 | 33 | 25 | 47 | 11 | 3 | 1 | 1 | 1 | 3 |
| 2003 | 29 | 0 | 0 | 0 | 0 | 14 | 30 | 28 | 29 | 49 | 50 | 23 | 11 | 3 | 2 | 0 | 3 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 | 30 | 0 | 0 | 0 | 0 | 1 | 8 | 22 | 23 | 33 | 52 | 19 | 5 | 7 | 2 | 2 | 5 |
| 2003 | 31 | 0 | 0 | 0 | 0 | 1 | 3 | 4 | 4 | 15 | 29 | 29 | 13 | 2 | 3 | 2 | 17 |
| 2003 | 32 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 2 | 7 | 15 | 10 | 8 | 6 | 2 | 3 | 5 |
| 2003 | 33 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 7 | 8 | 5 | 7 | 2 | 2 | 8 |
| 2003 | 34 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 1 | 3 | 6 | 2 | 2 | 0 | 4 | 4 |
| 2003 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 3 | 1 | 2 | 2 | 5 |
| 2003 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 2 | 1 | 0 | 0 | 8 |
| 2003 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 1 | 1 | 10 |
| 2003 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 8 |
| 2003 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2003 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 3 |
| 2004 | 21 | 0 | 0 | 0 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 22 | 0 | 0 | 17 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 23 | 0 | 0 | 52 | 126 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 24 | 0 | 0 | 51 | 186 | 14 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 25 | 0 | 0 | 29 | 164 | 44 | 27 | 6 | 3 | 2 | 2 | 2 | 0 | 0 | 0 | 0 | 0 |
| 2004 | 26 | 0 | 0 | 4 | 95 | 71 | 64 | 21 | 5 | 2 | 13 | 3 | 4 | 1 | 0 | 0 | 1 |
| 2004 | 27 | 0 | 0 | 2 | 28 | 65 | 108 | 35 | 9 | 6 | 10 | 11 | 4 | 0 | 0 | 0 | 1 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 | 28 | 0 | 0 | 1 | 2 | 36 | 73 | 50 | 9 | 9 | 21 | 5 | 7 | 0 | 1 | 0 | 2 |
| 2004 | 29 | 0 | 0 | 0 | 1 | 10 | 32 | 20 | 7 | 13 | 16 | 4 | 6 | 2 | 0 | 0 | 1 |
| 2004 | 30 | 0 | 0 | 0 | 0 | 2 | 4 | 11 | 5 | 8 | 8 | 12 | 3 | 4 | 0 | 1 | 2 |
| 2004 | 31 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 0 | 3 | 4 | 3 | 3 | 2 | 0 | 0 | 3 |
| 2004 | 32 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 3 | 1 | 1 | 3 | 1 | 1 | 1 | 6 |
| 2004 | 33 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 3 | 0 | 2 | 0 | 1 | 0 | 3 |
| 2004 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 1 | 2 | 1 | 0 | 7 |
| 2004 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 1 | 2 | 1 | 0 | 2 | 3 |
| 2004 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 0 | 2 | 1 | 1 | 5 |
| 2004 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 3 | 0 | 0 | 3 |
| 2004 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
| 2004 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 2 |
| 2004 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
| 2004 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2004 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2004 | 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2004 | 44 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2004 | 45 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 22 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 23 | 0 | 0 | 1 | 42 | 54 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 24 | 0 | 0 | 0 | 75 | 151 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 25 | 0 | 0 | 0 | 61 | 230 | 4 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 26 | 0 | 0 | 0 | 30 | 248 | 22 | 17 | 7 | 4 | 3 | 2 | 3 | 0 | 0 | 0 | 0 |
| 2005 | 27 | 0 | 0 | 0 | 18 | 160 | 40 | 35 | 7 | 8 | 7 | 7 | 6 | 2 | 0 | 2 | 1 |
| 2005 | 28 | 0 | 0 | 0 | 3 | 37 | 45 | 51 | 18 | 8 | 12 | 9 | 6 | 2 | 1 | 0 | 0 |
| 2005 | 29 | 0 | 0 | 0 | 0 | 3 | 21 | 39 | 26 | 8 | 19 | 20 | 10 | 3 | 0 | 0 | 3 |
| 2005 | 30 | 0 | 0 | 0 | 0 | 1 | 4 | 22 | 24 | 11 | 15 | 19 | 13 | 7 | 0 | 1 | 2 |
| 2005 | 31 | 0 | 0 | 0 | 0 | 0 | 1 | 10 | 12 | 6 | 6 | 15 | 14 | 2 | 0 | 2 | 3 |
| 2005 | 32 | 0 | 0 | 0 | 0 | 0 | 2 | 13 | 11 | 7 | 8 | 8 | 8 | 3 | 2 | 0 | 4 |
| 2005 | 33 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 0 | 2 | 9 | 5 | 3 | 2 | 0 | 9 |
| 2005 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 3 | 3 | 3 | 8 | 6 | 2 | 3 | 7 |
| 2005 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 | 1 | 5 | 6 | 5 | 1 | 11 |
| 2005 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 4 | 2 | 5 | 4 | 2 | 16 |
| 2005 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 1 | 2 | 3 | 0 | 1 | 15 |
| 2005 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 1 | 14 |
| 2005 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 0 | 3 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2005 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 3 |
| 2005 | 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2005 | 44 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2006 | 23 | 0 | 0 | 0 | 3 | 4 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 24 | 0 | 0 | 0 | 4 | 20 | 201 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 25 | 0 | 0 | 0 | 2 | 15 | 308 | 11 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 26 | 0 | 0 | 0 | 0 | 7 | 303 | 24 | 12 | 3 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2006 | 27 | 0 | 0 | 0 | 0 | 2 | 290 | 30 | 20 | 5 | 2 | 0 | 3 | 4 | 2 | 0 | 0 |
| 2006 | 28 | 0 | 0 | 0 | 0 | 1 | 129 | 67 | 34 | 31 | 5 | 1 | 6 | 8 | 7 | 0 | 0 |
| 2006 | 29 | 0 | 0 | 0 | 0 | 0 | 54 | 46 | 36 | 24 | 6 | 7 | 6 | 9 | 6 | 5 | 1 |
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| 2007 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 |
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| 2014 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 3 |
| 2014 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2014 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 |
| 2014 | 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2015 | 20 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 21 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 22 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 23 | 0 | 0 | 2 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 24 | 0 | 0 | 0 | 8 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 25 | 0 | 0 | 0 | 22 | 5 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 26 | 0 | 0 | 0 | 15 | 22 | 4 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 27 | 0 | 0 | 0 | 8 | 12 | 13 | 11 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2015 | 28 | 0 | 0 | 0 | 5 | 16 | 9 | 11 | 43 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 29 | 0 | 0 | 0 | 3 | 4 | 3 | 18 | 82 | 3 | 1 | 1 | 0 | 0 | 0 | 1 | 0 |
| 2015 | 30 | 0 | 0 | 0 | 0 | 1 | 5 | 15 | 85 | 8 | 2 | 2 | 1 | 1 | 1 | 5 | 1 |
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| 2015 | 32 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 36 | 10 | 6 | 1 | 5 | 9 | 5 | 34 | 5 |
| 2015 | 33 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 20 | 7 | 4 | 5 | 7 | 9 | 3 | 51 | 7 |
| 2015 | 34 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 9 | 0 | 10 | 6 | 5 | 10 | 4 | 43 | 12 |
| 2015 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 4 | 5 | 7 | 6 | 6 | 42 | 11 |
| 2015 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 1 | 2 | 1 | 32 | 9 |
| 2015 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 3 | 2 | 18 | 4 |
| 2015 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 1 | 5 | 5 |
| 2015 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 6 | 3 |
| 2015 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 1 |
| 2015 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 |
| 2016 | 17 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 18 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 19 | 0 | 0 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 20 | 0 | 0 | 25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2016 | 21 | 0 | 0 | 22 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 22 | 0 | 0 | 21 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 23 | 0 | 0 | 16 | 13 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 24 | 0 | 0 | 9 | 14 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 25 | 0 | 0 | 0 | 10 | 13 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 26 | 0 | 0 | 0 | 3 | 12 | 4 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 27 | 0 | 0 | 0 | 0 | 5 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 28 | 0 | 0 | 0 | 0 | 4 | 0 | 2 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 29 | 0 | 0 | 0 | 0 | 1 | 1 | 4 | 4 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 30 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 12 | 12 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 2016 | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 8 | 15 | 2 | 0 | 0 | 1 | 0 | 0 | 0 |
| 2016 | 32 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 15 | 4 | 1 | 1 | 2 | 2 | 7 | 4 |
| 2016 | 33 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 7 | 2 | 0 | 2 | 5 | 3 | 5 | 7 |
| 2016 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 3 | 2 | 5 | 5 | 5 | 7 |
| 2016 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 1 | 2 | 4 | 7 | 6 |
| 2016 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 3 | 6 | 5 | 7 |
| 2016 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 5 | 13 | 7 |
| 2016 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 9 | 3 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2016 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 7 |
| 2016 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 1 | 6 | 5 |
| 2016 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 |
| 2016 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 2017 | 9 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 10 | 10 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 11 | 10 | 91 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 12 | 10 | 46 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 13 | 10 | 60 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 14 | 4 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 15 | 0 | 29 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 16 | 0 | 22 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 17 | 0 | 23 | 74 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 18 | 0 | 19 | 79 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 19 | 0 | 7 | 40 | 70 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 20 | 0 | 1 | 22 | 98 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 21 | 0 | 0 | 8 | 97 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 22 | 0 | 0 | 4 | 104 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2017 | 23 | 0 | 0 | 0 | 112 | 23 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 24 | 0 | 0 | 1 | 105 | 53 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 25 | 0 | 0 | 0 | 69 | 112 | 44 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 26 | 0 | 0 | 1 | 47 | 88 | 128 | 39 | 5 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 27 | 0 | 0 | 0 | 27 | 50 | 145 | 83 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 28 | 0 | 0 | 0 | 6 | 29 | 117 | 136 | 50 | 4 | 7 | 1 | 0 | 0 | 0 | 0 | 2 |
| 2017 | 29 | 0 | 0 | 0 | 3 | 20 | 107 | 53 | 83 | 21 | 28 | 1 | 0 | 0 | 0 | 0 | 1 |
| 2017 | 30 | 0 | 0 | 0 | 0 | 6 | 73 | 24 | 27 | 99 | 74 | 11 | 0 | 0 | 0 | 1 | 2 |
| 2017 | 31 | 0 | 0 | 0 | 0 | 3 | 33 | 13 | 7 | 46 | 137 | 14 | 1 | 2 | 2 | 2 | 5 |
| 2017 | 32 | 0 | 0 | 0 | 0 | 2 | 7 | 3 | 11 | 40 | 97 | 80 | 7 | 2 | 3 | 8 | 6 |
| 2017 | 33 | 0 | 0 | 0 | 0 | 0 | 2 | 4 | 2 | 30 | 69 | 22 | 35 | 9 | 10 | 7 | 8 |
| 2017 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 10 | 47 | 16 | 20 | 31 | 16 | 15 | 6 |
| 2017 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 9 | 16 | 7 | 12 | 16 | 16 | 17 | 5 |
| 2017 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 14 | 6 | 10 | 6 | 9 | 27 | 4 |
| 2017 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 4 | 3 | 2 | 10 | 4 | 10 | 2 |
| 2017 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 3 | 2 | 0 | 1 | 2 | 1 |
| 2017 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 3 | 1 | 1 | 1 |
| 2017 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2017 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
| 2017 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 2018 | 5 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 6 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 7 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 8 | 9 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 9 | 13 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 10 | 14 | 118 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 11 | 3 | 160 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 12 | 2 | 68 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 13 | 18 | 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 14 | 18 | 76 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 15 | 11 | 83 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 16 | 0 | 54 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 17 | 0 | 56 | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 18 | 0 | 66 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 19 | 0 | 55 | 61 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 20 | 0 | 42 | 102 | 41 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2018 | 21 | 0 | 21 | 184 | 100 | 49 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 22 | 0 | 10 | 112 | 104 | 167 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 23 | 0 | 0 | 70 | 119 | 431 | 11 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 24 | 0 | 0 | 15 | 113 | 584 | 52 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 25 | 0 | 0 | 0 | 52 | 531 | 79 | 27 | 3 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 26 | 0 | 0 | 0 | 6 | 409 | 146 | 49 | 10 | 3 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 27 | 0 | 0 | 0 | 3 | 175 | 203 | 140 | 39 | 13 | 6 | 0 | 1 | 0 | 0 | 0 | 1 |
| 2018 | 28 | 0 | 0 | 0 | 0 | 81 | 145 | 217 | 93 | 15 | 15 | 4 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 29 | 0 | 0 | 0 | 0 | 24 | 74 | 177 | 158 | 54 | 12 | 19 | 1 | 1 | 0 | 0 | 0 |
| 2018 | 30 | 0 | 0 | 0 | 0 | 3 | 34 | 130 | 59 | 138 | 61 | 55 | 8 | 0 | 0 | 0 | 2 |
| 2018 | 31 | 0 | 0 | 0 | 0 | 3 | 15 | 78 | 25 | 43 | 139 | 121 | 30 | 9 | 4 | 3 | 13 |
| 2018 | 32 | 0 | 0 | 0 | 0 | 0 | 3 | 41 | 40 | 16 | 65 | 229 | 39 | 16 | 8 | 4 | 40 |
| 2018 | 33 | 0 | 0 | 0 | 0 | 0 | 2 | 13 | 12 | 14 | 40 | 192 | 116 | 33 | 10 | 8 | 62 |
| 2018 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 7 | 4 | 27 | 102 | 63 | 91 | 27 | 18 | 106 |
| 2018 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 16 | 62 | 21 | 70 | 47 | 32 | 115 |
| 2018 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 6 | 26 | 15 | 16 | 15 | 45 | 135 |
| 2018 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 8 | 8 | 7 | 11 | 128 |
| 2018 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 5 | 1 | 4 | 7 | 3 | 79 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2018 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 4 | 6 | 5 | 37 |
| 2018 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 32 |
| 2018 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 9 |
| 2018 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 |
| 2018 | 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 |
| 2018 | 44 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2018 | 45 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 |
| 2018 | 46 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2018 | 47 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2018 | 48 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2018 | 49 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2018 | 50 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |
| 2018 | 51 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 |
| 2019 | 6 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 7 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 8 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 9 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 10 | 27 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2019 | 11 | 12 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 12 | 6 | 68 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 13 | 2 | 63 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 14 | 0 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 15 | 0 | 25 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 16 | 0 | 29 | 33 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 17 | 0 | 17 | 47 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 18 | 0 | 23 | 52 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 19 | 0 | 26 | 52 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 20 | 0 | 25 | 80 | 23 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 21 | 0 | 19 | 99 | 63 | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 22 | 0 | 3 | 92 | 101 | 17 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 23 | 0 | 2 | 67 | 101 | 45 | 31 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 24 | 0 | 0 | 30 | 107 | 77 | 145 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 25 | 0 | 0 | 5 | 67 | 108 | 358 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 26 | 0 | 0 | 0 | 12 | 114 | 509 | 20 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 2019 | 27 | 0 | 0 | 0 | 1 | 83 | 526 | 80 | 18 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 3 |
| 2019 | 28 | 0 | 0 | 0 | 2 | 63 | 404 | 119 | 48 | 6 | 3 | 1 | 1 | 0 | 0 | 0 | 0 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2019 | 29 | 0 | 0 | 0 | 2 | 28 | 219 | 103 | 88 | 22 | 4 | 6 | 5 | 0 | 0 | 0 | 0 |
| 2019 | 30 | 0 | 0 | 0 | 1 | 7 | 98 | 78 | 93 | 78 | 38 | 8 | 26 | 3 | 0 | 0 | 3 |
| 2019 | 31 | 0 | 0 | 0 | 0 | 2 | 40 | 42 | 110 | 33 | 75 | 49 | 61 | 7 | 0 | 0 | 3 |
| 2019 | 32 | 0 | 0 | 0 | 0 | 0 | 14 | 24 | 75 | 19 | 22 | 110 | 96 | 12 | 5 | 2 | 14 |
| 2019 | 33 | 0 | 0 | 0 | 0 | 0 | 2 | 8 | 53 | 17 | 11 | 54 | 136 | 29 | 3 | 2 | 38 |
| 2019 | 34 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 25 | 15 | 8 | 17 | 88 | 68 | 22 | 7 | 56 |
| 2019 | 35 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 10 | 9 | 8 | 15 | 45 | 35 | 37 | 21 | 71 |
| 2019 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 1 | 8 | 24 | 10 | 12 | 34 | 60 |
| 2019 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 1 | 2 | 13 | 8 | 3 | 11 | 71 |
| 2019 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 2 | 8 | 2 | 4 | 2 | 54 |
| 2019 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 2 | 0 | 34 |
| 2019 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 18 |
| 2019 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 8 |
| 2019 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 9 |
| 2019 | 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2019 | 44 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 |
| 2020 | 7 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 8 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
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| 2020 | 9 | 50 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 10 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 11 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 12 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 13 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 14 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 15 | 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 16 | 4 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 17 | 0 | 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 18 | 0 | 38 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 19 | 0 | 56 | 29 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 20 | 0 | 24 | 107 | 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 21 | 0 | 4 | 203 | 40 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 22 | 0 | 4 | 136 | 75 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 23 | 0 | 0 | 97 | 111 | 10 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 24 | 0 | 0 | 21 | 109 | 16 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 25 | 0 | 0 | 12 | 89 | 66 | 23 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 26 | 0 | 0 | 0 | 58 | 76 | 35 | 83 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2020 | 27 | 0 | 0 | 1 | 24 | 69 | 60 | 185 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 28 | 0 | 0 | 0 | 1 | 40 | 101 | 333 | 25 | 3 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 29 | 0 | 0 | 0 | 3 | 6 | 121 | 321 | 31 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 30 | 0 | 0 | 0 | 0 | 5 | 58 | 322 | 68 | 24 | 2 | 4 | 0 | 4 | 0 | 0 | 0 |
| 2020 | 31 | 0 | 0 | 0 | 0 | 4 | 23 | 197 | 102 | 49 | 15 | 8 | 10 | 12 | 0 | 0 | 0 |
| 2020 | 32 | 1 | 0 | 0 | 0 | 0 | 4 | 74 | 62 | 113 | 18 | 10 | 19 | 41 | 5 | 0 | 6 |
| 2020 | 33 | 0 | 0 | 0 | 0 | 0 | 1 | 21 | 29 | 72 | 99 | 15 | 18 | 54 | 2 | 3 | 16 |
| 2020 | 34 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 21 | 39 | 35 | 77 | 24 | 56 | 8 | 4 | 28 |
| 2020 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 9 | 24 | 16 | 40 | 25 | 36 | 11 | 3 | 33 |
| 2020 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 15 | 9 | 19 | 8 | 27 | 24 | 15 | 4 | 39 |
| 2020 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 8 | 9 | 5 | 8 | 15 | 31 | 8 | 1 | 28 |
| 2020 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 2 | 6 | 3 | 6 | 6 | 13 | 10 | 16 |
| 2020 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 5 | 6 | 0 | 0 | 8 | 12 |
| 2020 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 4 | 4 | 0 | 0 | 10 |
| 2020 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 5 |
| 2020 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| 2021 | 5 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 6 | 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2021 | 7 | 24 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 8 | 23 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 9 | 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 10 | 32 | 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 11 | 31 | 66 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 12 | 28 | 122 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 13 | 4 | 189 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 14 | 0 | 152 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 15 | 4 | 96 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 16 | 7 | 98 | 31 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 17 | 7 | 56 | 30 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 18 | 1 | 31 | 60 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 19 | 1 | 24 | 81 | 33 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 20 | 0 | 15 | 125 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 21 | 0 | 5 | 115 | 71 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 22 | 0 | 5 | 66 | 120 | 18 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 23 | 0 | 0 | 41 | 111 | 87 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 24 | 0 | 0 | 12 | 123 | 160 | 55 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2021 | 25 | 0 | 0 | 2 | 85 | 264 | 69 | 25 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 26 | 0 | 0 | 0 | 20 | 261 | 138 | 55 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 27 | 0 | 0 | 0 | 13 | 117 | 170 | 159 | 45 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 28 | 0 | 0 | 1 | 16 | 93 | 73 | 191 | 139 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 29 | 0 | 0 | 0 | 9 | 41 | 79 | 153 | 230 | 16 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 30 | 0 | 0 | 0 | 1 | 15 | 58 | 79 | 402 | 58 | 19 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2021 | 31 | 0 | 0 | 0 | 0 | 11 | 39 | 98 | 295 | 115 | 84 | 15 | 2 | 5 | 10 | 1 | 0 |
| 2021 | 32 | 0 | 0 | 0 | 1 | 6 | 17 | 74 | 265 | 81 | 152 | 31 | 2 | 14 | 15 | 0 | 3 |
| 2021 | 33 | 0 | 0 | 0 | 0 | 0 | 10 | 38 | 146 | 38 | 118 | 68 | 15 | 4 | 23 | 1 | 3 |
| 2021 | 34 | 0 | 0 | 0 | 0 | 0 | 3 | 19 | 61 | 19 | 44 | 69 | 38 | 12 | 23 | 5 | 17 |
| 2021 | 35 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 17 | 17 | 23 | 46 | 54 | 30 | 30 | 6 | 24 |
| 2021 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 14 | 17 | 19 | 16 | 10 | 59 | 54 | 15 | 27 |
| 2021 | 37 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 7 | 14 | 7 | 9 | 26 | 75 | 45 | 23 |
| 2021 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 5 | 6 | 4 | 3 | 22 | 78 | 13 |
| 2021 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 10 | 2 | 51 | 9 |
| 2021 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 12 | 1 |
| 2021 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 6 |
| 2021 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2021 | 46 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 |
| 2022 | 7 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 8 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 9 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 10 | 49 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 11 | 41 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 12 | 17 | 23 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 13 | 6 | 61 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 14 | 2 | 91 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 15 | 0 | 106 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 16 | 0 | 62 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 17 | 0 | 84 | 33 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 18 | 1 | 83 | 53 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 19 | 0 | 49 | 59 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 20 | 0 | 25 | 37 | 25 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 21 | 0 | 19 | 44 | 30 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 22 | 0 | 17 | 38 | 29 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 23 | 0 | 10 | 30 | 46 | 32 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


|  | Length | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2022 | 24 | 0 | 0 | 15 | 51 | 77 | 21 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 25 | 0 | 0 | 1 | 30 | 117 | 53 | 4 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 26 | 0 | 0 | 0 | 16 | 87 | 135 | 21 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 27 | 0 | 0 | 0 | 6 | 60 | 192 | 35 | 5 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 28 | 0 | 0 | 0 | 2 | 43 | 162 | 49 | 18 | 21 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 29 | 0 | 0 | 0 | 1 | 16 | 123 | 56 | 57 | 106 | 5 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2022 | 30 | 0 | 0 | 0 | 0 | 10 | 63 | 57 | 46 | 220 | 11 | 3 | 1 | 0 | 1 | 0 | 0 |
| 2022 | 31 | 0 | 0 | 0 | 1 | 5 | 40 | 50 | 70 | 248 | 52 | 16 | 3 | 0 | 1 | 2 | 2 |
| 2022 | 32 | 0 | 0 | 0 | 0 | 2 | 19 | 27 | 71 | 229 | 54 | 86 | 14 | 3 | 5 | 6 | 4 |
| 2022 | 33 | 0 | 0 | 0 | 0 | 2 | 11 | 19 | 42 | 158 | 42 | 111 | 53 | 16 | 11 | 20 | 5 |
| 2022 | 34 | 0 | 0 | 0 | 1 | 0 | 1 | 5 | 14 | 64 | 25 | 66 | 82 | 44 | 26 | 35 | 18 |
| 2022 | 35 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 4 | 18 | 11 | 40 | 53 | 62 | 30 | 29 | 31 |
| 2022 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 2 | 8 | 11 | 16 | 47 | 69 | 45 |
| 2022 | 37 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 1 | 5 | 3 | 2 | 16 | 50 | 48 |
| 2022 | 38 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 1 | 3 | 20 | 55 |
| 2022 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 4 | 23 |
| 2022 | 40 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 |
| 2022 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |


|  | Length | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2022 | 43 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |

Table 7.2.4.5. Western horse mackerel. Catch-at-length distribution from the commercial fleet.

| year |  | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Timing |  | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 | 7 |
| Fleet |  | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| Sex |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sample number |  | 42 | 50 | 40 | 47 | 53 | 57 | 37 | 46 | 87 | 68 | 49 | 48 | 66 | 63 | 82 | 101 | 108 | 104 | 96 | 51 | 95 | 111 |
| Length bins (cm) | 5 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 6 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 7 | 0.004 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 8 | 0.003 | 0.003 | 0.000 | 0.000 | 0.000 | 0.000 | 0.003 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 9 | 0.001 | 0.006 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.030 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 10 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.059 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 11 | 0.009 | 0.007 | 0.000 | 0.002 | 0.002 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.037 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
|  | 12 | 0.035 | 0.034 | 0.000 | 0.010 | 0.004 | 0.002 | 0.001 | 0.003 | 0.000 | 0.002 | 0.000 | 0.000 | 0.001 | 0.000 | 0.020 | 0.004 | 0.000 | 0.001 | 0.004 | 0.002 | 0.000 | 0.001 |
|  | 13 | 0.014 | 0.055 | 0.001 | 0.018 | 0.003 | 0.002 | 0.002 | 0.003 | 0.002 | 0.005 | 0.000 | 0.000 | 0.004 | 0.000 | 0.016 | 0.007 | 0.002 | 0.007 | 0.011 | 0.016 | 0.002 | 0.002 |
|  | 14 | 0.008 | 0.045 | 0.002 | 0.016 | 0.007 | 0.004 | 0.002 | 0.004 | 0.044 | 0.006 | 0.001 | 0.001 | 0.020 | 0.000 | 0.010 | 0.009 | 0.028 | 0.016 | 0.017 | 0.015 | 0.007 | 0.005 |
|  | 15 | 0.016 | 0.039 | 0.007 | 0.022 | 0.017 | 0.007 | 0.001 | 0.033 | 0.054 | 0.010 | 0.003 | 0.002 | 0.048 | 0.001 | 0.012 | 0.014 | 0.017 | 0.026 | 0.016 | 0.003 | 0.009 | 0.014 |
|  | 16 | 0.024 | 0.040 | 0.011 | 0.029 | 0.014 | 0.010 | 0.004 | 0.045 | 0.012 | 0.009 | 0.004 | 0.005 | 0.067 | 0.002 | 0.012 | 0.012 | 0.010 | 0.010 | 0.009 | 0.004 | 0.012 | 0.010 |


| year |  | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 17 | 0.042 | 0.049 | 0.011 | 0.020 | 0.006 | 0.014 | 0.008 | 0.021 | 0.008 | 0.009 | 0.010 | 0.009 | 0.052 | 0.002 | 0.008 | 0.018 | 0.010 | 0.003 | 0.008 | 0.011 | 0.010 | 0.008 |
|  | 18 | 0.044 | 0.054 | 0.016 | 0.025 | 0.007 | 0.013 | 0.012 | 0.020 | 0.014 | 0.009 | 0.017 | 0.009 | 0.043 | 0.003 | 0.011 | 0.019 | 0.022 | 0.008 | 0.005 | 0.016 | 0.010 | 0.016 |
|  | 19 | 0.044 | 0.037 | 0.021 | 0.035 | 0.012 | 0.012 | 0.012 | 0.008 | 0.024 | 0.010 | 0.017 | 0.022 | 0.026 | 0.006 | 0.024 | 0.028 | 0.027 | 0.013 | 0.011 | 0.019 | 0.015 | 0.033 |
|  | 20 | 0.052 | 0.030 | 0.031 | 0.042 | 0.018 | 0.012 | 0.024 | 0.009 | 0.036 | 0.026 | 0.016 | 0.034 | 0.022 | 0.015 | 0.024 | 0.047 | 0.029 | 0.029 | 0.018 | 0.019 | 0.013 | 0.036 |
|  | 21 | 0.061 | 0.033 | 0.027 | 0.091 | 0.054 | 0.023 | 0.036 | 0.014 | 0.019 | 0.057 | 0.030 | 0.046 | 0.022 | 0.025 | 0.021 | 0.055 | 0.043 | 0.051 | 0.030 | 0.046 | 0.027 | 0.025 |
|  | 22 | 0.072 | 0.031 | 0.027 | 0.109 | 0.120 | 0.039 | 0.076 | 0.044 | 0.024 | 0.062 | 0.041 | 0.035 | 0.022 | 0.028 | 0.019 | 0.041 | 0.060 | 0.069 | 0.038 | 0.034 | 0.029 | 0.027 |
|  | 23 | 0.098 | 0.034 | 0.032 | 0.117 | 0.120 | 0.086 | 0.123 | 0.065 | 0.032 | 0.044 | 0.048 | 0.039 | 0.026 | 0.024 | 0.026 | 0.023 | 0.072 | 0.121 | 0.038 | 0.030 | 0.039 | 0.024 |
|  | 24 | 0.112 | 0.054 | 0.026 | 0.092 | 0.113 | 0.161 | 0.102 | 0.067 | 0.031 | 0.034 | 0.059 | 0.049 | 0.026 | 0.026 | 0.031 | 0.016 | 0.065 | 0.135 | 0.053 | 0.047 | 0.048 | 0.031 |
|  | 25 | 0.087 | 0.077 | 0.029 | 0.088 | 0.084 | 0.139 | 0.109 | 0.081 | 0.037 | 0.033 | 0.051 | 0.072 | 0.045 | 0.030 | 0.032 | 0.022 | 0.058 | 0.109 | 0.097 | 0.021 | 0.059 | 0.042 |
|  | 26 | 0.069 | 0.063 | 0.040 | 0.069 | 0.071 | 0.086 | 0.114 | 0.101 | 0.049 | 0.041 | 0.041 | 0.076 | 0.075 | 0.036 | 0.031 | 0.026 | 0.039 | 0.077 | 0.126 | 0.041 | 0.065 | 0.059 |
|  | 27 | 0.059 | 0.044 | 0.071 | 0.063 | 0.058 | 0.068 | 0.099 | 0.110 | 0.084 | 0.067 | 0.050 | 0.066 | 0.087 | 0.060 | 0.038 | 0.033 | 0.042 | 0.048 | 0.132 | 0.103 | 0.075 | 0.066 |
|  | 28 | 0.043 | 0.032 | 0.094 | 0.042 | 0.048 | 0.049 | 0.069 | 0.097 | 0.105 | 0.092 | 0.055 | 0.052 | 0.076 | 0.102 | 0.060 | 0.037 | 0.050 | 0.033 | 0.103 | 0.171 | 0.102 | 0.073 |
|  | 29 | 0.027 | 0.026 | 0.106 | 0.031 | 0.038 | 0.034 | 0.048 | 0.072 | 0.098 | 0.119 | 0.083 | 0.064 | 0.058 | 0.118 | 0.075 | 0.060 | 0.056 | 0.032 | 0.067 | 0.117 | 0.113 | 0.093 |
|  | 30 | 0.021 | 0.025 | 0.107 | 0.019 | 0.028 | 0.024 | 0.030 | 0.053 | 0.066 | 0.106 | 0.117 | 0.087 | 0.050 | 0.112 | 0.093 | 0.083 | 0.069 | 0.032 | 0.050 | 0.091 | 0.116 | 0.119 |
|  | 31 | 0.014 | 0.021 | 0.111 | 0.014 | 0.024 | 0.017 | 0.020 | 0.041 | 0.043 | 0.078 | 0.101 | 0.094 | 0.054 | 0.109 | 0.095 | 0.092 | 0.074 | 0.039 | 0.042 | 0.052 | 0.087 | 0.104 |
|  | 32 | 0.012 | 0.023 | 0.098 | 0.008 | 0.019 | 0.022 | 0.016 | 0.033 | 0.035 | 0.062 | 0.072 | 0.073 | 0.046 | 0.096 | 0.063 | 0.098 | 0.066 | 0.039 | 0.034 | 0.033 | 0.055 | 0.075 |
|  | 33 | 0.009 | 0.025 | 0.047 | 0.009 | 0.021 | 0.028 | 0.013 | 0.023 | 0.033 | 0.041 | 0.052 | 0.055 | 0.035 | 0.077 | 0.063 | 0.088 | 0.057 | 0.032 | 0.032 | 0.029 | 0.030 | 0.047 |
|  | 34 | 0.008 | 0.029 | 0.027 | 0.010 | 0.024 | 0.031 | 0.014 | 0.016 | 0.032 | 0.026 | 0.043 | 0.036 | 0.025 | 0.047 | 0.029 | 0.069 | 0.045 | 0.028 | 0.025 | 0.028 | 0.022 | 0.031 |



Table 7.2.4.6. Western horse mackerel. Catch-at-length distribution from the PELACUS survey (fleet 5).

| year |  | $\begin{array}{r} 199 \\ 5 \end{array}$ | $\begin{array}{r} 199 \\ 7 \end{array}$ | 199 | $\begin{array}{r} 199 \\ 9 \end{array}$ | $\begin{array}{r} 200 \\ 0 \end{array}$ | 200 | 200 | $\begin{array}{r} 200 \\ 3 \end{array}$ | 200 4 | 200 | 200 | 200 7 | 201 | 2014 | 201 | 201 | 201 7 | 201 8 | 201 9 | 202 | 202 | $\begin{array}{r} 202 \\ 3 \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Timing |  | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 | 5.08 |
| Sex |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Number samples |  | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 | 10 |
| Length bins (cm) | 5 | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | 0.00 0 | 0.00 0 | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | 0.00 0 | 0.00 0 | 0.00 0 | 0.00 0 | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | $\begin{array}{r} 0.00 \\ 0 \end{array}$ | 0.00 0 | 0.00 0 | $\begin{array}{r} 0.00 \\ 0 \end{array}$ |
| 6 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 8 |  | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 2 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 |
| 9 |  | 0.00 | 0.00 | 0.00 | 0.03 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 8 | 0 | 0 | 2 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 2 | 0 | 0 | 0 | 1 |
| 10 |  | 0.00 | 0.00 | 0.00 | 0.05 | 0.00 | 0.00 | 0.20 | 0.00 | 0.00 | 0.14 | 0.00 | 0.00 | 0.00 | 0.00 | 0.04 | 0.00 | 0.04 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 5 | 0 | 0 | 7 | 0 | 4 | 8 | 0 | 0 | 4 | 0 | 9 | 0 | 7 | 7 | 3 | 2 | 0 | 2 |
| 11 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.25 | 0.00 | 0.00 | 0.11 | 0.00 | 0.00 | 0.00 | 0.00 | 0.05 | 0.00 | 0.11 | 0.10 | 0.07 | 0.05 | 0.00 | 0.04 |
|  |  | 2 | 0 | 2 | 6 | 4 | 0 | 7 | 0 | 6 | 3 | 0 | 0 | 9 | 3 | 8 | 9 | 2 | 1 | 7 | 8 | 4 | 1 |
| 12 |  | 0.04 | 0.01 | 0.00 | 0.00 | 0.04 | 0.00 | 0.09 | 0.00 | 0.00 | 0.02 | 0.00 | 0.00 | 0.02 | 0.01 | 0.10 | 0.01 | 0.09 | 0.06 | 0.14 | 0.11 | 0.04 | 0.23 |
|  |  | 3 | 7 | 9 | 2 | 6 | 0 | 2 | 0 | 1 | 5 | 0 | 0 | 4 | 5 | 8 | 4 | 7 | 8 | 4 | 0 | 1 | 1 |
| 13 |  | 0.06 | 0.02 | 0.01 | 0.00 | 0.02 | 0.00 | 0.06 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.08 | 0.01 | 0.12 | 0.00 | 0.06 | 0.08 | 0.09 | 0.07 | 0.14 | 0.15 |
|  |  | 6 | 8 | 6 | 2 | 5 | 0 | 3 | 0 | 0 | 7 | 1 | 0 | 0 | 2 | 6 | 3 | 0 | 1 | 6 | 3 | 7 | 4 |
| 14 |  | 0.04 | 0.08 | 0.01 | 0.00 | 0.00 | 0.00 | 0.03 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.08 | 0.00 | 0.09 | 0.00 | 0.03 | 0.08 | 0.03 | 0.02 | 0.16 | 0.08 |
|  |  | 7 | 4 | 3 | 0 | 6 | 0 | 8 | 0 | 0 | 9 | 0 | 1 | 3 | 3 | 5 | 9 | 4 | 7 | 8 | 9 | 7 | 8 |
| 15 |  | 0.02 | 0.14 | 0.00 | 0.00 | 0.01 | 0.00 | 0.01 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.02 | 0.00 | 0.03 | 0.05 | 0.01 | 0.12 | 0.05 | 0.03 | 0.18 | 0.03 |
|  |  | 9 | 0 | 5 | 0 | 9 | 0 | 8 | 0 | 0 | 7 | 4 | 3 | 0 | 1 | 5 | 3 | 4 | 4 | 1 | 9 | 9 | 8 |
| 16 |  | 0.01 | 0.12 | 0.00 | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.03 | 0.02 | 0.00 | 0.02 | 0.01 | 0.00 | 0.16 | 0.01 | 0.18 | 0.06 | 0.05 | 0.06 | 0.02 |
|  |  | 8 | 3 | 0 | 0 | 5 | 0 | 5 | 0 | 1 | 4 | 0 | 4 | 7 | 1 | 7 | 5 | 7 | 4 | 8 | 2 | 1 | 3 |
| 17 |  | 0.07 | 0.08 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.01 | 0.00 | 0.02 | 0.01 | 0.00 | 0.02 | 0.03 | 0.01 | 0.14 | 0.10 | 0.13 | 0.08 | 0.06 | 0.04 | 0.02 |
|  |  | 9 | 9 | 1 | 0 | 8 | 0 | 2 | 7 | 0 | 0 | 8 | 1 | 3 | 9 | 2 | 4 | 6 | 0 | 1 | 2 | 3 | 1 |
| 18 |  | 0.14 | 0.04 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.02 | 0.00 | 0.01 | 0.01 | 0.00 | 0.02 | 0.06 | 0.02 | 0.05 | 0.12 | 0.03 | 0.09 | 0.06 | 0.07 | 0.05 |
|  |  | 8 | 5 | 5 | 0 | 3 | 0 | 4 | 4 | 0 | 2 | 9 | 3 | 1 | 6 | 0 | 9 | 0 | 9 | 1 | 9 | 3 | 9 |
| 19 |  | 0.16 | 0.07 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.01 | 0.01 | 0.02 | 0.08 | 0.02 | 0.05 | 0.07 | 0.02 | 0.07 | 0.05 | 0.06 | 0.07 |
|  |  | 3 | 3 | 5 | 0 | 1 | 0 | 2 | 9 | 1 | 1 | 7 | 2 | 0 | 1 | 2 | 9 | 6 | 9 | 2 | 5 | 0 | 3 |


| year |  | $\begin{array}{r} 199 \\ 5 \end{array}$ | $\begin{array}{r} 199 \\ 7 \end{array}$ | 199 | 199 | 200 | 200 | 200 | 200 3 | 200 4 | 200 | 200 6 | 200 7 | 201 3 | 201 4 | 201 5 | 201 | 201 7 | 201 8 | 201 9 | 202 1 | 202 2 | 202 3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 20 |  | 0.08 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 | 0.05 | 0.02 | 0.19 | 0.03 | 0.05 | 0.04 | 0.03 | 0.03 | 0.03 | 0.07 | 0.03 |
|  |  | 3 | 8 | 5 | 0 | 7 | 0 | 5 | 6 | 8 | 2 | 9 | 7 | 4 | 5 | 6 | 7 | 3 | 6 | 9 | 0 | 7 | 9 |
| 21 |  | 0.03 | 0.03 | 0.00 | 0.00 | 0.01 | 0.00 | 0.01 | 0.01 | 0.12 | 0.00 | 0.04 | 0.11 | 0.01 | 0.23 | 0.05 | 0.05 | 0.03 | 0.03 | 0.05 | 0.03 | 0.08 | 0.04 |
|  |  | 2 | 1 | 7 | 2 | 2 | 0 | 3 | 8 | 6 | 2 | 7 | 7 | 3 | 5 | 3 | 9 | 4 | 2 | 0 | 9 | 3 | 0 |
| 22 |  | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.03 | 0.12 | 0.00 | 0.08 | 0.17 | 0.01 | 0.08 | 0.05 | 0.05 | 0.03 | 0.02 | 0.03 | 0.02 | 0.01 | 0.02 |
|  |  | 2 | 7 | 3 | 7 | 7 | 2 | 0 | 0 | 3 | 8 | 7 | 1 | 1 | 9 | 9 | 2 | 1 | 8 | 2 | 6 | 5 | 9 |
| 23 |  | 0.01 | 0.02 | 0.00 | 0.03 | 0.02 | 0.00 | 0.00 | 0.05 | 0.12 | 0.02 | 0.07 | 0.14 | 0.02 | 0.03 | 0.08 | 0.07 | 0.03 | 0.02 | 0.01 | 0.02 | 0.01 | 0.02 |
|  |  | 4 | 6 | 7 | 5 | 3 | 4 | 4 | 6 | 9 | 6 | 3 | 2 | 2 | 9 | 3 | 3 | 5 | 4 | 9 | 7 | 1 | 3 |
| 24 |  | 0.02 | 0.03 | 0.01 | 0.06 | 0.06 | 0.02 | 0.00 | 0.07 | 0.07 | 0.03 | 0.07 | 0.07 | 0.02 | 0.00 | 0.10 | 0.06 | 0.03 | 0.01 | 0.02 | 0.05 | 0.00 | 0.01 |
|  |  | 8 | 2 | 1 | 6 | 4 | 5 | 8 | 3 | 8 | 5 | 2 | 0 | 6 | 9 | 0 | 1 | 1 | 2 | 7 | 8 | 2 | 3 |
| 25 |  | 0.04 | 0.05 | 0.00 | 0.07 | 0.12 | 0.10 | 0.04 | 0.09 | 0.08 | 0.06 | 0.07 | 0.06 | 0.02 | 0.03 | 0.06 | 0.05 | 0.02 | 0.00 | 0.02 | 0.05 | 0.00 | 0.00 |
|  |  | 2 | 3 | 3 | 6 | 5 | 9 | 7 | 8 | 3 | 3 | 1 | 4 | 4 | 4 | 8 | 3 | 1 | 1 | 4 | 6 | 6 | 6 |
| 26 |  | 0.04 | 0.04 | 0.00 | 0.03 | 0.12 | 0.24 | 0.08 | 0.17 | 0.13 | 0.08 | 0.09 | 0.08 | 0.03 | 0.02 | 0.02 | 0.04 | 0.02 | 0.00 | 0.02 | 0.03 | 0.01 | 0.00 |
|  |  | 2 | 0 | 8 | 9 | 3 | 4 | 3 | 9 | 6 | 7 | 0 | 6 | 8 | 8 | 6 | 5 | 8 | 0 | 0 | 3 | 0 | 7 |
| 27 |  | 0.02 | 0.04 | 0.02 | 0.02 | 0.10 | 0.29 | 0.07 | 0.13 | 0.14 | 0.09 | 0.13 | 0.08 | 0.04 | 0.02 | 0.01 | 0.03 | 0.02 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 |
|  |  | 5 | 2 | 9 | 9 | 9 | 3 | 4 | 4 | 1 | 1 | 6 | 3 | 8 | 7 | 1 | 9 | 7 | 0 | 3 | 6 | 8 | 4 |
| 28 |  | 0.02 | 0.03 | 0.09 | 0.04 | 0.08 | 0.14 | 0.03 | 0.09 | 0.05 | 0.08 | 0.10 | 0.07 | 0.07 | 0.01 | 0.00 | 0.01 | 0.02 | 0.00 | 0.01 | 0.02 | 0.00 | 0.02 |
|  |  | 3 | 0 | 9 | 4 | 4 | 1 | 7 | 8 | 8 | 8 | 3 | 6 | 7 | 6 | 7 | 7 | 2 | 1 | 3 | 6 | 2 | 3 |
| 29 |  | 0.03 | 0.04 | 0.21 | 0.14 | 0.09 | 0.08 | 0.01 | 0.09 | 0.03 | 0.06 | 0.07 | 0.05 | 0.12 | 0.02 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.02 | 0.00 | 0.02 |
|  |  | 1 | 4 | 2 | 6 | 4 | 9 | 5 | 7 | 7 | 9 | 7 | 1 | 7 | 7 | 7 | 9 | 3 | 1 | 9 | 5 | 0 | 1 |
| 30 |  | 0.02 | 0.04 | 0.27 | 0.17 | 0.10 | 0.06 | 0.00 | 0.06 | 0.02 | 0.05 | 0.05 | 0.03 | 0.13 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.03 | 0.00 | 0.01 |
|  |  | 9 | 7 | 5 | 9 | 0 | 2 | 8 | 1 | 9 | 9 | 6 | 9 | 4 | 1 | 3 | 2 | 7 | 1 | 2 | 2 | 0 | 8 |
| 31 |  | 0.01 | 0.01 | 0.16 | 0.12 | 0.06 | 0.02 | 0.00 | 0.04 | 0.02 | 0.03 | 0.04 | 0.01 | 0.08 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.03 | 0.00 | 0.01 |
|  |  | 7 | 6 | 6 | 0 | 7 | 1 | 1 | 1 | 2 | 3 | 2 | 4 | 0 | 3 | 6 | 0 | 2 | 0 | 2 | 2 | 0 | 3 |
| 32 |  | 0.00 | 0.01 | 0.07 | 0.06 | 0.01 | 0.00 | 0.00 | 0.02 | 0.00 | 0.01 | 0.04 | 0.00 | 0.04 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.01 |
|  |  | 9 | 7 | 8 | 2 | 6 | 8 | 1 | 8 | 5 | 7 | 0 | 4 | 7 | 6 | 5 | 3 | 3 | 0 | 5 | 4 | 0 | 3 |
| 33 |  | 0.00 | 0.00 | 0.02 | 0.02 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 5 | 0 | 4 | 9 | 0 | 2 | 0 | 6 | 3 | 9 | 4 | 2 | 4 | 8 | 3 | 2 | 4 | 0 | 1 | 4 | 0 | 3 |
| 34 |  | 0.00 | 0.00 | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 4 | 0 | 9 | 1 | 3 | 0 | 0 | 2 | 0 | 2 | 3 | 0 | 6 | 9 | 1 | 1 | 2 | 3 | 1 | 2 | 0 | 2 |
| 35 |  | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 4 | 0 | 4 | 2 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 1 | 2 | 1 | 4 | 1 | 0 | 0 | 0 | 0 | 2 |
| 36 |  | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 2 | 0 | 3 | 1 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 1 |
| 37 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 1 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 1 | 0 | 0 | 0 | 1 |


| year |  | $\begin{array}{r} 199 \\ 5 \end{array}$ | $\begin{array}{r} 199 \\ 7 \end{array}$ | $\begin{array}{r} 199 \\ 8 \\ \hline \end{array}$ | 199 | 200 | 200 1 | 200 | 200 3 | 200 4 | 200 5 | 200 6 | 200 7 | 201 3 | 201 4 | 201 | 201 6 | 201 7 | 201 8 | 201 9 | 202 1 | 202 2 | $\begin{array}{r} 202 \\ 3 \\ \hline \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 38 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 39 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 40 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 41 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 42 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 43 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 44 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 45 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 0 | 0 |
| 46 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 | 0 |
| 47 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 |
| 48 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 |
| 49 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |
| 50 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 51 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
|  |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |


| Q1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.k | 27.8.a | 27.8.b | 27.8.c | 27.8.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.000 | 0.000 | 0.021 | 0.021 | 0.021 | 0.028 | 0.050 | 0.021 | 0.028 | 0.021 |
| 2 | 0 | 0 | 0.079 | 0.079 | 0.079 | 0.079 | 0.062 | 0.079 | 0.079 | 0.093 | 0.079 | 0.079 | 0.079 | 0.062 | 0.045 | 0.089 | 0.051 | 0.085 | 0.062 | 0.059 | 0.063 |
| 3 | 0.000 | 0.000 | 0.115 | 0.115 | 0.115 | 0.115 | 0.082 | 0.146 | 0.115 | 0.101 | 0.115 | 0.067 | 0.115 | 0.097 | 0.078 | 0.116 | 0.084 | 0.098 | 0.097 | 0.084 | 0.099 |
| 4 | 0.139 | 0.139 | 0.139 | 0.136 | 0.139 | 0.139 | 0.123 | 0.166 | 0.139 | 0.119 | 0.139 | 0.129 | 0.139 | 0 | 0.137 | 0.130 | 0.143 | 0.132 | 0.143 | 0.143 | 0.138 |
| 5 | 0.166 | 0.166 | 0.158 | 0.166 | 0.158 | 0.158 | 0.158 | 0.160 | 0.158 | 0.149 | 0.158 | 0.157 | 0.158 | 0 | 0.171 | 0.147 | 0.170 | 0.164 | 0.171 | 0.174 | 0.160 |
| 6 | 0.221 | 0.221 | 0.201 | 0.202 | 0.201 | 0.201 | 0.199 | 0.214 | 0.201 | 0.201 | 0.201 | 0.199 | 0.201 | 0 | 0.198 | 0.178 | 0.193 | 0.188 | 0.201 | 0.196 | 0.204 |
| 7 | 0.234 | 0.234 | 0.225 | 0.226 | 0.225 | 0.225 | 0.225 | 0.225 | 0.225 | 0.225 | 0.225 | 0.225 | 0.225 | 0 | 0.219 | 0.219 | 0.221 | 0.213 | 0.219 | 0.214 | 0.226 |
| 8 | 0.247 | 0.247 | 0.222 | 0.232 | 0.222 | 0.222 | 0.221 | 0.221 | 0.222 | 0.222 | 0.222 | 0.219 | 0.222 | 0 | 0.243 | 0.253 | 0.238 | 0.233 | 0.253 | 0.250 | 0.230 |
| 9 | 0.277 | 0.277 | 0.253 | 0.253 | 0.253 | 0.253 | 0.253 | 0.253 | 0.253 | 0.253 | 0.253 | 0.253 | 0.253 | 0 | 0.268 | 0.263 | 0.272 | 0.266 | 0.263 | 0.257 | 0.258 |
| 10 | 0.307 | 0.307 | 0.311 | 0.301 | 0.311 | 0.311 | 0.311 | 0.311 | 0.311 | 0.311 | 0.311 | 0.316 | 0.311 | 0 | 0.299 | 0.303 | 0.301 | 0.299 | 0.303 | 0.303 | 0.307 |
| 11 | 0.324 | 0.324 | 0.352 | 0.341 | 0.352 | 0.352 | 0.352 | 0.352 | 0.352 | 0.352 | 0.352 | 0.355 | 0.352 | 0 | 0.335 | 0.342 | 0.347 | 0.335 | 0.342 | 0.342 | 0.342 |
| 12 | 0.345 | 0.345 | 0.346 | 0.341 | 0.346 | 0.346 | 0.346 | 0.346 | 0.346 | 0.346 | 0.346 | 0.347 | 0.346 | 0 | 0.369 | 0.372 | 0.375 | 0.368 | 0.372 | 0.372 | 0.354 |
| 13 | 0.330 | 0.330 | 0.400 | 0.336 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.410 | 0.400 | 0 | 0.399 | 0.404 | 0.406 | 0.397 | 0.404 | 0.404 | 0.372 |
| 14 | 0.344 | 0.344 | 0.270 | 0.301 | 0.270 | 0.270 | 0.270 | 0.270 | 0.270 | 0.270 | 0.270 | 0.265 | 0.270 | 0 | 0.436 | 0.440 | 0.440 | 0.446 | 0.440 | 0.440 | 0.310 |
| 15+ | 0.358 | 0.358 | 0.348 | 0.354 | 0.348 | 0.348 | 0.348 | 0.348 | 0.348 | 0.348 | 0.348 | 0.324 | 0.348 | 0 | 0.508 | 0.514 | 0.511 | 0.534 | 0.514 | 0.514 | 0.381 |


| Q2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.5.b | 27.6.a | 27.7.6 | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | 27.8.8. 2 | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | , | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | 0.000 | 0.000 | 0.000 | 0.120 | 0.120 | 0.120 | 0.120 | 0.120 | 0.120 | 0.120 | 0.120 | 0.120 | 0.120 | 0.087 | 0.085 | 0.099 | 0.092 | 0.086 | 0.087 | 0.068 | 0.087 |
| 4 | 0.000 | 0.000 | 0.000 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 | 0.143 | 0.130 | 0.135 | 0.130 | 0.139 | 0.117 | 0.130 | 0.145 | 0.131 |
| 5 | 0.000 | 0.000 | 0.000 | 0.160 | 0.160 | 0.160 | 0.160 | 0.160 | 0.160 | 0.160 | 0.160 | 0.160 | 0.160 | 0.158 | 0.162 | 0.162 | 0.160 | 0.151 | 0.158 | 0.169 | 0.158 |
| 6 | 0.000 | 0.000 | 0.000 | 0.204 | 0.204 | 0.204 | 0.204 | 0.204 | 0.204 | 0.204 | 0.204 | 0.204 | 0.204 | 0.180 | 0.187 | 0.192 | 0.181 | 0.177 | 0.180 | 0.191 | 0.182 |
| 7 | 0.481 | 0.481 | 0.481 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.211 | 0.219 | 0.224 | 0.209 | 0.218 | 0.211 | 0.198 | 0.213 |
| 8 | 0.502 | 0.502 | 0.502 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.229 | 0.233 | 0.245 | 0.239 | 0.229 | 0.240 | 0.233 | 0.239 | 0.232 |
| 9 | 0.511 | 0.511 | 0.511 | 0.255 | 0.255 | 0.255 | 0.255 | 0.255 | 0.255 | 0.255 | 0.255 | 0.255 | 0.255 | 0.270 | 0.270 | 0.260 | 0.269 | 0.272 | 0.270 | 0.270 | 0.275 |
| 10 | 0.519 | 0.519 | 0.519 | 0.306 | 0.306 | 0.306 | 0.306 | 0.306 | 0.306 | 0.306 | 0.306 | 0.306 | 0.306 | 0.303 | 0.305 | 0.290 | 0.303 | 0.300 | 0.303 | 0.308 | 0.309 |
| 11 | 0.526 | 0.526 | 0.526 | 0.350 | 0.350 | 0.350 | 0.350 | 0.350 | 0.350 | 0.350 | 0.350 | 0.350 | 0.350 | 0.350 | 0.342 | 0.288 | 0.358 | 0.346 | 0.350 | 0.353 | 0.352 |
| 12 | 0.533 | 0.533 | 0.533 | 0.348 | 0.348 | 0.348 | 0.348 | 0.348 | 0.348 | 0.348 | 0.348 | 0.348 | 0.348 | 0.376 | 0.370 | 0.376 | 0.379 | 0.377 | 0.376 | 0.376 | 0.380 |
| 13 | 0.539 | 0.539 | 0.539 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.404 | 0.405 | 0.404 | 0.404 | 0.406 | 0.404 | 0.404 | 0.419 |
| 14 | 0.545 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.271 | 0.271 | 0.440 | 0.439 | 0.481 | 0.442 | 0.436 | 0.440 | 0.437 | 0.437 |
| 15+ | 0.567 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.348 | 0.348 | 0.506 | 0.497 | 0.502 | 0.510 | 0.509 | 0.506 | 0.488 | 0.513 |


| Q3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.3.a | 27.3.a. 20 | 27.4.a | 27.6.a | 27.7.b | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | 0 | 0 | 0 | 0 | 0.318 | 0.171 | 0.171 | 0.171 | 0.171 | 0.171 | 0.179 | 0.174 | 0.167 | 0.138 | 0.138 | 0.132 | 0.132 | 0.139 | 0.138 | 0.138 | 0.139 |
| 4 | 0.300 | 0.300 | 0.300 | 0.300 | 0.300 | 0.188 | 0.188 | 0.188 | 0.188 | 0.188 | 0.191 | 0.194 | 0.188 | 0.165 | 0.165 | 0.164 | 0.165 | 0.165 | 0.165 | 0.165 | 0.179 |
| 5 | 0.294 | 0.294 | 0.294 | 0.294 | 0.294 | 0.216 | 0.216 | 0.216 | 0.216 | 0.216 | 0.219 | 0.243 | 0.214 | 0.180 | 0.181 | 0.165 | 0.180 | 0.180 | 0.180 | 0.180 | 0.200 |
| 6 | 0.315 | 0.315 | 0.315 | 0.315 | 0.315 | 0.235 | 0.235 | 0.235 | 0.235 | 0.235 | 0.218 | 0.232 | 0.236 | 0.208 | 0.208 | 0.208 | 0.205 | 0.209 | 0.208 | 0.208 | 0.230 |
| 7 | 0.324 | 0.324 | 0.324 | 0.324 | 0.324 | 0.244 | 0.244 | 0.244 | 0.244 | 0.244 | 0.266 | 0.228 | 0.242 | 0.233 | 0.235 | 0.240 | 0.227 | 0.234 | 0.233 | 0.233 | 0.272 |
| 8 | 0.326 | 0.326 | 0.326 | 0.326 | 0.326 | 0.268 | 0.268 | 0.268 | 0.268 | 0.268 | 0.268 | 0.253 | 0.268 | 0.255 | 0.263 | 0.266 | 0.254 | 0.255 | 0.255 | 0.255 | 0.271 |
| 9 | 0.336 | 0.336 | 0.336 | 0.336 | 0.336 | 0.271 | 0.271 | 0.271 | 0.271 | 0.271 | 0.282 | 0.271 | 0.269 | 0.287 | 0.298 | 0.278 | 0.294 | 0.286 | 0.287 | 0.288 | 0.293 |
| 10 | 0.341 | 0.341 | 0.341 | 0.341 | 0.341 | 0.291 | 0.291 | 0.291 | 0.291 | 0.291 | 0.330 | 0.291 | 0.290 | 0.329 | 0.330 | 0.335 | 0.338 | 0.326 | 0.329 | 0.329 | 0.325 |
| 11 | 0.347 | 0.347 | 0.347 | 0.347 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.359 | 0.378 | 0.364 | 0.350 | 0.354 | 0.354 | 0.353 |
| 12 | 0.349 | 0.349 | 0.349 | 0.349 | 0.349 | 0.421 | 0.421 | 0.421 | 0.421 | 0.421 | 0.448 | 0.421 | 0.395 | 0.389 | 0.387 | 0.393 | 0.391 | 0.388 | 0.389 | 0.388 | 0.380 |
| 13 | 0.355 | 0.355 | 0.355 | 0.355 | 0.355 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.400 | 0.421 | 0.410 | 0.408 | 0.424 | 0.420 | 0.421 | 0.420 | 0.400 |
| 14 | 0.358 | 0.358 | 0.358 | 0.358 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.454 | 0.457 | 0.460 | 0.462 | 0.461 | 0.461 | 0.435 |
| 15+ | 0.368 | 0.368 | 0.368 | 0.368 | 0.368 | 0.348 | 0.348 | 0.348 | 0.348 | 0.348 | 0.357 | 0.348 | 0.348 | 0.920 | 0.530 | 0.552 | 0.533 | 0.544 | 0.563 | 0.563 | 0.496 |


| Q4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.3.a | 27.4.a | 27.6.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | 27.8.8. 2 | Total |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0.186 | 0.186 | 0.186 | 0.186 | 0.156 | 0.132 | 0.132 | 0.141 | 0.126 | 0.121 | 0.133 | 0.132 | 0.196 | 0.187 | 0.167 | 0.173 | 0.164 | 0.164 | 0.173 | 0.157 | 0.140 |
| 5 | 0.213 | 0.213 | 0.262 | 0.205 | 0.177 | 0.185 | 0.185 | 0.199 | 0.131 | 0.128 | 0.198 | 0.185 | 0.206 | 0.219 | 0.190 | 0.198 | 0.185 | 0.181 | 0.198 | 0.176 | 0.197 |
| 6 | 0.238 | 0.238 | 0.275 | 0.229 | 0.197 | 0.226 | 0.226 | 0.231 | 0.217 | 0.232 | 0.241 | 0.226 | 0.223 | 0.231 | 0.210 | 0.217 | 0.208 | 0.204 | 0.217 | 0.192 | 0.227 |
| 7 | 0.279 | 0.279 | 0.288 | 0.260 | 0.228 | 0.256 | 0.256 | 0.257 | 0.000 | 0.309 | 0.269 | 0.256 | 0.246 | 0.271 | 0.258 | 0.263 | 0.234 | 0.224 | 0.263 | 0.263 | 0.263 |
| 8 | 0.271 | 0.271 | 0.292 | 0.247 | 0.223 | 0.251 | 0.251 | 0.253 | 0.000 | 0.314 | 0.259 | 0.251 | 0.246 | 0.275 | 0.277 | 0.269 | 0.258 | 0.249 | 0.269 | 0.293 | 0.258 |
| 9 | 0.265 | 0.265 | 0.265 | 0.265 | 0.278 | 0.267 | 0.267 | 0.268 | 0.000 | 0.335 | 0.284 | 0.267 | 0.252 | 0.306 | 0.301 | 0.300 | 0.290 | 0.278 | 0.300 | 0.305 | 0.277 |
| 10 | 0.300 | 0.300 | 0.302 | 0.293 | 0.266 | 0.280 | 0.280 | 0.280 | 0.000 | 0.321 | 0.283 | 0.280 | 0.278 | 0.331 | 0.336 | 0.333 | 0.332 | 0.327 | 0.333 | 0.337 | 0.302 |
| 11 | 0.306 | 0.306 | 0.304 | 0.316 | 0.295 | 0.295 | 0.295 | 0.295 | 0.000 | 0.505 | 0.297 | 0.295 | 0.269 | 0.359 | 0.362 | 0.358 | 0.361 | 0.330 | 0.358 | 0.366 | 0.327 |
| 12 | 0.322 | 0.322 | 0.322 | 0.322 | 0.319 | 0.319 | 0.319 | 0.319 | 0.000 | 0.379 | 0.292 | 0.319 | 0.401 | 0.386 | 0.383 | 0.386 | 0.389 | 0.371 | 0.386 | 0.387 | 0.372 |
| 13 | 0.262 | 0.262 | 0.262 | 0.262 | 0.415 | 0.415 | 0.415 | 0.415 | 0.432 | 0.406 | 0.405 | 0.415 | 0.356 | 0.412 | 0.404 | 0.413 | 0.420 | 0.403 | 0.413 | 0.408 | 0.390 |
| 14 | 0.337 | 0.337 | 0.337 | 0.337 | 0.317 | 0.317 | 0.317 | 0.317 | 0.000 | 0.371 | 0.325 | 0.317 | 0.269 | 0.452 | 0.448 | 0.453 | 0.456 | 0.456 | 0.453 | ${ }^{0.455}$ | 0.393 |
| 15+ | 0.335 | 0.335 | 0.328 | 0.351 | 0.308 | 0.619 | 0.619 | 0.619 | 0.000 | 0.669 | 0.619 | 0.619 | 0.394 | 0.511 | 0.483 | 0.512 | 0.518 | 0.543 | 0.512 | 0.478 | 0.495 |

## Table 7.2.5.1 cont. Western horse mackerel stock. Mean weight ( kg ) in catch-at-age by quarter and area in 2022 ( $15=15+$ group)

| $\begin{aligned} & \text { all Q } \\ & \text { Aggs } \end{aligned}$ | 27.2.a | 27.3.a | 27.3.a.20 | 27.4.a | 27.5.b | 27.6.a | 27.7.a | 27.7.b | 27.7. | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.8 | 27.7.h | 27.7.j | 27.7.7. 2 | 27.7.k | 27.7.k. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 |  |  |  |  | 0 |  |  |  |  |  | 0 | 0 |  |  |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | - |
| 3 | 0.159 | 0.152 | 0.318 | 0.192 | 0.000 | 0.151 | 0.115 | 0.117 | 0.110 | 0.115 | 0.106 | 0.112 | 0.108 | 0.111 | 0.134 | 0.131 | 0.115 | 0.120 | 0.093 | 0.889 | 0.106 | 0.095 | 0.094 | 0.089 | 0.127 | 0.098 |
| 4 | 0.178 | 0.188 | 0.300 | 0.208 | 0.000 | 0.175 | 0.139 | 0.151 | 0.136 | 0.145 | 0.138 | 0.140 | 0.122 | 0.131 | 0.142 | 0.141 | 0.139 | 0.143 | 0.150 | 0.143 | 0.136 | 0.141 | 0.136 | 0.134 | 0.160 | 0.140 |
| 5 | 0.193 | 0.213 | 0.294 | 0.239 | 0.000 | 0.189 | 0.158 | 0.169 | 0.164 | 0.164 | 0.180 | 0.152 | 0.144 | 0.188 | 0.159 | 0.167 | 0.158 | 0.160 | 0.179 | 0.171 | 0.163 | 0.163 | 0.170 | 0.163 | 0.177 | 0.172 |
| 6 | 0.237 | 0.244 | 0.315 | 0.279 | 0.000 | 0.224 | 0.201 | 0.201 | 0.206 | 0.208 | 0.220 | 0.214 | 0.208 | 0.236 | 0.202 | 0.205 | 0.201 | 0.204 | 0.211 | 0.196 | 0.192 | 0.185 | 0.197 | 0.191 | 0.200 | 0.209 |
| 7 | 0.278 | 0.289 | 0.324 | 0.302 | 0.481 | 0.240 | 0.225 | 0.227 | 0.234 | 0.234 | 0.244 | 0.225 | 0.242 | 0.267 | 0.226 | 0.232 | 0.225 | 0.229 | 0.237 | 0.224 | 0.224 | 0.214 | 0.222 | 0.216 | 0.221 | 0.242 |
| 8 | 0.259 | 0.273 | 0.326 | 0.288 | 0.502 | 0.247 | 0.222 | 0.231 | 0.228 | 0.236 | 0.242 | 0.222 | 0.239 | 0.257 | 0.223 | 0.231 | 0.222 | 0.229 | 0.267 | 0.253 | 0.240 | 0.239 | 0.249 | 0.247 | ${ }^{0.257}$ | 0.241 |
| 9 | 0.372 | 0.310 | 0.336 | 0.329 | 0.511 | 0.279 | 0.253 | 0.253 | 0.256 | 0.258 | 0.262 | 0.253 | 0.270 | 0.282 | 0.253 | 0.254 | 0.253 | 0.255 | 0.283 | 0.284 | 0.262 | 0.282 | 0.283 | 0.267 | 0.293 | 0.271 |
| 10 | 0.381 | 0.308 | 0.341 | 0.317 | 0.519 | 0.308 | 0.311 | 0.300 | 0.298 | 0.299 | 0.291 | 0.311 | 0.315 | 0.288 | 0.310 | 0.294 | 0.311 | ${ }_{0}^{0.306}$ | 0.314 | 0.317 | 0.294 | 0.320 | 0.321 | 0.304 | ${ }_{0} 0.321$ | 0.311 |
| 11 | 0.414 | 0.312 | 0.347 | 0.320 | ${ }_{0} .526$ | 0.326 | 0.352 | 0.341 | 0.340 | 0.345 | 0.328 | 0.352 | 0.422 | 0.300 | 0.352 | 0.332 | 0.352 | 0.350 | 0.351 | 0.349 | 0.320 | 0.358 | 0.348 | 0.347 | ${ }_{0} 0.357$ |  |
| 12 | 0.483 | 0.346 | 0.349 | 0.349 | 0.533 | 0.353 | 0.346 | 0.341 | 0.344 | 0.349 | 0.345 | 0.346 | 0.359 | 0.311 | 0.346 | 0.348 | 0.346 | 0.348 | 0.377 | 0.374 | 0.388 | 0.382 | 0.384 | 0.375 | 0.380 | 0.369 |
| 13 | 0.465 | 0.329 | 0.355 | 0.351 | 0.539 | 0.332 | 0.400 | 0.336 | 0.400 | 0.400 | 0.401 | 0.408 | 0.400 | 0.402 | 0.400 | 0.409 | 0.400 | 0.400 | 0.409 | 0.405 | 0.410 | 0.416 | 0.418 | 0.405 | 0.408 | 0.384 |
| 14 | 0.435 | 0.354 | 0.358 | 0.354 | 0.545 | 0.345 | 0.270 | 0.301 | 0.272 | ${ }^{0.273}$ | 0.278 | 0.270 | 0.309 | 0.294 | 0.270 | 0.266 | 0.270 | 0.271 | 0.444 | 0.442 | ${ }_{0} 0.461$ | 0.450 | 0.459 | 0.441 | 0.444 | 0.350 |
| 15+ | 0.503 | 0.364 | 0.368 | 0.366 | 0.567 | 0.365 | 0.348 | 0.354 | 0.373 | 0.355 | 0.394 | 0.348 | 0.579 | 0.491 | 0.348 | 0.334 | 0.348 | 0.348 | 0.605 | 0.498 | 0.512 | 0.520 | 0.542 | 0.511 | 0.511 | 0.451 |

Table 7.2.5.2. Western horse mackerel stock. Mean length (cm) in catch-at-age by quarter and area in 2022 ( $15=15+$ group)

| Q1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.k | 27.8.a | 27.8.b | 27.8.c | 27.8.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | total |
| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | NA |
| 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 13.1 | 12.9 | 13.1 | 14.5 | 17.5 | 13.1 | 14.4 | 13.1 |
| 2 | 0.0 | 0.0 | 22.1 | 22.1 | 22.1 | 22.1 | 20.3 | 22.1 | 22.1 | 23.5 | 22.1 | 22.1 | 22.1 | 18.8 | 17.0 | 21.1 | 17.7 | 21.1 | 18.8 | 18.6 | 19.0 |
| 3 | 0.0 | 0.0 | 24.7 | 24.7 | 24.7 | 24.7 | 22.6 | 26.7 | 24.7 | 24.1 | 24.7 | 20.6 | 24.7 | 22.1 | 20.5 | 23.5 | 21.0 | 22.2 | 22.1 | 21.1 | 22.5 |
| 4 | 26.4 | 26.4 | 26.5 | 26.3 | 26.5 | 26.5 | 25.5 | 27.6 | 26.5 | 25.6 | 26.5 | 26.1 | 26.5 | 27.2 | 24.9 | 24.5 | 25.2 | 24.5 | 25.3 | 25.3 | 26.3 |
| 5 | 27.9 | 27.9 | 27.8 | 28.0 | 27.8 | 27.8 | 27.8 | 27.8 | 27.8 | 27.3 | 27.8 | 27.8 | 27.8 | 28.4 | 27.0 | 25.6 | 26.7 | 26.4 | 27.3 | 27.4 | 27.8 |
| 6 | 30.5 | 30.5 | 29.8 | 29.8 | 29.8 | 29.8 | 29.7 | 30.2 | 29.8 | 29.8 | 29.8 | 29.8 | 29.8 | 30.2 | 28.3 | 27.4 | 27.9 | 27.7 | 28.8 | 28.4 | 29.8 |
| 7 | 31.1 | 31.1 | 30.6 | 30.8 | 30.6 | 30.6 | 30.6 | 30.6 | 30.6 | 30.6 | 30.6 | 30.5 | 30.6 | 30.6 | 29.3 | 29.8 | 29.2 | 28.9 | 29.8 | 29.5 | 30.6 |
| 8 | 31.6 | 31.6 | 31.0 | 31.1 | 31.0 | 31.0 | 31.0 | 30.9 | 31.0 | 31.0 | 31.0 | 30.9 | 31.0 | 32.1 | 30.3 | 31.2 | 30.0 | 29.8 | 31.2 | 30.7 | 31.2 |
| 9 | 32.9 | 32.9 | 31.9 | 31.9 | 31.9 | 31.9 | 31.9 | 31.9 | 31.9 | 31.9 | 31.9 | 31.9 | 31.9 | 31.4 | 31.2 | 31.3 | 31.3 | 31.1 | 31.3 | 30.9 | 32.0 |
| 10 | 33.9 | 33.9 | 34.9 | 33.8 | 34.9 | 34.9 | 34.9 | 34.9 | 34.9 | 34.9 | 34.9 | 35.5 | 34.9 | 33.8 | 32.4 | 32.8 | 32.4 | 32.4 | 32.8 | 32.8 | 34.0 |
| 11 | 34.7 | 34.7 | 35.3 | 35.0 | 35.3 | 35.3 | 35.3 | 35.3 | 35.3 | 35.3 | 35.3 | 35.4 | 35.3 | 33.8 | 33.6 | 33.8 | 34.0 | 33.6 | 33.8 | 33.8 | 34.8 |
| 12 | 35.3 | 35.3 | 34.6 | 35.0 | 34.6 | 34.6 | 34.6 | 34.6 | 34.6 | 34.6 | 34.6 | 34.5 | 34.6 | 34.8 | 34.7 | 34.8 | 34.9 | 34.7 | 34.8 | 34.8 | 34.8 |
| 13 | 34.7 | 34.7 | 37.1 | 35.1 | 37.1 | 37.1 | 37.1 | 37.1 | 37.1 | 37.1 | 37.1 | 37.4 | 37.1 | 35.8 | 35.6 | 35.8 | 35.8 | 35.6 | 35.8 | 35.8 | 36.0 |
| 14 | 35.2 | 35.2 | 33.5 | 33.9 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.4 | 33.5 | 36.8 | 36.7 | 36.8 | 36.8 | 37.0 | 36.8 | 36.8 | 34.4 |
| 15+ | 35.6 | 35.6 | 35.5 | 35.6 | 35.5 | 35.5 | 35.5 | 35.5 | 35.5 | 35.5 | 35.5 | 35.0 | 35.5 | 38.8 | 38.6 | 38.8 | 38.7 | 39.2 | 38.8 | 38.8 | 36.1 |


| Q2 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.5.b | 27.6.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.7.k. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | Total |
| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | NA |
| 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 15.5 | 12.9 | 17.7 | 15.7 | 18.5 | 15.5 | 15.4 | 15.5 |
| 2 | 0.0 | 0.0 | 0.0 | 22.1 | 22.1 | 22.1 | 22.1 | 22.1 | 22.1 | 22.1 | 22.1 | 22.1 | 22.1 | 19.9 | 18.3 | 21.2 | 19.6 | 20.0 | 19.9 | 17.7 | 19.9 |
| 3 | 0.0 | 0.0 | 0.0 | 24.9 | 24.9 | 24.9 | 24.9 | 24.9 | 24.9 | 24.9 | 24.9 | 24.9 | 24.9 | 21.3 | 21.1 | 22.3 | 21.7 | 21.2 | 21.3 | 19.5 | 21.3 |
| 4 | 0.0 | 0.0 | 0.0 | 26.6 | 26.6 | 26.6 | 26.6 | 26.6 | 26.6 | 26.6 | 26.6 | 26.6 | 26.6 | 24.4 | 24.7 | 24.4 | 25.0 | 23.5 | 24.4 | 25.4 | 24.5 |
| 5 | 0.0 | 0.0 | 0.0 | 27.8 | 27.8 | 27.8 | 27.8 | 27.8 | 27.8 | 27.8 | 27.8 | 27.8 | 27.8 | 26.1 | 26.3 | 26.3 | 26.2 | 25.7 | 26.1 | 26.7 | 26.4 |
| 6 | 0.0 | 0.0 | 0.0 | 29.9 | 29.9 | 29.9 | 29.9 | 29.9 | 29.9 | 29.9 | 29.9 | 29.9 | 29.9 | 27.3 | 27.6 | 27.9 | 27.3 | 27.1 | 27.3 | 27.8 | 27.4 |
| 7 | 35.9 | 35.9 | 35.9 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 30.5 | 28.7 | 29.1 | 29.3 | 28.7 | 29.1 | 28.7 | 28.2 | 28.9 |
| 8 | 37.1 | 37.1 | 37.1 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 | 31.1 | 29.7 | 30.3 | 30.0 | 29.6 | 30.1 | 29.7 | 30.0 | 30.3 |
| 9 | 37.3 | 37.3 | 37.3 | 32.0 | 32.0 | 32.0 | 32.0 | 32.0 | 32.0 | 32.0 | 32.0 | 32.0 | 32.0 | 31.3 | 31.3 | 30.9 | 31.2 | 31.3 | 31.3 | 31.3 | 31.5 |
| 10 | 37.5 | 37.5 | 37.5 | 34.4 | 34.4 | 34.4 | 34.4 | 34.4 | 34.4 | 34.4 | 34.4 | 34.4 | 34.4 | 32.5 | 32.6 | 32.0 | 32.5 | 32.4 | 32.5 | 32.7 | 32.7 |
| 11 | 37.7 | 37.7 | 37.7 | 35.2 | 35.2 | 35.2 | 35.2 | 35.2 | 35.2 | 35.2 | 35.2 | 35.2 | 35.2 | 34.1 | 33.8 | 32.0 | 34.4 | 34.0 | 34.1 | 34.2 | 34.1 |
| 12 | 37.9 | 37.9 | 37.9 | 34.6 | 34.6 | 34.6 | 34.6 | 34.6 | 34.6 | 34.6 | 34.6 | 34.6 | 34.6 | 34.9 | 34.7 | 34.9 | 35.0 | 35.0 | 34.9 | 34.9 | 35.0 |
| 13 | 38.1 | 38.1 | 38.1 | 37.1 | 37.1 | 37.1 | 37.1 | 37.1 | 37.1 | 37.1 | 37.1 | 37.1 | 37.1 | 35.8 | 35.8 | 35.8 | 35.8 | 35.8 | 35.8 | 35.8 | 36.1 |
| 14 | 38.2 | 38.2 | 38.2 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 36.8 | 36.8 | 37.9 | 36.9 | 36.7 | 36.8 | 36.8 | 36.7 |
| $15+$ | 38.8 | 38.8 | 38.8 | 35.5 | 35.5 | 35.5 | 35.5 | 35.5 | 35.5 | 35.5 | 35.5 | 35.5 | 35.5 | 38.6 | 38.3 | 38.5 | 38.7 | 38.6 | 38.6 | 38.1 | 38.6 |

Table 7.2.5.2 cont. Western horse mackerel stock. Mean length ( cm ) in catch-at-age by quarter and area in 2022 ( $\mathbf{1 5 = 1 5 + \text { group } ) ~}$

| Q3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.3.a | 27.3.a. 20 | 27.4.a | 27.6.a | 27.7.b | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | Total |
| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 12.2 | 10.7 | 11.0 | 14.8 | 12.2 | 12.2 | 10.8 | 12.2 |
| 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 17.3 | 16.9 | 21.8 | 16.7 | 21.1 | 17.3 | 17.2 | 17.3 |
| 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 22.7 | 22.7 | 23.0 | 20.7 | 23.6 | 22.7 | 22.7 | 22.7 |
| 3 | 33.0 | 33.0 | 33.0 | 33.0 | 33.0 | 26.8 | 26.8 | 26.8 | 26.8 | 26.8 | 27.0 | 27.5 | 26.0 | 24.9 | 24.9 | 24.6 | 24.5 | 25.0 | 24.9 | 24.9 | 25.0 |
| 4 | 32.5 | 32.5 | 32.5 | 32.5 | 32.5 | 28.0 | 28.0 | 28.0 | 28.0 | 28.0 | 28.0 | 28.4 | 28.0 | 26.5 | 26.5 | 26.5 | 26.5 | 26.5 | 26.5 | 26.5 | 27.4 |
| 5 | 32.2 | 32.2 | 32.2 | 32.2 | 32.2 | 29.1 | 29.1 | 29.1 | 29.1 | 29.1 | 29.0 | 30.4 | 29.0 | 27.3 | 27.3 | 26.5 | 27.3 | 27.3 | 27.3 | 27.3 | 28.3 |
| 6 | 33.2 | 33.2 | 33.2 | 33.2 | 33.2 | 30.0 | 30.0 | 30.0 | 30.0 | 30.0 | 29.0 | 30.5 | 30.0 | 28.6 | 28.6 | 28.6 | 28.5 | 28.7 | 28.6 | 28.6 | 29.7 |
| 7 | 33.5 | 33.5 | 33.5 | 33.5 | 33.5 | 30.2 | 30.2 | 30.2 | 30.2 | 30.2 | 31.9 | 30.5 | 30.0 | 29.7 | 29.8 | 30.0 | 29.5 | 29.8 | 29.7 | 29.7 | 31.3 |
| 8 | 33.7 | 33.7 | 33.7 | 33.7 | 33.7 | 32.0 | 32.0 | 32.0 | 32.0 | 32.0 | 32.0 | 31.4 | 32.0 | 30.7 | 31.0 | 31.1 | 30.6 | 30.7 | 30.7 | 30.7 | 31.9 |
| 9 | 34.0 | 34.0 | 34.0 | 34.0 | 34.0 | 32.0 | 32.0 | 32.0 | 32.0 | 32.0 | 32.0 | 32.0 | 32.0 | 31.9 | 32.3 | 31.6 | 32.1 | 31.9 | 31.9 | 31.9 | 32.4 |
| 10 | 34.2 | 34.2 | 34.2 | 34.2 | 34.2 | 33.1 | 33.1 | 33.1 | 33.1 | 33.1 | 34.7 | 33.1 | 33.0 | 33.4 | 33.4 | 33.6 | 33.7 | 33.3 | 33.4 | 33.4 | 33.4 |
| 11 | 34.3 | 34.3 | 34.3 | 34.3 | 34.3 | 34.0 | 34.0 | 34.0 | 34.0 | 34.0 | 32.0 | 34.0 | 36.0 | 34.2 | 34.4 | 35.0 | 34.5 | 34.1 | 34.2 | 34.2 | 34.2 |
| 12 | 34.5 | 34.5 | 34.5 | 34.5 | 34.5 | 37.5 | 37.5 | 37.5 | 37.5 | 37.5 | 38.0 | 37.5 | 37.0 | 35.3 | 35.3 | 35.5 | 35.4 | 35.3 | 35.3 | 35.3 | 35.1 |
| 13 | 34.7 | 34.7 | 34.7 | 34.7 | 34.7 | 37.0 | 37.0 | 37.0 | 37.0 | 37.0 | 37.0 | 37.0 | 37.0 | 36.3 | 35.9 | 35.9 | 36.3 | 36.2 | 36.3 | 36.3 | 35.8 |
| 14 | 34.8 | 34.8 | 34.8 | 34.8 | 34.8 | 35.0 | 35.0 | 35.0 | 35.0 | 35.0 | 35.8 | 35.0 | 35.0 | 37.4 | 37.2 | 37.3 | 37.4 | 37.4 | 37.4 | 37.4 | 36.8 |
| 15+ | 35.1 | 35.1 | 35.1 | 35.1 | 35.1 | 35.0 | 35.0 | 35.0 | 35.0 | 35.0 | 35.6 | 35.0 | 35.0 | 46.2 | 39.0 | 39.5 | 39.2 | 39.5 | 39.8 | 39.8 | 38.2 |


| Q4 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Ages | 27.2.a | 27.3.a | 27.4.a | 27.6.a | 27.7.b | 27.7.c | 27.7.c. 2 | 27.7.e | 27.7.f | 27.7.g | 27.7.h | 27.7.j | 27.7.j. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.c.e | 27.8.c.w | 27.8.d | 27.8.d. 2 | Total |
| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 12.0 | 10.7 | 11.0 | 12.3 | 17.4 | 12.3 | 11.1 | 11.7 |
| 1 | 0.0 | 0.0 | 0.0 | 0.0 | 19.0 | 19.0 | 19.0 | 19.0 | 19.1 | 18.5 | 19.0 | 19.0 | 19.0 | 20.5 | 18.7 | 20.5 | 21.0 | 20.8 | 20.7 | 16.7 | 20.1 |
| 2 | 0.0 | 0.0 | 0.0 | 0.0 | 22.7 | 22.1 | 22.1 | 22.1 | 22.1 | 22.2 | 22.1 | 22.1 | 22.1 | 22.2 | 20.7 | 21.9 | 22.2 | 22.2 | 22.2 | 20.5 | 22.1 |
| 3 | 26.5 | 26.5 | 26.5 | 26.5 | 25.0 | 24.6 | 24.6 | 24.7 | 24.6 | 24.6 | 24.8 | 24.6 | 26.5 | 24.8 | 23.4 | 23.8 | 24.8 | 24.8 | 24.8 | 24.6 | 24.7 |
| 4 | 28.5 | 28.5 | 28.5 | 28.5 | 27.4 | 26.1 | 26.1 | 26.3 | 25.8 | 25.5 | 26.1 | 26.1 | 29.0 | 28.1 | 26.7 | 27.1 | 26.5 | 26.5 | 27.1 | 26.1 | 26.4 |
| 5 | 29.5 | 29.5 | 30.7 | 29.4 | 28.6 | 28.5 | 28.5 | 28.4 | 26.1 | 25.8 | 29.1 | 28.5 | 29.5 | 29.6 | 27.8 | 28.4 | 27.5 | 27.3 | 28.4 | 27.1 | 28.9 |
| 6 | 30.6 | 30.6 | 31.1 | 30.4 | 29.7 | 30.4 | 30.4 | 29.5 | 29.1 | 29.4 | 31.0 | 30.4 | 30.6 | 30.3 | 28.8 | 29.3 | 28.6 | 28.5 | 29.3 | 27.9 | 30.2 |
| 7 | 31.5 | 31.5 | 31.6 | 31.4 | 31.2 | 31.8 | 31.8 | 31.0 | 0.0 | 33.2 | 32.3 | 31.8 | 31.4 | 32.3 | 31.3 | 31.8 | 29.8 | 29.4 | 31.8 | 31.8 | 31.7 |
| 8 | 31.4 | 31.4 | 31.8 | 31.0 | 31.0 | 31.8 | 31.8 | 31.6 | 0.0 | 33.1 | 32.0 | 31.8 | 31.7 | 32.5 | 31.7 | 32.0 | 30.8 | 30.5 | 32.0 | 32.2 | 31.7 |
| 9 | 32.0 | 32.0 | 32.0 | 32.0 | 33.3 | 32.1 | 32.1 | 31.5 | 0.0 | 34.3 | 32.8 | 32.1 | 31.5 | 33.2 | 32.4 | 32.8 | 32.0 | 31.6 | 32.8 | 32.6 | 32.3 |
| 10 | 32.3 | 32.3 | 32.2 | 32.5 | 32.9 | 32.5 | 32.5 | 32.5 | 0.0 | 33.7 | 32.6 | 32.5 | 32.5 | 33.8 | 33.6 | 33.6 | 33.5 | 33.3 | 33.6 | 33.7 | 32.8 |
| 11 | 32.2 | 32.2 | 32.1 | 32.7 | 34.6 | 34.6 | 34.6 | 34.6 | 0.0 | 37.5 | 34.1 | 34.6 | 34.4 | 34.4 | 34.5 | 34.4 | 34.4 | 33.4 | 34.4 | 34.6 | 33.9 |
| 12 | 34.0 | 34.0 | 34.0 | 34.0 | 35.0 | 35.0 | 35.0 | 35.0 | 0.0 | 35.8 | 34.6 | 35.0 | 36.8 | 35.2 | 35.2 | 35.2 | 35.3 | 34.8 | 35.2 | 35.3 | 35.1 |
| 13 | 31.8 | 31.8 | 31.8 | 31.8 | 37.0 | 37.0 | 37.0 | 37.0 | 37.5 | 37.0 | 36.6 | 37.0 | 35.0 | 36.0 | 35.8 | 36.0 | 36.3 | 35.7 | 36.0 | 35.9 | 35.5 |
| 14 | 34.6 | 34.6 | 34.6 | 34.6 | 33.9 | 33.9 | 33.9 | 33.9 | 0.0 | 35.6 | 34.1 | 33.9 | 32.3 | 37.2 | 37.0 | 37.2 | 37.3 | 37.3 | 37.2 | 37.2 | 35.8 |
| $15+$ | 33.7 | 33.7 | 33.2 | 34.7 | 34.5 | 39.9 | 39.9 | 39.9 | 0.0 | 40.6 | 39.9 | 39.9 | 38.1 | 38.7 | 38.0 | 38.7 | 38.9 | 39.5 | 38.7 | 37.9 | 38.0 |

## Table 7.2.5.2 cont. Western horse mackerel stock. Mean length ( cm ) in catch-at-age by quarter and area in 2022 ( $15=15+$ group)

| all C <br> Ages | 27.2.a | 27.3.a | 27.3.a.20 | 27.4.a | 27.5.b | 27.6.a | 27.7.a | 27.7.b | 27.7.c | 27.7.e.2 | 27.7.e | 27.7. | 27.7.8 | 27.7.h | 27.7.j | 27.7. 2 | 27.7.k | 27.7.k. 2 | 27.8.a | 27.8.b | 27.8.c | 27.8.e.e | 27.8.,.w | 27.8.d | 27.8.d. 2 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 12.0 | 10.7 | 11.0 | 14.8 | 15.6 | 12.3 | 11.0 | 11.8 |
| 1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 19.0 | 19.0 | 19.0 | 19.0 | 19.1 | 18.5 | 19.0 | 19.0 | 19.0 | 0.0 | 0.0 | 14.6 | 13.0 | 20.6 | 16.0 | 20.2 | 13.7 | 16.6 | 15.1 |
| 2 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 22.1 | 22.6 | 22.1 | 22.1 | 21.6 | 22.1 | 22.2 | 22.4 | 22.1 | 22.1 | 22.1 | 22.1 | 20.2 | 17.7 | 22.0 | 19.6 | 20.8 | 19.8 | 19.8 | 20.6 |
| 3 | 26.8 | 26.6 | 33.0 | 28.1 | 0.0 | 26.5 | 24.7 | 24.9 | 24.6 | 24.8 | 24.2 | 24.7 | 24.6 | 24.7 | 25.6 | 24.9 | 24.7 | 24.9 | 21.7 | 21.4 | 22.8 | 21.9 | 21.8 | 21.4 | 24.1 | 22.4 |
| 4 | 28.0 | 28.5 | 32.5 | 29.2 | 0.0 | 27.9 | 26.5 | 27.1 | 26.3 | 26.6 | 26.1 | 26.4 | 25.6 | 26.0 | 26.6 | 26.6 | 26.5 | 26.6 | 25.7 | 25.2 | 24.8 | 25.1 | 24.7 | 24.6 | 26.2 | 26.1 |
| 5 | 28.9 | 29.6 | 32.2 | 30.1 | 0.0 | 28.8 | 27.8 | 28.1 | 28.0 | 27.9 | 28.2 | 27.4 | 26.9 | 28.7 | 27.8 | 28.1 | 27.8 | 27.8 | 27.8 | 26.9 | 26.4 | 26.4 | 26.8 | 26.6 | 27.1 | 28.1 |
| 6 | 30.9 | 30.8 | 33.2 | 31.8 | 0.0 | 30.5 | 29.8 | 29.7 | 30.0 | 29.9 | 29.6 | 30.1 | 29.8 | 30.8 | 29.9 | 29.9 | 29.8 | 29.9 | 29.5 | 28.1 | 27.9 | 27.5 | 28.1 | 28.0 | 28.3 | 29.6 |
| 7 | 32.1 | 32.0 | 33.5 | 32.5 | 35.9 | 31.2 | 30.6 | 30.8 | 30.9 | 30.5 | 30.7 | 30.6 | 31.1 | 32.2 | 30.6 | 30.7 | 30.6 | 30.5 | 31.0 | 29.5 | 29.4 | 28.9 | 29.3 | 29.5 | 29.2 | 30.9 |
| 8 | 31.7 | 31.5 | 33.7 | 32.0 | 37.1 | 31.5 | 31.0 | 31.1 | 31.2 | 31.3 | 31.4 | 30.9 | 31.4 | 32.0 | 31.0 | 31.2 | 31.0 | 31.1 | 32.1 | 30.6 | 30.0 | 30.0 | 30.4 | 30.7 | 30.8 | 31.3 |
| 9 | 34.7 | 33.2 | 34.0 | 33.9 | 37.3 | 32.9 | 31.9 | 32.0 | 32.0 | 32.0 | 31.8 | 31.9 | 32.4 | 32.7 | 31.9 | 31.9 | 31.9 | 32.0 | ${ }^{32.3}$ | 31.8 | 31.0 | 31.7 | 31.7 | 31.3 | 32.1 | 32.2 |
| 10 | 35.0 | 32.6 | 34.2 | 33.2 | 37.5 | 33.8 | 34.9 | 33.8 | 33.9 | 33.9 | 33.3 | 34.9 | 34.3 | 33.0 | 34.8 | 33.6 | 34.9 | 34.4 | 33.5 | 33.0 | 32.2 | 33.1 | 33.1 | 32.7 | 33.1 | 33.5 |
| 11 | 35.7 | 32.5 | 34.3 | 33.1 | 37.7 | 34.6 | 35.3 | 35.0 | 35.2 | 35.1 | 35.0 | 35.3 | 36.3 | 33.8 | 35.3 | 35.1 | 35.3 | 35.2 | 34.1 | 34.0 | 33.0 | 34.4 | 34.0 | 34.0 | 34.3 | 34.4 |
| 12 | 37.1 | 34.5 | 34.5 | 34.6 | 37.9 | 35.4 | 34.6 | 35.0 | 34.6 | 34.7 | 34.8 | 34.6 | 35.1 | 34.9 | 34.6 | 34.5 | 34.6 | 34.6 | 35.0 | 34.9 | 35.3 | 35.1 | 35.2 | 34.9 | 35.0 | 35.0 |
| 13 | 36.8 | 33.9 | 34.7 | 34.7 | 38.1 | 34.7 | 37.1 | 35.1 | 37.1 | 37.1 | 37.1 | 37.2 | 37.1 | 36.9 | 37.1 | 37.4 | 37.1 | 37.1 | 35.9 | 35.8 | 35.9 | 36.1 | 36.2 | 35.8 | 35.9 | 35.9 |
| 14 <br> $15+$ | 36.5 377 | 34.8 34 | 34.8 35.1 | 34.9 35.1 | 38.2 388 | 35.3 357 | ${ }_{3}^{33.5}$ | 33.9 35 | 33.5 359 | 33.5 35.5 | 33.6 | ${ }^{33.5}$ | 34.3 | 33.8 378 | 33.5 355 | 33.4 35.1 | 33.5 35.5 | 33.5 35.5 | 36.9 | 36.9 <br> 8.4 | 37.4 38.7 | 37.1 38.9 | 37.3 39.4 | 36.9 38.7 | 36.9 38.6 | 35.1 37.4 |
| 15+ | 37.7 | 34.9 | 35.1 | 35.1 | 38.8 | 35.7 | 35.5 | 35.6 | 35.9 | 35.5 | 36.2 | 35.5 | 39.2 | 37.8 | 35.5 | 35.1 | 35.5 | 35.5 | 40.4 | 38.4 | 38.7 | 38.9 | 39.4 | 38.7 | 38.6 | 37.4 |

Table 7.2.5.3. Western horse mackerel. Catch weights-at-age (kg), from Q1 and Q3 data

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0.024 | 0.052 | 0.066 | 0.080 | 0.207 | 0.232 | 0.269 | 0.280 | 0.292 | 0.305 | 0.369 | 0.348 | 0.348 | 0.348 | 0.356 | 0.366 |
| 1983 | 0.024 | 0.052 | 0.066 | 0.080 | 0.171 | 0.227 | 0.257 | 0.276 | 0.270 | 0.243 | 0.390 | 0.348 | 0.348 | 0.348 | 0.356 | 0.366 |
| 1984 | 0.024 | 0.052 | 0.064 | 0.077 | 0.122 | 0.155 | 0.201 | 0.223 | 0.253 | 0.246 | 0.338 | 0.348 | 0.348 | 0.348 | 0.356 | 0.366 |
| 1985 | 0.024 | 0.052 | 0.066 | 0.081 | 0.148 | 0.140 | 0.193 | 0.236 | 0.242 | 0.289 | 0.247 | 0.241 | 0.251 | 0.314 | 0.346 | 0.321 |
| 1986 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.134 | 0.169 | 0.195 | 0.242 | 0.292 | 0.262 | 0.319 | 0.287 | 0.345 | 0.260 | 0.360 |
| 1987 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.126 | 0.150 | 0.171 | 0.218 | 0.254 | 0.281 | 0.336 | 0.244 | 0.328 | 0.245 | 0.373 |
| 1988 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.126 | 0.141 | 0.143 | 0.217 | 0.274 | 0.305 | 0.434 | 0.404 | 0.331 | 0.392 | 0.424 |
| 1989 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.103 | 0.131 | 0.159 | 0.127 | 0.210 | 0.252 | 0.381 | 0.400 | 0.421 | 0.448 | 0.516 |
| 1990 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.127 | 0.135 | 0.124 | 0.154 | 0.174 | 0.282 | 0.328 | 0.355 | 0.399 | 0.388 | 0.379 |
| 1991 | 0.024 | 0.052 | 0.066 | 0.080 | 0.121 | 0.137 | 0.143 | 0.144 | 0.150 | 0.182 | 0.189 | 0.303 | 0.323 | 0.354 | 0.365 | 0.330 |
| 1992 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.133 | 0.151 | 0.150 | 0.158 | 0.160 | 0.182 | 0.288 | 0.306 | 0.359 | 0.393 | 0.401 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1993 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.153 | 0.1661 | 0.173 | 0.172 | 0.170 | 0.206 | 0.238 | 0.308 | 0.327 | 0.376 | 0.421 |
| 1994 | 0.024 | 0.052 | 0.066 | 0.080 | 0.105 | 0.147 | 0.185 | 0.169 | 0.191 | 0.191 | 0.190 | 0.275 | 0.240 | 0.326 | 0.342 | 0.383 |
| 1995 | 0.024 | 0.052 | 0.059 | 0.066 | 0.119 | 0.096 | 0.152 | 0.166 | 0.178 | 0.187 | 0.197 | 0.222 | 0.215 | 0.246 | 0.237 | 0.298 |
| 1996 | 0.024 | 0.052 | 0.073 | 0.095 | 0.118 | 0.129 | 0.148 | 0.172 | 0.183 | 0.185 | 0.202 | 0.224 | 0.233 | 0.229 | 0.280 | 0.332 |
| 1997 | 0.024 | 0.052 | 0.066 | 0.080 | 0.112 | 0.124 | 0.162 | 0.169 | 0.184 | 0.188 | 0.208 | 0.241 | 0.229 | 0.268 | 0.286 | 0.266 |
| 1998 | 0.024 | 0.052 | 0.071 | 0.090 | 0.108 | 0.129 | 0.142 | 0.151 | 0.162 | 0.174 | 0.191 | 0.220 | 0.229 | 0.268 | 0.286 | 0.271 |
| 1999 | 0.024 | 0.052 | 0.081 | 0.110 | 0.120 | 0.130 | 0.160 | 0.170 | 0.180 | 0.190 | 0.210 | 0.241 | 0.233 | 0.268 | 0.286 | 0.274 |
| 2000 | 0.024 | 0.052 | 0.102 | 0.115 | 0.128 | 0.158 | 0.169 | 0.181 | 0.208 | 0.224 | 0.225 | 0.227 | 0.247 | 0.247 | 0.272 | 0.378 |
| 2001 | 0.020 | 0.048 | 0.077 | 0.109 | 0.133 | 0.160 | 0.169 | 0.176 | 0.187 | 0.205 | 0.220 | 0.241 | 0.265 | 0.244 | 0.266 | 0.308 |
| 2002 | 0.020 | 0.039 | 0.067 | 0.133 | 0.152 | 0.164 | 0.175 | 0.194 | 0.202 | 0.222 | 0.242 | 0.275 | 0.299 | 0.307 | 0.306 | 0.329 |
| 2003 | 0.022 | 0.060 | 0.089 | 0.114 | 0.142 | 0.160 | 0.175 | 0.178 | 0.194 | 0.205 | 0.226 | 0.249 | 0.267 | 0.286 | 0.278 | 0.317 |
| 2004 | 0.036 | 0.064 | 0.100 | 0.120 | 0.148 | 0.168 | 0.186 | 0.201 | 0.219 | 0.209 | 0.221 | 0.233 | 0.262 | 0.260 | 0.322 | 0.303 |
| 2005 | 0.023 | 0.053 | 0.071 | 0.114 | 0.136 | 0.158 | 0.184 | 0.196 | 0.197 | 0.202 | 0.222 | 0.230 | 0.247 | 0.281 | 0.268 | 0.344 |
| 2006 | 0.019 | 0.038 | 0.078 | 0.114 | 0.141 | 0.154 | 0.180 | 0.199 | 0.212 | 0.222 | 0.235 | 0.229 | 0.235 | 0.248 | 0.253 | 0.304 |
| 2007 | 0.024 | 0.048 | 0.067 | 0.092 | 0.130 | 0.150 | 0.163 | 0.186 | 0.210 | 0.233 | 0.248 | 0.256 | 0.264 | 0.286 | 0.310 | 0.347 |
| 2008 | 0.031 | 0.051 | 0.082 | 0.116 | 0.144 | 0.164 | 0.176 | 0.190 | 0.240 | 0.251 | 0.251 | 0.281 | 0.279 | 0.289 | 0.293 | 0.352 |
| 2009 | 0.025 | 0.047 | 0.070 | 0.107 | 0.156 | 0.177 | 0.187 | 0.203 | 0.225 | 0.252 | 0.270 | 0.292 | 0.306 | 0.322 | 0.316 | 0.370 |
| 2010 | 0.026 | 0.048 | 0.087 | 0.118 | 0.151 | 0.178 | 0.201 | 0.212 | 0.229 | 0.248 | 0.274 | 0.305 | 0.312 | 0.335 | 0.329 | 0.376 |


| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 0.028 | 0.051 | 0.079 | 0.112 | 0.151 | 0.172 | 0.192 | 0.211 | 0.223 | 0.243 | 0.261 | 0.288 | 0.305 | 0.324 | 0.329 | 0.330 |
| 2012 | 0.044 | 0.060 | 0.087 | 0.118 | 0.151 | 0.175 | 0.198 | 0.213 | 0.232 | 0.256 | 0.266 | 0.286 | 0.312 | 0.307 | 0.347 | 0.357 |
| 2013 | 0.040 | 0.058 | 0.102 | 0.130 | 0.154 | 0.172 | 0.195 | 0.228 | 0.243 | 0.249 | 0.248 | 0.288 | 0.288 | 0.321 | 0.348 | 0.355 |
| 2014 | 0.032 | 0.053 | 0.094 | 0.127 | 0.143 | 0.180 | 0.201 | 0.224 | 0.247 | 0.259 | 0.273 | 0.278 | 0.289 | 0.311 | 0.304 | 0.353 |
| 2015 | 0.021 | 0.082 | 0.083 | 0.137 | 0.144 | 0.176 | 0.200 | 0.219 | 0.235 | 0.256 | 0.279 | 0.285 | 0.297 | 0.313 | 0.312 | 0.348 |
| 2016 | 0.016 | 0.055 | 0.096 | 0.133 | 0.164 | 0.192 | 0.200 | 0.225 | 0.249 | 0.254 | 0.306 | 0.295 | 0.310 | 0.335 | 0.337 | 0.339 |
| 2017 | 0.016 | 0.039 | 0.077 | 0.098 | 0.124 | 0.173 | 0.199 | 0.216 | 0.249 | 0.266 | 0.286 | 0.307 | 0.333 | 0.334 | 0.337 | 0.370 |
| 2018 | 0.013 | 0.028 | 0.074 | 0.092 | 0.113 | 0.161 | 0.207 | 0.236 | 0.231 | 0.270 | 0.282 | 0.295 | 0.336 | 0.339 | 0.327 | 0.358 |
| 2019 | 0.011 | 0.032 | 0.074 | 0.108 | 0.156 | 0.159 | 0.205 | 0.237 | 0.268 | 0.277 | 0.304 | 0.309 | 0.346 | 0.386 | 0.400 | 0.402 |
| 2020 | 0.026 | 0.028 | 0.051 | 0.083 | 0.121 | 0.170 | 0.181 | 0.235 | 0.259 | 0.288 | 0.297 | 0.315 | 0.318 | 0.373 | 0.371 | 0.386 |
| 2021 | 0.027 | 0.042 | 0.072 | 0.090 | 0.122 | 0.169 | 0.205 | 0.219 | 0.262 | 0.282 | 0.301 | 0.328 | 0.340 | 0.334 | 0.383 | 0.408 |
| 2022 | 0.021 | 0.035 | 0.077 | 0.098 | 0.140 | 0.172 | 0.209 | 0.242 | 0.241 | 0.271 | 0.311 | 0.343 | 0.369 | 0.384 | 0.350 | 0.451 |

Table 7.2.6.1. Western horse mackerel. Maturity-at-age.

|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1983 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1984 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1985 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1986 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1987 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1988 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1989 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1990 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1991 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1992 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1993 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1994 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1995 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1996 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1997 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1998 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 1999 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2000 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2001 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2002 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2003 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2004 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2005 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2006 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2007 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2008 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |


|  | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2009 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2010 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2011 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2012 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2013 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2014 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2015 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2016 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2017 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2018 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2019 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2020 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2021 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |
| 2022 | 0.00 | 0.00 | 0.05 | 0.27 | 0.73 | 0.95 | 0.99 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 |

Table 7.2.8.1. Western horse mackerel. Potential fecundity ( $10^{6} \mathrm{eggs}$ ) per kg spawning female vs. weight in kg.

| $1987$ <br> w | pfec. | $1992$ <br> w | pfec. | $1995$ <br> w | pfec. | $1998$ <br> w | pfec. | $2000$ <br> w | pfec. | $2001$ <br> w | pfec. | 2001 (cont) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  | w | pfec. |
| 0.168 | 1.524 | 0.105 | 1.317 | 0.13 | 1.307 | 0.172 | 1.318 | 0.258 | 0.841 | 0.086 | 0.688 | 0.165 | 1.382 |
| 0.179 | 0.916 | 0.109 | 2.056 | 0.157 | 1.246 | 0.104 | 0.867 | 0.268 | 0.747 | 0.08 | 0.812 | 0.166 | 1.579 |
| 0.192 | 2.083 | 0.11 | 1.869 | 0.168 | 1.699 | 0.112 | 1.312 | 0.304 | 1.188 | 0.081 | 0.535 | 0.167 | 1.479 |
| 0.233 | 1.644 | 0.112 | 1.772 | 0.179 | 1.135 | 0.206 | 0.382 | 0.311 | 1.411 | 0.095 | 0.88 | 0.113 | 0.527 |
| 0.213 | 1.066 | 0.115 | 1.188 | 0.189 | 1.529 | 0.207 | 0.78 | 0.337 | 0.613 | 0.11 | 1.164 | 0.14 | 0.876 |
| 0.217 | 2.392 | 0.119 | 1.317 | 0.168 | 1.1 | 0.109 | 1.133 | 0.339 | 1.571 | 0.113 | 1.106 | 0.122 | 0.589 |
| 0.277 | 1.617 | 0.12 | 1.413 | 0.209 | 1.497 | 0.132 | 1.02 | 0.341 | 1.522 | 0.095 | 0.823 | 0.12 | 0.68 |
| 0.279 | 1.018 | 0.123 | 1.293 | 0.215 | 1.524 | 0.2 | 1.088 | 0.355 | 1.056 | 0.11 | 0.883 | 0.121 | 0.578 |
| 0.274 | 1.62 | 0.123 | 1.991 | 0.218 | 1.616 | 0.152 | 1.417 | 0.357 | 0.604 | 0.108 | 0.823 | 0.139 | 0.723 |
| 0.3 | 1.513 | 0.131 | 1.617 | 0.226 | 1.883 | 0.149 | 1.004 | 0.367 | 1.15 | 0.097 | 0.741 | 0.144 | 1.213 |
| 0.32 | 1.647 | 0.135 | 0.793 | 0.22 | 1.324 |  |  | 0.393 | 1.279 | 0.101 | 0.853 | 0.144 | 1.265 |


| $1987$ <br> w | pfec. | $1992$ <br> w | pfec. | $1995$ <br> w | pfec. | $1998$ <br> w | pfec. | $2000$ <br> w | pfec. | $2001$ <br> w | pfec. | 2001 (cont) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  | w | pfec. |
| 0.273 | 1.956 | 0.131 | 1.039 | 0.236 | 1.221 |  |  | 0.393 | 0.668 | 0.106 | 1.133 | 0.171 | 0.956 |
| 0.212 | 2.83 | 0.136 | 1.06 | 0.261 | 1.21 |  |  | 0.413 | 0.694 | 0.107 | 0.935 | 0.121 | 0.607 |
| 0.268 | 1.687 | 0.138 | 1.489 | 0.245 | 1.445 |  |  | 0.421 | 1.339 | 0.107 | 0.494 | 0.122 | 0.689 |
| 0.32 | 1.088 | 0.147 | 1.214 | 0.306 | 1.693 |  |  | 0.423 | 0.798 | 0.11 | 0.85 | 0.139 | 0.915 |
| 0.318 | 1.208 | 0.151 | 1.158 | 0.314 | 1.312 |  |  | 0.445 | 1.03 | 0.111 | 0.67 | 0.153 | 0.943 |
| 0.343 | 1.933 | 0.16 | 1.349 | 0.46 | 1.575 |  |  | 0.446 | 1.208 | 0.103 | 0.632 | 0.154 | 0.709 |
| 0.378 | 1.429 | 0.165 | 1.359 | 0.449 | 1.43 |  |  | 0.152 | 0.643 | 0.111 | 0.547 | 0.156 | 0.773 |
| 0.404 | 1.849 | 0.165 | 0.945 |  |  |  |  | 0.165 | 0.579 | 0.118 | 0.88 | 0.162 | 1.158 |
| 0.428 | 2.236 | 0.167 | 1 |  |  |  |  | 0.175 | 0.596 | 0.107 | 0.944 | 0.174 | 1.389 |
| 0.398 | 1.538 | 0.168 | 1.545 |  |  |  |  | 0.179 | 0.997 | 0.104 | 0.724 | 0.175 | 1.426 |
| 0.431 | 1.223 | 0.18 | 1.299 |  |  |  |  | 0.19 | 0.744 | 0.111 | 0.86 | 0.179 | 1.248 |
| 0.432 | 1.465 | 0.174 | 1.487 |  |  |  |  | 0.197 | 0.613 | 0.11 | 0.728 | 0.179 | 1.236 |
| 0.421 | 1.843 | 0.178 | 1.594 |  |  |  |  | 0.203 | 0.702 | 0.111 | 0.544 | 0.18 | 2.353 |
| 0.481 | 1.757 | 0.185 | 1.475 |  |  |  |  | 0.219 | 0.472 | 0.129 | 0.935 | 0.184 | 2.255 |
| 0.494 | 1.611 | 0.195 | 1.41 |  |  |  |  | 0.223 | 0.806 | 0.114 | 0.901 | 0.139 | 0.931 |
| 0.54 | 1.754 | 0.203 | 1.937 |  |  |  |  | 0.227 | 0.606 | 0.114 | 0.557 | 0.161 | 1.037 |
| 0.564 | 2.255 | 0.205 | 1.534 |  |  |  |  | 0.289 | 1.273 | 0.151 | 1.377 | 0.162 | 0.893 |
| 0.585 | 1.221 | 0.213 | 1.577 |  |  |  |  | 0.294 | 1.395 | 0.153 | 1.596 | 0.169 | 0.691 |
|  |  | 0.222 | 0.958 |  |  |  |  | 0.3 | 1.305 | 0.154 | 1.699 | 0.18 | 1.609 |
|  |  | 0.275 | 2.444 |  |  |  |  |  |  | 0.103 | 0.679 | 0.185 | 1.776 |
|  |  |  |  |  |  |  |  |  |  | 0.12 | 1.14 | 0.211 | 2.102 |
|  |  |  |  |  |  |  |  |  |  | 0.12 | 0.631 | 0.224 | 1.466 |
|  |  |  |  |  |  |  |  |  |  | 0.121 | 0.834 | 0.162 | 0.849 |
|  |  |  |  |  |  |  |  |  |  | 0.144 | 0.626 | 0.17 | 0.668 |
|  |  |  |  |  |  |  |  |  |  | 0.116 | 0.668 | 0.187 | 1.453 |
|  |  |  |  |  |  |  |  |  |  | 0.118 | 1.194 | 0.198 | 1.371 |
|  |  |  |  |  |  |  |  |  |  | 0.112 | 0.779 | 0.219 | 1.847 |
|  |  |  |  |  |  |  |  |  |  | 0.126 | 0.782 | 0.22 | 1.578 |


| $1987$ <br> w | pfec. | $1992$ <br> w | pfec. | $1995$ <br> w | pfec. | $1998$ <br> w | pfec. | $2000$ <br> w | pfec. | $2001$ <br> w | pfec. | 2001 (cont) |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |  |  |  |  |  |  |  | w | pfec. |
|  |  |  |  |  |  |  |  |  |  | 0.139 | 1.244 | 0.201 | 0.878 |
|  |  |  |  |  |  |  |  |  |  | 0.119 | 1.212 | 0.206 | 1.196 |
|  |  |  |  |  |  |  |  |  |  | 0.109 | 0.755 | 0.223 | 1.115 |
|  |  |  |  |  |  |  |  |  |  | 0.122 | 0.841 | 0.225 | 1.43 |
|  |  |  |  |  |  |  |  |  |  | 0.131 | 0.929 | 0.233 | 1.724 |
| 8 |  |  |  |  |  |  |  |  |  | 0.135 | 0.862 | 0.241 | 1.131 |
|  |  |  |  |  |  |  |  |  |  | 0.142 | 1.834 | 0.219 | 0.96 |
|  |  |  |  |  |  |  |  |  |  | 0.146 | 1.689 | 0.237 | 1.33 |
|  |  |  |  |  |  |  |  |  |  | 0.148 | 1.357 | 0.241 | 0.918 |
|  |  |  |  |  |  |  |  |  |  | 0.151 | 1.817 | 0.34 | 0.605 |
|  |  |  |  |  |  |  |  |  |  | 0.164 | 1.631 | 0.407 | 1.189 |
|  |  |  |  |  |  |  |  |  |  | 0.164 | 1.052 |  |  |

Table 7.3.1.1. Western horse mackerel. Final assessment. Numbers-at-age (thousands).

| ye <br> ar | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19 | 48926 | 96627 | 24887 | 59303 | 92549 | 12975 | 11865 | 63814 | 35327 | 31441 | 2991 | 3988 | 5497 | 7814 | 3938 | 2047 | 1840 | 1644 | 1470 | 1307 | 1070 |
| 82 | 600 | 7 | 20 | 40 | 2 | 50 | 20 | 2 | 7 | 3 | 59 | 06 | 79 | 14 | 43 | 64 | 75 | 41 | 82 | 23 | 810 |
| 19 | 12328 | 42075 | 82879 | 21245 | 50347 | 78193 | 10925 | 99695 | 53558 | 29632 | 2636 | 2508 | 3343 | 4609 | 6550 | 3301 | 1716 | 1543 | 1378 | 1233 | 1007 |
| 83 | 40 | 300 | 7 | 90 | 30 | 4 | 20 | 1 | 0 | 7 | 54 | 27 | 51 | 06 | 83 | 66 | 56 | 12 | 52 | 00 | 250 |
| 19 | 13742 | 10598 | 36044 | 70546 | 17948 | 42256 | 65321 | 91009 | 82920 | 44511 | 2461 | 2189 | 2083 | 2776 | 3827 | 5440 | 2741 | 1425 | 1281 | 1144 | 9388 |
| 84 | 00 | 80 | 000 | 1 | 90 | 00 | 9 | 5 | 2 | 4 | 79 | 93 | 19 | 73 | 64 | 11 | 82 | 49 | 45 | 76 | 30 |
| 19 | 19810 | 11815 | 90839 | 30715 | 59713 | 15103 | 35409 | 54599 | 75964 | 69164 | 3711 | 2052 | 1825 | 1736 | 2314 | 3190 | 4534 | 2285 | 1188 | 1068 | 8779 |
| 85 | 30 | 40 | 4 | 500 | 7 | 70 | 50 | 1 | 9 | 3 | 44 | 34 | 53 | 47 | 53 | 45 | 45 | 35 | 16 | 10 | 37 |
| 19 | 25487 | 17036 | 10134 | 77554 | 26080 | 50460 | 12719 | 29759 | $45835$ | 63735 | 5801 |  |  |  |  |  |  | 3802 | 1916 |  |  |
| 86 | 70 | 30 | 70 | 5 | 300 | 3 | 90 | $20$ | 1 | 0 | 30 | 61 | 08 | 82 | 11 | $81$ | 28 | 24 | 31 | $0$ | $25$ |
| 19 | 50187 | 21914 | 14602 | 86374 | 65658 | 21951 | 42297 | 10635 | 24848 | 38244 | 5316 | 4838 | 2595 | 1435 | 1276 | 1214 | 1618 | 2230 | 3170 | 1597 | 7715 |
| 87 | 80 | 70 | 10 | 2 | 3 | 700 | 2 | 40 | 30 | 9 | 24 | 14 | 62 | 15 | 47 | 15 | 29 | 70 | 36 | 85 | 68 |
| 19 | 22292 | 43140 | 18763 | 12413 | 72816 | 54949 | 18276 | 35104 | 88116 | 20569 | 3164 | 4397 | 4002 | 2146 | 1187 | 1055 | 1004 | 1338 | 1844 | 2622 | 7702 |
| 88 | 80 | 80 | 30 | 50 | 9 | 7 | 400 | 5 | 8 | 60 | 57 | 99 | 03 | 93 | 02 | 76 | 20 | 45 | 95 | 11 | 90 |
| 19 | $26246$ | 19159 | $36915$ | $15929$ | $10441$ | $60752$ | $45582$ | 15107 | $28962$ | $72630$ | 1694 | 2606 | 3622 |  | 1767 | 9774 | 8693 | 8269 | 1102 | 1519 | 8501 |
| 89 | 00 | 90 | 60 | 20 | 10 | 3 | 1 | 700 | 8 | 2 | 640 | 54 | 02 | 71 | 95 | 7 | 6 | 0 | 13 | 20 |  |
| 19 | 19862 | 22556 | 16393 | 31330 | 13391 | 87052 | 50354 | 37645 | 12453 | 23850 | 5977 | 1394 | 2144 | 2979 | 2711 | 1454 | 8041 | 7151 | 6802 | 9066 | 8243 |
| 90 | 80 | 80 | 00 | 30 | 50 | 9 | 9 | 9 | 000 | 1 | 98 | 470 | 58 | 89 | 34 | 44 | 2 | 8 | 5 | 6 | 75 |
| 19 | 37729 | 17060 | 19253 | 13835 | 26093 | 11024 | 71079 | 40910 | 30501 | 10075 | 1928 | 4831 | 1126 | 1732 | 2407 | 2190 | 1175 | 6497 | 5778 | 5496 | 7393 |
| 91 | 40 | 80 | 70 | 20 | 60 | 70 | 7 | 0 | 6 | 800 | 41 | 90 | 930 | 97 | 84 | 77 | 17 | 1 | 5 | 2 | 17 |
| 19 | 72521 | 32397 | 14546 | 16206 | 11471 | 21353 | 89379 | 57299 | 32877 | 24474 | 8078 | 1545 | 3871 | 9029 | 1388 | 1929 | 1755 | 9414 | 5204 | 4629 | 6362 |
| 92 | 00 | 90 | 10 | 50 | 70 | 20 | 7 | 4 | 0 | 0 | 430 | 54 | 81 | 17 | 41 | 02 | 08 | 5 | 9 | 1 |  |


| ye <br> ar | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19 | 69841 | 62195 | 27483 | 12098 | 13171 | 91369 | 16765 | 69565 | 44385 | 25406 | 1889 | 6231 | 1191 | 2985 | 6961 | 1070 | 1487 | 1352 | 7257 | 4012 | 5261 |
| 93 | 60 | 00 | 00 | 60 | 60 | 7 | 10 | 3 | 6 | 1 | 02 | 650 | 86 | 31 | 21 | 36 | 09 | 97 | 4 | 3 | 72 |
| 19 | 67925 | 59825 | 52505 | 22600 | 96473 | 10224 | 69580 | 12619 | 52033 | 33093 | 1891 | 1405 | 4633 | 8859 | 2218 | 5173 | 7954 | 1105 | 1005 | 5393 | 4208 |
| 94 | 20 | 50 | 60 | 10 | 4 | 10 | 4 | 50 | 1 | 1 | 24 | 09 | 380 | 9 | 91 | 73 | 8 | 16 | 46 | 3 | 31 |
| 19 | 41422 | 58170 | 50457 | 43081 | 17954 | 74510 | 77397 | 52035 | 93749 | 38525 | 2446 | 1396 | 1037 | 3419 | 6538 | 1637 | 3817 | 5869 | 8154 | 7419 | 3503 |
| 95 | 80 | 40 | 80 | 60 | 50 | 0 | 7 | 0 | 7 | 5 | 17 | 82 | 32 | 870 | 6 | 44 | 77 | 8 | 7 | 0 | 05 |
| 19 | 21707 | 35385 | 48570 | 40430 | 32891 | 13141 | 52923 | 53980 | 35933 | 64415 | 2640 | 1674 | 9555 | 7094 | 2338 | 4470 | 1119 | 2609 | 4012 | 5574 | 2901 |
| 96 | 40 | 60 | 30 | 30 | 80 | 10 | 7 | 8 | 3 | 0 | 52 | 53 | 9 | 2 | 390 | 4 | 42 | 87 | 5 | 4 | 66 |
| 19 | 14538 | 18561 | 29657 | 39265 | 31330 | 24563 | 95593 | 37888 | 38310 | 25390 | 4541 | 1859 | 1178 | 6724 | 4991 | 1645 | 3144 | 7874 | 1835 | 2822 | 2433 |
| 97 | 10 | 00 | 40 | 30 | 60 | 40 | 5 | 4 | 9 | 2 | 65 | 72 | 73 | 6 | 4 | 110 | 8 | 6 | 88 | 5 | 14 |
| 19 | 24019 | 12397 | 15387 | 23367 | 29146 | 22075 | 16678 | 63460 | 24846 | 24967 | 1649 | 2946 | 1205 | 7637 | 4356 | 3233 | 1065 | 2036 | 5099 | 1188 | 1758 |
| 98 | 00 | 30 | 70 | 80 | 60 | 30 | 20 | 6 | 3 | 8 | 66 | 34 | 54 | 8 | 3 | 1 | 490 | 7 | 7 | 91 | 41 |
| 19 | 26655 | 20543 | 10402 | 12472 | 18189 | 21898 | 16173 | 12033 | 45409 | 17703 | 1775 | 1171 | 2091 | 8556 | 5420 | 3091 | 2294 | 7559 | 1445 | 3618 | 2091 |
| 99 | 10 | 30 | 10 | 90 | 30 | 10 | 50 | 80 | 2 | 9 | 35 | 80 | 76 | 3 | 1 | 1 | 0 | 75 | 0 | 1 | 01 |
| 20 | 20168 | 22796 | 17233 | 84274 | 97005 | 13650 | 16021 | 11652 | 85973 | 32304 | 1256 | 1259 | 8305 | 1482 | 6062 | 3839 | 2189 | 1624 | 5354 | 1023 | 1737 |
| 00 | 40 | 70 | 50 | 5 | 9 | 10 | 90 | 30 | 4 | 2 | 82 | 03 | 7 | 21 | 0 | 7 | 7 | 9 | 80 | 5 | 33 |
| 20 | 12509 | 17277 | 19249 | 14179 | 67263 | 75393 | 10410 | 12080 | 87306 | 64213 | 2409 | 9365 | 9378 | 6185 | 1103 | 4513 | 2858 | 1630 | 1209 | 3986 | 1369 |
| 01 | 500 | 00 | 80 | 60 | 3 | 7 | 60 | 00 | 9 | 5 | 03 | 3 | 1 | 3 | 68 | 6 | 8 | 2 | 8 | 60 | 60 |
| 20 | 17902 | 10704 | 14523 | 15671 | 11117 | 51033 | 55882 | 76076 | 87596 | 63062 | 4629 | 1735 | 6741 | 6749 | 4450 | 7941 | 3247 | 2056 | 1172 | 8703 | 3853 |
| 02 | 80 | 200 | 20 | 10 | 60 | 3 | 7 | 7 | 9 | 6 | 26 | 05 | 9 | 3 | 8 | 2 | 4 | 8 | 9 |  | 34 |
| 20 | 10294 | 15326 | 90156 | 11878 | 12382 | 85237 | 38298 | 41395 | 55956 | 64198 | 4613 | 3383 | 1267 | 4924 | 4929 | 3250 | 5799 | 2371 | 1501 | 8564 | 2877 |
| 03 | 00 | 50 | 90 | 00 | 40 | 1 | 4 | 2 | 3 | 8 | 62 | 77 | 67 | 6 | 4 | 4 | 1 | 4 | 9 |  | 23 |
| 20 | 16799 | 88138 | 12916 | 73832 | 94059 | 95217 | 64193 | 28479 | 30570 | 41179 | 4716 | 3386 | 2482 | 9298 | 3611 | 3615 | 2383 | 4252 | 1739 | 1101 | 2172 |
| 04 | 60 | 7 | 10 | 60 | 7 | 4 | 0 | 3 | 5 | 7 | 42 | 54 | 71 | 9 | 9 | 1 | 7 | 7 | 0 | 3 | 63 |


| ye ar | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 20 | 12701 | 14400 | 74616 | 10691 | 59528 | 74111 | 73807 | 49266 | 21739 | 23271 | 3130 | 3583 | 2571 | 1885 | 7060 | 2742 | 2744 | 1809 | 3228 | 1320 | 1732 |
| 05 | 20 | 30 | 4 | 70 | 80 | 7 | 6 | 7 | 4 | 8 | 57 | 12 | 91 | 15 | 0 | 1 | 5 | 6 | 3 | 1 | 86 |
| 20 | 10928 | 10883 | 12173 | 61561 | 85725 | 46552 | 56937 | 56095 | 37224 | 16376 | 1750 | 2353 | 2692 | 1932 | 1416 | 5302 | 2059 | 2061 | 1359 | 2424 | 1400 |
| 06 | 90 | 50 | 90 | 3 | 4 | 40 | 6 | 3 | 6 | 9 | 56 | 18 | 35 | 15 | 06 | 9 | 6 | 3 | 1 | 6 | 57 |
| 20 | 20385 | 93708 | 92248 | 10105 | 49870 | 67979 | 36360 | 44063 | 43194 | 28591 | 1256 | 1342 | 1803 | 2063 | 1480 | 1084 | 4062 | 1577 | 1579 | 1041 | 1258 |
| 07 | 30 | 4 | 1 | 90 | 9 | 7 | 50 | 7 | 7 | 0 | 28 | 03 | 44 | 03 | 38 | 90 | 6 | 8 | 1 | 2 | 67 |
| 20 | 46835 | 17491 | 79654 | 77096 | 82798 | 40156 | 54067 | 28702 | 34642 | 33888 | 2240 | 9841 | 1051 | 1412 | 1615 | 1159 | 8494 | 3180 | 1235 | 1236 | 1066 |
| 08 | 80 | 60 | 9 | 5 | 9 | 6 | 3 | 90 | 3 | 9 | 85 | 3 | 03 | 19 | 34 | 08 | 1 | 7 | 3 | 3 | 92 |
| 20 | 12195 | 40158 | 14825 | 66122 | 62453 | 65655 | 31363 | 41839 | 22100 | 26605 | 2599 | 1717 | 7541 | 8053 | 1081 | 1237 | 8879 | 6506 | 2436 | 9463 | 9119 |
| 09 | 30 | 70 | 70 | 5 | 9 | 9 | 2 | 8 | 50 | 8 | 46 | 79 | 7 | 0 | 92 | 49 | 2 | 8 | 5 |  | 8 |
| 20 | 92733 | 10440 | 33826 | 12127 | 52259 | 47898 | 49291 | 23242 | 30788 | 16204 | 1947 | 1900 | 1255 | 5511 | 5884 | 7905 | 9041 | 6487 | 4753 | 1780 | 7353 |
| 10 | 2 | 50 | 20 | 10 | 3 | 8 | 7 | 4 | 3 | 80 | 38 | 98 | 65 | 5 | 3 | 0 | 2 | 0 | 7 | 0 | 6 |
| 20 | 36639 | 79291 | 87501 | 27343 | 93963 | 39019 | 34835 | 35279 | 16491 | 21749 | 1142 | 1371 | 1337 | 8833 | 3876 | 4138 | 5559 | 6358 | 4562 | 3343 | 6422 |
| 11 | 9 | 0 | 1 | 50 | 7 | 5 | 7 | 6 | 2 | 1 | 230 | 18 | 77 | 8 | 8 | 7 | 5 | 4 | 0 | 0 | 9 |
| 20 | 22068 | 31316 | 66348 | 70470 | 21055 | 69571 | 28096 | 24662 | 24747 | 11514 | 1515 | 7947 | 9535 | 9300 | 6140 | 2694 | 2876 | 3863 | 4418 | 3170 | 6786 |
| 12 | 70 | 5 | 6 | 0 | 20 | 4 | 9 | 7 | 9 | 3 | 05 | 74 | 2 | 0 | 0 | 3 | 2 | 4 | 4 | 1 | 0 |
| 20 | 10053 | 18865 | 26223 | 53526 | 54420 | 15650 | 50328 | 19992 | 17392 | 17373 | 8064 | 1060 | 5557 | 6665 | 6499 | 4291 | 1882 | 2009 | 2699 | 3087 | 6956 |
| 13 | 90 | 80 | 9 | 6 | 1 | 20 | 2 | 6 | 2 | 0 | 9 | 00 | 45 | 5 | 9 | 0 | 8 | 8 | 6 | 4 |  |
| 20 | 31895 | 85885 | 15752 | 21011 | 40863 | 39826 | 11114 | 35096 | 13804 | 11948 | 1190 | 5520 | 7250 | 3800 | 4557 | 4443 | 2933 | 1286 | 1373 | 1845 | 6864 |
| 14 | 80 | 1 | 00 | 5 | 3 | 5 | 80 | 6 | 2 | 4 | 56 | 1 | 6 | 15 | 0 | 3 | 1 | 9 | 7 | 1 | 6 |
| 20 | 20883 | 27257 | 71821 | 12667 | 16139 | 30153 | 28564 | 78347 | 24507 | 95933 | 8284 | 8244 | 3820 | 5016 | 2628 | 3152 | 3073 | 2028 | 8900 | 9500 | 6023 |
| 15 | 80 | 50 | 2 | 20 | 1 | 9 | 0 | 6 | 3 |  | 1 | 8 | 4 | 6 | 78 | 0 | 1 | 5 |  |  | 3 |
| 20 | 22237 | 17873 | 22930 | 58572 | 99613 | 12294 | 22457 | 20982 | 57124 | 17801 | 6955 | 6000 | 5969 | 2765 | 3630 | 1902 | 2280 | 2223 | 1467 | 6440 | 5045 |
| 16 | 30 | 20 | 10 | 8 | 5 | 2 | 2 | 6 | 4 | 1 | 2 | 4 | 1 | 2 | 5 | 29 | 8 | 7 | 8 |  |  |


| $\begin{aligned} & \text { ye } \\ & \text { ar } \end{aligned}$ | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 20 | 26929 | 19027 | 15022 | 18661 | 45901 | 75527 | 91055 | 16396 | 15201 | 41223 | 1282 | 5004 | 4315 | 4291 | 1988 | 2609 | 1367 | 1639 | 1598 | 1055 | 4089 |
| 17 | 90 | 50 | 50 | 70 | 7 | 5 |  | 7 | 7 | 4 | 11 | 6 | 5 | 9 | 0 | 8 | 37 | 4 | 3 | 0 | 1 |
| 20 | 20562 | 23066 | 16060 | 12348 | 14870 | 35597 | 57454 | 68458 | 12249 | 11319 | 3064 | 9524 | 3716 | 3203 | 3185 | 1475 | 1937 | 1014 | 1216 | 1186 | 3817 |
| 18 | 30 | 80 | 40 | 50 | 70 | 2 | 0 |  | 2 | 7 | 75 | 3 | 2 | 8 | 9 | 6 | 0 | 87 | 7 | 2 | 8 |
| 20 | 11880 | 17598 | 19406 | 13100 | 97144 | 11333 | 26527 | 42232 | 49948 | 89036 | 8212 | 2221 | 6900 | 2691 | 2320 | 2307 | 1068 | 1402 | 7348 | 8810 | 3623 |
| 19 | 70 | 40 | 70 | 80 | 1 | 80 | 2 | 8 |  |  | 7 | 49 | 5 | 8 | 3 | 1 | 5 | 6 | 6 |  | 2 |
| 20 | 11975 | 10157 | 14742 | 15670 | 10132 | 72356 | 82190 | 18926 | 29866 | 35164 | 6254 | 5762 | 1557 | 4837 | 1886 | 1626 | 1617 | 7488 | 9830 | 5149 | 3156 |
| 20 | 70 | 40 | 40 | 60 | 50 | 4 | 3 | 4 | 4 |  | 4 | 8 | 93 | 9 | 9 | 3 | 0 |  |  | 9 | 5 |
| 20 | 10010 | 10267 | 86051 | 12223 | 12670 | 80139 | 56339 | 63390 | 14521 | 22855 | 2687 | 4777 | 4400 | 1189 | 3692 | 1440 | 1241 | 1234 | 5715 | 7502 | 6339 |
| 21 | 50 | 20 | 7 | 90 | 20 | 8 | 7 | 6 | 9 | 9 | 6 | 1 | 2 | 34 | 9 | 2 | 3 | 2 |  |  | 6 |
|  | 10315 | 85819 | 86963 | 71317 | 98754 | 10010 | 62319 | 43391 | 48566 | 11096 | 1744 | 2049 | 3641 | 3353 | 9064 | 2814 | 1097 | 9459 | 9405 | 4355 | 5402 |
|  | 00 | 5 | 7 | 0 | 7 | 10 | 6 | 4 | 5 | 4 | 17 | 6 | 9 | 9 | 6 | 4 | 6 |  |  |  | 6 |

Table 7.3.1.2. Western horse mackerel. Final assessment. Fishing mortality-at-age

| year | $\mathbf{0}$ | $\mathbf{1}$ | $\mathbf{2}$ | $\mathbf{3}$ | $\mathbf{4}$ | $\mathbf{5}$ | $\mathbf{6}$ | $\mathbf{7}$ | $\mathbf{8}$ | $\mathbf{9}$ | $\mathbf{1 0}$ | $\mathbf{1 1}$ | $\mathbf{1 2}$ | $\mathbf{1 3}$ | $\mathbf{1 4}$ | $\mathbf{1 5}$ | $\mathbf{1 6}$ | $\mathbf{1 7}$ | $\mathbf{1 8}$ | $\mathbf{1 9}$ | $\mathbf{2 0}$ |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 1982 | 0.001 | 0.003 | 0.008 | 0.014 | 0.019 | 0.022 | 0.024 | 0.025 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 |
| 1983 | 0.001 | 0.005 | 0.011 | 0.019 | 0.025 | 0.030 | 0.033 | 0.034 | 0.035 | 0.035 | 0.036 | 0.036 | 0.036 | 0.036 | 0.036 | 0.036 | 0.036 | 0.036 | 0.036 | 0.036 | 0.036 |
| 1984 | 0.001 | 0.004 | 0.010 | 0.017 | 0.023 | 0.027 | 0.029 | 0.031 | 0.031 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 |
| 1985 | 0.001 | 0.003 | 0.008 | 0.014 | 0.018 | 0.022 | 0.024 | 0.025 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 | 0.026 |
| 1986 | 0.001 | 0.004 | 0.010 | 0.017 | 0.022 | 0.026 | 0.029 | 0.030 | 0.031 | 0.031 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 | 0.032 |
| 1987 | 0.001 | 0.005 | 0.012 | 0.021 | 0.028 | 0.033 | 0.036 | 0.038 | 0.039 | 0.039 | 0.040 | 0.040 | 0.040 | 0.040 | 0.040 | 0.040 | 0.040 | 0.040 | 0.040 | 0.040 | 0.040 |
| 1988 | 0.001 | 0.006 | 0.014 | 0.023 | 0.031 | 0.037 | 0.040 | 0.042 | 0.043 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 | 0.044 |


| year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989 | 0.001 | 0.006 | 0.014 | 0.024 | 0.032 | 0.038 | 0.041 | 0.043 | 0.044 | 0.045 | 0.045 | 0.045 | 0.045 | 0.045 | 0.045 | 0.045 | 0.045 | 0.045 | 0.045 | 0.045 | 0.045 |
| 1990 | 0.002 | 0.008 | 0.020 | 0.033 | 0.044 | 0.053 | 0.058 | 0.060 | 0.062 | 0.063 | 0.063 | 0.063 | 0.063 | 0.063 | 0.063 | 0.063 | 0.063 | 0.063 | 0.063 | 0.063 | 0.063 |
| 1991 | 0.002 | 0.009 | 0.022 | 0.037 | 0.050 | 0.060 | 0.066 | 0.069 | 0.070 | 0.071 | 0.071 | 0.072 | 0.072 | 0.072 | 0.072 | 0.072 | 0.072 | 0.072 | 0.072 | 0.072 | 0.072 |
| 1992 | 0.004 | 0.015 | 0.034 | 0.057 | 0.078 | 0.092 | 0.101 | 0.105 | 0.108 | 0.109 | 0.110 | 0.110 | 0.110 | 0.110 | 0.110 | 0.110 | 0.110 | 0.110 | 0.110 | 0.110 | 0.110 |
| 1993 | 0.005 | 0.019 | 0.046 | 0.076 | 0.103 | 0.122 | 0.134 | 0.140 | 0.144 | 0.145 | 0.146 | 0.146 | 0.147 | 0.147 | 0.147 | 0.147 | 0.147 | 0.147 | 0.147 | 0.147 | 0.147 |
| 1994 | 0.005 | 0.020 | 0.048 | 0.080 | 0.108 | 0.128 | 0.141 | 0.147 | 0.151 | 0.152 | 0.153 | 0.153 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 | 0.154 |
| 1995 | 0.008 | 0.030 | 0.072 | 0.120 | 0.162 | 0.192 | 0.210 | 0.220 | 0.225 | 0.228 | 0.229 | 0.230 | 0.230 | 0.230 | 0.230 | 0.230 | 0.230 | 0.230 | 0.230 | 0.230 | 0.230 |
| 1996 | 0.007 | 0.027 | 0.063 | 0.105 | 0.142 | 0.168 | 0.184 | 0.193 | 0.197 | 0.199 | 0.201 | 0.201 | 0.201 | 0.202 | 0.202 | 0.202 | 0.202 | 0.202 | 0.202 | 0.202 | 0.202 |
| 1997 | 0.009 | 0.037 | 0.088 | 0.148 | 0.200 | 0.237 | 0.260 | 0.272 | 0.278 | 0.281 | 0.283 | 0.284 | 0.284 | 0.284 | 0.284 | 0.284 | 0.284 | 0.284 | 0.284 | 0.285 | 0.285 |
| 1998 | 0.006 | 0.025 | 0.060 | 0.101 | 0.136 | 0.161 | 0.176 | 0.185 | 0.189 | 0.191 | 0.192 | 0.193 | 0.193 | 0.193 | 0.193 | 0.193 | 0.193 | 0.193 | 0.193 | 0.193 | 0.193 |
| 1999 | 0.006 | 0.026 | 0.061 | 0.101 | 0.137 | 0.162 | 0.178 | 0.186 | 0.191 | 0.193 | 0.194 | 0.194 | 0.194 | 0.195 | 0.195 | 0.195 | 0.195 | 0.195 | 0.195 | 0.195 | 0.195 |
| 2000 | 0.005 | 0.019 | 0.045 | 0.075 | 0.102 | 0.121 | 0.132 | 0.139 | 0.142 | 0.143 | 0.144 | 0.145 | 0.145 | 0.145 | 0.145 | 0.145 | 0.145 | 0.145 | 0.145 | 0.145 | 0.145 |
| 2001 | 0.006 | 0.024 | 0.056 | 0.093 | 0.126 | 0.149 | 0.164 | 0.171 | 0.175 | 0.177 | 0.178 | 0.179 | 0.179 | 0.179 | 0.179 | 0.179 | 0.179 | 0.179 | 0.179 | 0.179 | 0.179 |
| 2002 | 0.005 | 0.022 | 0.051 | 0.086 | 0.116 | 0.137 | 0.150 | 0.157 | 0.161 | 0.163 | 0.163 | 0.164 | 0.164 | 0.164 | 0.164 | 0.164 | 0.164 | 0.164 | 0.164 | 0.164 | 0.164 |
| 2003 | 0.005 | 0.021 | 0.050 | 0.083 | 0.113 | 0.134 | 0.146 | 0.153 | 0.157 | 0.158 | 0.159 | 0.160 | 0.160 | 0.160 | 0.160 | 0.160 | 0.160 | 0.160 | 0.160 | 0.160 | 0.160 |
| 2004 | 0.004 | 0.017 | 0.039 | 0.065 | 0.088 | 0.105 | 0.115 | 0.120 | 0.123 | 0.124 | 0.125 | 0.125 | 0.125 | 0.125 | 0.126 | 0.126 | 0.126 | 0.126 | 0.126 | 0.126 | 0.126 |
| 2005 | 0.004 | 0.018 | 0.042 | 0.071 | 0.096 | 0.114 | 0.124 | 0.130 | 0.133 | 0.135 | 0.135 | 0.136 | 0.136 | 0.136 | 0.136 | 0.136 | 0.136 | 0.136 | 0.136 | 0.136 | 0.136 |
| 2006 | 0.004 | 0.015 | 0.036 | 0.061 | 0.082 | 0.097 | 0.106 | 0.111 | 0.114 | 0.115 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 |


| year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2007 | 0.003 | 0.012 | 0.029 | 0.049 | 0.067 | 0.079 | 0.086 | 0.091 | 0.093 | 0.094 | 0.094 | 0.094 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 | 0.095 |
| 2008 | 0.004 | 0.015 | 0.036 | 0.061 | 0.082 | 0.097 | 0.106 | 0.111 | 0.114 | 0.115 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | 0.116 | 0.117 | 0.117 | 0.117 | 0.117 | 0.117 |
| 2009 | 0.005 | 0.022 | 0.051 | 0.085 | 0.115 | 0.137 | 0.150 | 0.157 | 0.160 | 0.162 | 0.163 | 0.163 | 0.164 | 0.164 | 0.164 | 0.164 | 0.164 | 0.164 | 0.164 | 0.164 | 0.164 |
| 2010 | 0.007 | 0.027 | 0.063 | 0.105 | 0.142 | 0.168 | 0.184 | 0.193 | 0.198 | 0.200 | 0.201 | 0.201 | 0.202 | 0.202 | 0.202 | 0.202 | 0.202 | 0.202 | 0.202 | 0.202 | 0.202 |
| 2011 | 0.007 | 0.028 | 0.066 | 0.111 | 0.151 | 0.178 | 0.195 | 0.205 | 0.209 | 0.212 | 0.213 | 0.213 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 | 0.214 |
| 2012 | 0.007 | 0.027 | 0.065 | 0.108 | 0.147 | 0.174 | 0.190 | 0.199 | 0.204 | 0.206 | 0.207 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 | 0.208 | 0.209 |
| 2013 | 0.008 | 0.030 | 0.072 | 0.120 | 0.162 | 0.192 | 0.210 | 0.220 | 0.225 | 0.228 | 0.229 | 0.230 | 0.230 | 0.230 | 0.230 | 0.230 | 0.231 | 0.231 | 0.231 | 0.231 | 0.231 |
| 2014 | 0.007 | 0.029 | 0.068 | 0.114 | 0.154 | 0.182 | 0.200 | 0.209 | 0.214 | 0.216 | 0.217 | 0.218 | 0.218 | 0.219 | 0.219 | 0.219 | 0.219 | 0.219 | 0.219 | 0.219 | 0.219 |
| 2015 | 0.006 | 0.023 | 0.054 | 0.090 | 0.122 | 0.145 | 0.158 | 0.166 | 0.170 | 0.172 | 0.173 | 0.173 | 0.173 | 0.173 | 0.173 | 0.174 | 0.174 | 0.174 | 0.174 | 0.174 | 0.174 |
| 2016 | 0.006 | 0.024 | 0.056 | 0.094 | 0.127 | 0.150 | 0.165 | 0.172 | 0.176 | 0.178 | 0.179 | 0.180 | 0.180 | 0.180 | 0.180 | 0.180 | 0.180 | 0.180 | 0.180 | 0.180 | 0.180 |
| 2017 | 0.005 | 0.020 | 0.046 | 0.077 | 0.104 | 0.124 | 0.135 | 0.142 | 0.145 | 0.146 | 0.147 | 0.148 | 0.148 | 0.148 | 0.148 | 0.148 | 0.148 | 0.148 | 0.148 | 0.148 | 0.148 |
| 2018 | 0.006 | 0.023 | 0.054 | 0.090 | 0.122 | 0.144 | 0.158 | 0.165 | 0.169 | 0.171 | 0.172 | 0.172 | 0.173 | 0.173 | 0.173 | 0.173 | 0.173 | 0.173 | 0.173 | 0.173 | 0.173 |
| 2019 | 0.007 | 0.027 | 0.064 | 0.107 | 0.145 | 0.171 | 0.188 | 0.196 | 0.201 | 0.203 | 0.204 | 0.205 | 0.205 | 0.205 | 0.205 | 0.205 | 0.205 | 0.206 | 0.206 | 0.206 | 0.206 |
| 2020 | 0.004 | 0.016 | 0.037 | 0.063 | 0.085 | 0.100 | 0.110 | 0.115 | 0.118 | 0.119 | 0.119 | 0.120 | 0.120 | 0.120 | 0.120 | 0.120 | 0.120 | 0.120 | 0.120 | 0.120 | 0.120 |
| 2021 | 0.004 | 0.016 | 0.038 | 0.063 | 0.086 | 0.101 | 0.111 | 0.116 | 0.119 | 0.120 | 0.121 | 0.121 | 0.122 | 0.122 | 0.122 | 0.122 | 0.122 | 0.122 | 0.122 | 0.122 | 0.122 |
| 2022 | 0.003 | 0.013 | 0.032 | 0.053 | 0.072 | 0.085 | 0.093 | 0.098 | 0.100 | 0.101 | 0.101 | 0.102 | 0.102 | 0.102 | 0.102 | 0.102 | 0.102 | 0.102 | 0.102 | 0.102 | 0.102 |

Table 7.3.1.3. Western horse mackerel. Final assessment. Stock summary table.

| Year | Recruit (thousands) | Total <br> Biomass | Spawning <br> Biomass | Catch | Yield/SSB | Fbar(1-3) | Fbar(4-8) | Fbar(1-10) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 48926600 | 2886380 | 2214000 | 20000 | 0.0260 | 0.008 | 0.023 | 0.019 |
| 1983 | 1232840 | 3417510 | 2355730 | 61197 | 0.0365 | 0.011 | 0.031 | 0.026 |
| 1984 | 1374200 | 4083800 | 2493800 | 90442 | 0.0370 | 0.010 | 0.028 | 0.024 |
| 1985 | 1981030 | 4690830 | 2953590 | 96244 | 0.0316 | 0.008 | 0.023 | 0.019 |
| 1986 | 2548770 | 5121830 | 4243370 | 96343 | 0.0319 | 0.010 | 0.028 | 0.023 |
| 1987 | 5018780 | 5318200 | 4987120 | 137499 | 0.0372 | 0.013 | 0.035 | 0.029 |
| 1988 | 2229280 | 5301870 | 5027500 | 187338 | 0.0416 | 0.014 | 0.039 | 0.032 |
| 1989 | 2624600 | 5136480 | 4820510 | 210989 | 0.0431 | 0.015 | 0.040 | 0.033 |
| 1990 | 1986280 | 4881420 | 4554180 | 209583 | 0.0600 | 0.020 | 0.055 | 0.046 |
| 1991 | 3772940 | 4509120 | 4228340 | 275968 | 0.0671 | 0.023 | 0.063 | 0.053 |
| 1992 | 7252100 | 4129240 | 3847640 | 287438 | 0.1003 | 0.035 | 0.097 | 0.081 |
| 1993 | 6984160 | 3706390 | 3344520 | 393631 | 0.1316 | 0.047 | 0.129 | 0.108 |
| 1994 | 6792520 | 3320740 | 2817850 | 453246 | 0.1407 | 0.049 | 0.135 | 0.113 |
| 1995 | 4142280 | 3061080 | 2448980 | 412291 | 0.2098 | 0.074 | 0.202 | 0.169 |
| 1996 | 2170740 | 2720820 | 2133210 | 538950 | 0.1870 | 0.065 | 0.177 | 0.148 |
| 1997 | 1453810 | 2485770 | 2009990 | 422396 | 0.2510 | 0.091 | 0.249 | 0.208 |
| 1998 | 2401900 | 2105910 | 1786560 | 534673 | 0.1724 | 0.062 | 0.169 | 0.142 |
| 1999 | 2665510 | 1895400 | 1670020 | 325340 | 0.1707 | 0.063 | 0.171 | 0.143 |
| 2000 | 2016840 | 1690470 | 1487800 | 298992 | 0.1302 | 0.047 | 0.127 | 0.106 |
| 2001 | 12509500 | 1612350 | 1350840 | 202732 | 0.15 | 0.058 | 0.157 | 0.131 |
| 2002 | 1790280 | 1580000 | 1201490 | 229081 | 0.19 | 0.053 | 0.144 | 0.121 |
| 2003 | 1029400 | 1616600 | 1117590 | 196120 | 0.18 | 0.051 | 0.140 | 0.117 |
| 2004 | 1679960 | 1642320 | 1139280 | 191856 | 0.17 | 0.040 | 0.110 | 0.092 |
| 2005 | 1270120 | 1653470 | 1358210 | 159742 | 0.12 | 0.044 | 0.120 | 0.100 |
| 2006 | 1092890 | 1590610 | 1428260 | 182001 | 0.13 | 0.037 | 0.102 | 0.085 |
| 2007 | 2038530 | 1511380 | 1371930 | 155827 | 0.11 | 0.030 | 0.083 | 0.069 |
| 2008 | 4683580 | 1452440 | 1300820 | 123356 | 0.09 | 0.037 | 0.102 | 0.085 |
| 2009 | 1219530 | 1384180 | 1189100 | 143349 | 0.12 | 0.053 | 0.144 | 0.120 |


| Year | Recruit |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| (thousands) |  |

Table 7.4.1. Western Horse Mackerel. Short term prediction: INPUT DATA. *geometric mean of the recruitment time series from 1983 to 2022. ** from assessment output

| Age | N | Mat | M | PF | PM | Stock weight at age** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 2088520* | 0.000 | 0.150 | 0 | 0 | 0.0041 |
| 1 | 884863 | 0.000 | 0.150 | 0 | 0 | 0.0180 |
| 2 | 728781 | 0.050 | 0.150 | 0 | 0 | 0.0421 |
| 3 | 725136 | 0.250 | 0.150 | 0 | 0 | 0.0730 |
| 4 | 582070 | 0.700 | 0.150 | 0 | 0 | 0.1069 |
| 5 | 791065 | 0.950 | 0.150 | 0 | 0 | 0.1407 |
| 6 | 791264 | 1.000 | 0.150 | 0 | 0 | 0.1726 |
| 7 | 488647 | 1.000 | 0.150 | 0 | 0 | 0.2014 |
| 8 | 338739 | 1.000 | 0.150 | 0 | 0 | 0.2267 |
| 9 | 378294 | 1.000 | 0.150 | 0 | 0 | 0.2484 |
| 10 | 86337 | 1.000 | 0.150 | 0 | 0 | 0.2668 |
| 11 | 135634 | 1.000 | 0.150 | 0 | 0 | 0.2821 |
| 12 | 15934 | 1.000 | 0.150 | 0 | 0 | 0.2948 |


| Age | $\mathbf{N}$ | Mat | $\mathbf{M}$ | PF | PM | Stock weight at age** |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 13 | 28309 | 1.000 | 0.150 | 0 | 0 | 0.3053 |
| 14 | 26068 | 1.000 | 0.150 | 0 | 0 | 0.3138 |
| 15 | 70451 | 1.000 | 0.150 | 0 | 0 | 0.3208 |
| 16 | 21873 | 1.000 | 0.150 | 0 | 0 | 0.3264 |
| 17 | 7352 | 1.000 | 0.150 | 0 | 0 | 0.3309 |
| 18 | 7309 | 1.000 | 0.150 | 0 | 0 | 0.3346 |
| 19 | 45371 | 0.150 | 0 | 0 | 0.3376 |  |
| 20 |  | 0.150 | 0 | 0 | 0.3419 |  |

Table 7.4.2. Western Horse Mackerel. Short term prediction; single area management option table. Assumption: Catch 2023: $15 \mathbf{2 7 7}$ t (100\% of 2023 TOTAL TAC).
Scenarios $\quad F_{\text {factor }} \quad F_{\text {bar }} \quad$ Catch_2023 Catch_2024 SSB_2024 SSB_2025 Change_SSB_2024-2025(\%) Change_Catch_2023-2024(\%)
$S S B_{2025}=M S Y B_{\text {trigger }}=B_{p a}=B_{\text {lim }} \quad$ The $B_{p a}, B_{\text {lim }}$ and MSY $B_{\text {trigger }}$ options were left blank because $B_{p a}, B_{\text {lim }}$ and MSY $B_{\text {trigger }}$ cannot be achieved in 2025 , even with a zero catch in 2024.

| $\mathrm{F}=\mathrm{F}_{\mathrm{MSY}}$ | 0.99 | 0.074 | 15277 | 73124 | 744547 | 715711 | -3.9 | 378.7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\mathrm{F}=\mathrm{F}_{\mathrm{P} 05}=\mathrm{F}_{\mathrm{pa}}$ | 1.06 | 0.079 | 15277 | 77837 | 744547 | 711377 | -4.5 | 409.5 |
| $\mathrm{F}=\mathrm{F}_{\text {lim }}$ | 1.38 | 0.103 | 15277 | 100070 | 744547 | 690950 | -7.2 | 555.0 |
| $F=0$ | 0.00 | 0.000 | 15277 | 0 | 744547 | 783126 | 5.2 | -100.0 |
| $F=F_{2023}$ | 0.21 | 0.016 | 15277 | 16082 | 744547 | 768273 | 3.2 | 5.3 |
| PelAC proposed HCR | 0.20 | 0.015 | 15277 | 15348 | 744547 | 768951 | 3.3 | 0.5 |

## Table 7.4.3. Western Horse Mackerel. Landings inside and outside the NEAFC Regulatory Area (RA), as estimated by ICES, as well as total landings. Weights are in tonnes.

| Year | Inside the NEAFC RA | Outside the NEAFC RA | Total catches |
| :--- | :--- | :--- | :--- |
| 2020 | 0 | 76422 | 76422 |
| 2021 | 10 | 81547 | 81557 |
| 2022 | 0 | 70114 | 70114 |

### 7.15 Figures



Figure 7.1.1.1: Western horse mackerel. Catch by quarter and year for 2000--2022.

North Sea HOM catches by Category


Figure 7.1.2.1. Western horse mackerel. Catch categories since $\mathbf{2 0 0 0}$ (green bars indicate when countries have submitted catch data without specifying landings/discards).


Figure 7.1.3.1: Western horse mackerel. Catch by ICES Division and year for 1982-2022.


Figure 7.2.1.1The total annual horse mackerel egg production for $1992 \mathbf{- 2 0 2 2}$ for the western horse mackerel stock



Figure 7.2.1.3. Egg production by period for the western horse mackerel stock since 2007


Figure 7.2.2.1: Western horse mackerel. Trend of the fisheries independent indices of abundance used in the assessment of Western Horse mackerel. Top: Spawning index from egg survey; middle: recruitment index from IBTS survey; bottom: biomass estimates from PELACUS acoustic survey. Confidence intervals are shown as well.

2022 Western Stock: cat@ge by division


Figure 7.2.4.1: Western horse mackerel. Catch-at-age (millions) by ICES division in 2022.

Western Stock: cat@ge by Year


Figure 7.2.4.2: Western horse mackerel. Catch-at-age (millions) by Year.


Figure 7.2.4.3: Western horse mackerel. Catch-at-age - the area of bubbles is proportional to the catch number. Age 15 is a plus group.

Weight at age - 1st \& 2nd quarter
$-1-4-7-10-13$
$-2-5-8-11-14$
$-3-6-9-12-15$


Figure 7.2.5.1: Western horse mackerel. Weight at age in the catch (kg) by year.


Figure 7.2.5.2: Western horse mackerel. Weight at length in the stock (kg) as estimated by the stock assessment.


Figure 7.2.6.1: Western horse mackerel. Maturity at age as used in the assessment model.


Figure 7.2.10.1: Western horse mackerel. Length frequency distribution of the landing data as used in the assessment model.


Figure 7.2.10.2: Western horse mackerel. Stacked length frequency distribution of the landing data as used in the assessment model.


Figure 7.2.10.3: Western horse mackerel. Within-cohort consistency in the catch-at-age matrix, shown by plotting the log-catch of a cohort at a particular age against the log-catch of the same cohort at subsequent ages.


Figure 7.2.10.4: Western horse mackerel. Catch numbers at age composition by decade (year specifies start of decade i.e., $1980=1980-1999$, also note that 2020 only includes years 2020-2022).


Figure 7.2.10.5: Western horse mackerel. Data exploration. Correlation plots between indices of abundance (including 2022 data points). Size and shade of circle indicates magnitude of correlation and color indicated sign (blue positive, red negative).


Figure 7.2.11.1: Western horse mackerel. Model fitting. Fitting of the model to the fisheries-independent indices. From top to bottom: IBTS, egg survey, PELACUS. Dots represent observations (with confidence intervals) and blue line the model.


Figure 7.2.11.1 cont.: Western horse mackerel. Model fitting. Fitting of the model to the catch at age matrix from 1982 to 2002. Black joined dots represent observations and green line represents model.


Figure 7.2.11.1 cont.: Western horse mackerel. Model fitting. Fitting of the model to the length composition of the landing data from 2000 to 2022.


Figure 7.2.11.1 cont.: Western horse mackerel. Model fitting. Fitting of the model to the length composition of the acoustic survey.



Age (yr)

Figure 7.2.11.1 cont.: Western horse mackerel. Model fitting. Fitting of the model to the Age length comp of the catch.


Figure 7.2.11.2: Western horse mackerel. Model results. Spawning stock biomass ( 0.5 of the overall SSB only is shown; plot on the top) and recruitment estimates (plot on the bottom) from the assessment model from 1982 to $2023.95 \% \mathrm{Cl}$ are shown. Note this figure is a standard SS output. Whilst the $y$-axis denotes spawning biomass in mt , the axis values reflect the actual (data) values. Therefore, the axis values should be between 0 and 6 to correspond to the axis title.


Figure 7.2.11.2 cont.: Western horse mackerel. Model results. Fishing mortality estimates (Fbar ages 1-10) from the assessment model from 1982 to $\mathbf{2 0 2 2}$. $95 \% \mathrm{Cl}$ are shown.


Figure 7.2.11.3: Western horse mackerel. 5 years of retrospective analysis for SSB, F and Recruitment. Dash lines are the 2023 assessment confidence intervals.


Figure 7.2.11.4: Western horse mackerel. Historical model assessment results.


Figure 7.10.1. Western horse mackerel. Top: comparison of (max) scientific advice, TAC (or sum of unilateral quota) and Total Catch. Bottom: percentage deviation from ICES advice, CoA is Catch over Advice, ToA is TAC over Advice.


Figure 7.13.1. Sensitivity of the model to new data. Spawning biomass and fishing mortality (ages 1-10) as estimated in the model conducted in 2022 (in blue) and in 2023 (in red).

## 8 Mackerel in Northeast Atlantic and adjacent waters

## Scomber scombrus in subareas 1-8 and 14 and division 9.a (mac.27.nea)

### 8.1 ICES advice and international management applicable to 2022

From 2001 to 2007, the internationally agreed TACs covered most of the distribution area of the Northeast Atlantic mackerel. From 2008 to 2014, no agreement was reached among the Coastal States on the sharing of the mackerel quotas. In 2014, three of the Coastal States (European Union, Norway and the Faroe Islands) agreed on a Management Strategy for 2014 to 2018. In November 2018, the 2014 agreement was extended for a further two years until 2020. No agreement on the share of the stock has been reached since 2021. Despite various agreements, the total declared quotas in each of the years 2015 to 2022 all exceeded the TAC advised by ICES. An overview of declared quotas and transfers for 2023, as available to WGWIDE, is given in the text table below. An estimate of the expected quota uptake in 2023 was carried out based on the declared quotas by nation including transfer from previous year, and an estimation of uptake for the remainder of the year. This was based on the fishery up to the end of August 2023 and knowledge of the fishery development from September to December in the last two years. Total removals of mackerel are expected to be approximately 1.15 million tonnes in 2023, exceeding the ICES advice for 2023 by approximately 365341 tonnes ( $47 \%$ ).
The quota figures and transfers in the text table below were based on national regulations as reported to NEAFC and discard estimates.

| Estimation of 2023 catch ( t ) | Declared quotas ${ }^{1}$ | Transfer from previous year ${ }^{1}$ | Sum | Expected uptake ${ }^{2}$ | Justification |
| :---: | :---: | :---: | :---: | :---: | :---: |
| EU | 177,048 |  | 177,048 | 177,048 | Full uptake |
| Faroe Islands | 153,285 | 66,700 | 219,985 | 200,000 | Faroese expected catch ${ }^{3}$ |
| Greenland | 50,834 |  | 50,834 | 34,000 | Greenland expected catch ${ }^{3}$ |
| Iceland | 125,311 | 18,728 | 144,039 | 140,000 | Icelandic expected catch ${ }^{3}$ |
| Norway | 249,870 | 14,119 | 263,989 | 263,989 | Full uptake |
| Russia | 118,000 |  | 118,000 | 118,000 | Full uptake |
| UK | 210,814 |  | 210,814 | 210,814 | Full uptake |
| Discards ${ }^{4}$ | 3,556 |  | 3,556 | 3,556 | Full uptake |
| Total expected catch (incl. discards) ${ }^{5}$ |  |  |  | 1,147,407 | tonnes |

[^5]
### 8.2 The fishery

### 8.2.1 Fleet Composition in 2022

The total fleet can be considered to consist of the following components: freezer trawlers, purse seiners, pelagic trawlers, lines and jigging, and gillnets (see stock annex for detailed description of each component).

### 8.2.2 The fishery in 2022

The 2022 fishery includes catch from 18 countries. Catches per ICES statistical rectangle are shown in Figures 8.2.2.1 to 8.2.2.4. It should be noted that these figures are a combination of official catches and ICES estimates and may not indicate the true location of the catches or represent the location of the entire stock. These data are based on catches reported by all the major catching nations and represents almost the entire ICES estimated catch.
Of the total catch in 2022, Norway accounted for the greatest proportion ( $28 \%$ ) followed by Scotland ( $15 \%$ ), Faroes ( $13 \%$ ), Iceland ( $12 \%$ ) and Russia ( $10 \%$ ) (Table 8.2.2.1). In the absence of sharing arrangements, the fishing parties declared unilateral quotas for 2022.

In 2022, catches in the northern areas (Subareas 1, 2, 5, 14) remained high and amounted to 571 403, a decrease from the high of 663111 tonnes in 2021 (see Table 8.2.2.1). Norwegian catches were over 290 kt and Icelandic, Russian, Scottish and Faroese catches were all over 100 kt . Catches from Division 2.a accounted for $54 \%$ of the total catch in 2022. The wide geographical distribution of the fishery noted in previous years has continued. In 2022 there was only a small amount of catch from southern Iceland with the fishery moving further east and north into the International zone. There were low catches of around 100 tonnes from Subarea 14 and a reduced fishery in Subarea 5.

The time series of catches by country from the North Sea, Skagerrak and Kattegat (Subarea 4, Division 3.a) is given in Table 8.2.2.2. Catches in 2022 increased to over 353 kt from 221340 tonnes in 2021. The majority of the catch is from Subarea 4. Small catches were also reported in Divisions 3.a-d.

Catches in the western area (Subareas 6, 7 and Divisions 8.a, b, d and e) decreased again in 2022 to under 90000 t . This is a decrease of around 75000 t from 2021. The catches are detailed in Table 8.2.2.3.

Table 8.2.2.4 details the catches in the southern areas (Divisions 8.c and 9.a) which are taken almost exclusively by Spain and Portugal. The reported catch in 2022 of 32264 is similar to the 2021 catch of 31928 tonnes.

### 8.2.3 Quarterly Distribution of the catch

The proportion of the catches by quarter (\%) is presented in Table 8.2.3.1 and figure 8.2.3.1.
The quarterly distribution of catch from 2010-2022 is similar to recent years with the northern summer fishery in Q3 accounting for the greatest proportion of the total catch. The average proportion taken in quarter 3 from 2010-2022 is $48 \%$. In 2022 this proportion increased to $58 \%$ and is higher than the quarter 1 and quarter 4 catches which when combined account for $36 \%$ of the total. The proportion of the catch taken in quarter 2 has remained stable.

- First quarter 2022 (188 172 tonnes - 18\%)

The distribution of catches in the first quarter is shown in Figure 8.2.2.1. The proportion of the fishery taken in quarter 1 has decreased in 2022 with the Scottish and Irish pelagic fleets targeting mackerel in Divisions 6.a, 7.b and 7.j. Substantial catches are also taken by the Dutch owned freezer trawler fleet. The largest catches were taken in Division 6.a, as in recent years. The Spanish fisheries also take significant catches along the north coast of Spain during the first quarter.

- $\quad$ Second quarter 2022 (57 159 tonnes - 6\%)

The distribution of catches in the second quarter is shown in Figure 8.2.2.2. The quarter 2 fishery is traditionally the smallest and this was also the case in 2022. Quarter 2 is the start of the summer fishery in northern waters by Icelandic, Norwegian and Russian fleets in Division 2.a. There are also significant catches taken in Division 8.c.

- Third quarter 2022 (610 648 tonnes - 58\%)

Figure 8.2.2.3 shows the distribution of the quarter 3 catches. Large catches were taken throughout Division 2.a, with high concentrations in international waters. Fishing was carried out mainly by vessels from Russia, Norway, Iceland, the Faroes and Greenland. There were also catches from Division 4.a but very little from Division 5.a.

- Fourth quarter 2022 (190 741 tonnes - 18\%)

The fourth quarter distribution of catches is shown in Figure 8.2.2.4. The proportion of the catch taken in the fourth quarter has decreased to $18 \%$ 2022. The summer fishery in northern waters has largely finished with some catches reported from Division 2.a. The largest catches in quarter 4 are taken by Scotland around the Shetland Isles in Division 4.a.

### 8.3 Quality and adequacy of sampling data from commercial fishery

The sampling of the commercial catch of Northeast Atlantic mackerel in table 8.3.1. Overall sampling effort in 2022 was higher than 2021 with $88 \%$ of the catch sampled. Nations with large, directed fisheries are capable of sampling $100 \%$ of their catch which may conceal deficiencies in sampling elsewhere.

The 2022 sampling levels by country are shown in table 8.3.2. The majority of countries achieved a high level of sampling coverage. Belgian catches consist of by-catch in the demersal fisheries in the North Sea. France supplied a quantity of length-frequency data to the working group which can be utilised to characterise the selection of the fleet but requires an allocation of catch at age proportions from another sampled fleet in order to raise the data for use in the assessment. Belgium, Denmark, Sweden, Greenland, Northern Ireland and Poland did not supply sampling information in 2022. Portugal sampled landings from 9.a only. England sampled landings from the handline fleet operating off the Cornish coast as well as from freezer trawlers. Cooperation between the Dutch and German sampling programmes is designed to provide complete coverage for the freezer trawlers operating under these national flags and also those of England and France.

Catch sampling levels per ICES Division (for those with a WG catch of $>100 \mathrm{t}$ ) are shown in table 8.3.3. $94 \%$ of the total catch is taken in divisions $2 . a, 4 . a$ and $6 . a$ and these areas are well sampled. In general, areas with insufficient sampling have relatively low levels of catch.

### 8.4 Catch data

### 8.4.1 ICES Catch Estimates

## Total Catch 2022

The total ICES estimated catch for 2022 was 1046720 tonnes which represents a decrease from 1 081540 tonnes in 2021.

The combined 2022 estimated catch uptake, arising from agreements and autonomous quotas, and transfers amounted to 1131416 tonnes. The ICES catch estimate ( 1046720 tonnes) represents an undershoot of this and is still above the ICES advice of 794920 tonnes for 2022. The combined fishable TAC for 2023, as best ascertained by the Working Group (see Section 8.1), amounts to 1147407 tonnes.

Catches reported for 2022 and in previous Working Group reports are considered to be best estimates. In most cases, catch information comes from official logbook records. Other sources of information include catch processors. Some countries provide information on discards and slipped catch from observer programs and compliance reports. In several countries discarding is illegal. Spanish data is based on the official data supplied by the Fisheries General Secretary (SGP) but supplemented by scientific estimates which are recorded as unallocated catch in the ICES estimates. A detailed basis for the ICES catch estimates is presented in the stock annex.

The total catch as estimated by ICES is shown in Table 8.4.1.1. It is broken down by ICES area group and illustrates the development of the fishery since 1969.

## Discard Estimates

With a few exceptions, estimates of discards have been provided to the Working Group for the ICES Subareas and Divisions 6, 7/8.a,b,d,e and 3/4 (see Table 8.4.1.1) since 1978. Historical discard estimates were revised during the data compilation exercise undertaken for the 2014 benchmark assessment (ICES, 2014). The Working Group considers that the estimates for these areas are incomplete. In 2022, discard data for mackerel were provided by France, Germany, Ireland, Spain, Denmark, England, Scotland and Sweden. Total discards amounted to 3556 tonnes which is a small increase from 2021 ( 3129 tonnes). The German, Dutch and Portuguese pelagic discard monitoring programmes did not record any instances of discarding of mackerel. Estimates from the other countries supplying data include results from the sampling of demersal fleets.

### 8.4.2 Catch at Age

This catch in numbers relates to a total ICES estimated catch of 1046720 tonnes. These figures have been appended to the catch-at-age assessment table (see Table 8.7.1.2).

Age distributions of commercial catch were provided by England, Germany, Faroes, Iceland, Ireland, the Netherlands, Norway, Portugal, Scotland and Spain. There remain gaps in the age sampling of catches, notably from France (length samples were provided), Belgium, Sweden, Denmark, Greenland, Northern Ireland and Poland.

Catches for which there were no sampling data were converted into numbers-at-age using data from the most appropriate fleets. Accurate national fleet descriptions are required for the allocation of sample data to unsampled catches.

The catch numbers at age show a number of strong year classes in this fishery. Over $81 \%$ of the catch in numbers in 2022 consists of 2 to 10-year olds with the 2016-year class again being the
strongest. The 2016 year-class was strong in the fishery in previous years and accounted for $14 \%$ of the catch numbers at age in 2021. The 2019-year class, which are now 3 years old accounts for $11 \%$ of the catch numbers at age and were also strong as 2 year olds in the fishery in 2021 . The 2015 year-class does not look as strong as the other year classes and represents $7 \%$ of the total. In 2022 there was a significant increase in the proportion of fish in the plus group. The plus group now represents $12 \%$ of the catch numbers at age, an increase from $7 \%$ in 2021. The big 2010 year class is now in the plus group and contributes to this increase.

There is a small presence of juvenile (age 0) fish within the 2022 catch. As in previous years catches from Divisions 8.c and 9.a have contained a proportion of juveniles.

### 8.5 Biological data

### 8.5.1 Length Composition of Catch

The mean length-at-age in the catch for 2022 is given in Table 8.5.1.1.
For the most common ages which are well sampled there is little difference to recent years. The length of juveniles is traditionally rather variable. The range of lengths recorded in 2021 for 0 group mackerel is $16.5 \mathrm{~cm}-27.8 \mathrm{~cm}$. The rapid growth of 0 -group fish combined with variations in sampling between northern and southern areas will contribute to the observed variability in the observed size of 0-group fish. Growth is also affected by fish density as indicated by a recent study which demonstrated a link between growth of juveniles and adults ( $0-4$ years) and the abundance of juveniles and adults (Jansen and Burns, 2015). A similar result was obtained for mature 3- to 8-year-old mackerel where a study over 1988-2014 showed declining growth rate since the mid-2000s to 2014, which was negatively related to both mackerel stock size and the stock size of Norwegian spring spawning herring (Ólafsdóttir et al., 2015).

### 8.5.2 Weights at Age in the Catch and Stock

The mean weight-at-age in the catch for 2022 are given in Table 8.7.1.3. There is a trend towards lighter weight-at-age for the most age classes (except 0 to 2 years old) starting around 2005, continuing until 2013 (Figure 8.5.2.1). This decrease in the catch mean weight-at-age seems to have stopped since 2013 and values for the last seven years do not show any particular trend for the older ages (age 6 and older) and are slightly increasing for younger ages (ages 1 to 5). In 2022, however, the mean weight-at-age in the catches have decreased for all ages compared to 2021.

The Working Group used weight-at-age in the stock calculated as the average of the weight-atage in three areas, corresponding to definition of the former spawning components, weighted by the relative mackerel biomass in each area (as estimated by the 2022 egg survey for the southern and western areas and the 2017 egg survey for the North Sea area). Mean weight-at-age in 2022 for the western area are estimated from Dutch, Irish and German commercial catch data, the biological sampling data taken during the egg surveys and during the Norwegian tagging survey. Only samples corresponding to mature fish, from areas and periods corresponding to spawning, as defined at the 2014 benchmark assessment (ICES, 2014) and laid out in the Stock Annex, were used to compute the mean weight-at-age in the western spawning area. For the North Sea area, mean weight-at-age in 2022 were calculated from samples of the commercial catches collected from Divisions 4.a and 4.b in the second quarter of 2022. Stock weights for the southern area, are based on samples from the Spanish catches and surveys in Divisions 8.c and 9 a in the $2^{\text {nd }}$ quarter of the year. The mean weights in the three area and in the stock in 2022 are shown in table 8.5.2.

As for the catch weights, the decreasing trend observed since 2005 for fish of age 3 and older seems to have stopped in 2013 and values in the last 8 years show an increasing trend (except for weights of ages 0 and 1 which have been stable, Figure 8.5.2.2). The mean weight in the stock do not show, contrary to the mean weights in the stock, any specific decrease from 2021 to 2022.

### 8.5.3 Natural Mortality and Maturity Ogive

Natural mortality is assumed to be 0.15 for all age groups and constant over time.
The maturity ogive for 2022 was calculated as the average of the ogives of the three areas weighted by the relative biomass of mackerel in these areas, calculated as described above for the stock weights. The ogives for the North Sea and Southern areas are fixed over time. For the Western area the ogive is updated every year, using maturity data from commercial catch samples from Germany, Ireland, the Netherlands and the UK collected during the first and second quarters (ICES, 2014 and Stock Annex). The 2022 maturity ogives for the three areas and for the mackerel stock are shown in the table 8.5.3.

A trend towards earlier maturation (increasing proportion mature at age 2) has been observed from around 2008 to 2015 . A change in the opposite direction has been observed after that year until 2018. Since then, there is again a trend towards earlier maturation (Figure 8.5.3.1).

### 8.6 Fishery independent data

### 8.6.1 International Mackerel Egg Survey

### 8.6.1.1 Annual egg production for mackerel in 2022

The Mackerel and Horse Mackerel Egg Survey is an ICES co-ordinated international survey in the North-East Atlantic. The survey is a combined plankton and fisheries survey which has been conducted triennially since the late 1970s under the coordination of the ICES Working Group on Mackerel and Horse Mackerel Egg Surveys (WGMEGS).

The main objective is to estimate a Northeast Atlantic mackerel biomass index and a Western horse mackerel egg abundance index using the Annual Egg Production Method (AEPM) during this series of individual cruises from January to July.

Plankton and fecundity samples were analysed according to sampling protocols described in the WGMEGS survey manual (ICES, 2019a) and fecundity manual (ICES, 2019b).

Mackerel egg production (using stage I mackerel eggs) is used in combination with AEPM fecundity parameters to estimate SSB.

The 2022 survey was divided into 6 separate sampling periods. A total of 16 individual cruises have been conducted during 2022. More details can be found in the working paper (WD10: Costas et al. 2023) presented at the WGWIDE 2023 meeting and in the WGMEGS 2023 report (ICES, 2023).

Maps of the distribution of daily egg production for the western and southern areas and the area covered are shown in Figure 8.6.1.1.1. The distribution and timing of mackerel spawning peaks have remained fairly stable in recent surveys. In 2022, spawning peak occurred during the period 5 (Figure 8.6.1.1.2). In 2022 the western and northwestern expansion observed in periods 5 and 6 in 2016 was reported again. (WD10 Costas et al., ICES, 2023),

The total annual egg production (TAEP) for the western area in 2022 was calculated to be 1.80 * $10^{15}$, which is an increase of $47 \%$ compared to the TAEP in 2019 in that area. The TAEP for the southern area in 2022 was calculated to be $2.93^{*} 10^{14}$ (Table 8.6.1.1.1). That is a reduction of $30 \%$ compared to the 2019 TAEP in that area. TAEP for both the western and southern areas combined in 2022 is $2.093^{*} 10^{15}$. This is an increase in production of $28 \%$ compared to 2019 (Figure 8.6.1.1.3) (Table 8.6.1.1.1).

### 8.6.1.2 Mackerel fecundity and atresia estimation

Thanks to the cooperation of the pelagic fishing industry (PFA), the number of fecundity samples collected in 2022 (2312) was significantly increased compared to previous surveys. Additional samples for DEPM analysis were collected in periods 3 and 4 (details in ICES, 2023).

Fecundity estimates are presented as realised fecundity, which is the potential fecundity minus the atresia rate. Relative potential fecundity in 2022 was 1313 oocytes/ g. female fish, slightly higher than the estimates for 2019 (Table 8.6.1.2.1). Realised fecundity was estimated using an average $3 \%$ atresia rate, which gives an estimate of 1268 eggs /g female fish in the year 2022 (Table 8.6.1.2.1). (see WD10Costas et al. 2023 for details)

### 8.6.1.3 Mackerel biomass estimates.

The total spawning stock biomass (SSB) using the AEPM, resulted in an estimated SSB of 3,065 Mt for the western area and 0.499 Mt for the southern area (see Table 8.6.1.1.1). Combining these two areas gives a total of 3.563 Mt , which represents an $18 \%$ increase compared to the SSB estimate in 2019 (Figure 8.6.1.3.1).

### 8.6.1.4 2022 North Sea mackerel egg survey

In 2022, a DEPM North Sea mackerel egg survey was carried out in June by Denmark, England and Norway (Figure 8.6.1.4.1). The survey covered the entire spawning area of the North Sea. The spatial distribution of stage Ia eggs was similar to that of 2021, but egg numbers appeared to be more evenly distributed. No clear pattern in egg densities was observed in the survey. (Figure 8.6.1.4.2) The total survey area in 2022 was slightly smaller than in 2021, with an estimated spawning area of 371126 km 2 . The total daily egg production for 2022 was $0.6909^{*} 10^{13}$ eggs/day (Table 8.6.1.4.1). This is a decrease of $50 \%$ compared to the number of eggs reported in 2021 (Table 8.6.1.4.2), see WD10 Costas et al. 2023 and ICES, 2023 report for more details. The results for the DEPM fecundity parameters should be considered as very preliminary as a full review of the application of the daily method to mackerel will be carried out at WKMADE in November 2023.

### 8.6.2 Demersal trawl surveys in October - March (IBTS Q4 and Q1)

### 8.6.2.1 The data and the model

An index of survivors in the first autumn-winter (recruitment index) was derived from a geostatistical model fitted to catch data from bottom trawl surveys conducted during autumn and winter. A complete description of the data and model can be found in Jansen et al. (2015) and the NEA mackerel stock annex.

The data were compiled from several bottom trawl surveys conducted between October and March from 1998-2023 by research institutes in (Denmark, England, France, Germany, Ireland, Nederland, Norway, Scotland and Sweden). Surveys conducted on the European shelf in the first and fourth quarters are collectively known as the International Bottom Trawl Survey (IBTS), although several of the surveys use different names. All surveys sample the fish community on
the continental shelf and upper shelf slope. IBTS Q4 covers the shelf from the Bay of Biscay to North of Scotland, excluding the North Sea, while IBTS Q1 covers the shelf waters from north of Ireland, around Scotland, the North Sea, Skagerrak and Kattegat.

Trawl operations during the IBTS have largely been standardized through the relevant ICES working group (ICES, 2013b). Furthermore, the effects of variation in wing-spread and trawl speed were included in the model (Jansen et al., 2015). Trawling speed was generally $3.5-4.0$ knots, and trawl gear is also standardized and collectively known as the Grande Ouverture Verticale (GOV) trawl. Some countries use modified trawl gear to suit the particular conditions in the respective survey areas, although this was not expected to change catchability significantly. However, in other cases, the trawl design deviated more significantly from the standard GOV type, namely the Spanish BAKA trawl, the French GOV trawl, and the Irish mini-GOV trawl. The BAKA trawl had a vertical opening of only $2.1-2.2 \mathrm{~m}$ and was towed at only 3 knots. This was considered substantially less suitable for catching juvenile mackerel and, therefore, was excluded from the analysis. The French GOV trawl was rigged without a kite and typically had a reduced vertical opening, which may have reduced the catchability of pelagic species like mackerel. Catchability was assumed to equal the catchability of the standard GOV trawl because testing has shown that the recruitment index was not very sensitive to this assumption (Jansen et al., 2015). Finally, the Irish mini-GOV trawl, used during 1998-2002, was a GOV trawl in reduced dimensions which was accounted for by inclusion of the wing-spread parameter in the model.

No estimation of the 2021 cohort was done because of cancellation of essential Scottish surveys. All surveys in 2022 Q4 and 2023 Q1 were conducted according to standards, so the 2022 cohort could be estimated. Figure 8.6.2.1.1 provides an overview of the distribution and number of samples.

A geostatistical log-Gaussian Cox process model (LGC) with spatiotemporal correlations was used to estimate the catch rates of mackerel recruits through space and time.

### 8.6.2.2 Results

The index of survivors in the first autumn-winter (recruitment index) was updated with data from surveys in 2022 Q4 and 2023 Q1. Parameter estimates and standard errors in the final model are listed in Table 8.6.2.1. The modelled average recruitment index (squared CPUE) surfaces were mapped (Figure 8.6.2.2.1) and the time series of spatially integrated recruitment index values (Table 8.6.2.2) is used in the assessment as a relative abundance index of mackerel at age 0 (recruits). The estimated index value for the 2022 year class is below the previous 5 values, closer to the mean of the time series (Figure 8.6.2.2.2).

### 8.6.2.3 Discussion

The model is being refitted to the entire dataset during the update each year. All index values will therefore be slightly different from year to year, but with the same interannual pattern. This does not affect the stock assessment because it is used in the assessment as a relative abundance index.

The combined demersal surveys have incomplete spatial coverage in some areas that can be important for the estimation of age-0 mackerel abundance, namely: (i) Since 2011, the English survey (covering the Irish sea and the central-eastern part of the Celtic sea including the area around Cornwall) has been discontinued, (ii) the Scottish survey has not consistently covered the area around Donegal Bay, (iii) the IBTS has observed high catch rates in some years at the northeastern edge of the survey area (towards the Norwegian trench) in winter. It is therefore possible that some recruits are also overwintering on the other side of the trench along the south western shelf edge of Norway. Consequently, the NS-IBTS in Q1 should be extended to include the southwestern Norwegian shelf and shelf edge in proximity to the Norwegian trench.

Finally, WGWIDE encourage studies of vertical distribution and catchability of age-0 mackerel in the Q4 and Q1 surveys, to evaluate if it is comparable in all areas (see acoustic information in Jansen et al., 2015).

### 8.6.3 International Ecosystem Summer Survey in Nordic Seas (IESSNS, A7806)

IESSNS is the only annual survey providing data used in the assessment and covers summer feeding distribution of mackerel age 3+ in Nordic Seas. The survey was successfully conducted in 2023. Major survey results worth mentioning is that survey coverage retracted $19 \%$ compared to 2022 as Greenlandic waters were not survey, strata 10 and 11, and mackerel distribution south of Iceland was limited to latitude $62^{\circ} \mathrm{N}$ in the Iceland basin and on the Reykjanes ridge. The zeroline for mackerel abundance was reached for the part of the survey area located north of latitude $60^{\circ} \mathrm{N}$, but not on the European shelf, where mackerel are present west of the British Isles and in the southern North Sea. Value of the mackerel index was not impacted by extremely large catches as occurred in 2022. Western part of the Norwegian Sea (stratum 9) five surface trawl stations were added during the survey at a shorter distance than used between other stations in that strata to find the boundary of the mackerel distribution. However, these additional stations were excluded when calculating the age segregated abundance index which is opposite to what was done in 2022. Furthermore, part of Faroese waters, northern Norwegian Sea by Bear island, and west of Iceland were under sampled as four predetermined surface stations were not sampled due to lack of survey time and mechanical vessel issues, see details in the cruise report. The IESSNS cruise report is available as a working document to this report (WD01) and a detailed survey description is available in the mackerel Stock Annex.

The main results are that estimated total stock abundance and total biomass declined $42 \%$ compared to 2022. Most of the surveyed mackerel are still distributed in the Norwegian Sea. However, they were more easterly and northeasterly distributed compared to 2022. The distribution of mackerel in the Norwegian Sea retracted compared to the last decade, particularly withdrawal from the northernmost part was observed. Internal consistency increased compared to 2022, particularly for ages 5-8 years which had lower consistency than other ages in 2022. Abundance estimates by age are displayed in input data for the assessment (Table 8.7.1.9). Figures 8.6.3.1-2 display estimates of total stock abundance and stock biomass with confidence intervals for the time series. Figures 8.6.3.3-4 show the internal consistency and catch curves for abundance at age from 2012 to 2022. Figures 8.6.3.5-6 display swept area trawl catch rate and mean mackerel density per rectangle for 2023, and mean mackerel density per rectangle for years 2010 and from 2012 to 2023.

Analytical work to develop index calculation method less sensitive to extreme catches was undertaken at ICES Working Group on Improving use of Survey Data for Assessment and Advice (WGISDAA) at their annual meeting in October 2022. The report from the meeting has not yet been published. The meeting did not provide an alternative method for index calculation, therefore the StoX framework was used for index calculation in 2023 as in previous years.

### 8.6.4 Tag Recapture data

Information from steel tagging experiments conducted by Institute of Marine Research in Bergen (IMR) on mackerel at spawning grounds west of Ireland and British Isles in May-June and the respective recaptures at Norwegian factories with metal detectors (Tenningen et al. 2011) was introduced to the mackerel assessment during ICES WKPELA 2014 (ICES, 2014). Data from release years 1980-2004, and recapture years 1986-2006 have been used in the update assessments following this benchmark. From 2011 onwards IMR changed tagging methodology to radio-frequency identification (RFID), more specifically passive integrated transponder tags (PIT-tags). This allowed for more automated data processing with recaptures from scanned landings at factories also outside Norway now being updated in real time to an IMR database over internet.

The data format is the same for both tag types; a table containing the numbers of tagged fish per year class in each release year, and the corresponding numbers scanned and recaptured of the same year classes in all years after release. The RFID data is a separate time series with a different scaling factor (survival) than the steel tags, and it has been used in update assessments following the ICES WKWIDE2017 benchmark (ICES, 2017). For steel tags data from ages 2-11 and all recapture years are used in the assessment. During the 2017 benchmark it was decided to use the same filtering for the RFID data from release year 2011 onwards. However, following decisions made during ICES IBPNEAMac 2019 (ICES 2019c) update assessments are now only using RFID data from release years 2013 onwards, ages 5-11 and recapture year 1 and 2 after release for reasons described below.

During 2011-2019, most of the mackerel were tagged off Ireland, only a minor part farther north of The British Isles, whereas in the years 2020-2023 most of the fish were tagged off the British Isles (Figure 8.6.4.1). The scanned landings and recaptures in time and space changed from 2014 onwards when Iceland and Scotland installed scanners and a larger part of the fishery was covered (Figures 8.6.4.2-3). The numbers recaptured have increased significantly over the years. This is a result of an increasing tagged population, higher biomass scanned as well as a reduced stock. The recaptures are stemming from a wide range of factories, some contributing with more recaptures, which is mainly linked to the biomass scanned (Figure 8.6.4.4). Note that the exclusion of recapture years 3 and longer after release is due to potential tag loss over time. The exclusion of release years 2011-2012 is mainly based on the change with more scanners from 2014 onwards. The exclusion of ages 1-4, was mainly because the tagged numbers of these age groups were relatively few early in the time series compared with the scanned fish year 1 and 2 after release, leading to some noise in the data. However, the age structure of tagged and scanned fish year 12 after release has developed over time series to be more overlapping, and high proportions of tagged mackerel are now at ages 2-4 (Figure 8.6.4.5), implying a potential to also include data from the younger ages in future assessments.

It is difficult to grasp any trends in the raw data of released, scanned and recaptured mackerel used as input to the stock assessment. Hence, we also describe the trend in these data as indices of year class abundance and biomass. The trends in year class abundance indices from RFID data based on recaptures year 1 and 2 after release now seem consistent and informative for assessment from ages 2-12 (Figures 8.6.4.6-7). Note that an alternative assessment at WGWIDE 2023 using these indices for the selected ages 5-11 instead of the raw data table resulted in negligible differences in final SSB, but an improved retrospective pattern. Hence, in a future benchmark it should be explored if using such indices rather than the raw data table as input to stock assessment is more meaningful. By doing that we may also have more comparable diagnostics with the other input data. Translating these abundance indices into different age-aggregated biomass indices also show comparable time trend with SSB from WGWIDE 2023 from release years 2013 onwards (Figure 8.6.4.8), yet with a tendency to be lower in the two last years used in the assessment 2020-2021.

The overall conclusion is that the RFID time series seems informative for stock assessment. However, these data show higher total mortality rate signals than the other input data, resulting in higher SSB if excluded from the assessment. Such conflicting trends suggest that year to year variations in assessment and leave out effects may frequently occur in coming years when time series are short. Finally, the new development of the time series suggests that the current filtering of RFID data for use in stock assessment should be revised in a future benchmark. This especially counts for the inclusion of younger ages 2-4 now contributing with a large proportion among the tagged fish.

### 8.6.5 Other surveys

### 8.6.5.1 International Ecosystem survey in the Norwegian Sea (IESNS, A3675)

After the mid-2000s an increasing amount of NEA mackerel has been observed in catches in the Norwegian Sea during the International Ecosystem survey in the Norwegian Sea in May (IESNS) targeting herring and blue whiting (Salthaug et al. 2019; 2020).

Mackerel was present in the southern and eastern part of the Norwegian Sea in the beginning of May 2023. The spatial distribution of catches in 2023 was similar to 2022, i.e. a lower northward extent than in the period 2008-2021. No further quantitative information can be drawn from these data as this survey is not designed to monitor mackerel.

### 8.6.5.2 Acoustic estimates of mackerel in the Iberian Peninsula and Bay of Biscay (PELACUS, A2548)

PELACUS survey data was not available for WGWIDE 2023 and therefore, no new information from the Bay of Biscay on mackerel distribution and abundance during spawning time is available.

### 8.7 Stock assessment

### 8.7.1 Update assessment in 2023

The update assessment was carried out by fitting the state-space assessment model SAM (Nielsen and Berg, 2014) using the R library stockassessment (downloadable at install_github("fishfollower/SAM/stockassessment")) and adopting the configuration described in the Stock Annex.

The assessment model is fitted to catch-at-age data for ages 0 to 12 (plus group) for the period 1980 to 2022 (with a strong down-weighting of the catches for the period 1980-1999) and three surveys: 1) the SSB estimates from the triennial Mackerel Egg survey (every three years in the period 1992-2022), 2) the recruitment index from the western Europe bottom trawl IBTS Q1 and Q4 surveys $(1998-2020,2022)$ and 3) the abundance estimates for ages 3 to 11 from the IESSNS survey (2010, 2012-2023). The model also incorporates tagging-recapture data from the Norwegian tagging program (for fish recaptured between 1980 and 2005 for the steel tags time series, and fish recaptured between 2014 and 2022 (age 5 and older at release) for the radio frequency tags time series).
Fishing mortality-at-age and recruitment are modelled as random walks, and there is a process error term on abundances at ages 1-11.

The differences in the new data used in this assessment compared to the last year's assessment were:

- Addition of the 2023 abundance-at-age index from IESSNS.
- Revision of the 2022 SSB index from the mackerel egg survey. The final index value is $8 \%$ lower than the preliminary value used in 2022.
- Addition of the 2022 catch-at-age, weights-at-age in the catch and in the stock and maturity ogive, proportions of natural and fishing mortality occurring before spawning.
- The inclusion of the tag recaptures from 2022

Input parameters and configurations are summarized in Table 8.7.1.1. The input data are given in Tables 8.7.1.2 to 8.7.1.9. Given the size of the data base, only the data from the last year of recaptures is given in this report (table 8.7.1.10).

### 8.7.2 Model diagnostics

## Parameter estimates

The estimated parameters and their uncertainty estimates are shown in Table 8.7.2.1 and Figure 8.7.2.1. The model estimates different observation standard deviations for young fish and for older fish. Reflecting the suspected high uncertainty in the catches of age 0 fish (mainly discards), the model gives a very poor fit to this data (large observation standard deviation). The standard deviation of the observation errors on catches of age 1 is lower, though still high, indicating a better fit. For the age 2 and older, the fit to the catch data is very good, with a very low observation standard deviation.

The observation standard deviations for the egg survey and the IESSNS surveys ages 4 to 11 are higher indicating that the assessment gives a lower weight to the information coming from these surveys compared to the catches. The IESSNS age 3 is very poorly fitted in the assessment (high observation standard deviation). Overdispersion of the tag recaptures has the same meaning as the observation standard deviations, but is not directly comparable.

The catchability of the egg survey is 1.13 , larger than 1 , which implies that the assessment considers the egg survey index to be an overestimate. The catchabilities at age for the IESSNS increase from 0.81 for age 3 to 1.65 for age 9 . Since the IESSNS index is expressed as fish abundance, this also means that the assessment considers the IESSNS to provide over-estimated abundance values for the oldest ages. The post tagging mortality estimate is higher for the steel tags (around $40 \%$ ) than for the RFID tags (around 18\%).

The process error standard deviation (ages 1-11) is moderate as well as the standard deviation of the F and recruitment random walks.

The catchability parameters for the egg survey, recruitment index and post tagging survival appear to be estimated more precisely than other parameters (table 8.7.2.1). The catchability for the IESSNS have a slightly higher standard deviation, except for the catchability of the IESSNS at age 3 which has a much higher standard deviation. Uncertainty on the observation standard deviations is larger for the egg survey, the IESSNS age 3 , for the recruitment index and for the catches at age 1 than for the other observations. The uncertainty on the observation variance estimates is not particularly high, especially for the data sources with the lowest observation variances, which are the most influential on the assessment (Figure 8.7.2.2). Uncertainty on the overdispersion of the tag data is high. The standard deviation on the estimate of process error is low, and the standard deviations for the estimates of $F$ random walk variances of age 0 and 1 are both very high. The uncertainty on the random walk variance for recruitment is very large, indicating that the parameter was poorly estimated. Overall, parameters related to age groups 0 and 1 are estimated with larger uncertainty than other parameters in the model, indicating that the model has difficulties representing the dynamics of these age groups based on the data available.

The estimated AR1 error correlation structure for the observations from the IESSNS survey age 3 to 11 has a high correlation between the errors of adjacent ages ( $\mathrm{r}=0.78$ ), then decreasing
exponentially with age difference (Figure 8.7.2.3.). This high error correlation implies that the weight of this survey in the assessment in lower than for a model without correlation structure, which is also reflected in the high observation standard deviation for this survey.

There are some correlations between parameter estimates (Figure 8.7.2.4):

- Catchabilities are positively correlated (especially for the IESSNS age 4 to 11), and negatively correlated to the survival rate for the RFID tags. This simply represents the fact that all scaling parameters are linked, which is to be expected.
- The observation variance for the recruitment index is inversely correlated to the variance of the random walk of the recruitment. This implies that when the model relies less on the recruitment index, the estimated recruitment time series becomes smoother.
- The parameter related to the magnitude of the correlation in the AR1 matrix for the IESSNS is correlated to the observation variance for this survey, which reflects the fact that a strong correlation for the errors is linked to lower weight of the surveys.


## Residuals

The "one step ahead" (uncorrelated) residuals for the catches showed some weak pattern in the residuals, with a prevalence of positive residuals for ages 3 to 10 between 2008 and 2014 (Figure 8.7.2.5). Empirical correlation plot in the residuals shows positive correlations between ages 3 to 12 (Figure 8.7.2.6) which suggest that incorporating a correlation structure in the observation error for the catches might be appropriate. Residuals are of a similar size for all ages, indicating that the model configuration with respect to the decoupling of the observation variances for the catches is appropriate.

The residuals for the egg survey shows a strong temporal pattern with large positive residuals for the period 2007-2010-2013, followed by large negative residuals for 2016, 2019 and 2022. This pattern reflects the fact that the model, based on all the information available, does not follow the trend present in the egg survey, with a steep decline between 2013 and 2016 (when the stock was at its highest according to the assessment model) and the very low 2019 value. The relatively high observation variance for this survey indicates a poor fit with the egg survey due mainly to these observations which point towards a very different direction from the other observations.

Residuals for the IESSNS indices are relatively well balanced for most of the years. Despite the strong drop in the abundances at age in 2018 and 2021, the residuals for these years do not indicate any year effect (e.g. no large residuals of the same sign observed across ages). Correlations between age in the observation errors for this survey are explicitly modelled in the assessment, and as a result, empirical correlations in the one step ahead residuals between ages are low (Figure 8.7.2.7).

Residuals to the recruitment index show no particular pattern, and appear to be relatively randomly distributed in the earlier years, but positive residuals are consistently observed over the last 6 years of data, indicating that the model has difficulties agreeing with this period of sustained high values in the index.

Finally, inspection of the residuals for the tag recaptures (Figure 8.7.2.8) suggest that there is a tendency to have more positive residuals in the RFID tag recaptures for the fish tagged at a young age (age 5 or 6 ) and more negative residuals for fish tagged at older age ( 10 and 12 year olds). This pattern may potentially be explained by age related differences in survival after release, or in tag loss. For the steel tags, there is a tendency to have more positive residuals at the end of the period which could indicate that using a constant survival rate for this dataset may not be appropriate.

## Leave one out runs

In order to visualise the respective impact of the different surveys on the estimated stock trajectories, the assessment was run leaving out successively each of the data sources (Figure 8.7.2.9).

All leave one out runs showed parallel trajectories in SSB and Fbar, except the one leaving out the RFID tag information, which shows a less steep decline in the SSB after 2014 that the other runs and a higher fishing mortality between 2006 and 2017. For recruitment, all runs also resulted in similar trajectories, expect the run without the recruitment index, which recruitment decreased from high levels in the mid-2010s to historical low levels currently.

Removing the IESSNS resulted estimates very close to the base case run, with only some minor difference for Fbar over the period covered by the survey. This behaviour, already observed in 2022, contrasts with what was observed in previous years, where removing the IESSNS index resulted in a markedly smaller SSB estimated, indicating that this survey was pulling stock estimates upwards. This suggests that this survey is becoming increasingly in agreement with the information provided by the catches-at-age, which are the most influential data source.

Removing the recruitment index resulted in lower estimates of SSB over the last 5 years, and, to a lesser extent, higher estimates of Fbar. Without the recruitment index, the estimates of age 0 abundance are only informed by the catch data. As a result, the estimated recruitment for this specific run has a very different trajectory, which - despite the adjustments of year-class strength through the process error as fish become older - has a certain effect on SSB estimates. However, the current model configuration is not appropriate when run without the recruitment index. The remaining data (catch-at-age 0 and 1 ) is not informative enough to allow for a good estimation of the various parameters specific to ages 0 and 1 (observation and random walk variances).

Removing the egg survey resulted in a larger estimated stock, exploited with a lower fishing mortality. The run leaving out the RFID also resulted in a higher SSB than in the assessment using all data for the years after 2017, and a slightly higher fishing mortality between 2011 and 2015, but lower after 2019.

## Additional sensitivity runs

There was a small ( $-8 \%$ ) revision of the 2022 SSB index from the mackerel egg survey. In order to test the sensitivity of the model to such a revision, and potentially explain any revision of stock perception between last year and this year's assessments, the model was run again with the value used in 2022. Model fit and model output were found to be insensitive to this small revision of the SSB index (less than $0.5 \%$ and $0.3 \%$ in SSB and $\mathrm{F}_{\mathrm{bar}}$ respectively).

### 8.7.3 State of the Stock

The stock summary is presented in Figure 8.7.3.1 and Table 8.7.3.1. The stock numbers-at-age and fishing mortality-at-age are presented in Tables 8.7.3.2-3. The spawning stock biomass is estimated to have increased almost continuously from just above 2 million tonnes in the late 1990s and early 2000s to above 6 million tonnes between 2014 and 2016 and subsequently declined to reach a level just under 4 million tonnes in 2022. The fishing mortality has declined from levels between Fpa (0.36) and Flim (0.46) in the early-2000s to levels at or below 0.20 between 2016 to 2019 and increased sharply after that to 0.30 in 2022. The recruitment time series from the assessment is not considered a reliable indicator of year-class strength (see section 8.7.5.1).

There are clear indications of changes in the selectivity of the fishery over the last 30 years (Figure 8.7.3.2.). In the years 1990s, the fishery seems to have had a steeper selection pattern (more rapid increase in fishing mortality with age). Between the end of the 1990s and the end of the 2000s, the selection on the ages 1 to 5 decreased, and selection of older fish ( 7 and older) increased. After 2008, the pattern started reversing towards a steeper selection pattern, until 2017. Since
then, selection on age 2 to 5 decreased sharply, has the fishery targeted more the older part of the population (age 6 and older).

### 8.7.4 Quality of the assessment

## Parametric uncertainty

Large confidence intervals are associated with the SSB in the years before 1992 (Figure 8.7.3.1). This results from the absence of information from the egg survey index, the down-weighting of the information from the catches and the assessment being only driven by the tagging data and natural mortality in the early period. The uncertainty in SSB and Fbar decreases from the early 1990s to the mid-2000s, corresponding to the period where information is available from the egg survey index, the tagging data and (partially) catches (Figure 8.7.4.1 and Figure 8.7.3.1). The uncertainty increases slightly after that, but more sharply in the final assessment year and the SSB estimate for 2022 is estimated with a precision of $+/-24.8 \%$ (Figure 8.7.3.1 and table 8.7.3.1). There is generally also a corresponding large uncertainty on the fishing mortality, especially before 1995. The estimate of Fbar4-8 in 2022 has a precision of $+/-27.0 \%$.

## Model instability

The retrospective analysis was carried out for 7 retro years, (or peels) by fitting the assessment using the 2023 data, removing successively 1 year of data (Figure 8.7.4.2.). There was a systematic retrospective pattern found in Fbar for the older retrospective peels (current year -7 to current year -5 ) with a systematic downwards revision. The was also a pattern in the opposite direction for the SSB (current year -6 to current year -5 ). However this pattern is not apparent in the most recent peals (current -1 to -4 ), and the Mohn's rho value calculated over the last 5 years is of 0.25 for Fbar and -0.15 for SSB. Recruitment at age 0 appears to be quite consistently estimated for the older peals (current -7 to -4 ), but the perception changed for the last 3 peals. This change was associated to an increase in the observation variance for the recruitment index (from 0.17 in the older peel to 0.27 in the current assessment), meaning that the recruitment estimates were more influenced by the recruitment index in the older peels, which was less the case in the last 3 peels.

## Model behaviour

The realisation of the process error in the model was also inspected. The process error expressed as annual deviations in abundances-at-age (Figure 8.7.4.3) shows indications of some pattern across time and ages. There is a predominance of positive deviations in the recent years for ageclasses 5 to 8 . While process error is assumed to be independent and identically distributed, there is clear evidence of correlations in the realisation of the process error in the mackerel assessment, which appears to be correlated both across age-classes and temporally.

The temporal autocorrelation can also be visualised if the process error is expressed in term of biomass (process error expressed as deviations in abundances-at-age multiplied by weight at age and summed over all age classes, Figure 8.7.4.4). Periods with positive values (when the model globally estimates larger abundances-at-age than corresponding to the survival equation) have been alternating with periods with negative values (1991-1994, 2004-2005, and 2017-2019). For the years between 2007 and 2016, the biomass cumulated process error remains positive, and large (reaching in 2013 almost the magnitude of the catches ). The reason for this misbehaviour of the model could not be identified.

### 8.7.5 Exploratory runs

### 8.7.5.1 Assessment starting at age 2

The age 0 estimates in the current assessment mainly rely on the recruitment index; the catch-atage 0 information is considered by the model as uninformative (large observation variance). Catch-at-age information becomes influential at age 2 (very low observation variance). The recruitment signal provided by abundances estimated at age 2 or 3 (when the fish enters the fishery), is different from the signal in the age 0 abundance (Figure 8.7.5.1). Age 0 abundances are less variable than abundances at age 2 and 3 . For the period before 2012, there is a broad agreement in the perception of year class strength, although some year classes that do not appear particularly large at age 0 are perceived as very large at age 2 and 3 (e.g. 2002 year-class). For the more recent period (since the 2013 year-class), there is a greater discrepancy between recruitment at age 0 and at older ages. While the age 0 abundances indicate very high recruitment for the year-classes 2012 to 2020, number of those year-classes appear as particularly poor based on age 2 and 3 abundances (2015, 2017 and 2018). As very little fishing occurs between age 0 and 2 and 3, exploitation is not likely to explain these changes in perception of cohort strength. Such variations could be possibly due to variations in natural mortality (e.g. the strength of a cohort may not be fully determined at age 0 and processes occurring during the first years of life may still be determining year-class strength). However, some cohorts increase in size as they become older (e.g. 2002, 2010 and 2011), which clearly indicates that this is more likely a model artefact. The cohort strength at age 0 , based on the recruitment index, is progressively revised, due to the process error occurring on annual survival, so that cohort strength at age 2 corresponds to the information coming from the catches.

This discrepancy between the recruitment estimates at age 0 and the actual size of the cohort when entering the fishery implies that the age 0 recruitment does not give an accurate indication on year-class strength, and should not be used to make speculations on future stock development.

As very little fishing occurs on 0 and 1 year olds, and catch-at-age data is considered very noisy, and since there appears to be a disagreement between the recruitment index at age 0 and at older ages in the recent years, it does not seem relevant to start the assessment at age 0 or 1 . An exploratory run was conducted starting the assessment at age 2 (and hence removing catch-at-age information for age 0 and 1 and the recruitment index, while leave the rest of the data and model configuration unchanged).

This exploratory run gave a slightly different perception of the stock in the recent years, with a lower SSB for the years 2020 to 2023 and a slightly higher Fbar on the same period (Figure 8.7.5.2). The recruitment at age 2 (in blue on Figure 8.7.5.2, note that the curve should be shifted backwards by 2 years to compare year-class strength with the recruitment at age 0 , red curve) shows a much variable year-class strength signal, with the same perception of year class strength as the age 0 recruitment for some years (broadly between year-classes 2000 and 2012), but a much lower estimated year-class strength since 2012. The recruitment at age 2 shows similar variations as the age 2 abundances in the update assessment (Figure 8.7.5.1). A retrospective analysis conducted on the assessment starting at age 2 shows similar Mohn's rho values as for the update assessment (-0.16 and 0.25 for SSB and Fbar respectively).
In conclusion, both models broadly agree both in terms of fit to the data and in terms of stock trajectories, and the model starting at age 2 could be considered as potential alternative to the current model at the next benchmark for this stock.

### 8.7.5.2 Assessment using the tagging information expressed as a survey index

During the last interbenchmark on the mackerel stock (ICES 2019c), alternative ways of using the tagging data were investigated. One of them consisted in computing an abundance-at-age index using the tag recaptures after one year at liberty. The advantage of this approach is that it offers more consistency in the way the difference data sources are used (same distribution used for the observation errors, thereby making it easier to compare how the different data sources are fitted in the model) and a similar number of data points (with the current approach the tagging data growth by 14 data point each year, compared to 9 for the IESSNS and 1 for each of the other surveys).

The model using the tagging data as index gives a very similar perception of stock development over time compared to the update assessment (Figure 8.7.5.3). The SSB over the years 2014-2020 is slightly lower and the Fbar value is slightly higher, but overall the model as very comparable (including parameter values). The model using the tags as index has lower Mohn's rho values (0.10 and 0.16 for SSB and Fbar respectively (Figure 8.7.5.4). However the residuals for the tagging index show some temporal structure indicating that alternative model configuration (with correlated observation error) should be tested. This is in line with the conclusions from the 2019 interbenchmark.

### 8.8 Short term forecast

The short-term forecast provides estimates of SSB and catch in 2024 and 2025 (given an assumed catch for the current (intermediate) year) and a range of management options for 2024.

All procedures used this year follow those used in the benchmark of 2014 as described in the stock annex.

### 8.8.1 Intermediate year catch estimation

Estimation of catch in the intermediate year (2023) is based on declared quotas, interannual transfers and information from the fisheries shown in the text table in Section 8.1.

### 8.8.2 Initial abundances at age

The recruitment estimate at age 0 from the assessment in the terminal assessment year (2022) is considered too uncertain to be used directly, because this year-class has not yet fully recruited into the fishery. The last recruitment estimate is therefore replaced by predictions from the RCT3 software (Shepherd, 1997). The RCT3 software evaluates the historical performance of the IBTS recruitment index as predictor for the SAM recruitment, by performing a linear regression between the index and the SAM estimates over the period 1998 to the year before the terminal year (2021). The recruitment is then calculated as a weighted mean of the prediction from this linear regression based on the IBTS index value in the corresponding year (2022), and a time tapered weighted mean of the SAM estimates from 1990 to the year before the terminal year (2021). The time tapered weighted mean gives the latest years more weight than a geometric mean. This is done because the recent productivity of the stock appears different than in the 1990's.

The RCT3 prediction done this year lead to a recruitment for 2022 of 5635503 millions, based for $40 \%$ on the time tapered weighted mean (5759032) and for $60 \%$ on the prediction from the linear regression (5 555 006).

The assumption for the subsequent recruitments in 2023 and 2024 was 4498424 million, the geometric mean calculated based on the SAM estimates for the period 1990-2021.

### 8.8.3 Short term forecast

A deterministic short-term forecast was conducted using FLR (www.flr-project.org). Table 8.8.3.1 lists the input data to the forecast and tables 8.8.3.2 and 8.8.3.3 provide projections for various fishing mortality multipliers and catch options in 2024.

Assuming catches for 2023 of 1147407 t , Fbar in 2023 is forecasted to be at 0.38 (above FMSY) and SSB at 3.68 Mt (above $\mathrm{B}_{\mathrm{pa}}$ ) in spring 2023. If catches in 2024 equal to the assumed catch for 2023, Fbar is expected to increase to 0.43 (above $\mathrm{F}_{\mathrm{pa}}$ ) in 2024 with a corresponding decrease in SSB to 3.37 Mt in spring 2023. Assuming an F of 0.40 again in 2025, the SSB will further decrease to 3.10 Mt in spring 2025 (Figure 8.8.1).

Following the MSY approach, the target fishing mortality in 2024 shall be at Fmsy (0.26). This would result in catches of 739386 t and a decrease in SSB to 3.46 Mt in spring 2024 ( $2 \%$ decrease). During the subsequent year, SSB would remain at a similar level ( 3.43 Mt ) in spring 2024.

The forecasted population is partly based on cohorts for which there are abundance estimates from the assessment, and on cohorts for which abundance is based on assumption (RCT3 or geometric mean). The majority of the catches and SSB in 2024 for management option based on the FMSY approach is based on cohorts with SAM estimates. However the abundance of the 2022 year class predicted by the RCT3 software also contributes to a non-negligible proportion of the catches and SSB in the advice year ( $16 \%$ and $24 \%$ respectively). The cohorts based on the geometric mean assumption contribute to a negligible part of the 2024 values (Figure 8.8.2).

### 8.9 Biological reference points

A management strategy evaluation Workshop on northeast Atlantic mackerel (WKMSEMAC) was conducted during 2020 (ICES, 2020) which resulted in the adoption of new reference points for NEA mackerel stock by ICES.

The table below summarises the currently used reference points.

| Framework | Reference point | Value | Technical basis | Source |
| :---: | :---: | :---: | :---: | :---: |
| MSY approach | MSY $\mathrm{B}_{\text {trigger }}$ | 2.58 million tonnes | $\mathrm{B}_{\mathrm{pa}}$ | ICES (2020) |
|  | $\mathrm{F}_{\text {MSY }}$ | 0.26 | Stochastic simulations | ICES (2020) |
| Precautionary approach | $\mathrm{Blim}^{\text {lim }}$ | 2.00 million tonnes | $B_{\text {loss }}$ in 2003 from the 2019 WGWIDE assessment (ICES, 2019d | ICES (2020) |
|  | $\mathrm{B}_{\mathrm{pa}}$ | 2.58 million tonnes | $B_{\lim } \mathrm{x} \exp (1.645 \times \sigma)$, with $\sigma_{S S B}=0.15$ | ICES (2020) |
|  | $\mathrm{F}_{\text {lim }}$ | 0.46 | F that, on average, leads to $\mathrm{Bl}_{\text {lim }}$ | ICES (2020) |
|  | $\mathrm{F}_{\mathrm{pa}}$ | 0.36 | $\mathrm{F}_{\mathrm{p} 05}$ (the F that leads to $\mathrm{SSB} \geq$ Blim with $95 \%$ probability) | ICES (2020) |

### 8.10 Comparison with previous assessments and forecast

## Historical assessment results

The inclusion of a new year of data modifies the relative weight of the different data sources in the assessment, which leads to revisions of the perception of the stock. The relative weights of the data sources in the assessment are dependent on both the length of the time series and the consistency of the information. The recent assessments have systematically revised the estimates of SSB upwards and Fbar downwards for the years prior to 2018 (Figure 8.10.1). Such systematic revisions do not happen for estimates of the most recent years in the assessment period and the assessment results are more consistent. In 2023, there has been an upwards revision of SSB and a downwards revision of fishing mortality

## Comparison of the last two assessments

The last available assessment used to provide advice was carried out in 2022 during the WGWIDE. The new 2023 WGWIDE assessment gives a slightly different perception of the development of the stock, with a higher SSB estimated for the period after 2014 and a lower Fbar estimated since 2009 (Figure 8.10.2). For the latest comparable year between the 2 assessments, 2021, the differences TSB, SSB and Fbar estimates between the previous and the present assessments are of $11.1 \%, 7.58 \%$ and $-13.6 \%$ respectively, similar to the differences observed last year (table below). The upward revision of stock size in 2021 is mainly due to the revision of the abundance of the older age classes (7 year olds and older, Figure 8.10.3). The younger age classes have been revised downwards, especially age 2 (2019 year class).

|  | TSB 2021 | SSB 2021 | Fbar4-8 2021 |
| :--- | :--- | :--- | :--- |
| 2022 WGWIDE | 4710861 tonnes | 3891546 tonnes | 0.308 |
| 2023 WGWIDE | 5238323 tonnes | 4186470 tonnes | 0.266 |
| $\%$ difference | $11.1 \%$ | $7.58 \%$ | $-13.6 \%$ |

The addition of a new year of data modified only marginally the model parameters compared to last year (Figure 8.10.4). The observation standard deviation has decreased slightly for the IESSNS survey while it increased (only marginally) for the other data sources. Process variances were similar as last year, except for the random walk variance of fishing mortality on age 1 , that increased sharply. There was also a decrease in the catchabilities for the IESSNS survey, which results from the revised perception of stock size for the period covered by the survey.

The uncertainty on the estimates of the model parameters is very similar to last year, except for the uncertainty on the observation variance on the catches at age 1 that have increased markedly. The uncertainty on SSB and Fbart-8 in this year's assessment is slightly higher than last year for the period 2002-2019, but is similar for the last 2 years of the assessment period (Figure 8.7.4.1).

## Short term forecast

The estimation for the intermediate year (2022) catch used for the short-term forecast in the advice given last year was $8.09 \%$ higher than the actual catches (table below). This is a discrepancy of the same order of magnitude as for 2021 (for years before that, the intermediate year catch estimate was always very close to the actual catch value). It should however be noted, that there is a $+-10 \%$ flexibility clause in the management strategy that the countries implement, although
the strategy is not formally accepted by the parties. It is impossible to anticipate whether the quota will be utilised in full or $10 \%$ less or more. The above mentioned overshoot is within this range.

The intermediate year assumption is made by summing the unilateral TAC declared, taking interannual transfers into account, and adding anticipated discards. During the WGWIDE, participants may provide information from their national administration and industry on the expected rate of use of the TAC (most often $100 \%$ ), which can lead to a modification of the expected national catches. In 2021 and 2022, several countries were not able to catch their national TAC, due to access restrictions, and this could not be anticipated at the time of the working group. This undershoot of some of the national TACs therefore lead to a too high assumption on the intermediate year catch. In the 2023 estimation of expected catches, some parties provided a lower than declared quota catches to pre-empt a more precise the outtake for 2023 (see text table in section 8.1).

Running the 2022 forecast with the correct value of the intermediate year catch would have resulted in a lower Fbar ( $-8.33 \%$ ) but a similar SSB value in 2022 (only $20 \%$ of the fishing mortality occurs before spawning time, table below). The corresponding catch advice for 2023 would have been 798744 t instead for 782066 t ( $2 \%$ difference).

The SSB from the 2022 forecast (both using the 2022 estimate of the 2022 catch, or the 2023 data) is slightly lower than the SSB estimated by the 2023 assessment. The main reason is a revision of the 2022 abundance-at-age (final year estimates in the 2022, second last year estimates for the 2023 assessment). The abundances of younger age groups ( 3 to 5 ) are revised downwards, but the abundances of the older age groups ( 7 to $12+$ ) are revised upwards (Figure 8.10.3).

|  | Catch (2022) | SSB (2022) | F bar4-8 (2022) |
| :--- | :--- | :--- | :--- | :--- |
| 2022 WGWIDE forecast | 1131416 t | 3769326 t | 0.36 |
| 2022 WGWIDE forecast with correct 2022 catch assumption | 1046720 t | 3790078 t | 0.33 |
| Difference with 2022 forecast | $8.09 \%$ | $0.55 \%$ | $-8.33 \%$ |


| 2023 WGWIDE assessment | 1046720 t | 3951017 t | 0.30 |
| :--- | :--- | :--- | :--- |
| Difference with 2022 forecast | $8.09 \%$ | $4.82 \%$ | $-15.9 \%$ |

Despite this upwards revision of the 2022 stock size with the new assessment, the forecasted catch corresponding to $\mathrm{F}_{\text {MSY }}$ for 2024 is $5 \%$ lower than the previous advice. This is mainly a consequence of the decline in the exploitable biomass of the stock, that is forecasted to continue in 2024 (Figure 8.10.5). This decline is explained by a the decrease in abundance of the oldest age groups on the stock (on which the fishery currently relies) which is not yet compensated by the entry of the good 2019 and 2020 cohorts in the exploitable stock (fish of lower weight, with lower fisheries selection).

### 8.11 Management considerations

Details and discussion on quality issues in this year's assessment is given in Section 8.7 above.

From 2001 to 2007, the internationally agreed TACs covered most of the distribution area of the Northeast Atlantic mackerel. From 2008 to 2014, no agreement was reached among the Coastal States on the sharing of the mackerel quotas. In 2014, three of the Coastal States (EU, NO and FO) agreed on a Management Strategy and sharing arrangement for 2014 to 2018. In November 2018, the agreement from 2014 was extended for two more years until 2020. In January 2020 the UK separated from the EU and assumed the status of a separate Coastal State. There has been no new agreement on a new Management Strategy or share of the stock since 2020. Despite agreeing to abide by the ICES advice, the total declared quotas in each of the years 2015 to 2022 all exceed the advised catch by ICES (Figure 8.11.1).

The mackerel in the Northeast Atlantic was traditionally characterised as three distinct 'spawning components': the southern component, the western component and the North Sea component. The basis for the components was derived from tagging experiments (ICES, 1974). However, the methods normally used to identify stocks or components (e.g., ectoparasite infections, blood phenotypes, otolith shapes and genetics) were not been able to demonstrate significant differences between animals from different components. A review of the mackerel in the North Sea, carried out during WKWIDE 2017 (ICES, 2017) concluded that Northeast Atlantic mackerel should be considered as a single population (stock) with individuals that show stronger or weaker affinity for spawning in certain parts of the spawning area.

In 2023 a Workshop which was tasked with an Evaluation of NEA Mackerel stock components (WKEVALMAC) presented the following conclusions (ICES WKEVALMAC 2023):

1. The workshop acknowledges that there are spatial and temporal patterns in demography (e.g., size and age) which are primarily related to the highly migratory and dynamic nature of the stock, and not due to separate components.
2. Recent directed genetics, tagging, otolith chemistry investigations; ongoing surveys, catch data, fisher perspective and preliminary modelling studies, all failed to support the three-component concept for NEA mackerel.
3. The Workshop therefore rejected the current three-component structure and accepted a single NEA mackerel stock concept to advance towards assessment, management and advice.
4. Given conclusion three, the workshop recommends that the current 'headline advice', based on a single stock assumption should continue.
5. The advice sheet should give no reference to components within the Northeast Atlantic Mackerel stock.

Further to the above conclusions the Workshop considers that within the frame of stock structure, there is no justifiable scientific reason for considering closed areas, closed seasons or spatial quota measures since this is one widely ranging stock. In addition, any reference to minimum landing sizes should apply to the whole stock equally.

## Conservation aspects

The most recent stock assessment shows that the SSB is well above the $B_{m s y}, B_{p a}$ and $B_{l i m}\left(B_{m s y}=\right.$ $B_{p a}=2580000$ tonnes, $\mathrm{B}_{\lim }=2000000$ tonnes). Atlantic Mackerel was by IUCN in 2015 evaluated as a species of least concern (https://www.iucnredlist.org/) and it is neither found on OSPAR's list of threatened and declining species nor on the HELCOM Red List. Furthermore, the WGWIDE has not identified any anthropogenic pressure threatening the stock. Thus ICES has not identified any conservation aspects.

### 8.12 Ecosystem considerations

An overview of the main ecosystem drivers possibly affecting the different life-stages of Northeast Atlantic mackerel and relevant observations are given in the Stock Annex. The discussion here is limited to recent features of relevance.

## Production (recruitment and growth)

Since 2012 the recruitment index (age 0) has been estimating substantially larger year-classes than what is later estimated at age 3 when they enter the fishery and the other surveys. It is not known if this mismatch is a sampling bias or altered mortality of the juveniles between age 0 and 3.

The rapid increase in stock size up until around 2015 was suggested to drive the recent expansion of the spawning northward into new areas (Jansen, 2016). There are several indications of a northward shift and/or expansion in spawning and nursery area towards northern and northeastern areas since 2016 (ICES, 2016; Nøttestad et al., 2018; Bjørdal, 2019; Bjørdal et al. 2022). This northerly shift seems to have continued (Nøttestad et al., 2018). However, spawning in the Norwegian Sea was shown to be of little quantitative significance in 2021 (Burns and O'Hea, WD 15 to WGWIDE 2021 (ICES, 2021b)). Nevertheless, both increased biomass and extended distribution based on a more northward extension of the mackerel egg survey in Norwegian offshore and coastal waters in 2022 (Anon, 2022) and poleward spawning of NEA mackerel, up to $75^{\circ} \mathrm{N}$ (dos Santos Schmidt et al. 2023) clearly document that mackerel seem to be spawning in increasing numbers in recent years in the Norwegian Sea. An on-going extension of spawning activities into the Norwegian Sea feeding area (62-75 \& DEG;N), reached stable levels around 2012 onwards. This poleward expansion of spawning increased as more fish entered the area, whilst the maximum proportions of spawners concurrently dropped from about 75 to $15 \%$ from May to July (dos Santos Schmidt et al. 2023).

Growth (i.e. length- and weight-at-age) have declined substantially in recent times for all ages (e.g. 0-3 year-old in 1998-2012, Jansen and Burns, 2015; all ages in 2005-2015, Jansen and Burns, 2015; Ólafsdóttir et al., 2015). The variations in growth of mackerel in all ages are correlated with mackerel density, e.g. mean weight-at-length have been shown to be positively related to location, day-of-year, temperature and SSB. Furthermore, the density dependent regulation of growth from juvenile to adult mackerel, appears to reflect the spatial dynamics observed in the migration patterns during the feeding season. As such, growth rates of the juveniles were tightly correlated with the density of juveniles in the nursery areas (Jansen and Burns, 2015) and growth for adults (age 3-8) were correlated with the combined effects of mackerel and herring stock sizes (Ólafsdóttir et al., 2015). Conspecific density-dependence was most likely mediated via intensified competition associated with greater mackerel density, possibly also coinciding with decreased prey availability. Nevertheless, weight-at-age of mackerel both from the catches and the surveys have increased during the last few years, particularly for the younger year classes from 1 to 6 years of age (ICES, 2019c; 2020; 2021b, 2022), coinciding with reduced abundance of mackerel in recent years.

## Drivers of the spatial distribution of mackerel

In the mid-2000s, the summer feeding distribution of Northeast Atlantic mackerel (Scomber scombrus) in Nordic Seas began expanding into new areas (Nøttestad et al., 2016). During the period 2007-2016 the mackerel distribution range increased three-fold and the centre-of-gravity shifted westward by 1650 km and northward by 400 km (Ólafsdóttir et al., 2019). Distribution range peaked in 2014/2015 and was positively correlated to SSB (ICES, 2022; Ólafsdóttir et al., 2019). During this period mackerel stock expansion during the feeding season in summer increased from 1.3 mill $\mathrm{km}^{2}$ in 2007 to at least $2.9 \mathrm{mill} \mathrm{km}^{2}$ in 2014, both westwards into Icelandic
and Greenlandic waters and northward in the Norwegian Sea (Nøttestad et al., 2016). The distribution area was stable around 2.8-2.9 mill $\mathrm{km}^{2}$ during the years 2017-2019 (Nøttestad et al., 2017; 2019; ICES, 2018a). However, we witnessed a substantial shift in mackerel concentrations and distribution during summers of 2020-2023, when no mackerel were registered in Greenland waters, and a substantial decline was documented in Icelandic waters, whereas increased biomasses of mackerel were distributed in the central and northern part of the Norwegian Sea (Nøttestad et al., 2020a; WD09 in ICES 2021b). Overall, we have witnessed that mackerel had a much more eastern distribution in 2018-2023 compared to 2014-2017 (ICES, 2018a; Nøttestad et al., 2019; 2020a; 2021).
Ólafsdóttir et al. (2018) modelled (GAM) IESSNS data (2007-2016) and found that mackerel was present in temperatures ranging from $5^{\circ} \mathrm{C}$ to $15^{\circ} \mathrm{C}$, but preferred areas between $9^{\circ} \mathrm{C}$ and $13^{\circ} \mathrm{C}$. The model showed that both mackerel occurrence and density were positively related to location, temperature, meso-zooplankton density and SSB. Thus, geographical expansion of mackerel during the summer feeding season in Nordic Seas was driven by increasing mackerel stock size and constrained by availability of preferred temperature and abundance of meso-zooplankton. However, these results are limited by time-series length (1997-2016; Olafsdottir et al., 2019). Notably, this seems to have changed during the most recent period from 2019 and onwards (e.g. high mackerel concentrations in 2020 at lower temperatures of $7-8{ }^{\circ} \mathrm{C}$; Nøttestad et al., 2019; 2020a; WD09 in ICES 2021b). It is not clear what causes this distributional shift, but the SST were $1-2^{\circ} \mathrm{C}$ lower in the western and south-western areas as compared to a 20-years mean (1999-2009), and substantially lower zooplankton concentrations in Icelandic and Greenland waters in 2019 and 2020 might partly explain such changes (ICES, 2018a; Nøttestad et al., 2019; 2020a). Marine climate with multi-decadal variability might also have affected the observed distributional changes but were not evaluated. Furthermore, tagging data suggest that when mackerel grow older and larger, they seem to cover a progressively wider area of its annual migration cycle, although large inter-annual variability in the recapture patterns likely reflect changes in environmental conditions such as prey availability and ocean currents (Ono et al. 2022)

## Trophic interactions

There are strong indications for interspecific competition for food between mackerel, NSS-herring and blue whiting (Huse et al., 2012), where the competition between mackerel and herring being the best studied relationship. Both higher stomach fullness and prey shift for mackerel compared to herring during low stock size periods indicates that herring may suffer from this competition. Thus, an opportunistic (i.e. rapid shift in diet) and more generalist diet (i.e. wider range of prey) may be advantageous for mackerel in periods with low zooplankton abundances (Langøy et al. 2012; Debes et al. 2012; Óskarsson et al. 2015; Bachiller et al. 2016). Feeding activity seem to be highest in areas associated with colder water masses (Bachiller et al., 2016), and bioenergetics indicate that mackerel consumption may be as high as both herring and blue whiting in some years (122-135 mill t year ${ }^{-1}$, Bachiller et al. 2018). Distribution overlap between mackerel and NSS herring during the summer feeding season is generally highest in the south-western part of the Norwegian Sea (Faroe and east Icelandic area) (Nøttestad et al., 2016; 2017; Ólafsdóttir et al., 2017). This spatiotemporal overlap between mackerel and herring have been present from 2016-2019 (ICES, 2018a, Nøttestad et al., 2016; 2017; 2019). In addition, increasing distribution overlaps in the north-western parts of the Norwegian Sea have also been observed since 2019 and onwards, which is in contrast to previous years (Nøttestad et al., 2019; 2020a; WD09 in ICES 2021b). Overlapping distributions of mackerel and Norwegian spring-spawning herring (NSSH) were particularly present in the western and north-western part of the Norwegian Sea in 2022.

Recently, a number of predators have been highlighted as potential sources of mortality for mackerel. Although limited spatial overlap between marine mammals and mackerel during summers in the Nordic Seas (Nøttestad et al., 2019; Løviknes, 2019), orcas have been observed to actively search and hunt for mackerel schools (Nøttestad et al., 2014; Nøttestad et al., 2020a; 2021,
2022). Furthermore, the increases of 0- and 1-groups mackerel found along major coastlines of Norway (2016-2018, Nøttestad et al., 2018; Bjørdal, 2019) have coincided with predation by increasing numbers of adult Atlantic bluefin tuna (Thynnus thunnus, Boge, 2019; Nøttestad et al., 2020b). Additionally, stomach samples from several species document that smaller sized mackerel is now eaten by different predators in northern waters (e.g. cod, saithe, marine mammals and seabirds; Bjørdal, 2019). Although, fewer 1-groups have been observed in coastal Norway waters in recent years (2019-2023, IESSNS; Nøttestad et al., 2019; 2020b; 2021; 2022; 2023) predation by the Atlantic bluefin tuna is still evident. The predation pressure and associated mortality from various predators on NEA mackerel (both juveniles and adults) are still unknown, but could have ecological impact in both time (i.e. population) and space (i.e. local and regional) (ICCAT, 2019; Nøttestad et al., 2020b).

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### 8.14 Tables

Table 8.2.2.1. NE Atlantic Mackerel. ICES estimated catch (t) in Subareas 1, 2, 5 and 14, 2000-2022 (Data submitted by Working Group members).

| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Denmark | 1375 | 7 | 1 |  |  |  |  |  |  |  | 4845 |
| Estonia | 2673 | 219 |  |  |  |  |  |  |  |  |  |
| Faroe Islands | 5546 | 3272 | 4730 |  | 650 | 30 |  | 278 | 123 | 2992 | 66312 |
| France |  |  |  |  | 2 | 1 |  |  |  |  |  |
| Germany |  |  |  |  |  |  |  | 7 |  |  |  |
| Greenland |  |  |  |  |  |  |  |  |  |  |  |
| Iceland |  |  | 53 | 122 |  | 363 | 4222 | 36706 | 112286 | 116160 | 121008 |
| Ireland |  |  |  | 495 | 471 |  |  |  |  |  |  |
| Latvia |  |  |  |  |  |  |  |  |  |  |  |
| Lithuania | 2085 |  |  |  |  |  |  |  |  |  |  |
| Netherlands |  |  | 569 | 44 | 34 | 2393 |  | 10 | 72 |  | 90 |
| Norway | 31778 | 21971 | 22670 | 125481 | 10295 | 13244 | 8914 | 493 | 3474 | 3038 | 104858 |
| Poland |  |  |  |  |  |  |  |  |  |  |  |
| Sweden |  | 8 |  |  |  |  |  |  |  |  |  |
| United Kingdom |  | 54 | 665 | 692 | 2493 |  |  |  | 4 |  |  |


| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Russia | 49101 | 41566 | 45811 | 40026 | 49489 | 40491 | 33580 | 35408 | 32728 | 414141 | 58613 |  |
| Misreported |  |  | -570 |  | -553 |  |  |  |  |  |  |  |
| Unallocated |  |  |  | -44 | 32 | -2393 |  | -10 | -18 |  |  |  |
| Discards |  |  |  |  | 9 |  |  |  | 112 |  | 5 |  |
| Total | 92557 | 67097 | 73929 | 53883 | 62922 | 54129 | 46716 | 72891 | 148781 | 163604 | 355729 |  |
| Year | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| Denmark | 269 |  | 391 | 2345 | 4321 | 1 | 2 | 289 |  |  | 0.691 | 0.09 |
| Estonia |  |  | 13671 |  |  |  |  |  |  |  |  |  |
| Faroe Islands | 121499 | 107198 | 142976 | 103896 | 76889 | 61901 | 66194 | 52061 | 37418 | 33291 | 105096 | 122237 |
| France | 2 |  | 197 | 8 | 36 |  |  | 733 |  | 8 | 0.2 |  |
| Germany |  | 107 | 74 |  | 2963 | 3499 | 4064 | 577 | 190 | 206 | 9 |  |
| Greenland | 621 | 74021 | 541481 | 875811 | 30351 | 36142 | 46388 | 62973 | 30241 | 26555 | 33360 | 18039 |
| Iceland | 159263 | 149282 | 151103 | 172960 | 169333 | 170374 | 167366 | 168330 | 128008 | 151534 | 132109 | 129975 |
| Ireland | 90 |  |  | 1725 | 6 | 2 |  |  |  |  |  |  |
| Latvia |  |  |  |  |  |  |  |  |  |  |  |  |
| Lithuania |  |  |  | 1082 |  | 1931 |  |  |  | 2 |  |  |


| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Netherlands | 178 | 5 | 1 | 5887 | 6996 | 8599 | 7671 | 2697 | 13 | 0.73 |  | 119 |
| Norway | 43168 | 110741 | 33817 | 192322 | 204574 | 153228 | 167739 | 46853 | 22605 | 15937 | 256124 | 198801 |
| Poland |  |  |  |  |  |  |  | 2 |  | 0.044 | 8.2 | 102 |
| Sweden |  | 4 | 825 | 3310 | 740 | 730 | 1720 | 910 |  | 220 | 228 |  |
| United Kingdom |  |  | 2 | 5534 | 7851 | 5240 | 4601 | 2009 |  | 426 |  | 2 |
| Russia | 73601 | 74587 | 80812 | 116433 | 128433 | 121614 | 138061 | 118255 | 126543 | 128805 | 136176 | 102129 |

Misreported

Unallocated

| Discards | 28 | 1 | 151 | 911 | 78 | 54 | 62 | 51 | 18 | 0.05 |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Total | 398160 | 449326 | 465729 | 684173 | 632571 | 563315 | 603869 | 455740 | 345036 | 356985 | 663111 | 571403 |  |

Table 8.2.2.2. NE Atlantic Mackerel. ICES estimated catch ( $t$ ) in the North Sea, Skagerrak and Kattegat (Subarea 4 and Division 3.a), 2000-2022 (Data submitted by Working Group members).

| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 146 | 97 | 22 | 2 | 4 | 1 | 3 | 1 | 2 | 3 | 27 |
| Denmark | 27720 | 21680 | 343751 | 275081 | 25665 | 232121 | 242191 | 252171 | 26716 | 23491 | 36552 |
| Faroe Islands | 10614 | 18751 | 12548 | 11754 | 11705 | 9739 | 12008 | 11818 | 7627 | 6648 | 4639 |
| France | 1588 | 1981 | 2152 | 1467 | 1538 | 1004 | 285 | 7549 | 490 | 1493 | 686 |
| Germany Fed. Rep. | 78 | 4514 | 3902 | 4859 | 4515 | 4442 | 2389 | 5383 | 4668 | 5158 | 25621 |


| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Iceland |  |  |  |  |  |  |  |  |  |  |  |  |
| Ireland | 9956 | 10284 | 20715 | 17145 | 18901 | 15605 | 4125 | 13337 | 11628 | 12901 | 14639 |  |
| Lithuania |  |  |  |  |  |  |  |  |  |  |  |  |
| Netherlands | 2262 | 2441 | 11044 | 6784 | 6366 | 3915 | 4093 | 5973 | 1980 | 2039 | 1300 |  |
| Norway | 142320 | 158401 | 161621 | 150858 | 147068 | 106434 | 113079 | 131191 | 114102 | 118070 | 129064 |  |
| Poland |  |  |  |  |  | 109 |  |  |  |  |  |  |
| Sweden | 49941 | 5090 | 52321 | 4450 | 4437 | 3204 | 3209 | 38581 | 36641 | 73031 | 34291 |  |
| United Kingdom | 58282 | 52988 | 61781 | 67083 | 62932 | 37118 | 28628 | 46264 | 37055 | 47863 | 52563 |  |
| Russia | 1672 | 1 |  |  |  | 4 |  |  |  |  | 696 |  |
| Misreported (Area 6.a) | 8591 | 39024 | 49918 | 62928 | 23692 | 37911 | 8719 |  | 17280 | 1959 |  |  |
| Unallocated | 34761 | 24873 | 22985 | -730 | -783 | 7043 | 171 | 2421 | 2039 | -629 | 660 |  |
| Discards | 1912 | 24 | 8583 | 11785 | 11329 | 4633 | 8263 | 4195 | 8862 | 8120 | 883 |  |
| Total | 304896 | 339970 | 394878 | 365894 | 317369 | 254374 | 209192 | 257208 | 236111 | 235049 | 247700 |  |
| Year | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| Belgium | 21 | 39 | 62 | 56 | 38 | 99 | 107 | 110 | 13 | 75 | 77 | 185 |
| Denmark | 32800 | 36492 | 31924 | 21340 | 35809 | 21696 | 27457 | 22207 | 25374 | 34375 | 28295 | 16537 |


| Year | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Faroe Islands | 543 | 432 | 25 | 42919 | 25672 | 18193 | 12915 | 15475 | 17460 | 32860 |  | 10896 |
| France | 1416 | 5736 | 1788 | 4912 | 7827 | 3448 | 5942 | 6714 | 5455 | 8959 | 5041 | 6342 |
| Germany Fed. Rep. | 52911 | 4560 | 5755 | 4979 | 6056 | 10172 | 11185 | 12091 | 7778 | 15946 | 9939 | 11903 |
| Iceland |  |  |  |  |  |  |  |  |  |  |  |  |
| Ireland | 15810 | 20422 | 13523 | 45167 | 34167 | 24437 | 35957 | 24567 | 1678 | 15395 | 11021 | 11818 |
| Lithuania |  |  |  | 8340 |  | 596 |  |  |  | 813 | 6655 |  |
| Netherlands | 9881 | 6018 | 4863 | 24536 | 17547 | 11434 | 17401 | 13844 | 8957 | 18425 | 15983 | 21892 |
| Norway | 162878 | 64181 | 130056 | 85409 | 36344 | 55089 | 51960 | 135715 | 135083 | 195515 | 14518 | 95284 |
| Poland |  |  |  |  | 24 |  | 0.721 | 4041 | 1394 | 16 | 559 | 3 |
| Sweden | 32481 | 4560 | 2081 | 1112 | 3190 | 2933 | 1981 | 3056 | 2152 | 3451 | 3277 | 3244 |
| United Kingdom | 69858 | 75959 | 70840 | 145119 | 129203 | 99945 | 104499 | 103707 | 101890 | 130650 | 125553 | 172657 |
| Russia |  |  | 4 |  |  |  |  |  | 0.12 |  |  |  |
| Misreported (Area 6.a) |  |  |  |  |  |  |  |  |  |  |  |  |
| Unallocated |  |  |  |  |  |  |  |  |  |  |  |  |
| Discards | 1906 | 1089 | 337 | 334 | 34 | 559 | 400 | 620 | 812 | 732 | 423 | 2470 |
| Total | 303652 | 219489 | 261258 | 384221 | 295911 | 248611 | 269804 | 342147 | 308047 | 457211 | 221340 | 353231 |

Table 8.2.2.3. NE Atlantic Mackerel. ICES estimated catch (t) in the Western area (Subareas 6 and 7 and Divisions 8.a,b,d,e), 2000-2022 (Data submitted by Working Group members).

| Country | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium |  |  |  |  | 1 |  |  |  |  | 1 | 2 |
| Denmark | 82 | 835 |  | 113 |  |  |  | 6 | 10 |  | 48 |
| Faroe Islands | 4863 | 2161 | 2490 | 2260 | 674 |  | 59 | 1333 | 3539 | 4421 | 36 |
| France | 17857 | 18975 | 19726 | 21213 | 18549 | 15182 | 14625 | 12434 | 14944 | 16464 | 10301 |
| Germany | 22901 | 20793 | 22630 | 19200 | 18730 | 14598 | 14219 | 12831 | 10834 | 17545 | 16493 |
| Guernsey |  |  |  |  |  |  | 10 |  |  |  |  |
| Ireland | 61277 | 60168 | 51457 | 49715 | 41730 | 30082 | 36539 | 35923 | 33132 | 48155 | 43355 |
| Isle of Man |  |  |  |  |  |  |  |  |  |  | 14 |
| Jersey |  |  |  |  |  | 9 | 8 | 6 | 7 | 8 | 6 |
| Lithuania |  |  |  |  |  |  | 95 | 7 |  |  |  |
| Netherlands | 30123 | 33654 | 21831 | 23640 | 21132 | 18819 | 20064 | 18261 | 17920 | 20900 | 21699 |
| Norway |  |  |  |  |  |  |  | 7 | 3948 | 121 | 30 |
| Poland |  |  |  |  |  | 461 | 1368 | 978 |  |  |  |
| Russia |  |  |  |  |  |  |  |  |  |  | 1 |
| Spain | 4500 | 4063 | 3483 |  |  | 4795 | 4048 | 2772 | 7327 | 8462 | 6532 |
| United Kingdom | 126620 | 139589 | 131599 | 167246 | 149346 | 115586 | 67187 | 87424 | 768821 | 109147 | 107840 |
| Misreported (Area 4.a) | -3775 | -39024 | -43339 | -62928 | -23139 | -37911 | -8719 |  | -17280 | -1959 |  |


| Country | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Unallocated | 31564 | 37952 | 27558 | 5587 | 9714 | 13412 | 4783 | 10042 | -952 | 490 | 4503 |  |
| Discards | 1920 | 1164 | 15191 | 7111 | 7696 | 20359 | 14723 | 10177 | 27351 | 6848 | 7518 |  |
| Total | 297932 | 280553 | 252620 | 233157 | 244432 | 190597 | 169009 | 192201 | 177662 | 230603 | 218377 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |
| Country | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| Belgium |  |  |  |  | 14 | 44 | 21 | 58 | 53 | 49 | 34 | 58 |
| Denmark | 2889 | 8 | 903 | 18538 | 6741 | 19443 | 12569 | 8194 | 5189 | 4110 | 4463 | 4897 |
| Faroe Islands | 8 |  |  | 3421 | 5851 | 13173 | 20559 | 13543 | 7787 | 2913 |  | 624 |
| France | 11304 | 14448 | 12438 | 16627 | 17820 | 16634 | 16925 | 13974 | 12371 | 12816 | 11308 | 10929 |
| Germany | 18792 | 14277 | 15102 | 23478 | 19238 | 9740 | 9608 | 7214 | 8936 | 8878 | 2049 | 2225 |
| Greenland |  |  |  |  |  |  |  |  |  | 22 |  |  |
| Guernsey | 10 | 5 | 9 | 9 | 4 |  |  | 12 | 9 |  |  |  |
| Iceland |  |  |  |  |  |  |  |  | 69 |  |  | 1 |
| Ireland | 45696 | 42627 | 42988 | 56286 | 54571 | 52087 | 48957 | 42181 | 51635 | 58720 | 49731 | 40455 |
| Isle of Man | 11 | 11 | 8 | 3 |  | 8 | 2 | 3 | 3 | 2 |  | 1 |
| Jersey | 7 | 8 | 8 | 7 | 3 | 3 | 0.003 | 3 | 2 | 5 |  | 0.1 |


| Country | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Lithuania | 23 |  |  | 176 | 554 | 13 |  |  |  |  |  |  |
| Netherlands | 18336 | 19794 | 16295 | 16242 | 15264 | 17896 | 18694 | 13851 | 13727 | 11895 | 8611 | 4374 |
| Norway | 2019 | 1101 | 734 |  | 1313 | 1035 | 2657 | 4639 | 1420 | 221 | 11 | 1 |
| Poland |  |  |  |  |  |  |  | 14 | 2312 | 5286 | 1155 |  |
| Portugal |  |  |  |  |  |  |  |  | 46 | 35 | 32 |  |
| Russia |  |  |  |  |  | 30 |  |  | 1 | 10 |  | 0.06 |
| Spain | 1257 | 773 | 635 | 1796 | 951 | 1253 | 786 | 4471 | 1220 | 1784 | 704 | 1008 |
| Sweden |  |  |  |  |  |  |  |  | 805 |  |  |  |
| United Kingdom | 111103 | 93775 | 92957 | 137195 | 110932 | 112268 | 116308 | 84309 | 50253 | 72637 | 84323 | 24342 |
| Unallocated | 399 | 16 | -144 |  | 34 |  |  | 13 |  |  |  |  |
| Discards | 7153 | 10654 | 2105 | 1742 | 3185 | 2126 | 2142 | 1701 | 6046 | 8405 | 2640 | 907 |
| Total | 219007 | 197496 | 183857 | 275519 | 236475 | 245754 | 249229 | 194180 | 161883 | 187788 | 165060 | 89822 |

Table 8.2.2.4. NE Atlantic Mackerel. ICES estimated catch ( t ) in Divisions 8.c and 9.a, 2000-2022 (Data submitted by Working Group members). 9.b is included in 2020.

| Country | Div | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| France | 8.6 | 177 | 151 | 43 | 55 | 168 | 383 | 392 | 44 | 283 |
| Portugal | 8.c |  |  |  |  |  |  | 1758 | 2302 | 4867.944 |
| Portugal | 9.a | 2289 | 1509 | 2620 | 2605 | 2381 | 1753 | 2363 | 962 | 824 |


| Country | Div | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Spain | 8.c |  |  | 43063 | 53401 | 50455 | 91043 | 38858 | 14709 | 17768 |  |
| Spain | 9.a |  |  | 7025 | 6773 | 6855 | 14569 | 7347 | 2759 | 845 |  |
| Discards | 8.c | 928 | 391 | 3606 | 156 | 73 | 725 | 4408 | 563 | 2187 |  |
| Discards | 9.a |  | 405 | 1 | 916 | 677 | 241 | 232 | 1245 | 1244 |  |
| Unallocated | 8.c | 28429 | 42851 |  |  |  |  |  | 4691 | 4144 |  |
| Unallocated | 9.a | 3946 | 5107 |  |  |  |  | 108 | 871 | 1076 |  |
| Total | 9.a | 6234 | 7021 | 9646 | 10293 | 9913 | 16562 | 10049 | 5836 | 3989 |  |
| Total |  | 35768 | 50414 | 56358 | 63906 | 60609 | 108713 | 55466 | 28146 | 33239 |  |
| Country | Div | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| France | 8.c | 220 | 171 | 21 | 106 | 83 | 50 | 43 | 96 | 93 | 103 |
| Portugal | 8.c | 5134 | 7334 | 6836 | 6069 | 3697 | 3709 | 3188 | 4189 | 3738 | 2738 |
| Portugal | 9.a | 254 | 618 | 1456 | 619 | 634 | 855 | 706 | 575 | 953 | 829 |
| Spain | 8.c | 14617 | 33783 | 29726 | 26553 | 30893 | 27190 | 19148 | 31143 | 25272 | 26999 |
| Spain | 9.a | 1162 | 2227 | 3853 | 2229 | 1206 | 1656 | 747 | 1379 | 1807 | 1417 |
| Russia |  |  |  |  |  |  |  |  | 2 |  |  |
| Discards | 8.c | 1428 | 2821 | 4724 | 2469 | 84 | 324 | 760 | 28 | 18 | 35 |
| Discards | 9.a | 1027 | 1463 | 2409 | 751 | 143 | 194 | 172 | 115 | 47 | 143 |


| Country | Div | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Unallocated | $8 . c$ | -573 | 8795 | 11 | 1357 |  | 300 |  |  |  |  |
| Unallocated | 9.a | 4053 | 662 | 1831 | 2123 |  |  |  |  |  |  |
| Total | 9.a | 6497 | 4308 | 9550 | 5722 | 1983 | 2736 | 1625 | 2070 | 2807 | 2389 |
| Total |  | 27322 | 57874 | 50867 | 42276 | 36740 | 34279 | 24764 | 37529 | 31928 | 32264 |

Table 8.2.3.1. NE Atlantic Mackerel. Proportion of the catch by quarter 1990-2022

| Year | Q1 | Q2 | Q3 | Q4 |
| :---: | :---: | :---: | :---: | :---: |
| 1990 | 28 | 6 | 26 | 40 |
| 1991 | 38 | 5 | 25 | 32 |
| 1992 | 34 | 5 | 24 | 37 |
| 1993 | 29 | 7 | 25 | 39 |
| 1994 | 32 | 6 | 28 | 34 |
| 1995 | 37 | 8 | 27 | 28 |
| 1996 | 37 | 8 | 32 | 23 |
| 1997 | 34 | 11 | 33 | 22 |
| 1998 | 38 | 12 | 24 | 27 |
| 1999 | 36 | 9 | 28 | 27 |
| 2000 | 41 | 4 | 21 | 33 |
| 2001 | 40 | 6 | 23 | 30 |
| 2002 | 37 | 5 | 29 | 28 |
| 2003 | 36 | 5 | 22 | 37 |
| 2004 | 37 | 6 | 28 | 29 |
| 2005 | 46 | 6 | 25 | 23 |
| 2006 | 41 | 5 | 18 | 36 |
| 2007 | 34 | 5 | 21 | 40 |
| 2008 | 34 | 4 | 35 | 27 |
| 2009 | 38 | 11 | 31 | 20 |
| 2010 | 26 | 5 | 54 | 15 |
| 2011 | 22 | 7 | 54 | 17 |
| 2012 | 22 | 6 | 48 | 24 |
| 2013 | 19 | 5 | 52 | 24 |
| 2014 | 20 | 4 | 46 | 30 |
| 2015 | 20 | 5 | 44 | 31 |
| 2016 | 23 | 4 | 44 | 29 |
| 2017 | 24 | 3 | 45 | 28 |
| 2018 | 20 | 3 | 40 | 37 |


| Year | Q1 | Q2 | Q3 | Q4 |
| :--- | :---: | :---: | :---: | :---: |
| 2019 | 28 | 5 | 42 | 26 |
| 2020 | 31 | 4 | 34 | 31 |
| 2021 | 19 | 5 | 56 | 20 |
| 2022 | 18 | 6 | 58 | 18 |

Table 8.3.1. NE Atlantic Mackerel. Sampling coverage from 1992-2022

| Year | WG Total Catch (t) | \% catch covered by sampling programme | No. Samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1992 | 760000 | 85 | 920 | 77000 | 11800 |
| 1993 | 825000 | 83 | 890 | 80411 | 12922 |
| 1994 | 822000 | 80 | 807 | 72541 | 13360 |
| 1995 | 755000 | 85 | 1008 | 102383 | 14481 |
| 1996 | 563600 | 79 | 1492 | 171830 | 14130 |
| 1997 | 569600 | 83 | 1067 | 138845 | 16355 |
| 1998 | 666700 | 80 | 1252 | 130011 | 19371 |
| 1999 | 608928 | 86 | 1109 | 116978 | 17432 |
| 2000 | 667158 | 76 | 1182 | 122769 | 15923 |
| 2001 | 677708 | 83 | 1419 | 142517 | 19824 |
| 2002 | 717882 | 87 | 1450 | 184101 | 26146 |
| 2003 | 617330 | 80 | 1212 | 148501 | 19779 |
| 2004 | 611461 | 79 | 1380 | 177812 | 24173 |
| 2005 | 543486 | 83 | 1229 | 164593 | 20217 |
| 2006 | 472652 | 85 | 1604 | 183767 | 23467 |
| 2007 | 579379 | 87 | 1267 | 139789 | 21791 |
| 2008 | 611063 | 88 | 1234 | 141425 | 24350 |
| 2009 | 734889 | 87 | 1231 | 139867 | 28722 |
| 2010 | 877272 | 91 | 1241 | 124695 | 29462 |
| 2011 | 948963 | 88 | 923 | 97818 | 22817 |
| 2012 | 899551 | 89 | 1216 | 135610 | 38365 |


| Year | WG Total Catch (t) | \% catch covered by sampling pro- <br> gramme | No. Samples | No. Measured | No. Aged |
| :--- | :--- | :--- | :--- | :--- | :--- |
| 2013 | 938299 | 89 | 1092 | 115870 | 25178 |
| 2014 | 1401788 | 90 | 1506 | 117250 | 43475 |
| 2015 | 1215827 | 88 | 2132 | 137871 | 24283 |
| 2016 | 1100135 | 89 | 2183 | 151548 | 24104 |
| 2017 | 1159641 | 1023144 | 87 | 1858 | 139590 |
| 2018 | 839727 | 87 | 1039 | 1430 | 141561 |

Table 8.3.2. NE Atlantic Mackerel. 2022 sampling by country.

| Country | ICES catch | \% WG catch covered by sampling programme | No. Samples | No. Measured | No. Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Belgium | 242 | 0\% | 0 | 0 | 0 |
| Denmark | 21446 | 0\% | 0 | 0 | 0 |
| Faroe Islands | 133757 | 98\% | 18 | 921 | 892 |
| France | 17785 | 0\% | 0 | 0 | 0 |
| Germany | 14473 | 52\% | 138 | 3875 | 697 |
| Greenland | 18039 | 0\% | 8 | 1600 | 0 |
| Iceland | 129976 | 99\% | 97 | 4005 | 2344 |
| Ireland | 52303 | 99\% | 45 | 8114 | 1958 |
| Netherlands | 26385 | 31\% | 21 | 1467 | 517 |
| Norway | 294086 | 98\% | 79 | 2015 | 1877 |
| Poland | 105 | 0\% | 0 | 0 | 0 |
| Portugal | 3567 | 23\% | 35 | 1768 | 833 |
| Russia | 102129 | 99\% | 136 | 43244 | 1160 |
| Spain | 29898 | 100\% | 1062 | 34890 | 3264 |


| Country | ICES catch | \% WG catch covered by sampling <br> programme | No. Sam- <br> ples | No. Meas- <br> ured | No. <br> Aged |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Sweden | 3244 | $0 \%$ | 0 | 0 | 0 |
| UK (England \& Wales) | 22987 | $92 \%$ | 154 | 11426 | 1199 |
| UK (Northern Ireland) | 15117 | $0 \%$ | 0 | 0 | 0 |
| UK (Scotland) | 161182 | $94 \%$ | 69 | 5248 | 2030 |

Table 8.3.3. NE Atlantic Mackerel. 2022 sampling by ICES division.

| Division | Official Catch (t) | WG Catch (t) | No. Samples | No. Measured | No Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2.a | 567564 | 567564 | 309 | 50992 | 5546 |
| 2.b | 5 | 5 | 0 | 0 | 0 |
| 3.a | 551 | 551 | 0 | 0 | 0 |
| 3.b | 26 | 26 | 0 | 0 | 0 |
| 3.c | 3 | 3 | 0 | 0 | 0 |
| 3.d | 4 | 4 | 0 | 0 | 0 |
| 4.a | 349349 | 349349 | 215 | 12998 | 3863 |
| 4.b | 2296 | 2296 | 0 | 0 | 0 |
| 4.c | 1002 | 1002 | 17 | 489 | 92 |
| 5.a | 573 | 573 | 1 | 50 | 25 |
| 5.b | 3150 | 3150 | 2 | 104 | 100 |
| 6.a | 64049 | 64049 | 38 | 5256 | 1348 |
| 6.b | 23 | 23 | 0 | 0 | 0 |
| 7.a | 4 | 4 | 0 | 0 | 0 |
| 7.b | 10272 | 10272 | 12 | 1595 | 346 |
| 7.c | 24 | 24 | 0 | 0 | 0 |
| 7.d | 5724 | 5724 | 7 | 721 | 175 |
| 7.e | 799 | 799 | 53 | 2870 | 434 |
| 7.f | 331 | 331 | 75 | 5480 | 269 |
| 7.9 | 27 | 27 | 0 | 0 | 0 |
| 7.h | 24 | 24 | 0 | 0 | 0 |


| Division | Official Catch (t) | WG Catch (t) | No. Samples | No. Measured | No Aged |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 7.j | 2932 | 2932 | 36 | 1360 | 476 |
| 7.k | 0.33 | 0.33 | 0 | 0 | 0 |
| 8.a | 2980 | 2980 | 11 | 204 | 4 |
| 8.b | 2618 | 2618 | 132 | 2608 | 41 |
| $8 . \mathrm{c}$ | 29875 | 29875 | 587 | 27539 | 2453 |
| 8.d | 15 | 15 | 7 | 100 | 3 |
| 9.a | 2389 | 2389 | 360 | 6207 | 1596 |
| 14.a | 110.7 | 110.7 | 0 | 0 | 0 |

Table 8.4.1.1. NE Atlantic Mackerel. ICES estimated catches by area ( t ). Discards not estimated prior to 1978 (data submitted by Working Group members).

| Year | Subarea 6 |  |  | Subarea 7 and divisions 8.a, 8.b, 8.d, and 8.e |  |  | Subareas 3 and 4 |  |  | Subareas 1, 2, 5, and 14 |  |  | Divisions 8.c and 9.a |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Land. | Disc. | Catch | Land. | Disc. | Catch | Land. | Disc. | Catch | Land. | Disc. | Catch | Land. | Disc. | Catch | Land. | Disc. | Catch |
| 1969 | 4800 |  | 4800 | 47404 |  | 47404 | 739175 |  | 739175 | 7 |  | 7 | 42526 |  | 42526 | 833912 |  | 833912 |
| 1970 | 3900 |  | 3900 | 72822 |  | 72822 | 322451 |  | 322451 | 163 |  | 163 | 70172 |  | 70172 | 469508 |  | 469508 |
| 1971 | 10200 |  | 10200 | 89745 |  | 89745 | 243673 |  | 243673 | 358 |  | 358 | 32942 |  | 32942 | 376918 |  | 376918 |
| 1972 | 13000 |  | 13000 | 130280 |  | 130280 | 188599 |  | 188599 | 88 |  | 88 | 29262 |  | 29262 | 361229 |  | 361229 |
| 1973 | 52200 |  | 52200 | 144807 |  | 144807 | 326519 |  | 326519 | 21600 |  | 21600 | 25967 |  | 25967 | 571093 |  | 571093 |
| 1974 | 64100 |  | 64100 | 207665 |  | 207665 | 298391 |  | 298391 | 6800 |  | 6800 | 30630 |  | 30630 | 607586 |  | 607586 |
| 1975 | 64800 |  | 64800 | 395995 |  | 395995 | 263062 |  | 263062 | 34700 |  | 34700 | 25457 |  | 25457 | 784014 |  | 784014 |
| 1976 | 67800 |  | 67800 | 420920 |  | 420920 | 305709 |  | 305709 | 10500 |  | 10500 | 23306 |  | 23306 | 828235 |  | 828235 |
| 1977 | 74800 |  | 74800 | 259100 |  | 259100 | 259531 |  | 259531 | 1400 |  | 1400 | 25416 |  | 25416 | 620247 |  | 620247 |
| 1978 | 151700 | 15100 | 166800 | 355500 | 35500 | 391000 | 148817 |  | 148817 | 4200 |  | 4200 | 25909 |  | 25909 | 686126 | 50600 | 736726 |
| 1979 | 203300 | 20300 | 223600 | 398000 | 39800 | 437800 | 152323 | 500 | 152823 | 7000 |  | 7000 | 21932 |  | 21932 | 782555 | 60600 | 843155 |
| 1980 | 218700 | 6000 | 224700 | 386100 | 15600 | 401700 | 87931 |  | 87931 | 8300 |  | 8300 | 12280 |  | 12280 | 713311 | 21600 | 734911 |
| 1981 | 335100 | 2500 | 337600 | 274300 | 39800 | 314100 | 64172 | 3216 | 67388 | 18700 |  | 18700 | 16688 |  | 16688 | 708960 | 45516 | 754476 |
| 1982 | 340400 | 4100 | 344500 | 257800 | 20800 | 278600 | 35033 | 450 | 35483 | 37600 |  | 37600 | 21076 |  | 21076 | 691909 | 25350 | 717259 |
| 1983 | 320500 | 2300 | 322800 | 235000 | 9000 | 244000 | 40889 | 96 | 40985 | 49000 |  | 49000 | 14853 |  | 14853 | 660242 | 11396 | 671638 |
| 1984 | 306100 | 1600 | 307700 | 161400 | 10500 | 171900 | 43696 | 202 | 43898 | 98222 |  | 98222 | 20208 |  | 20208 | 629626 | 12302 | 641928 |
| 1985 | 388140 | 2735 | 390875 | 75043 | 1800 | 76843 | 46790 | 3656 | 50446 | 78000 |  | 78000 | 18111 |  | 18111 | 606084 | 8191 | 614275 |
| 1986 | 104100 |  | 104100 | 128499 |  | 128499 | 23309 | 7431 | 243740 | 101000 |  | 101000 | 24789 |  | 24789 | 594697 | 7431 | 602128 |
| 1987 | 183700 |  | 183700 | 100300 |  | 100300 | 290829 | 10789 | 301618 | 47000 |  | 47000 | 22187 |  | 22187 | 644016 | 10789 | 654805 |
| 1988 | 115600 | 3100 | 118700 | 75600 | 2700 | 78300 | 308550 | 29766 | 338316 | 120404 |  | 120404 | 24772 |  | 24772 | 644926 | 35566 | 680492 |
| 1989 | 121300 | 2600 | 123900 | 72900 | 2300 | 75200 | 279410 | 2190 | 281600 | 90488 |  | 90488 | 18321 |  | 18321 | 582419 | 7090 | 589509 |
| 1990 | 114800 | 5800 | 120600 | 56300 | 5500 | 61800 | 300800 | 4300 | 305100 | 118700 |  | 118700 | 21311 |  | 21311 | 611911 | 15600 | 627511 |
| 1991 | 109500 | 10700 | 120200 | 50500 | 12800 | 63300 | 358700 | 7200 | 365900 | 97800 |  | 97800 | 20683 |  | 20683 | 637183 | 30700 | 667883 |
| 1992 | 141906 | 9620 | 151526 | 72153 | 12400 | 84553 | 364184 | 2980 | 367164 | 139062 |  | 139062 | 18046 |  | 18046 | 735351 | 25000 | 760351 |
| 1993 | 133497 | 2670 | 136167 | 99828 | 12790 | 112618 | 387838 | 2720 | 390558 | 165973 |  | 165973 | 19720 |  | 19720 | 806856 | 18180 | 825036 |
| 1994 | 134338 | 1390 | 135728 | 113088 | 2830 | 115918 | 471247 | 1150 | 472397 | 72309 |  | 72309 | 25043 |  | 25043 | 816025 | 5370 | 821395 |
| 1995 | 145626 | 74 | 145700 | 117883 | 6917 | 124800 | 321474 | 730 | 322204 | 135496 |  | 135496 | 27600 |  | 27600 | 748079 | 7721 | 755800 |
| 1996 | 129895 | 255 | 130150 | 73351 | 9773 | 83124 | 211451 | 1387 | 212838 | 103376 |  | 103376 | 34123 |  | 34123 | 552196 | 11415 | 563611 |
| 1997 | 65044 | 2240 | 67284 | 114719 | 13817 | 128536 | 226680 | 2807 | 229487 | 103598 |  | 103598 | 40708 |  | 40708 | 550749 | 18864 | 569613 |
| 1998 | 110141 | 71 | 110212 | 105181 | 3206 | 108387 | 264947 | 4735 | 269682 | 134219 |  | 134219 | 44164 |  | 44164 | 658652 | 8012 | 666664 |
| 1999 | 116362 |  | 116362 | 94290 |  | 94290 | 313014 |  | 313014 | 72848 |  | 72848 | 43796 |  | 43796 | 640311 |  | 640311 |
| 2000 | 187595 | 1 | 187595 | 115566 | 1918 | 117484 | 285567 | 165 | 304898 | 92557 |  | 92557 | 36074 |  | 36074 | 736524 | 2084 | 738608 |
| 2001 | 143142 | 83 | 143142 | 142890 | 1081 | 143971 | 327200 | 24 | 339971 | 67097 |  | 67097 | 43198 |  | 43198 | 736274 | 1188 | 737462 |
| 2002 | 136847 | 12931 | 149778 | 102484 | 2260 | 104744 | 375708 | 8583 | 394878 | 73929 |  | 73929 | 49576 |  | 49576 | 749131 | 23774 | 772905 |


| Year | Subarea 6 |  |  | Subarea 7 and divisions 8.a, 8.b, 8.d, and 8.e |  |  | Subareas 3 and 4 |  |  | Subareas 1, 2, 5, and 14 |  |  | Divisions 8.c and 9.a |  |  | Total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Land. | Disc. | Catch | Land. | Disc. | Catch | Land. | Disc. | Catch | Land. | Disc. | Catch | Land. | Disc. | Catch | Land. | Disc. | Catch |
| 2003 | 135690 | 1399 | 137089 | 90356 | 5712 | 96068 | 354109 | 11785 | 365894 | 53883 |  | 53883 | 25823 | 531 | 26354 | 659831 | 19427 | 679288 |
| 2004 | 134033 | 1705 | 134738 | 103703 | 5991 | 109694 | 306040 | 11329 | 317369 | 62913 | 9 | 62922 | 34840 | 928 | 35769 | 640529 | 19962 | 660491 |
| 2005 | 79960 | 8201 | 88162 | 90278 | 12158 | 102436 | 249741 | 4633 | 254374 | 54129 |  | 54129 | 49618 | 796 | 50414 | 523726 | 25788 | 549514 |
| 2006 | 88077 | 6081 | 94158 | 66209 | 8642 | 74851 | 200929 | 8263 | 209192 | 46716 |  | 46716 | 52751 | 3607 | 56358 | 454587 | 26594 | 481181 |
| 2007 | 110788 | 2450 | 113238 | 71235 | 7727 | 78962 | 253013 | 4195 | 257208 | 72891 |  | 72891 | 62834 | 1072 | 63906 | 570762 | 15444 | 586206 |
| 2008 | 76358 | 21889 | 98247 | 73954 | 5462 | 79416 | 227252 | 8862 | 236113 | 148669 | 112 | 148781 | 59859 | 750 | 60609 | 586090 | 37075 | 623165 |
| 2009 | 135468 | 3927 | 139395 | 88287 | 2921 | 91208 | 226928 | 8120 | 235049 | 163604 |  | 163604 | 107747 | 966 | 108713 | 722035 | 15934 | 737969 |
| 2010 | 106732 | 2904 | 109636 | 104128 | 4614 | 108741 | 246818 | 883 | 247700 | 355725 | 5 | 355729 | 50826 | 4640 | 55466 | 864229 | 13045 | 877272 |
| 2011 | 160756 | 1836 | 162592 | 51098 | 5317 | 56415 | 301746 | 1906 | 303652 | 398132 | 28 | 398160 | 26337 | 1807 | 28144 | 938070 | 10894 | 948963 |
| 2012 | 121115 | 952 | 122067 | 65728 | 9701 | 75429 | 218400 | 1089 | 219489 | 449325 | 1 | 449326 | 29809 | 3431 | 33240 | 884377 | 15174 | 899551 |
| 2013 | 132062 | 273 | 132335 | 49871 | 1652 | 51523 | 260921 | 337 | 261258 | 465846 | 15 | 465861 | 24867 | 2455 | 27322 | 933567 | 4732 | 938299 |
| 2014 | 180068 | 340 | 180408 | 93709 | 1402 | 95111 | 383887 | 334 | 384221 | 684082 | 91 | 684173 | 53591 | 4284 | 57875 | 1395337 | 6451 | 1401788 |
| 2015 | 134728 | 30 | 134757 | 98563 | 3155 | 101718 | 295877 | 34 | 295911 | 632493 | 78 | 632571 | 43735 | 7133 | 50869 | 1205396 | 10431 | 1215827 |
| 2016 | 206326 | 200 | 206526 | 37300 | 1927 | 39227 | 248041 | 570 | 248611 | 563440 | 54 | 563494 | 39056 | 3220 | 42276 | 1094163 | 5971 | 1100135 |
| 2017 | 225959 | 151 | 226110 | 21128 | 1992 | 23119 | 269404 | 400 | 269804 | 603806 | 62 | 603869 | 36512 | 227 | 36739 | 1156809 | 2832 | 1159641 |
| 2018 | 157239 | 90 | 157329 | 32037 | 1611 | 33649 | 341527 | 620 | 342147 | 455689 | 51 | 455740 | 33761 | 518 | 34279 | 1020254 | 2890 | 1023144 |
| 2019 | 122995 | 144 | 123139 | 32840 | 5902 | 38742 | 307235 | 812 | 308047 | 345019 | 18 | 345037 | 23832 | 931 | 24763 | 831920 | 7807 | 839727 |
| 2020 | 130577 | 341 | 130918 | 48806 | 8065 | 56871 | 456479 | 732 | 457211 | 356985 |  | 356985 | 37386 | 143 | 37529 | 1030233 | 9280 | 1039513 |
| 2021 | 146519 | 117 | 146635 | 15901 | 2524 | 18425 | 221019 | 423 | 221442 | 663111 |  | 663111 | 31862 | 65 | 31928 | 1078411 | 3129 | 1081540 |
| 2022 | 63850 | 222 | 64072 | 25065 | 685 | 25750 | 350760 | 2470 | 353231 | 571403 |  | 571403 | 32086 | 178 | 32264 | 1043164 | 3556 | 1046720 |

Table 8.5.1.1. NE Atlantic Mackerel. Mean length (cm) -at-age by area for 2022 (Q1-Q4).

| Age |  | 2.a |  | 2.6 |  | 3.a |  | 3.b |  | 3.c |  | 3.d |  | 4.a |  | 4.6 |  | $4 . c$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 25.14 |
|  | 1 |  | 28.23 |  | 28.00 |  | 28.28 |  | 28.28 |  | 28.28 |  | 30.57 |  | 28.52 |  | 29.69 |  | 29.71 |
|  | 2 |  | 31.64 |  | 30.94 |  | 31.25 |  | 31.29 |  | 29.86 |  | 31.03 |  | 31.51 |  | 31.18 |  | 31.36 |
|  | 3 |  | 33.12 |  | 32.75 |  | 33.01 |  | 32.78 |  | 32.78 |  | 32.38 |  | 33.26 |  | 32.20 |  | 33.07 |
|  | 4 |  | 33.48 |  | 34.94 |  | 34.72 |  | 34.65 |  | 34.69 |  | 34.30 |  | 34.71 |  | 34.16 |  | 34.42 |
|  | 5 |  | 34.61 |  | 35.60 |  | 35.45 |  | 35.39 |  | 35.42 |  | 35.40 |  | 35.54 |  | 35.24 |  | 35.60 |
|  | 6 |  | 35.89 |  | 36.39 |  | 36.33 |  | 36.31 |  | 36.39 |  | 37.03 |  | 36.57 |  | 35.42 |  | 36.74 |
|  | 7 |  | 36.54 |  | 37.60 |  | 37.30 |  | 37.34 |  | 37.35 |  | 37.43 |  | 37.29 |  | 36.75 |  | 37.44 |
|  | 8 |  | 37.39 |  | 37.85 |  | 37.64 |  | 37.65 |  | 37.68 |  | 37.92 |  | 37.72 |  | 37.27 |  | 38.19 |
|  | 9 |  | 37.86 |  | 38.47 |  | 38.36 |  | 38.15 |  | 38.23 |  | 38.29 |  | 38.10 |  | 37.94 |  | 38.49 |
|  | 10 |  | 37.98 |  | 38.61 |  | 38.41 |  | 38.31 |  | 38.36 |  | 38.56 |  | 38.33 |  | 38.44 |  | 39.01 |
|  | 11 |  | 38.58 |  | 38.67 |  | 38.81 |  | 38.62 |  | 38.57 |  | 39.29 |  | 38.78 |  | 38.26 |  | 38.44 |
|  | 12 |  | 38.95 |  | 39.28 |  | 39.16 |  | 39.03 |  | 39.05 |  | 39.02 |  | 39.14 |  | 39.16 |  | 39.14 |
|  | 13 |  | 39.00 |  | 39.37 |  | 39.64 |  | 39.62 |  | 39.71 |  | 39.22 |  | 39.62 |  | 39.61 |  | 39.58 |
|  | 14 |  | 39.30 |  | 39.56 |  | 40.18 |  | 40.23 |  | 40.27 |  | 38.64 |  | 40.05 |  | 40.18 |  | 40.15 |
|  | 15 |  | 39.92 |  | 39.39 |  | 40.03 |  | 38.76 |  | 38.46 |  | 39.44 |  | 40.33 |  | 40.12 |  | 39.21 |


| Age |  | 5.a |  | 5.b |  | 6.a |  | 6.b |  | 6.62 |  | 7.a |  | 7.b |  | 7.c |  | 7.d |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 |  |  |  |  |  | 16.5 |  |  |  |  |  | 16.5 |  |  |  |  | 25.14 |
|  | 1 |  |  |  |  |  | 18.693 |  |  |  |  |  | 18.5 |  |  |  |  | 29.7 |
|  | 2 |  | 33.44 |  | 30.38 |  | 29.98 |  | 30.10 |  | 28.96 |  | 29.26 |  | 29.75 |  | 29.80 | 29.82 |
|  | 3 |  | 33.70 |  | 32.78 |  | 32.63 |  | 32.75 |  | 31.62 |  | 32.50 |  | 32.73 |  | 32.69 | 32.26 |
|  | 4 |  | 35.40 |  | 34.60 |  | 34.79 |  |  |  | 34.20 |  | 35.00 |  | 34.92 |  | 34.92 | 33.90 |
|  | 5 |  | 36.11 |  | 35.79 |  | 35.86 |  | 35.79 |  | 34.76 |  | 36.13 |  | 35.96 |  | 35.96 | 36.02 |
|  | 6 |  | 36.86 |  | 36.68 |  | 36.69 |  | 36.69 |  | 36.02 |  | 36.58 |  | 36.53 |  | 36.50 | 35.78 |
|  | 7 |  | 37.75 |  | 37.18 |  | 37.68 |  | 37.18 |  | 37.51 |  | 37.71 |  | 37.85 |  | 37.88 | 38.20 |
|  | 8 |  | 37.96 |  | 38.27 |  | 38.15 |  | 38.28 |  | 37.53 |  | 38.12 |  | 38.28 |  | 38.26 | 38.65 |
|  | 9 |  | 38.69 |  | 37.82 |  | 38.45 |  | 37.82 |  | 37.89 |  | 38.52 |  | 38.90 |  | 38.84 |  |
|  | 10 |  | 38.63 |  | 38.72 |  | 38.71 |  | 38.72 |  | 38.01 |  | 38.79 |  | 39.44 |  | 39.59 |  |
|  | 11 |  | 38.63 |  | 38.48 |  | 38.87 |  | 38.48 |  | 38.72 |  | 38.90 |  | 39.20 |  | 39.31 |  |
|  | 12 |  | 39.47 |  | 38.32 |  | 39.12 |  | 38.30 |  | 38.68 |  | 39.27 |  | 39.17 |  | 39.23 |  |
|  | 13 |  | 39.57 |  | 38.71 |  | 39.88 |  | 38.70 |  | 39.64 |  | 39.88 |  | 39.63 |  | 39.63 |  |
|  | 14 |  | 39.63 |  | 39.71 |  | 39.62 |  | 39.70 |  | 40.61 |  | 39.22 |  |  |  | 40.10 |  |
|  | 15 |  | 39.59 |  | 43.24 |  | 40.30 |  | 43.30 |  | 40.28 |  | 40.30 |  |  |  | 40.70 |  |


| Age |  | $7 . e$ |  | 7.f |  | 7.9 |  | 7.h |  | 7.j |  | 7.k |  | 8.a |  | 8.b |  | 8.c |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 |  | 25.86 |  | 25.92 |  |  |  |  |  |  |  |  |  |  |  | 16.73 | 17.36 |
|  | 1 |  | 26.31 |  | 26.76 |  |  |  |  |  |  |  |  |  | 20.12 |  | 19.98 | 28.24 |
|  | 2 |  | 29.40 |  | 30.00 |  | 29.21 |  | 29.85 |  | 30.22 |  | 28.39 |  | 30.25 |  | 29.42 | 31.35 |
|  | 3 |  | 32.26 |  | 32.33 |  | 32.53 |  | 32.35 |  | 32.60 |  | 31.81 |  | 31.46 |  | 31.70 | 33.37 |
|  | 4 |  | 33.99 |  | 34.03 |  | 34.85 |  | 33.45 |  | 34.75 |  | 34.22 |  | 33.86 |  | 34.72 | 35.60 |
|  | 5 |  | 35.89 |  | 35.89 |  | 35.96 |  | 36.19 |  | 36.13 |  | 36.96 |  | 36.74 |  | 36.88 | 36.53 |
|  | 6 |  | 35.77 |  | 36.96 |  | 36.55 |  | 37.48 |  | 36.81 |  | 37.91 |  | 38.36 |  | 38.10 | 37.38 |
|  | 7 |  | 38.17 |  | 37.37 |  | 37.89 |  | 38.44 |  | 38.10 |  | 38.28 |  | 38.27 |  | 38.21 | 37.93 |
|  | 8 |  | 38.65 |  | 37.24 |  | 38.35 |  | 38.48 |  | 38.37 |  | 38.34 |  | 38.28 |  | 38.37 | 38.13 |
|  | 9 |  | 42.50 |  | 37.42 |  | 38.69 |  | 38.61 |  | 38.68 |  | 38.42 |  | 38.44 |  | 38.79 | 38.76 |
|  | 10 |  | 38.40 |  | 38.01 |  | 39.21 |  | 39.18 |  | 39.21 |  | 38.65 |  | 38.97 |  | 39.24 | 39.20 |
|  | 11 |  |  |  | 37.83 |  | 39.08 |  | 39.08 |  | 39.08 |  | 38.50 |  | 39.01 |  | 39.35 | 39.39 |
|  | 12 |  | 39.00 |  | 38.00 |  | 38.37 |  | 38.38 |  | 38.37 |  | 39.50 |  | 40.37 |  | 40.34 | 40.38 |
|  | 13 |  | 41.00 |  |  |  | 40.25 |  | 40.25 |  | 40.25 |  |  |  | 40.03 |  | 40.10 | 40.76 |
|  | 14 |  |  |  |  |  |  |  |  |  |  |  |  |  | 41.50 |  | 43.09 | 43.19 |
|  | 15 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Age | 8.d |  | 9.a |  | 9.aN |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :---: |
| 0 | 18.93 | 27.81 | 18.59 |  | All |  |
| 1 | 18.85 | 29.16 | 26.91 |  | 20.47 |  |
| 2 | 29.94 | 29.42 | 28.88 | 33.44 | 31.41 |  |
| 3 | 31.68 | 33.68 | 31.08 | 33.70 | 33.11 |  |
| 4 | 35.57 | 37.15 | 33.63 | 35.40 | 34.14 |  |
| 5 | 36.63 | 38.72 | 35.37 | 36.11 | 35.12 |  |
| 6 | 37.36 | 39.22 | 37.03 | 36.86 | 36.25 |  |
| 7 | 37.95 | 39.09 | 37.69 | 37.75 | 37.01 |  |
| 8 | 38.06 | 38.98 | 38.05 | 37.96 | 37.64 |  |
| 9 | 38.78 | 39.12 | 38.84 | 38.69 | 38.07 |  |
| 10 | 39.20 | 40.00 | 39.20 | 38.63 | 38.22 |  |
| 11 | 39.42 | 39.00 | 39.46 | 38.63 | 38.69 |  |
| 12 | 40.02 | 41.00 | 40.16 | 39.47 | 39.03 |  |
| 13 | 40.27 |  | 41.20 | 39.57 | 39.33 |  |
| 14 | 43.31 |  | 44.45 | 39.63 | 39.46 |  |
| 15 |  |  | 39.59 | 40.00 |  |  |

Table 8.5.2. NE Atlantic Mackerel. Mean weight-at-age in the stock for 2022 per area and for the whole NEA mackerel stock

|  | North Sea area | Western area | Southern area | NEA Mackerel $2022$ |
| :---: | :---: | :---: | :---: | :---: |
| Age |  |  |  | Weighted mean* |
| 0 |  |  |  | 0.000 |
| 1 |  | 0.123 | 0.109 | 0.121 |
| 2 | 0.184 | 0.202 | 0.157 | 0.194 |
| 3 | 0.282 | 0.244 | 0.193 | 0.239 |
| 4 | 0.346 | 0.305 | 0.296 | 0.306 |
| 5 | 0.399 | 0.335 | 0.340 | 0.340 |
| 6 | 0.392 | 0.354 | 0.365 | 0.358 |
| 7 | 0.425 | 0.396 | 0.391 | 0.398 |
| 8 | 0.442 | 0.407 | 0.400 | 0.408 |
| 9 | 0.492 | 0.426 | 0.428 | 0.431 |
| 10 | 0.482 | 0.435 | 0.436 | 0.439 |
| 11 | 0.491 | 0.434 | 0.460 | 0.442 |
| 12+ | 0.494 | 0.484 | 0.477 | 0.484 |
| Area Weighting | 7.1\% | 78.6\% | 14.3\% |  |
| Number of fish sampled | 148 | 1224 | 559 |  |

[^6]Table 8.5.3. NE Atlantic Mackerel. Mean proportion of mature fish-at-age for 2022 per area and for the whole NEA mackerel stock

| Age | North Sea | Western | Southern | NEA |
| :--- | :--- | :--- | :--- | :--- |
| area | area | 0 | 0 |  |
| 0 | 0 | 0 | 0.02 | 0.13 |
| 1 | 0 | 0.162 | 0.54 | 0.61 |
| 2 | 0.37 | 1 | 0.970 | 0.70 |
| 4 | 1 | 1 | 1 | 1 |
| 5 | 1 | 1 | 1 | 1 |
| 7 | 1 | 1 | 1 | 1 |

Table 8.6.1.1.1 Total annual egg production and SSB for the western and southern areas and for combined area. Temporal series

| Area | Year | TAEP ( ${ }^{*} 10^{12}$ ) | SSB (kt) |
| :---: | :---: | :---: | :---: |
| Combined | 1992 | 2570 | 3874 |
| Combined | 1995 | 2230 | 3766 |
| Combined | 1998 | 2020 | 4199 |
| Combined | 2001 | 1670 | 3234 |
| Combined | 2004 | 1500 | 3107 |
| Combined | 2007 | 1770 | 3783 |
| Combined | 2010 | 2380 | 4811 |
| Combined | 2013 | 2700 | 4832 |
| Combined | 2016 | 1770 | 3524 |
| Combined | 2019 | 1640 | 3088 |
| Combined | 2022 | 2090 | 3563 |
| Southern | 1992 | 336 | 507 |
| Southern | 1995 | 186 | 370 |
| Southern | 1998 | 479 | 883 |
| Southern | 2001 | 318 | 417 |
| Southern | 2004 | 138 | 309 |
| Southern | 2007 | 348 | 745 |
| Southern | 2010 | 459 | 926 |
| Southern | 2013 | 506 | 904 |
| Southern | 2016 | 225 | 447 |
| Southern | 2019 | 423 | 797 |
| Southern | 2022 | 293 | 499 |
| Western | 1992 | 2230 | 3367 |
| Western | 1995 | 2050 | 3396 |
| Western | 1998 | 1540 | 3316 |
| Western | 2001 | 1350 | 2816 |
| Western | 2004 | 1360 | 2798 |
| Western | 2007 | 1420 | 3038 |


| Area | Year | TAEP (*10 ${ }^{\mathbf{1 2}} \mathbf{)}$ | SSB (kt) |
| :--- | :--- | :--- | :--- |
| Western | 2010 | 1920 | 3884 |
| Western | 2013 | 2200 | 3928 |
| Western | 2016 | 1550 | 3077 |
| Western | 2019 | 1220 | 2291 |
| Western | 2022 | 1800 | 3064 |

Table 8.6.1.2.1 Time series of estimated fecundity parameters for adults

| Parameter | Y1998 | Y2001 | Y2004 | Y2007 | Y2010 | Y2013 | Y2016 | Y2019 | Y2022 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fecundity samples (n) | 96 | 187 | 205 | 176 | 74 | 132 | 97 | 62 | 396 |
| Prevalence of atresia (n) | 112 | 290 | 348 | 416 | 511 | 735 | 713 | 252 | 243 |
| Intensity of atresia (n) | 112 | 290 | 348 | 416 | 511 | 56 | 66 | 64 | 74 |
| Relative potential fecundity ( $\mathrm{n} / \mathrm{g}$ ) | 1206 | 1097 | 1127 | 1098 | 1140 | 1257 | 1159 | 1191 | 1313 |
| Prevalence of atresia | 0.55 | 0.2 | 0.28 | 0.38 | 0.33 | 0.22 | 0.3 | 0.28 | 0.28 |
| Geometric mean intensity of atresia (n/g | 46 | 40 | 33 | 30 | 26 | 27 | 30 | 19 | 20 |
| Potential fecundity lost per day ( $\mathrm{n} / \mathrm{g}$ ) | 3.37 | 1.07 | 1.25 | 1.48 | 1.16 | 0.8 | 1.2 | 0.71 | 0.75 |
| Potential fecundity lost ( $\mathrm{n} / \mathrm{g}$ ) | 202 | 64 | 75 | 89 | 70 | 48 | 72 | 43 | 45 |
| Relative potential fecundity lost (\%) | 17 | 6 | 7 | 9 | 6 | 4 | 6 | 4 | 3 |
| Realised fecundity ( $\mathrm{n} / \mathrm{g}$ ) | 1002 | 1033 | 1052 | 1009 | 1070 | 1209 | 1087 | 1147 | 1268 |

Table 8.6.1.4.1. Daily egg production estimate for mackerel (stage 1a)(PO) and spawning area in the North Sea in 2022.

| Year | PO <br> $($ eggs/m2/day) | Spawning area <br> $(\mathbf{k m 2})$ | Ptot <br> $\left(\right.$ eggs/day)(*10 ${ }^{\mathbf{1 3})}$ |
| :--- | :--- | :--- | :--- |
| 2022 | 18.6 | 371126 | 0.69 |

Table 8.6.1.4.2. Comparison of Total Daily Egg production (Ptot) between 2022 and 2021 in the North Sea.

|  | 2022 | 2021 |
| :--- | :--- | :--- |
| Ptot ${ }^{*} 10^{13}$ | 0.69 | 1.28 |

Table 8.6.2.1 Parameter estimates and standard errors used in the final model.

| Parameter | Estimate | Std error |
| :--- | :--- | :--- |
| T | 2 | 0.33 |
| H $(\log ($ WingSpread $[m] / 1852))$ | 452.9 | 79.64 |
| Log Nugget | -1 | 0.56 |
| Variance (YS) | 3.7 | NA |
| Variance (IS) | 5.6 | NA |

Table 8.6.2.2 Time series of spatially integrated recruitment index values of mackerel.

| Cohort | Index |
| :---: | :---: |
| 1998 | 0.010865 |
| 1999 | 0.016203 |
| 2000 | 0.011533 |
| 2001 | 0.017867 |
| 2002 | 0.022812 |
| 2003 | 0.011024 |
| 2004 | 0.025715 |
| 2005 | 0.033349 |
| 2006 | 0.030347 |
| 2007 | 0.019732 |
| 2008 | 0.017922 |
| 2009 | 0.012704 |
| 2010 | 0.019502 |
| 2011 | 0.032806 |
| 2012 | 0.023792 |
| 2013 | 0.026081 |
| 2014 | 0.019404 |
| 2015 | 0.021442 |
| 2016 | 0.037628 |
| 2017 | 0.038159 |
| 2018 | 0.033926 |
| 2019 | 0.038053 |
| 2020 | 0.031681 |
| 2021 | 0.025708 |

Table 8.7.1.1. NE Atlantic mackerel. Input data and parameters and the model configurations for the assessment.
Input data types and characteristics:

| Name |  | Year range |  | Age range | Variable from year to year |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Catch in tonnes |  | 1980-2022 | Yes |  |  |  |
| Catch-at-age in numbers |  | 1980-2022 |  | 0-12+ | Yes |  |
| Weight-at-age in the comm | mercial catch | 1980-2022 |  | 0-12+ | Yes |  |
| Weight-at-age of the spaw spawning time. | nning stock at | 1980-2022 |  | 0-12+ | Yes |  |
| Proportion of natural mor spawning | tality before | 1980-2022 |  | 0-12+ | Yes |  |
| Proportion of fishing mor spawning | tality before | 1980-2022 |  | 0-12+ | Yes |  |
| Proportion mature-at-age |  | 1980-2022 |  | 0-12+ | Yes |  |
| Natural mortality |  | 1980-2022 |  | 0-12+ | No, fixed at 0.15 |  |
| Tuning data: |  |  |  |  |  |  |
| Type | Name |  | Year range |  |  | Age range |
| Survey (SSB) | ICES Triennial Mackerel and Horse Mackerel Egg Survey |  | $\begin{aligned} & \text { 1992, 1995, 1998, 2001, 2004, 2007, } \\ & 2010,2013,2016,2019,2022 . \end{aligned}$ |  |  | Not applicable (gives SSB) |
| Survey <br> (abundance index) | IBTS Recruitment index (log transformed) |  | 1998-2020, 2022 |  |  | Age 0 |
| Survey <br> (abundance index) | International Ecosystem Summer Survey in the Nordic Seas (IESSNS) |  | 2010, 2012-2023 |  |  | Ages 3-11 |
| Tagging/recapture | Norwegian tagging program |  | Steal tags : 1980 (release year)-2006 (recapture years) <br> RFID tags : 2013 (release year) 2022 <br> (recapture year) |  |  | Ages 5 and older (age at release) |

## SAM parameter configuration :

| Setting | Value | Description |
| :---: | :---: | :---: |
| Coupling of fishing mortality states | 1/2/3/4/5/6/7/8/8/8/8/8/8 | Different $F$ states for ages 0 to 6, one same F state for ages 7 and older |
| Correlated random walks for the fishing mortalities | 0 | F random walk of different ages are independent |
| Coupling of catchability parameters | 0/0/0/0/0/0/0/0/0/0/0/0/0 <br> 1/0/0/0/0/0/0/0/0/0/0/0/0 <br> 2/0/0/0/0/0/0/0/0/0/0/0/0 <br> 0/0/0/3/4/5/6/7/8/9/10/10/0 | No catchability parameter for the catches One catchability parameter estimated for the egg <br> One catchability parameter estimated for the recruitment index <br> One catchability parameter for each age group estimated for the IESSNS (age 3 to11) |
| Power law model | 0 | No power law model used for any of the surveys |
| Coupling of fishing mortality random walk variances | 1/2/3/3/3/3/3/3/3/3/3/3/3 1-1 | Separate F random walk variances for age 0 , age 1 and a same variance for older ages |
| Coupling of log abundance random walk variances | 1/2/2/2/2/2/2/2/2/2/2/2/2 | Same variance used for the $\log$ abundance random walk of all ages except for the recruits (age 0) |
| Coupling of the observation variances | 1/2/3/3/3/3/3/3/3/3/3/3/3 0/0/0/0/0/0/0/0/0/0/0/0/0 4/0/0/0/0/0/0/0/0/0/0/0/0 | Separate observation variances for age 0 and 1 than for the older ages in the catches One observation variance for the egg survey |

\(\left.$$
\begin{array}{lll} & \text { 0/0/0/5/6/6/6/6/6/6/6/6/0 } & \begin{array}{l}\text { One observation variance for the recruitment } \\
\text { index }\end{array}
$$ <br>
2 observation variances for the IESSNS (age 3 <br>

and ages 4 and older)\end{array}\right\}\)| No stock-recruiment model |
| :--- |
| Stock recruitment model |
| Correlation structure |$\quad 0 \quad$ "ID", "ID", "ID", "AR" $\quad$| Auto-regressive correlation structure for the |
| :--- |
| IESSNS index, independent observations |
| assumed for the other data sources |

Table 8.7.1.2. NE Atlantic Mackerel. CATCH IN NUMBER

|  | year |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 0 | 33101 | 56682 | 11180 | 7333 | 287287 | 81799 | 49983 | 7403 | 57644 | 65400 |
| 1 | 411327 | 276229 | 213936 | 47914 | 31901 | 268960 | 58126 | 40126 | 152656 | 64263 |
| 2 | 393025 | 502365 | 432867 | 668909 | 86064 | 20893 | 424563 | 156670 | 137635 | 312739 |
| 3 | 64549 | 231814 | 472457 | 433744 | 682491 | 58346 | 38387 | 663378 | 190403 | 207689 |
| 4 | 328206 | 32814 | 184581 | 373262 | 387582 | 445357 | 76545 | 56680 | 538394 | 167588 |
| 5 | 254172 | 184867 | 26544 | 126533 | 251503 | 252217 | 364119 | 89003 | 72914 | 362469 |
| 6 | 142978 | 173349 | 138970 | 20175 | 98063 | 165219 | 208021 | 244570 | 87323 | 48696 |
| 7 | 145385 | 116328 | 112476 | 90151 | 22086 | 62363 | 126174 | 150588 | 201021 | 58116 |
| 8 | 54778 | 125548 | 89672 | 72031 | 61813 | 19562 | 42569 | 85863 | 122496 | 111251 |
| 9 | 130771 | 41186 | 88726 | 48668 | 47925 | 47560 | 13533 | 34795 | 55913 | 68240 |
| 10 | 39920 | 146186 | 27552 | 49252 | 37482 | 37607 | 32786 | 19658 | 20710 | 32228 |
| 11 | 56210 | 31639 | 91743 | 19745 | 30105 | 26965 | 22971 | 25747 | 13178 | 13904 |
| 12 | 104927 | 199615 | 156121 | 132040 | 69183 | 97652 | 81153 | 63146 | 57494 | 35814 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 24246 | 10007 | 43447 | 19354 | 25368 | 14759 | 37956 | 36012 | 61127 | 67003 |
| 1 | 140534 | 58459 | 83583 | 128144 | 147315 | 81529 | 119852 | 144390 | 99352 | 73597 |
| 2 | 209848 | 212521 | 156292 | 210319 | 221489 | 340898 | 168882 | 186481 | 229767 | 132994 |
| 3 | 410751 | 206421 | 356209 | 266677 | 306979 | 340215 | 333365 | 238426 | 264566 | 223639 |
| 4 | 208146 | 375451 | 266591 | 398240 | 267420 | 275031 | 279182 | 378881 | 323186 | 261778 |
| 5 | 156742 | 188623 | 306143 | 244285 | 301346 | 186855 | 177667 | 246781 | 361945 | 281041 |
| 6 | 254015 | 129145 | 156070 | 255472 | 184925 | 197856 | 96303 | 135059 | 207619 | 244212 |
| 7 | 42549 | 197888 | 113899 | 149932 | 189847 | 142342 | 119831 | 84378 | 118388 | 159019 |
| 8 | 49698 | 51077 | 138458 | 97746 | 106108 | 113413 | 55812 | 66504 | 72745 | 86739 |
| 9 | 85447 | 43415 | 51208 | 121400 | 80054 | 69191 | 59801 | 39450 | 47353 | 50613 |
| 10 | 33041 | 70839 | 36612 | 38794 | 57622 | 42441 | 25803 | 26735 | 24386 | 30363 |
| 11 | 16587 | 29743 | 40956 | 29067 | 20407 | 37960 | 18353 | 13950 | 16551 | 17048 |
| 12 | 27905 | 52986 | 68205 | 68217 | 57551 | 39753 | 30648 | 24974 | 22932 | 32446 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 0 | 36345 | 26034 | 70409 | 14744 | 11553 | 12426 | 75651 | 19302 | 25886 | 17615 |
| 1 | 102407 | 40315 | 222577 | 187997 | 31421 | 46840 | 149425 | 88439 | 59899 | 36514 |
| 2 | 142898 | 158943 | 70041 | 275661 | 453133 | 135648 | 173646 | 190857 | 167748 | 113574 |
| 3 | 275376 | 234186 | 367902 | 91075 | 529753 | 668588 | 159455 | 220575 | 399086 | 455113 |
| 4 | 390858 | 297206 | 350163 | 295777 | 147973 | 293579 | 470063 | 215655 | 284660 | 616963 |
| 5 | 295516 | 309937 | 262716 | 235052 | 258177 | 120538 | 195594 | 455131 | 260314 | 319465 |
| 6 | 241550 | 231804 | 237066 | 183036 | 145899 | 121477 | 97061 | 203492 | 255675 | 224848 |
| 7 | 175608 | 195250 | 151320 | 133595 | 89856 | 63612 | 73510 | 77859 | 124382 | 194326 |
| 8 | 106291 | 120241 | 118870 | 94168 | 65669 | 38763 | 33399 | 59652 | 57297 | 73171 |

```
    9 523994 72205 79945 75701 40443 (23947 18961 30494 32343 
```



```
    18918 20546 21611 25797 16430 7955 8334 11416 6798 
    34202 40706 40280
        year
age 2010 2011 2012 2013 2014 2015 2016 2017 2018 2019
```



```
    78636
    137351 115701 204121 235494 138880 127940 205840 89101 236207 126073
    304647 323847 216711 400036 672022 250575 258176 461621 136779 350611
    740816 471564 417953 371713 832975 583694 427212 353230 376312 114606
    613418 656507458718445515 568835 651786 593046 398273 257069 295731
    285438490219 514489433533 554367 453084 534943 505073 294539 226640
    143537 244725 325982 340686 506804 416897 341408 432242 424715 229725
    102446 113277 143643 190660 341618 356936 270586 262799 316779 267491
    45963 53512 69962 113220 142398 206045 170574 189449 197761 204818
    21268}225081 30761 46269 63871 107830 94849 138347 140403 102991
    6272 12322 11657 19025
```



```
age 2020 2021 2022
    6443 2332 2410
    52637 29202 8215
    107302 326976 240690
    182163217298 270019
    266760 281925 177836
    166627 366644 237153
    270154 182783 291996
    246268 300014 178966
    274182 208961 257886
    3 1 1 2 1 5 2 2 8 2 3 6 ~ 1 9 3 0 2 5
    241775 198134 177568
    1 2 8 2 9 4 1 2 8 9 5 7 1 8 5 3 5 4
    179703 118150 290793
```

Table 8.7.1.3. NE Atlantic Mackerel. WEIGHTS AT AGE IN THE CATCH

```
Units : Kg
    year
age 1980
    0}00.057 0.060 0.053 0.050 0.031 0.055 0.039 0.076 0.055 0.049 0.085 0.068
    1}00.131 0.132 0.131 0.168 0.102 0.144 0.146 0.179 0.133 0.136 0.156 0.156
    2 0.249 0.248}0.249 0.219 0. 184 0.262 0.245 0.223 0.259 0.237 0.233 0.253
    3 0.285 0.287 0.285 0.276 0.295 0.357 0.335 0.318 0.323 0.320 0.336 0.327
    4 0.345 0.344 0.345 0.310}00.326 0.418 0.423 0.399 0.388 0.377 0.379 0.394
    5 0.378 0.377 0.378 0.386 0.344 0.417 0.471 0.474 0.456 0.433 0.423 0.423
    6
    70.498 0.499 0.496 0.435 0.542 0.521 0.457 0.493 0.555 0.543 0.528 0.506
```




```
    10 0.574 0.573 0.574 0.606 0.628 0.629 0.552 0.634 0.613 0.581 0.606 0.630
    11 0.590 0.576 0.574 0.608 0.636 0.679 0.694 0.635 0.624 0.648 0.591 0.649
    12 0.580 0.584 0.582 0.614 0.663 0.710 0.688 0.718 0.697 0.739 0.713 0.708
```

```
    year
age 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003
    0 0.051 0.061 0.046 0.072 0.058 0.076 0.065 0.062 0.063 0.069 0.052 0.081
    100.167 0.134 0.136 0.143 0.143 0.143 0.157 0.176 0.135 0.172 0.160 0.170
    2 0.239 0.240 0.255 0.234 0.226 0.230 0.227 0.235 0.227 0.224 0.256 0.267
    0.333 0.317 0.339 0.333 0.313 0. 295 0.310 0.306 0.306 0.305 0.307 0.336
    0.397 0.376 0.390 0.390 0.377 0.359 0.354 0.361 0.363 0.376 0.368 0.385
    llllllllllllllllll
    0.495 0.483 0.512 0.501 0.484 0.453 0.452 0.452 0.463 0.474 0.461 0.477
    lllllllllllllllllllll
    0.555 0.548 0.590 0.577 0.551 0.524 0.518 0.536 0.534 0.540}00.536 0.572
```



```
    10}00.651 0.595 0.627 0.606 0.596 0.577 0.573 0.586 0.586 0.603 0.600 0.631
    11
    12 0.669 0.679 0.713 0.672 0.670}0.6.636 0.631 0.687 0.644 0.666 0.665 0.715
        year
age 2004 2005 2006 2007 2008 2009 2010 2011 2012 2013 2014 2015
    0.067 0.048 0.038 0.089 0.051 0.104 0.048 0.029 0.089 0.091 0.043 0.051
    0.156 0.151 0.071 0.120 0.105 0.153 0.118 0.113 0.123 0.173 0.126 0.154
    llllllllllllllllllll
    0.323 0.306 0.307 0.292 0.292 0.283 0.291 0.282 0.284 0.277 0.282 0.294
    0.400 0.366 0.357 0.372 0.370 0.331 0.331 0.334 0.340 0.336 0.324 0.320
    0.419 0.434 0.428 0.408 0.418 0.389 0.365 0.368 0.374 0.360 0.362 0.351
    0.485 0.440 0.479 0.456 0.444 0.424 0.418 0.411 0.401 0.386 0.394 0.392
    0.519 0.496 0.494 0.512 0.497 0.450 0.470 0.451 0.431 0.405 0.422 0.420
    0.554 0.539 0.543 0.534 0.551 0.497 0.487 0.494 0.469 0.431 0.443 0.443
    0.573 0.556 0.584 0.573 0.571 0.538 0.515 0.540 0.503 0.454 0.467 0.465
    0.595 0.583 0.625 0.571 0.620 0.586 0.573 0.580 0.537 0.472 0.482 0.489
    0.630 0.632 0.636 0.585 0.595 0.599 0.603 0.611 0.537 0.493 0.523 0.522
    0.684 0.655 0.689 0.666 0.662 0.630 0.630 0.664 0.585 0.554 0.589 0.561
```

        year
    age $201620172018 \quad 2019 \quad 2020 \quad 20212022$
$0.0350 .018 \quad 0.066 \quad 0.057 \quad 0.057 \quad 0.049 \quad 0.075$
$\begin{array}{lllllll}0.154 & 0.178 & 0.147 & 0.112 & 0.174 & 0.163 & 0.144\end{array}$
$0.240 \quad 0.266 \quad 0.247 \quad 0.260 \quad 0.285 \quad 0.277 \quad 0.271$
$0.2970 .3110 .320 \quad 0.297 \quad 0.322 \quad 0.338 \quad 0.316$
$\begin{array}{lllllll}0.329 & 0.356 & 0.355 & 0.360 & 0.360 & 0.374 & 0.345\end{array}$
$\begin{array}{llllllll}0.356 & 0.377 & 0.397 & 0.388 & 0.389 & 0.406 & 0.378\end{array}$
$\begin{array}{llllllll}0.383 & 0.397 & 0.410 & 0.429 & 0.417 & 0.441 & 0.413\end{array}$
$\begin{array}{lllllll}0.411 & 0.415 & 0.426 & 0.441 & 0.444 & 0.457 & 0.434\end{array}$
$\begin{array}{lllllll}0.438 & 0.444 & 0.446 & 0.453 & 0.459 & 0.477 & 0.456\end{array}$
$\begin{array}{lllllll}0.453 & 0.465 & 0.469 & 0.472 & 0.471 & 0.486 & 0.473\end{array}$
$\begin{array}{llllllll}10 & 0.479 & 0.484 & 0.492 & 0.497 & 0.495 & 0.501 & 0.488\end{array}$
$\begin{array}{llllllll}1 & 0.499 & 0.497 & 0.507 & 0.514 & 0.519 & 0.514 & 0.507\end{array}$
$\begin{array}{llllllll}2 & 0.520 & 0.531 & 0.537 & 0.537 & 0.554 & 0.548 & 0.538\end{array}$

## Table 8.7.1.4. NE Atlantic Mackerel. WEIGHTS AT AGE IN THE STOCK

```
Units : Kg
    year
age 1980
    0.063 0.063 0.063 0.063 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    0.114 0.112 0.112 0.111 0.108 0.111 0.104 0.075 0.099 0.058 0.096 0.174
    0.205 0.179 0.159 0.179 0.204 0.244 0.184 0.157 0.181 0.162 0.166 0.184
    0.287 0.258 0.217 0.233 0.251 0.281 0.269 0.234 0.238 0.230}0.24,2470.243
    0.322 0.312 0.300 0.282 0.293 0.308 0.301 0.318 0.298 0.272 0.290 0.303
    0.356 0.335 0.368 0.341 0.326 0.336 0.350 0.368 0.348 0.338 0.332 0.347
    0.377 0.376 0.362 0.416 0.395 0.356 0.350 0.414 0.392 0.392 0.383 0.392
    0.402 0.415 0.411 0.404 0.430}0.4070.474 0.415 0.445 0.388 0.435 0.423
    0.434 0.431 0.456 0.438 0.455 0.455 0.434 0.431 0.442 0.449 0.447 0.492
    9}00.438 0.454 0.455 0.475 0.489 0.447 0.428 0.483 0.466 0.432 0.494 0.500
```



```
    11 0.520 0.524 0.536 0.544 0.513 0.538 0.506 0.492 0.567 0.482 0.495 0.526
    12 0.532 0.530 0.542 0.528 0.566 0.590 0.541 0.581 0.594 0.556 0.536 0.619
        year
age 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003
    0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    0.130}0.1450.114 0.116 0.097 0.084 0.083 0.087 0.093 0.113 0.109 0.112
    0.201 0.190 0.163 0.200 0.185 0.196 0.170 0.210 0.194 0.190 0.206 0.181
    0.260 0.266 0.240 0.278 0.250 0.257 0.251 0.260 0.253 0.246 0.245 0.251
    0.308 0.323 0.306 0.327 0.322 0.310 0.300 0.317 0.301 0.303 0.288 0.277
    0.360 0.359 0.368 0.385 0.372 0.356 0.348 0.356 0.357 0.342 0.333 0.341
    0.397 0.410 0.418 0.432 0.425 0.401 0.384 0.392 0.394 0.398 0.360 0.401
    0.419 0.432 0.459 0.458 0.446 0.460 0.409 0.424 0.415 0.417 0.418 0.407
    lllllllllllllllll
    llllllllllllllllllll
    lllllllllllllllllllll
    110.543 0.547 0.592 0.560 0.538}0.5.546 0.500 0.545 0.514 0.535 0.523 0.521
    12 0.572 0.580 0.608 0.603 0.573 0.583 0.549 0.575 0.551 0.572 0.558 0.540
        year
age 2004 2005 2006 2007 2008 2009 2010 2011 2012 2013 2014 2015
    0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    0.112 0.114 0.114 0.095 0.133 0.112 0.096 0.080}0.0.089 0.076 0.107 0.078
```



```
    0.258 0.221 0.236 0.206 0.207 0.214 0.199 0.223 0.216 0.179 0.199 0.247
    0.319 0.328 0.291 0.285 0.260 0.268 0.246 0.274 0.255 0.249 0.238}00.254
    0.356 0.378 0.333 0.329 0.346 0.295 0.296 0.332 0.288 0.280}0.3910.288
    0.406 0.403 0.400 0.363 0.354 0.351 0.345 0.369 0.312 0.319 0.321 0.336
    0.449 0.464 0.413 0.448 0.393 0.386 0.389 0.389 0.360 0.341 0.341 0.350
    0.482 0.481 0.437 0.452 0.448}0.4.437 0.407 0.430 0.390 0.375 0.387 0.381
    lllllllllllllllllllll
    0}0.5190.538 0.469 0.538 0.478 0.517 0.489 0.495 0.498 0.441 0.466 0.447
    0.579 0.509 0.531 0.542 0.487 0.548 0.532 0.518 0.503 0.496 0.472 0.485
    20.588}0.6030.566 0.585 0.510 0.557 0.572 0.525 0.558 0.522 0.517 0.551
        year
age 2016 2017 2018 2019 2020 2021 2022
    0.000 0.000 0.000 0.000 0.000 0.000 0.000
    0.059 0.058 0.064 0.070 0.104 0.134 0.121
    0.182 0.204 0.190 0.191 0.209 0.186 0.194
    3 0.238 0.237 0.266 0.250 0.252 0.261 0.239
```

[^7]
## Table 8.7.1.5. NE Atlantic Mackerel. NATURAL MORTALITY

```
Units : NA
    year
```




```
    1 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    2 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    3 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    4 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    5 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    6 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    7 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    8 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    9 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    10 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    11 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    12 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
        year
age 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005 2006 2007 2008 2009
```




```
    2 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    3 0.15
    4 0.15
    5 0.15
    6}00.1
```



```
    8
```






```
        year
age 2010 2011 2012 2013 2014 2015 2016 2017 2018 2019 2020 2021 2022
    0}00.1
```








```
    7}0.1
```




```
    10}00.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    11}00.150.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
    12 0.15}0.150.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15 0.15
```


## Table 8.7.1.6. NE Atlantic Mackerel. PROPORTION MATURE

```
ear
age 19080 1981 1982 1983
    0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    0.093 0.097 0.097 0.098 0.102 0.102 0.102 0.102 0.102 0.102 0.102 0.102
    0.521 0.497 0.498 0.485 0.467 0.516 0.522 0.352 0.360 0.372 0.392 0.435
    0.872 0.837 0.857 0.863 0.853 0.885 0.926 0.922 0.901 0.915 0.909 0.912
    0.949 0.934 0.930 0.940 0.938 0.940 0.983 0.994 0.989 0.994 0.996 0.991
    0.972 0.976 0.969 0.972 0.966 0.966 0.965 0.997 0.994 0.996 0.998 0.996
    0.984 0.984 0.987 0.999 1.000 1.000 1.000 1.000 1.000 1.000 1.000 0.996
    0.990 0.987 0.985 0.984 0.975 0.976 1.000 1.000 1.000 1.000 1.000 1.000
    1.000 0.999 0.999 0.999 0.999 0.999 0.991 0.992 0.991 0.993 0.995 1.000
    1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    2 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
        year
age 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003
    0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    0.102 0.102 0.102 0.102 0.102 0.097 0.097 0.097 0.104 0.104 0.104 0.106
    0.520 0.534 0.621 0.599 0.586 0.621 0.688 0.669 0.692 0.675 0.710 0.690
    0.928 0.934 0.938 0.931 0.936 0.880 0.886 0.876 0.909 0.909 0.937 0.940
    0.996 0.996 0.994 0.993 1.000 0.993 0.994 0.989 0.989 0.987 0.992 0.988
    0.997 0.997 0.997 0.994 1.000 0.998 0.999 0.999 0.998 0.998 1.000 1.000
    0.994 0.994 0.993 0.987 0.994 0.999 0.999 0.999 0.999 0.999 1.000 1.000
    1.000 1.000 0.999 0.999 0.999 1.000 1.000 1.000 1.000 0.999 1.000 0.999
    1.000 1.000 1.000 1.000 1.000 0.994 0.995 0.996 0.997 0.997 1.000 1.000
    1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
        year
        2004 2005 2006 2007 2008 2009 2010 2011 2012 2013 2014 2015
        0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
        0.106 0.106 0.095 0.095 0.095 0.096 0.096 0.096 0.094 0.092 0.092 0.104
        0.761 0.616 0.589 0.546 0.524 0.541 0.667 0.655 0.604 0.683 0.675 0.763
        0.962 0.959 0.928 0.921 0.917 0.919 0.930}0.9.927 0.926 0.921 0.916 0.944
        0.993 0.993 0.994 0.994 0.999 0.999 0.999 0.999 0.999 0.998 0.999 0.998
        0.999 0.999 1.000 1.000 0.999 1.000 1.000 1.000 0.999 1.000 1.000 0.999
        1.000 1.000 1.000 1.000 1.000 1.000 0.999 0.999 0.999 0.999 0.999 1.000
        0.999 0.999 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 0.999 0.999
        1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
        1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
    1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000 1.000
        year
age 2016 2017 2018 2019 2020 2021 2022
    0.000 0.000 0.000 0.000 0.000 0.000 0.000
    0.138 0.136 0.114 0.114 0.114 0.130 0.130
    0.633 0.606 0.464 0.534 0.559 0.591 0.608
    0.937 0.945 0.902 0.910 0.913 0.940 0.934
    0.997 0.998 0.998 0.999 0.998 0.998 0.998
```

[^8]
## Table 8.7.1.7. NE Atlantic Mackerel. FRACTION OF HARVEST BEFORE SPAWNING

```
    year
```



```
    0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    0.166 0.166 0.166 0.166 0.166 0.166 0.166 0.166 0.166 0.166 0.139 0.111
    0.209 0.209 0.209 0.209 0.209 0.209 0.209 0.209 0.209 0.209 0.240 0.272
    0.209 0.209 0.209 0.209 0.209 0.209 0.209 0.209 0.209 0.209 0.240 0.272
    0.209 0.209 0.209 0.209 0.209 0.209 0.209 0.209 0.209 0.209 0.240 0.272
    0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.393 0.406
    0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.393 0.406
    0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.393 0.406
    0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.393 0.406
    0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.393 0.406
    0 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.393 0.406
    0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.393 0.406
    20.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.380 0.393 0.406
        year
age 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003
    0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
    0.084 0.165 0.249 0.331 0.269 0.206 0.144 0.125 0.106 0.088 0.142 0.197
    0.304 0.301 0.298 0.296 0.295 0.295 0.295 0.320 0.347 0.373 0.360 0.347
    0.304 0.301 0.298 0.296 0.295 0.295 0.295 0.320 0.347 0.373 0.360 0.347
    0.304 0.301 0.298 0.296 0.295 0.295 0.295 0.320 0.347 0.373 0.360 0.347
    0.419 0.444 0.469 0.494 0.494 0.494 0.495 0.461 0.426 0.392 0.408 0.425
    0.419 0.444 0.469 0.494 0.494 0.494 0.495 0.461 0.426 0.392 0.408 0.425
    0.419 0.444 0.469 0.494 0.494 0.494 0.495 0.461 0.426 0.392 0.408 0.425
    0.419 0.444 0.469 0.494 0.494 0.494 0.495 0.461 0.426 0.392 0.408 0.425
    0.419 0.444 0.469 0.494 0.494 0.494 0.495 0.461 0.426 0.392 0.408 0.425
    0.419 0.444 0.469 0.494 0.494 0.494 0.495 0.461 0.426 0.392 0.408 0.425
    0.419 0.444 0.469 0.494 0.494 0.494 0.495 0.461 0.426 0.392 0.408 0.425
    2 0.419 0.444 0.469 0.494 0.494 0.494 0.495 0.461 0.426 0.392 0.408 0.425
        year
        2004 2005 2006 2007 2008 2009 2010 2011 2012 2013 2014 2015
        0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000 0.000
        0.251 0.262 0.274 0.285 0.206 0.125 0.047 0.092 0.138 0.183 0.170}00.156
        0.334 0.317 0.300 0.284 0.266 0.249 0.232 0.176 0.119 0.064 0.117 0.171
        0.334 0.317 0.300 0.284 0.266 0.249 0.232 0.176 0.119 0.064 0.117 0.171
        0.334 0.317 0.300 0.284 0.266 0.249 0.232 0.176 0.119 0.064 0.117 0.171
        0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
        0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
        0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
        0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
        0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
    0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
    0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
    0.441 0.409 0.376 0.344 0.310 0.275 0.242 0.233 0.225 0.216 0.203 0.189
        year
age
    0
    0.143 0.232 0.393 0.581 0.533 0.187 0.336
    0.224 0.152 0.179 0.183 0.184 0.090 0.131
    0.224 0.152 0.179 0.183 0.184 0.090 0.131
    4 0.224 0.152 0.179 0.183 0.184 0.090 0.131
```

```
5 0.180 0.297 0.197 0.301 0.312 0.227 0.207
```






```
10 0.180 0.297 0.197 0.301 0.312 0.227 0.207
11}00.180 0.297 0.197 0.301 0.312 0.227 0.207
12 0.180 0.297 0.197 0.301 0.312 0.227 0.207
```


## Table 8.7.1.8. NE Atlantic Mackerel. FRACTION OF NATURAL MORTALITY BEFORE SPAWNING

```
    year
age 1980 1981 1982 1983 1984 1984985
    0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388}0.30.378 0.369 0.357 0.345
    0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345
    0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345
    0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345
    0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345
    0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345
    0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345
    0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345
    0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345
    0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345
    10 0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345
    11 0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345
    12 0.397 0.396 0.394 0.392 0.394 0.396 0.397 0.388 0.378 0.369 0.357 0.345
        year
age 1992 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003
    0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    0}0.3330.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
    2 0.333 0.341 0.349 0.357 0.339 0.322 0.304 0.325 0.346 0.366 0.361 0.355
        year
age
    0
    1
    2
    3
    4
    5
    6
    7
    8
    9
    10}00.3500.346 0.342 0.339 0.311 0.283 0. 255 0.252 0.249 0.246 0.278 0.311
    11 0.350 0.346 0.342 0.339 0.311 0. 283 0. 255 0.252 0.249 0.246 0.278 0.311
    12 0.350 0.346 0.342 0.339 0.311 0.283 0.255 0.252 0.249 0.246 0.278 0.311
        year
age 2016 2017 2018 2019 2020 2021 2022
    0.343 0.327 0.312 0.396 0.312 0.329 0.345
    0.343 0.327 0.312 0.396 0.312 0.329 0.345
    0.343 0.327 0.312 0. 096 0.312 0.329 0.345
    0.343 0.327 0.312 0.296 0.312 0.329 0.345
    4 0.343 0.327 0.312 0.296 0.312 0.329 0.345
```

[^9]
## Table 8.7.1.9. NE Atlantic MackereI. SURVEY INDICES



```
10.011024186
1 0.025714749
1 0.033349294
1 0.030346953
1 0.019732337
1 0.017922417
1 0.012704437
1 0.019501867
1 0.03280643
1 0.023792277
1 0.02608089
1 0.019404049
1 0.021441518
1 0.037627731
1 0.038158683
1 0.033925745
1 0.038052761
1 0.03168093
1 -1
1 0.025707856
Swept-idx
2 0 1 0 2 0 2 3
1 -
3
    1 1
```

| 1617005 | 4035646 | 3059146 | 1591100 | 691936 | 413253 | 198106 | 65803 | 24747 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| 1283247 | 2383260 | 2164365 | 2850847 | 1783942 | 740361 | 299490 | 149282 | 84344 |
| 9201746 | 2456618 | 3073772 | 3218990 | 2540444 | 1087937 | 377406 | 144695 | 146826 |
| 7034162 | 4896456 | 2659443 | 2630617 | 2768227 | 1910160 | 849010 | 379745 | 95304 |
| 2539963 | 6409324 | 4802298 | 1795564 | 1628872 | 1254859 | 727691 | 270562 | 72410 |
| 1374705 | 2635033 | 5243607 | 4368491 | 1893026 | 1658839 | 1107866 | 754993 | 450100 |
| 3562908 | 1953609 | 3318099 | 4680603 | 4653944 | 1754954 | 1944991 | 626406 | 507546 |
| 496595 | 2384310 | 1200541 | 1408582 | 2330520 | 1787503 | 1049868 | 499295 | 557573 |
| 3814661 | 1211770 | 2920591 | 2856932 | 1948653 | 3906891 | 3824410 | 1499778 | 1248160 |
| 1430995 | 3361778 | 2134411 | 2528651 | 2525460 | 2032783 | 2904239 | 3835479 | 1495649 |
| 709444 | 1220543 | 1527964 | 367017 | 1291607 | 811226 | 1051955 | 969868 | 927410 |
| 2355905 | 944385 | 1307793 | 1043409 | 598182 | 956129 | 995936 | 1862024 | 1605735 |
| 3539644 | 1703188 | 549158 | 460261 | 786940 | 321450 | 483156 | 387485 | 450229 |

Table 8.7.1.10. NE Atlantic Mackerel. RFID recapture data for the year 2022

| Release Yr | Recapture Yr | Year-class | age at re- <br> lease | Numbers scanned <br> in recapture Yr | Numbers Released <br> in Release Year | Numbers re- <br> captured |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2020 | 2022 | 2009 | 11 | 23221600 | 2191 | 19 |
| 2020 | 2022 | 2010 | 10 | 63969402 | 5001 | 44 |
| 2020 | 2022 | 2011 | 9 | 82262728 | 5081 | 45 |
| 2020 | 2022 | 2012 | 8 | 53631347 | 5474 | 49 |
| 2020 | 2022 | 2013 | 7 | 43891854 | 2665 | 25 |
| 2020 | 2022 | 2014 | 6 | 86709988 | 4339 | 37 |
| 2020 | 2022 | 2015 | 5 | 47572116 | 1509 | 14 |
| 2021 | 2022 | 2010 | 11 | 63969402 | 5810 | 72 |
| 2021 | 2022 | 2011 | 10 | 82262728 | 5039 | 65 |
| 2021 | 2022 | 2012 | 9 | 53631347 | 2815 | 36 |
| 2021 | 2022 | 2013 | 8 | 43891854 | 2588 | 35 |
| 2021 | 2022 | 2014 | 7 | 86709988 | 3970 | 53 |
| 2021 | 2022 | 2015 | 6 | 47572116 | 1384 | 20 |
| 2021 | 2022 | 2016 | 5 | 100784920 | 9976 | 142 |

Table 8.7.2.1. NE Atlantic Mackerel. SAM parameter estimates for the 2023 update.

|  | estimate | std.dev | confidence interval lower bound | confidence interval upper bound |
| :---: | :---: | :---: | :---: | :---: |
| observation standard deviations |  |  |  |  |
| Catches age 0 | 0.87 | 0.17 | 0.61 | 1.23 |
| Catches age 1 | 0.36 | 0.33 | 0.19 | 0.71 |
| Catches age 2-12 | 0.13 | 0.12 | 0.10 | 0.16 |
| Egg survey | 0.35 | 0.24 | 0.22 | 0.57 |
| Recruitment index | 0.27 | 0.29 | 0.15 | 0.48 |
| IESSNS age 3 | 0.60 | 0.22 | 0.38 | 0.94 |
| IESSNS ages 4-11 | 0.41 | 0.12 | 0.32 | 0.52 |
| Recapture overdispersion tags | 1.26 | 0.23 | 1.41 | 1.17 |
| random walk standard deviation |  |  |  |  |
| $F$ age 0 | 0.29 | 0.41 | 0.13 | 0.66 |
| F age 1 | 0.27 | 0.44 | 0.11 | 0.66 |
| F age 2+ | 0.15 | 0.13 | 0.12 | 0.20 |
| N@age0 | 0.19 | 0.58 | 0.06 | 0.62 |
| process error standard deviation |  |  |  |  |
| N@age1-12+ | 0.21 | 0.08 | 0.18 | 0.24 |
| catchabilities |  |  |  |  |
| egg survey | 1.13 | 0.12 | 0.89 | 1.44 |
| recruitment index | 4.39E-09 | 0.11 | 3.50E-09 | 5.50E-09 |
| IESSNS age 3 | 0.81 | 0.20 | 0.54 | 1.21 |
| IESSNS age 4 | 1.14 | 0.15 | 0.84 | 1.55 |
| IESSNS age 5 | 1.50 | 0.15 | 1.10 | 2.04 |
| IESSNS age 6 | 1.46 | 0.16 | 1.07 | 1.99 |
| IESSNS age 7 | 1.63 | 0.16 | 1.19 | 2.23 |
| IESSNS age 8 | 1.55 | 0.16 | 1.12 | 2.12 |
| IESSNS age 9 | 1.65 | 0.16 | 1.20 | 2.26 |
| IESSNS ages 10-11 | 1.60 | 0.16 | 1.17 | 2.18 |
| post tagging survival steal tags | 0.40 | 0.11 | 0.34 | 0.45 |
| post tagging survival RFID tags | 0.18 | 0.11 | 0.15 | 0.21 |

Table 8.7.3.1. NE Atlantic Mackerel. STOCK SUMMARY.

| Year | Recruitment |  |  | SSB |  |  | Total <br> Catch | F |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Low Re-cruitment | Midpoint <br> Recruit- <br> ment | High Re-cruitment | Low SSB | Midpoint | High SSB |  | Low | Mid- <br> point | High |
|  | thousands |  |  | tonnes |  |  | tonnes |  |  |  |
| 1980 | 738556 | 2236661 | 6773556 | 1929717 | 4089931 | 8668385 | 734950 | 0.126 | 0.21 | 0.33 |
| 1981 | 1482481 | 3935162 | 10445665 | 1906223 | 3606692 | 6824085 | 754045 | 0.129 | 0.21 | 0.33 |
| 1982 | 2255093 | 4250320 | 8010857 | 2031684 | 3472009 | 5933427 | 716987 | 0.133 | 0.21 | 0.32 |
| 1983 | 2477846 | 4181532 | 7056617 | 2405542 | 3678958 | 5626478 | 672283 | 0.137 | 0.21 | 0.31 |
| 1984 | 1055106 | 1953170 | 3615633 | 2791467 | 3979779 | 5673949 | 641928 | 0.141 | 0.21 | 0.31 |
| 1985 | 914783 | 1763044 | 3397881 | 2941137 | 4008008 | 5461878 | 614371 | 0.146 | 0.21 | 0.31 |
| 1986 | 2413122 | 4062330 | 6838664 | 2703953 | 3602549 | 4799773 | 602201 | 0.153 | 0.22 | 0.31 |
| 1987 | 1645845 | 2707304 | 4453334 | 2684762 | 3569947 | 4746984 | 654992 | 0.161 | 0.22 | 0.31 |
| 1988 | 1705889 | 2665714 | 4165588 | 2714143 | 3521030 | 4567797 | 680491 | 0.170 | 0.23 | 0.32 |
| 1989 | 2621896 | 3850196 | 5653929 | 2577521 | 3288174 | 4194763 | 585920 | 0.181 | 0.24 | 0.33 |
| 1990 | 1614577 | 2351096 | 3423590 | 2695479 | 3373583 | 4222278 | 626107 | 0.193 | 0.26 | 0.34 |
| 1991 | 1807124 | 2641378 | 3860763 | 2642309 | 3273562 | 4055622 | 675665 | 0.21 | 0.27 | 0.36 |
| 1992 | 1256626 | 1957547 | 3049428 | 2448205 | 3002776 | 3682970 | 760690 | 0.22 | 0.29 | 0.38 |
| 1993 | 1664834 | 2394688 | 3444507 | 2189003 | 2663959 | 3241967 | 824568 | 0.24 | 0.31 | 0.39 |
| 1994 | 1933108 | 2783885 | 4009096 | 1921814 | 2321191 | 2803563 | 819087 | 0.25 | 0.32 | 0.40 |
| 1995 | 1451810 | 2073325 | 2960908 | 1913380 | 2290216 | 2741269 | 756277 | 0.25 | 0.31 | 0.39 |
| 1996 | 1466815 | 2063895 | 2904021 | 1825198 | 2174716 | 2591166 | 563472 | 0.25 | 0.30 | 0.37 |
| 1997 | 1218996 | 1735510 | 2470881 | 1823874 | 2144903 | 2522439 | 573029 | 0.24 | 0.30 | 0.36 |
| 1998 | 1451094 | 2299761 | 3644767 | 1797995 | 2118203 | 2495437 | 666316 | 0.25 | 0.30 | 0.37 |
| 1999 | 1255030 | 1906540 | 2896261 | 1955252 | 2297701 | 2700127 | 640309 | 0.27 | 0.32 | 0.38 |
| 2000 | 1631376 | 2231701 | 3052937 | 1963930 | 2261229 | 2603532 | 738606 | 0.29 | 0.33 | 0.39 |
| 2001 | 1905587 | 2546612 | 3403273 | 1870992 | 2145615 | 2460547 | 737463 | 0.31 | 0.37 | 0.43 |
| 2002 | 858620 | 1155583 | 1555255 | 1776826 | 2053280 | 2372746 | 771422 | 0.33 | 0.39 | 0.46 |
| 2003 | 3661266 | 4783942 | 6250872 | 1727292 | 2006047 | 2329787 | 679287 | 0.34 | 0.41 | 0.48 |
| 2004 | 4904673 | 6577625 | 8821210 | 2201692 | 2609154 | 3092025 | 660491 | 0.31 | 0.37 | 0.45 |
| 2005 | 1760759 | 2399484 | 3269909 | 2032428 | 2433808 | 2914455 | 549514 | 0.26 | 0.30 | 0.36 |
| 2006 | 2549599 | 3480919 | 4752431 | 1927087 | 2296686 | 2737170 | 481181 | 0.24 | 0.28 | 0.33 |
| 2007 | 3664024 | 5095654 | 7086659 | 2064174 | 2441423 | 2887618 | 586206 | 0.26 | 0.31 | 0.36 |
| 2008 | 3714761 | 5197356 | 7271667 | 2419214 | 2891383 | 3455707 | 623165 | 0.24 | 0.29 | 0.35 |
| 2009 | 2691002 | 3749368 | 5223987 | 2998720 | 3599313 | 4320194 | 737969 | 0.22 | 0.26 | 0.31 |


| 2010 | 3016440 | 4193134 | 5828850 | 3401426 | 4057815 | 4840869 | 877272 | 0.21 | 0.25 | 0.30 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2011 | 2556669 | 3556127 | 4946294 | 3957832 | 4736686 | 5668809 | 948963 | 0.20 | 0.24 | 0.29 |
| 2012 | 4281527 | 5939342 | 8239066 | 3751514 | 4500692 | 5399482 | 899551 | 0.181 | 0.22 | 0.27 |
| 2013 | 4942486 | 6867546 | 9542402 | 4203695 | 5057375 | 6084418 | 938299 | 0.176 | 0.22 | 0.26 |
| 2014 | 2765711 | 3834283 | 5315712 | 5239637 | 6303439 | 7583226 | 1401788 | 0.176 | 0.22 | 0.26 |
| 2015 | 2447345 | 3387247 | 4688118 | 5310901 | 6418653 | 7757462 | 1215827 | 0.162 | 0.20 | 0.25 |
| 2016 | 3615580 | 5004656 | 6927402 | 4996244 | 6056742 | 7342338 | 1100135 | 0.147 | 0.183 | 0.23 |
| 2017 | 1523690 | 2120346 | 2950643 | 4830993 | 5859454 | 7106864 | 1159641 | 0.150 | 0.186 | 0.23 |
| 2018 | 3323155 | 4681101 | 6593949 | 4187213 | 5076992 | 6155849 | 1023144 | 0.154 | 0.191 | 0.24 |
| 2019 | 1964768 | 2814162 | 4030761 | 3530572 | 4306452 | 5252841 | 839727 | 0.158 | 0.197 | 0.24 |
| 2020 | 1748546 | 2544660 | 3703247 | 3305920 | 4038542 | 4933521 | 1039513 | 0.199 | 0.25 | 0.31 |
| 2021 | 3700171 | 5528431 | 8260036 | 3398574 | 4186470 | 5157025 | 1081540 | 0.21 | 0.27 | 0.33 |
| 2022 | 3116177 | 4953325 | 7873566 | 3091402 | 3951017 | 5049663 | 1046720 | 0.23 | 0.30 | 0.40 |
| 2023 | 1549418 | 3269823 | 6900492 |  | $3681064+$ |  |  |  |  |  |
| Av- <br> er- <br> age | 2280090 | 3402852 | 5217379 | 2792719 | 3532891 | 4536291 | 783747 | 0.21 | 0.26 | 0.34 |

$\dagger$ Estimated value from the forecast.

## Table 8.7.3.2. NE Atlantic Mackerel. ESTIMATED POPULATION ABUNDANCE

| Unit | s:Thousan |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 |
| 0 | 4993795 | 4604685 | 3797410 | 3627594 | 4124798 | 4019352 | 3998993 | 4091728 | 3716101 | 3544801 |
| 1 | 4756420 | 4674563 | 4449821 | 2755669 | 2556728 | 4152007 | 3336161 | 3299389 | 4020954 | 2991042 |
| 2 | 2236661 | 3935162 | 4250320 | 4181532 | 1953170 | 1763044 | 4062330 | 2707304 | 2665714 | 3850196 |
| 3 | 908772 | 1804561 | 3293937 | 4024415 | 4215998 | 1348926 | 1232760 | 3985118 | 2145255 | 2326513 |
| 4 | 1556855 | 703232 | 1365141 | 2780216 | 3656294 | 3972222 | 1007201 | 840192 | 3691712 | 1666522 |
| 5 | 3392903 | 1162259 | 509463 | 943935 | 2127641 | 2996480 | 3117457 | 788673 | 530559 | 2947938 |
| 6 | 2621608 | 2408066 | 842135 | 378398 | 653611 | 1594903 | 2205898 | 2151188 | 603020 | 344817 |
| 7 | 823436 | 1782822 | 1634657 | 575372 | 268131 | 457848 | 1072888 | 1495225 | 1407653 | 463670 |
| 8 | 309595 | 573410 | 1242900 | 1136803 | 396300 | 194834 | 310709 | 766018 | 1041604 | 1062944 |
| 9 | 849547 | 215588 | 398854 | 866074 | 788871 | 277686 | 138902 | 210017 | 544322 | 728955 |
| 10 | 233888 | 591940 | 150046 | 277134 | 602658 | 546476 | 196327 | 96148 | 141046 | 374104 |
| 11 | 341930 | 162895 | 411773 | 104401 | 192481 | 417907 | 376503 | 135066 | 66068 | 92214 |
| 12 | 691584 | 720410 | 614275 | 711119 | 564497 | 523704 | 646553 | 697062 | 563917 | 423905 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 1990 | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 |
| 0 | 3296690 | 3320330 | 3331859 | 3102977 | 2966742 | 2857519 | 2949914 | 2913000 | 2972230 | 3251063 |
| 1 | 3101668 | 2571518 | 2857170 | 3087322 | 2573261 | 2510057 | 2258953 | 2641902 | 2391332 | 2607841 |
| 2 | 2351096 | 2641378 | 1957547 | 2394688 | 2783885 | 2073325 | 2063895 | 1735510 | 2299761 | 1906540 |
| 3 | 3869183 | 2107804 | 2529093 | 1615862 | 1962776 | 2374987 | 2149403 | 1921972 | 1228406 | 2345975 |
| 4 | 1822699 | 3017243 | 1510622 | 2012102 | 1085678 | 1414842 | 1796396 | 1771019 | 1626367 | 1229880 |
| 5 | 1073317 | 1242251 | 1911691 | 978289 | 1369952 | 674441 | 964620 | 1200357 | 1502045 | 1264800 |
| 6 | 1962651 | 774645 | 942991 | 1152046 | 584416 | 963652 | 489065 | 723483 | 853197 | 899973 |
| 7 | 215155 | 1217902 | 473982 | 568446 | 647643 | 344411 | 571827 | 320366 | 478247 | 612014 |
| 8 | 351224 | 137739 | 729875 | 309316 | 335059 | 280904 | 212429 | 344689 | 261577 | 310193 |
| 9 | 728116 | 248271 | 88656 | 410067 | 181481 | 175430 | 136189 | 149900 | 211121 | 180565 |
| 10 | 477091 | 495048 | 159188 | 52777 | 214699 | 107923 | 92125 | 85631 | 101280 | 131744 |
| 11 | 251205 | 302048 | 310677 | 96684 | 29468 | 129039 | 61967 | 49195 | 53050 | 63056 |
| 12 | 343426 | 388693 | 440253 | 464220 | 337416 | 217410 | 207613 | 167601 | 139152 | 124229 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 |
| 0 | 3111203 | 4389093 | 5043317 | 4042636 | 4959806 | 6098294 | 6166509 | 5211531 | 4832639 | 4710740 |
| 1 | 2975840 | 1845140 | 4979934 | 6070821 | 2909754 | 3983490 | 5962938 | 5711013 | 4436918 | 4272735 |
| 2 | 2231701 | 2546612 | 1155583 | 4783942 | 6577625 | 2399484 | 3480919 | 5095654 | 5197356 | 3749368 |
| 3 | 1771718 | 1732818 | 2455311 | 813382 | 3935118 | 5320042 | 1787244 | 2633659 | 4699279 | 5347442 |
| 4 | 1823904 | 1273154 | 1508857 | 1565163 | 753503 | 1945485 | 3294304 | 1549549 | 2116110 | 4177835 |
| 5 | 993678 | 1246480 | 967594 | 904320 | 1002265 | 547341 | 1091047 | 2164268 | 1299684 | 1737520 |
| 6 | 861901 | 648281 | 808643 | 568995 | 474680 | 492802 | 383418 | 773503 | 1183153 | 964198 |
| 7 | 618805 | 597414 | 399006 | 382446 | 268251 | 236713 | 289691 | 261814 | 442503 | 733724 |
| 8 | 374774 | 413668 | 343720 | 236868 | 184628 | 136236 | 136142 | 187464 | 182714 | 280012 |
| 9 | 191184 | 239635 | 226796 | 191079 | 116149 | 88903 | 75309 | 96372 | 104843 | 115566 |
| 10 | 114039 | 125262 | 126860 | 115440 | 90595 | 62673 | 53883 | 47829 | 59922 | 55504 |
| 11 | 70927 | 68986 | 62614 | 65355 | 47189 | 32353 | 32859 | 34601 | 22893 | 30682 |
| 12 | 121908 | 125471 | 110198 | 80294 | 57035 | 41447 | 39651 | 40541 | 32654 | 22419 |
| year |  |  |  |  |  |  |  |  |  |  |
| age | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 | 2018 | 2019 |
| 0 | 5605915 | 6429152 | 5734036 | 5470180 | 5529895 | 5047612 | 5839707 | 6128054 | 6146977 | 6857188 |
| 1 | 4336537 | 5724823 | 6564180 | 4632674 | 4239607 | 5579064 | 3434145 | 5089660 | 4179739 | 4085434 |
| 2 | 4193134 | 3556127 | 5939342 | 6867546 | 3834283 | 3387247 | 5004656 | 2120346 | 4681101 | 2814162 |
| 3 | 3600369 | 3960768 | 2959315 | 5858933 | 7316370 | 3118790 | 2712386 | 4350708 | 1353985 | 3391438 |

```
5068330 3335577 3358824 2674386 5541704 5233421 2930306 2163318 2903339 940422
3180406 3798337 2718361 2873748 2710128 4199358 3959586 2334086 1468854 1590813
1406509 2355237 2822336 2651438 2786983 2298245 3418982 3461790 1650074 1090025
    616194 1009091 15461291927603 2383142 2180015 1830017 3078698 2602387 1162159
    408409 451129 680309 1024509 1593249 1789912 1598989 1481687 2170698 1847241
    183166 228872 299367 468474 717702 1159140 1080442 1275224 1157749 1533436
    78709 104192 140474 189118 315531
    27837 50274 60085 94692 108251 166776 280838 440284 553097 502967
```



```
year
20202021 2022 2023
6354080 5326079 54114045411404
6352981 5630840 3814513 4653840
2544660 55284314953325 3269823
1956909 1934161 3316479 4548413
2161348 17118331146833 2369765
796428 1500772 1126092 733023
1254771 621457 1083888 646205
944893 1212405 542038 764105
983003 783212 827062 325610
1330291 891399 601710 468196
1135584 946266 670247 362107
508044 743797 643854 404920
2731508 590979 964063 955340
```


## Table 8.7.3.3. NE Atlantic Mackerel. ESTIMATED FISHING MORTALITY

```
    year
age 1980
    0.009 0.009 0.009 0.009 0.009 0.009 0.009 0.009 0.009 0.009 0.009 0.008 0.008
    0.035 0.035 0.035 0.035 0.035 0.034 0.034 0.034 0.034 0.034 0.034 0.034 0.034
    0.058 0.057 0.057 0.057 0.057 0.057 0.057 0.057 0.057 0.057 0.058 0.058 0.059
    0.105 0.105 0.105 0.105 0.106 0.108 0.110 0.112 0.115 0.118 0.122 0.125 0.129
    0.173 0.173 0.174 0.174 0.175 0.179 0.183 0.191 0.196 0.204 0.210 0.217 0.221
    0.192 0.193 0.193 0.195 0.197 0.200 0.205 0.209 0.216 0.222 0.226 0.233 0.243
    0.237 0.238 0.239 0.240 0.244 0.249 0.254 0.261 0.267 0.280 0.292 0.304 0.314
    0.212 0.212 0.212 0.213 0.214 0.217 0.222 0.229 0.239 0.253 0.275 0.304 0.336
    0.212 0.212 0.212 0.213 0.214 0.217 0.222 0.229 0.239 0.253 0.275 0.304 0.336
    0.212 0.212 0.212 0.213 0.214 0.217 0.222 0.229 0.239 0.253 0.275 0.304 0.336
    10 0.212 0.212 0.212 0.213 0.214 0.217 0.222 0.229 0.239 0.253 0.275 0.304 0.336
    110.212 0.212 0.212 0.213 0.214 0.217 0.222 0.229 0.239 0.253 0.275 0.304 0.336
    2 0.212 0.212 0.212 0.213 0.214 0.217 0.222 0.229 0.239 0.253 0.275 0.304 0.336
        year
age 1993 1994 1995 1996 1997 1998 1999 2000 2001 2002 2003 2004 2005
    0.008 0.008 0.008 0.008 0.008 0.008 0.008 0.008 0.007 0.007 0.006 0.005 0.005
    0.035 0.035 0.035 0.035 0.034 0.034 0.033 0.033 0.030 0.031 0.026 0.019 0.017
    0.060 0.061 0.062 0.063 0.064 0.066 0.067 0.069 0.069 0.068 0.066 0.069 0.061
    0.133 0.135 0.138 0.140 0.144 0.149 0.157 0.167 0.161 0.161 0.142 0.146 0.134
    0.224 0.227 0.227 0.227 0.226 0.231 0.241 0.256 0.268 0.265 0.237 0.220 0.190
    0.250 0.253 0.258 0.267 0.280 0.299 0.316 0.340}0.3.326 0.333 0.326 0.310 0.271
    0.322 0.326 0.325 0.323 0.324 0.329 0.343 0.366 0.411 0.405 0.409 0.384 0.336
    0.368 0.387 0.378 0.349 0.327 0.328 0.342 0.355 0.412 0.477 0.530}00.478 0.358
    0.368 0.387 0.378 0.349 0.327 0.328 0.342 0.355 0.412 0.477 0.530}00.478 0.358
    0.368 0.387 0.378 0.349 0.327 0.328 0.342 0.355 0.412 0.477 0.530}00.478 0.358
    0.368 0.387 0.378 0.349 0.327 0.328 0.342 0.355 0.412 0.477 0.530}00.478 0.358
    10.368 0.387 0.378 0.349}0.327 0.328 0.342 0.355 0.412 0.477 0.530 0.478 0.358
    2 0.368 0.387 0.378 0.349 0.327 0.328 0.342 0.355 0.412 0.477 0.530}00.478 0.358
        year
        2006 2007 2008 2009 2010 2011 2012 2013 2014 2015 2016 2017 2018
        0.005 0.005 0.005 0.004 0.004 0.004 0.003 0.003 0.003 0.002 0.001 0.001 0.001
        0.019 0.017 0.015 0.013 0.014 0.013 0.012 0.012 0.013 0.015 0.014 0.011 0.010
        0.053 0.043 0.037 0.035 0.036 0.036 0.037 0.038 0.040 0.042 0.045 0.047 0.052
        0.109 0.099 0.096 0.095 0.094 0.090 0.085 0.084 0.095 0.096 0.106 0.115 0.116
        0.173 0.165 0.162 0.167 0.166 0.161 0.152 0.158}0.161 0.147 0.163 0.172 0.157
        0.241 0.247 0.237 0.225 0.222 0.208 0.200 0.196 0.213 0.192 0.186 0.199 0.211
        0.320 0.313 0.283 0.274 0.252 0.242 0.222 0.209 0.222 0.217 0.194 0.190}0.212
        0.331 0.404 0.389 0.324 0.308 0.299 0.263 0.257 0.241 0.222 0.185 0.185 0.188
        0.331 0.404 0.389 0.324 0.308 0.299 0.263 0.257 0.241 0.222 0.185 0.185 0.188
        0.331 0.404 0.389 0.324 0.308 0.299 0.263 0.257 0.241 0.222 0.185 0.185 0.188
    llllllllllllllllllllll
    11 0.331 0.404 0.389 0.324 0.308 0.299 0.263 0.257 0.241 0.222 0.185 0.185 0.188
    20.331 0.404 0.389 0.324 0.308 0.299 0.263 0.257 0.241 0.222 0.185 0.185 0.188
        year
age 2019 2020 2021 2022
    0.001 0.001 0.001 0.001
    0.009 0.008 0.006 0.004
    0.051 0.051 0.058 0.055
    0.115 0.111 0.115 0.101
    40.150 0.155 0.178 0.182
```

[^10]Table 8.8.3.1. NE Atlantic Mackerel. Short-term prediction: INPUT DATA

|  |  | $\Sigma$ | 글 |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2023 |  |  |  |  |  |  |  |  |
| 0 | 4498424 | 0.15 | 0.000 | 0.000 | 0.329 | 0.000 | 0.001 | 0.060 |
| 1 | 4846566 | 0.15 | 0.125 | 0.352 | 0.329 | 0.120 | 0.006 | 0.160 |
| 2 | 3269823 | 0.15 | 0.586 | 0.135 | 0.329 | 0.196 | 0.055 | 0.278 |
| 3 | 4548413 | 0.15 | 0.929 | 0.135 | 0.329 | 0.251 | 0.110 | 0.325 |
| 4 | 2369765 | 0.15 | 0.998 | 0.135 | 0.329 | 0.292 | 0.172 | 0.360 |
| 5 | 733023 | 0.15 | 1.000 | 0.249 | 0.329 | 0.337 | 0.260 | 0.391 |
| 6 | 646205 | 0.15 | 1.000 | 0.249 | 0.329 | 0.358 | 0.309 | 0.424 |
| 7 | 764105 | 0.15 | 1.000 | 0.249 | 0.329 | 0.388 | 0.311 | 0.445 |
| 8 | 325610 | 0.15 | 1.000 | 0.249 | 0.329 | 0.406 | 0.311 | 0.464 |
| 9 | 468196 | 0.15 | 1.000 | 0.249 | 0.329 | 0.418 | 0.311 | 0.477 |
| 10 | 362107 | 0.15 | 1.000 | 0.249 | 0.329 | 0.436 | 0.311 | 0.495 |
| 11 | 404920 | 0.15 | 1.000 | 0.249 | 0.329 | 0.448 | 0.311 | 0.513 |
| 12+ | 955340 | 0.15 | 1.000 | 0.249 | 0.329 | 0.489 | 0.311 | 0.547 |
| 2024 and 2025 |  |  |  |  |  |  |  |  |
| 0 | 4498424 | 0.15 | 0.000 | 0.000 | 0.329 | 0.000 | 0.001 | 0.060 |
| 1 | - | 0.15 | 0.125 | 0.352 | 0.329 | 0.120 | 0.006 | 0.160 |
| 2 | - | 0.15 | 0.586 | 0.135 | 0.329 | 0.196 | 0.055 | 0.278 |
| 3 | - | 0.15 | 0.929 | 0.135 | 0.329 | 0.251 | 0.110 | 0.325 |
| 4 | - | 0.15 | 0.998 | 0.135 | 0.329 | 0.292 | 0.172 | 0.360 |
| 5 | - | 0.15 | 1.000 | 0.249 | 0.329 | 0.337 | 0.260 | 0.391 |
| 6 | - | 0.15 | 1.000 | 0.249 | 0.329 | 0.358 | 0.309 | 0.424 |
| 7 | - | 0.15 | 1.000 | 0.249 | 0.329 | 0.388 | 0.311 | 0.445 |
| 8 | - | 0.15 | 1.000 | 0.249 | 0.329 | 0.406 | 0.311 | 0.464 |
| 9 | - | 0.15 | 1.000 | 0.249 | 0.329 | 0.418 | 0.311 | 0.477 |
| 10 | - | 0.15 | 1.000 | 0.249 | 0.329 | 0.436 | 0.311 | 0.495 |
| 11 | - | 0.15 | 1.000 | 0.249 | 0.329 | 0.448 | 0.311 | 0.513 |
| 12+ | - | 0.15 | 1.000 | 0.249 | 0.329 | 0.489 | 0.311 | 0.547 |

Table 8.8.3.2. NE Atlantic Mackerel. Short-term prediction: Multi-option table for 1131416 t catch in 2022 and a range of F-values in 2023.

| 2023 |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| TSB | SSB | $F_{\text {bar }}$ | Catch |  |  |  |
| 4965336 | 3681064 | 0.38 | 1147407 |  |  |  |
| 2024 |  |  |  | 2025 |  |  |
| TSB | SSB | Fbar | Catch | TSB | SSB | Implied change in the catch |
| 4565496 | 3593282 | 0 | 0 | 5127054 | 4174174 | -100\% |
| - | 3587818 | 0.01 | 31447 | 5100357 | 4141755 | -97\% |
| - | 3582366 | 0.02 | 62634 | 5073885 | 4109684 | -95\% |
| - | 3576926 | 0.03 | 93564 | 5047636 | 4077959 | -92\% |
| - | 3571499 | 0.04 | 124241 | 5021608 | 4046576 | -89\% |
| - | 3566084 | 0.05 | 154665 | 4995799 | 4015529 | -87\% |
| - | 3560681 | 0.06 | 184839 | 4970206 | 3984815 | -84\% |
| - | 3555290 | 0.07 | 214766 | 4944829 | 3954430 | -81\% |
| - | 3549911 | 0.08 | 244449 | 4919663 | 3924369 | -79\% |
| - | 3544544 | 0.09 | 273889 | 4894708 | 3894630 | -76\% |
| - | 3539189 | 0.1 | 303088 | 4869962 | 3865208 | -74\% |
| - | 3533846 | 0.11 | 332050 | 4845422 | 3836099 | -71\% |
| - | 3528516 | 0.12 | 360775 | 4821087 | 3807300 | -69\% |
| - | 3523197 | 0.13 | 389267 | 4796954 | 3778807 | -66\% |
| - | 3517890 | 0.14 | 417528 | 4773022 | 3750616 | -64\% |
| - | 3512595 | 0.15 | 445560 | 4749289 | 3722723 | -61\% |
| - | 3507312 | 0.16 | 473365 | 4725753 | 3695126 | -59\% |
| - | 3502040 | 0.17 | 500945 | 4702411 | 3667819 | -56\% |
| - | 3496781 | 0.18 | 528302 | 4679263 | 3640801 | -54\% |
| - | 3491533 | 0.19 | 555438 | 4656306 | 3614067 | -52\% |
| - | 3486297 | 0.2 | 582356 | 4633539 | 3587615 | -49\% |
| - | 3481072 | 0.21 | 609057 | 4610959 | 3561440 | -47\% |
| - | 3475860 | 0.22 | 635544 | 4588565 | 3535539 | -45\% |
| - | 3470659 | 0.23 | 661818 | 4566355 | 3509910 | -42\% |
| - | 3465469 | 0.24 | 687882 | 4544328 | 3484548 | -40\% |
| - | 3460291 | 0.25 | 713737 | 4522481 | 3459451 | -38\% |
| - | 3455125 | 0.26 | 739386 | 4500813 | 3434616 | -36\% |
| - | 3449970 | 0.27 | 764830 | 4479323 | 3410039 | -33\% |
| - | 3444827 | 0.28 | 790071 | 4458008 | 3385718 | -31\% |
| - | 3439695 | 0.29 | 815112 | 4436867 | 3361649 | -29\% |
| - | 3434575 | 0.3 | 839953 | 4415898 | 3337829 | -27\% |
| - | 3429466 | 0.31 | 864598 | 4395100 | 3314256 | -25\% |
| - | 3424368 | 0.32 | 889047 | 4374471 | 3290926 | -23\% |
| - | 3419282 | 0.33 | 913302 | 4354009 | 3267837 | -20\% |
| - | 3414207 | 0.34 | 937366 | 4333713 | 3244986 | -18\% |
| - | 3409143 | 0.35 | 961240 | 4313582 | 3222370 | -16\% |
| - | 3404090 | 0.36 | 984926 | 4293613 | 3199985 | -14\% |
| - | 3399049 | 0.37 | 1008425 | 4273806 | 3177830 | -12\% |
| - | 3394019 | 0.38 | 1031740 | 4254159 | 3155902 | -10\% |
| - | 3389000 | 0.39 | 1054872 | 4234669 | 3134198 | -8\% |
| - | 3383993 | 0.4 | 1077823 | 4215337 | 3112716 | -6\% |
| - | 3378996 | 0.41 | 1100593 | 4196159 | 3091452 | -4\% |
| - | 3374011 | 0.42 | 1123187 | 4177136 | 3070404 | -2\% |
| - | 3369036 | 0.43 | 1145603 | 4158265 | 3049571 | 0\% |
| - | 3364073 | 0.44 | 1167845 | 4139544 | 3028948 | 2\% |
| - | 3359120 | 0.45 | 1189914 | 4120974 | 3008534 | 4\% |
| - | 3354179 | 0.46 | 1211812 | 4102552 | 2988326 | 6\% |
| - | 3349248 | 0.47 | 1233540 | 4084276 | 2968323 | 8\% |
| - | 3344329 | 0.48 | 1255099 | 4066146 | 2948520 | 9\% |
| - | 3339420 | 0.49 | 1276492 | 4048160 | 2928917 | 11\% |


| - | 3334522 | 0.5 | 1297719 | 4030316 | 2909511 | 13\% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| - | 3329635 | 0.51 | 1318783 | 4012614 | 2890299 | 15\% |
| - | 3324759 | 0.52 | 1339684 | 3995052 | 2871279 | 17\% |
| - | 3319894 | 0.53 | 1360425 | 3977629 | 2852450 | 19\% |
| - | 3315039 | 0.54 | 1381006 | 3960344 | 2833808 | 20\% |
| - | 3310195 | 0.55 | 1401430 | 3943194 | 2815352 | 22\% |
| - | 3305362 | 0.56 | 1421697 | 3926180 | 2797079 | 24\% |
| - | 3300539 | 0.57 | 1441809 | 3909299 | 2778988 | 26\% |
| - | 3295727 | 0.58 | 1461768 | 3892551 | 2761076 | 27\% |
| - | 3290926 | 0.59 | 1481575 | 3875934 | 2743341 | 29\% |
| - | 3286135 | 0.6 | 1501230 | 3859447 | 2725781 | 31\% |
| - | 3281355 | 0.61 | 1520737 | 3843089 | 2708394 | 33\% |
| - | 3276585 | 0.62 | 1540095 | 3826858 | 2691178 | 34\% |
| - | 3271826 | 0.63 | 1559307 | 3810754 | 2674131 | 36\% |
| - | 3267077 | 0.64 | 1578374 | 3794776 | 2657252 | 38\% |
| - | 3262339 | 0.65 | 1597296 | 3778921 | 2640537 | 39\% |
| - | 3257611 | 0.66 | 1616076 | 3763190 | 2623987 | 41\% |
| - | 3252894 | 0.67 | 1634714 | 3747580 | 2607597 | 42\% |
| - | 3248187 | 0.68 | 1653212 | 3732091 | 2591367 | 44\% |
| - | 3243490 | 0.69 | 1671572 | 3716722 | 2575295 | 46\% |
| - | 3238804 | 0.7 | 1689793 | 3701472 | 2559379 | 47\% |
| - | 3234128 | 0.71 | 1707878 | 3686339 | 2543617 | 49\% |
| - | 3229462 | 0.72 | 1725829 | 3671322 | 2528007 | 50\% |
| - | 3224807 | 0.73 | 1743645 | 3656421 | 2512548 | 52\% |
| - | 3220161 | 0.74 | 1761328 | 3641634 | 2497238 | 54\% |
| - | 3215526 | 0.75 | 1778880 | 3626960 | 2482075 | 55\% |
| - | 3210901 | 0.76 | 1796301 | 3612398 | 2467058 | 57\% |
| - | 3206286 | 0.77 | 1813594 | 3597948 | 2452184 | 58\% |
| - | 3201681 | 0.78 | 1830758 | 3583608 | 2437453 | 60\% |
| - | 3197087 | 0.79 | 1847795 | 3569377 | 2422863 | 61\% |
| - | 3192502 | 0.8 | 1864707 | 3555254 | 2408411 | 63\% |
| - | 3187928 | 0.81 | 1881494 | 3541238 | 2394097 | 64\% |
| - | 3183363 | 0.82 | 1898158 | 3527329 | 2379919 | 65\% |
| - | 3178808 | 0.83 | 1914699 | 3513525 | 2365875 | 67\% |
| - | 3174264 | 0.84 | 1931119 | 3499825 | 2351965 | 68\% |
| - | 3169729 | 0.85 | 1947418 | 3486229 | 2338185 | 70\% |
| - | 3165204 | 0.86 | 1963599 | 3472735 | 2324535 | 71\% |
| - | 3160689 | 0.87 | 1979661 | 3459342 | 2311014 | 73\% |
| - | 3156184 | 0.88 | 1995606 | 3446050 | 2297620 | 74\% |
| - | 3151689 | 0.89 | 2011435 | 3432858 | 2284351 | 75\% |
| - | 3147204 | 0.9 | 2027150 | 3419765 | 2271207 | 77\% |
| - | 3142728 | 0.91 | 2042750 | 3406769 | 2258185 | 78\% |
| - | 3138262 | 0.92 | 2058237 | 3393871 | 2245284 | 79\% |
| - | 3133806 | 0.93 | 2073613 | 3381069 | 2232503 | 81\% |
| - | 3129359 | 0.94 | 2088877 | 3368362 | 2219841 | 82\% |
| - | 3124923 | 0.95 | 2104031 | 3355749 | 2207297 | 83\% |
| - | 3120495 | 0.96 | 2119077 | 3343230 | 2194868 | 85\% |
| - | 3116078 | 0.97 | 2134014 | 3330804 | 2182554 | 86\% |
| - | 3111670 | 0.98 | 2148844 | 3318470 | 2170354 | 87\% |
| - | 3107272 | 0.99 | 2163569 | 3306227 | 2158265 | 89\% |
| - | 3102883 | 1 | 2178187 | 3294074 | 2146288 | 90\% |
| - | 3098504 | 1.01 | 2192702 | 3282010 | 2134420 | 91\% |
| - | 3094134 | 1.02 | 2207113 | 3270036 | 2122661 | 92\% |
| - | 3089774 | 1.03 | 2221422 | 3258149 | 2111009 | 94\% |
| - | 3085423 | 1.04 | 2235629 | 3246349 | 2099463 | 95\% |
| - | 3081082 | 1.05 | 2249735 | 3234636 | 2088021 | 96\% |
| - | 3076750 | 1.06 | 2263742 | 3223008 | 2076684 | 97\% |
| - | 3072427 | 1.07 | 2277649 | 3211465 | 2065449 | 99\% |
| - | 3068114 | 1.08 | 2291459 | 3200006 | 2054316 | 100\% |
| - | 3063810 | 1.09 | 2305171 | 3188630 | 2043283 | 101\% |
| - | 3059516 | 1.1 | 2318787 | 3177337 | 2032349 | 102\% |

Table 8.8.3.3. NE Atlantic Mackerel. Short-term prediction: specific management options.

| Rationale | $\begin{aligned} & \text { Catch } \\ & \text { (2024 } \\ & \text { ) } \end{aligned}$ | $F_{\text {bar }}$ (2024) | $\begin{aligned} & \text { SSB } \\ & \text { (2024) } \end{aligned}$ | $\begin{aligned} & \text { SSB } \\ & \text { (2025) } \end{aligned}$ | \% SSB chang e* | \% catch change ** | \% advice change *** |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\begin{aligned} & \operatorname{Fbar}(2024)=0.26 \\ & \text { (Fmsy) } \end{aligned}$ | $\begin{array}{r} 73938 \\ 6 \end{array}$ | 0.26 | 3455125 | $\begin{array}{r} 343461 \\ 6 \end{array}$ | -1\% | -36\% | -5\% |
| Catch(2024) $=$ Zero | 0 | 0 | 3593282 | $\begin{array}{r} 417417 \\ 4 \end{array}$ | 16\% | -100\% | -100\% |
| $\begin{aligned} & \text { Catch(2024) }=2023 \\ & \text { catch }-20 \% \end{aligned}$ | $\begin{array}{r} 91792 \\ 6 \end{array}$ | 0.33 | 3418309 | $\begin{array}{r} 327784 \\ 3 \end{array}$ | -4\% | -20\% | 17\% |
| $\begin{aligned} & \text { Catch }(2024)=2023 \\ & \text { catch } \end{aligned}$ | $\begin{array}{r} 11474 \\ 07 \end{array}$ | 0.43 | 3368635 | $\begin{array}{r} 310783 \\ 0 \end{array}$ | -8\% | 0\% | 47\% |
| $\begin{aligned} & \text { Catch }(2024)=2023 \\ & \text { catch }+25 \% \end{aligned}$ | $\begin{array}{r} 14342 \\ 59 \end{array}$ | 0.57 | 3302353 | $\begin{array}{r} 289582 \\ 5 \end{array}$ | -12\% | 25\% | 83\% |
| $\operatorname{Fbar}(2024)=$ $\operatorname{Fbar}(2023)$ | $\begin{array}{r} 10398 \\ 88 \end{array}$ | 0.38 | 3392255 | $\begin{array}{r} 314825 \\ 1 \end{array}$ | -7\% | -9\% | 33\% |
| $\operatorname{Fbar}(2024)=0.36$ (Fpa) | $\begin{array}{r} 98492 \\ 6 \end{array}$ | 0.36 | 3404090 | $\begin{array}{r} 319998 \\ 5 \end{array}$ | -6\% | -14\% | 26\% |
| $\begin{aligned} & \operatorname{Fbar}(2024)=0.46 \\ & \text { (Flim) } \end{aligned}$ | $\begin{array}{r} 12118 \\ 12 \end{array}$ | 0.46 | 3354179 | $\begin{array}{r} 298832 \\ 6 \end{array}$ | -11\% | 6\% | 55\% |
| SSB2025 = Blim | $\begin{array}{r} 23590 \\ 65 \end{array}$ | 1.13 | 3046688 | $\begin{array}{r} 200013 \\ 2 \end{array}$ | -34\% | 106\% | 202\% |
| SSB2025=Bpa | $\begin{array}{r} 16715 \\ 72 \end{array}$ | 0.69 | 3243490 | $\begin{array}{r} 257529 \\ 5 \end{array}$ | -21\% | 46\% | 114\% |

* SSB 2025 relative to SSB 2024.
** Catch in 2024 relative to assumed catches in 2023 (1 147 40 t). There is no internationally agreed TAC for 2023.
*** Catch in 2024 relative to the advice value for 2023 ( $\mathbf{7 8 2} \mathbf{0 6 6 t}$ ).


### 8.15 Figures



Figure 8.2.2.1. NE Atlantic Mackerel. Commercial catches in 2022, quarter 1.


Figure 8.2.2.2. NE Atlantic Mackerel. Commercial catches in 2022, quarter 2.


Figure 8.2.2.3. NE Atlantic Mackerel. Commercial catches in 2022, quarter 3.


Figure 8.2.2.4. NE Atlantic Mackerel. Commercial catches in 2022, quarter 4.


Figure 8.2.3.1. NE Atlantic Mackerel. Proportion of the catch by quarter by year.


Figure 8.5.2.1. NE Atlantic mackerel. Weights-at-age in the catch.


Figure 8.5.2.2. NE Atlantic mackerel. Weights-at-age in the stock.


Figure 8.5.3.1. NE Atlantic mackerel. Proportion of mature fish at age.


Figure 8.6.1.1.1. Mackerel egg production by half rectangle and survey period. Colour scale represents mackerel stage I eggs/m2/day by ICES half rectangle. Crosses represent zero values


Figure 8.6.1.1.2. Egg production by period for the southern (left) and western (right) areas for the NEA mackerel stock since 2004


Figure 8.6.1.1.3. Combined mackerel TAEP estimates (southern and western areas)(* ${ }^{10}{ }^{13}$ ) - 1992 - 2022.


Figure 8.6.1.3.1. SSB estimates using AEPM for NEA mackerel. 1992-2022


Figure 8.6.1.4.1. Survey coverage for the North Sea, 2022.Norway (Green), England (Blue), Denmark (Red).


Figure 8.6.1.4.2. Heat map of Stage la mackerel egg production (eggs. m2. day-1) by half rectangle for the North Sea, 2022. Grey circles represent observed values, crosses represent observed zeros.


Figure 8.6.2.1.1. Overview of the distribution of trawl hauls and number of samples in IBTS in the period 1998-2023. a) Geographical positions of hauls in Q4. b) Geographical positions of hauls in Q1 hauls. c) Number of hauls from Q1 and Q4. d) Vertical distribution of all hauls.


Figure 8.6.2.2.1. Spatial distribution of mackerel juveniles at age $\mathbf{0}$ in October to March. a) Average for cohorts from 1998-2022. b) $\mathbf{2 0 2 2}$ cohort. Mackerel squared catch rates by trawl haul (circle areas represent catch rates in $\mathbf{k g} / \mathrm{km}^{2}$ ) overlaid on modelled squared catch rates per $10 \times 10 \mathrm{~km}$ rectangle. Each rectangle is coloured according to the expected squared catch rate in percent of the highest value for that year. See Jansen et al. (2015) for details.


Figure 8.6.2.2.2. Index of mackerel juveniles at age $\mathbf{0}$ in October to March proxied by annual integration of square root of expected catch in demersal trawl surveys (Blue lines). See Jansen et al. (2015) for details.


Figure 8.6.3.1. Estimated total stock numbers (TSN) of mackerel from IESSNS calculated using StoX for the years 2007 and from 2010 to 2023. Displayed is StoX baseline estimate (red dot) and a bootstrap estimate (black dot), calculated using 1000 replicates, with $90 \%$ confidence intervals (vertical line) based on the bootstrap. Analysis excludes the North Sea and survey coverage was incomplete in 2007 and 2011.


Figure 8.6.3.2. Estimated total stock biomass of mackerel from IESSNS calculated using StoX for the years 2007 and from 2010 to 2023. Displayed is StoX baseline estimate (red dot) and a bootstrap estimate (black dot), calculated using 1000 replicates, with $90 \%$ confidence intervals (vertical line) based on the bootstrap. Analysis excludes the North Sea and survey coverage was incomplete in 2007 and 2011.


Figure 8.6.3.3. Internal consistency of the mackerel abundance index from the IESSNS survey including data from 2012 to 2023, excluding North Sea. Ages indicated by white numbers in grey diagonal cells. Statistically significant positive correlations ( $p<0.05$ ) are indicated by regression lines and red cells in upper left half. Correlation coefficients ( $r$ ) are given in the lower right half.


Figure 8.6.3.4. Mackerel catch curves from the estimate stock size at age from the IESSNS in 2010 and from 2012 to 2023, excluding the North Sea. Each cohort is marked by a uniquely coloured line that connects the estimates indicated by the respective ages.


Figure 8.6.3.5. Mackerel catch rates from predetermined surface trawl stations (circle size represents catch rate in $\mathrm{kg} / \mathrm{km} 2$ ) overlaid on mean catch rate per standardized rectangle ( $2^{\circ}$ lat. $\times 4^{\circ}$ lon.) from the 2023 IESSNS, including North Sea. Zero mackerel catches are displayed as grey crosses.



Figure 8.6.3.6. Mackerel annual distribution proxied by the absolute distribution of mean mackerel catch rates per standardized rectangles ( $2^{\circ}$ lat. x $4^{\circ}$ lon.), from predetermined surface trawl stations from IESSNS in 2010 and from 2012 to 2023, including North Sea. Colour scale goes from white $(=0)$ to red (= maximum value for the given year).


Figure 8.6.4.1. Number and distribution of RFID tagged mackerel from experiments west of Ireland and British Isles during 2011-2023. Note that data from releases 2011-2012 are not used in the stock assessment, based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019c), and data from experiments in 2022-2023 are not included as there are no full years with recaptures yet.


Figure 8.6.4.2. Biomass and distribution of catches scanned for RFID tagged mackerel during 2012-2022. Note that data from scanned catches in 2012-2013 are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019c).


Figure 8.6.4.3. Distribution of recaptures of RFID tagged mackerel west off Ireland and British Isles during 20112022. Note that data on recaptures in 2012-2013 are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019c).


Figure 8.6.4.4. Number of recaptures per factory over the years 2012-2023 (22.Aug)

8.6.4.5. Overview of the relative year class distribution among RFID tagged mackerel per release year 20112023 from experiments west of Ireland and British Isles in April-June, compared with the number scanned and recaptured in year 1 and 2 after release of the same year classes. Note that data from releases in 2011-2012 are not used in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019c). Note also that it was decided to only use ages 5-11 in updated assessments, and limits for this age span is marked (vertical grey dotted lines) for each release year.


Figure 8.6.4.6. Trends in year class abundance ( $\mathbf{N}=$ numbers released/numbers recaptured*numbers scanned) from RFID tag-recapture data based on aggregated data on recaptures and scanned numbers in year 1 and 2 after each release year. Data that are excluded in the stock assessment based on decisions in the ICES IBPNEAMac 2019 meeting (ICES 2019c), i.e., release years 2011-2012 and ages 2-4 and 12+, are marked with dotted lines in year class trends.


Figure 8.6.4.7. Consistency plot showing correlations between age groups in the from the RFID indices of year class abundance (Figure 8.6.4.6)


Figure 8.6.4.8. Trends in various age aggregated biomass indices from RFID tag-recapture data compared with the SSB ( $\pm 95$ confidence intervals) from the WGWIDE2023 stock assessment. Data are based on a combination of estimated numbers by year class (Figure 8.6.4.6) scaled by survival parameter ( 0.1778 ) and weight at age in stock from WGWIDE2021. Note that data from release years 2011-2012, and ages 2-4 and 12+ are excluded from the stock assessment.


Figure 8.7.2.1. NE Atlantic mackerel. Parameter estimates from the SAM model (and associated confidence intervals) for the WGWIDE 2023 update assessment. top left : estimated standard deviation for the observation errors, top centre: estimated overdispersion for the errors on the tag recaptures, top right: standard deviation for the processes, bottom: survey catchabilities and post-release survival of tagged fish.


Figure 8.7.2.2. NE Atlantic mackerel. Parameter uncertainty (standard deviation of estimate) versus parameter value for the observation variances.


Figure 8.7.2.3. NE Atlantic mackerel. Estimated AR1 error correlation structure for the observations from the IESSNS survey age 3 to 11.


Figure 8.7.2.4. NE Atlantic mackerel. Correlation between parameter estimates from the SAM model for the WGWIDE 2023 update assessment


Figure 8.7.2.5. NE Atlantic mackerel. One Step Ahead Normalized residuals for the fit to the catch data (catch data prior to 2000 are not used to fit the model). Blue circles indicate positive residuals (observation larger than predicted) and filled red circles indicate negative residuals.


Figure 8.7.2.6. NE Atlantic mackerel. Empirical correlations between ages in the One Step Ahead residuals for the catch-at-age data.


Figure 8.7.2.7. NE Atlantic mackerel. Empirical correlations between ages in the One Step Ahead residuals for the IESSNS abundances-at-age.


Figure 8.7.2.8. NE Atlantic mackerel. One step ahead residuals for the fit to the recaptures of tags in the final assessment. The $x$-axis represents the release year, and the $y$-axis is the number of years between tagging and recapture. Each panel correspond to a given age at release. Blue circles indicate positive residuals (observation larger than predicted) and filled red circles indicate negative residuals.


Figure 8.7.2.9. NE Atlantic mackerel. Leave one out assessment runs. SAM estimates of SSB , Fbar and recruitment, for assessments runs leaving out one of the observation data sets.


Figure 8.7.3.1. NE Atlantic mackerel. Perception of the NEA mackerel stock, showing the SSB, $\mathrm{F}_{\mathrm{bar}} 4-8$ and recruitment (with $95 \%$ confidence intervals) from the SAM assessment. crosses in the catch panel represent the catch data.


Figure 8.7.3.2. NE Atlantic mackerel. Estimated exploitation pattern for the period 1990 to 2022, calculated as the ratio of the estimated fishing mortality-at-age and the Fbar4-8 value in the corresponding year.


Figure 8.7.4.1. NE Atlantic mackerel. Uncertainty (standard deviation of the log values) of the estimates of SSB and Fbar from the SAM for the 2022 and 2023 WGWIDE assessments.


Figure 8.7.4.2. NE Atlantic mackerel. Analytical retrospective patterns (8 years back) of SSB, $\mathrm{F}_{\mathrm{bar}} 4-8$ and recruitment from the WGWIDE 2022 update assessment. the Mohn's rho values are calculated based on 5 retro years .


Figure 8.7.4.3. NE Atlantic mackerel. Process error expressed as annual deviations of abundances at age, for the 2023 WGWIDE assessment and from the 2022 WGWIDE assessment.


Figure 8.7.4.4. NE Atlantic mackerel. Model process error expressed in biomass cumulated across age-group for the 2023 WGWIDE assessment and for the 2022 WGWIDE assessment.


Figure 8.7.5.1. NE Atlantic mackerel. Model. comparison of the cohort signal based on SAM estimates at age 0, 2 and3.


Figure 8.7.5.2. NE Atlantic mackerel. Model. comparison of the perception of the stocks from the WGWIDE 2023 assessment, and the assessment starting at age 2 .


Figure 8.7.5.3. NE Atlantic mackerel. Model. comparison of the perception of the stocks from the WGWIDE 2023 assessment, and an assessment in which the tagging data is used as a survey index instead of modelling the recaptures. Crosses in the catch panel represent the actual catch data.


Figure 8.7.5.4. NE Atlantic mackerel. Analytical retrospective patterns (8 years back) of SSB, Fbar4-8 and recruitment from an assessment using the tagging information as survey index instead of modelling the recaptures. The Mohn's rho values are calculated based on 5 retro years


Figure 8.8.1. NE Atlantic mackerel. Recent and forecasted future SSB trajectories for a number of management options


Figure 8.8.2. NE Atlantic mackerel. Contribution of the different age groups to the catches and SSB in 2024 in the mackerel forecast. Year classes that are derived from abundances estimated by the stock assessment model in 2023 are depicted in blue. Year class based on assumption are depicted in red (geometric mean recruitment assumption for the recruitment in 2023 and 2024) and in green (RCT3 prediction for the recruitment in 2022)

SSB (million t)

$F$ (ages 4-8)


Rec (age 2; Billi


Figure 8.10.1. NE Atlantic mackerel. Historical assessment results. Prior to the 2022 assessment, recruitment was presented as age 0 and is not shown.


Figure 8.10.2. NE Atlantic mackerel. Comparison of the stock trajectories between the 2023 WGWIDE assessment and the 2022 WGWIDE assessment.


Figure 8.10.3. NE Atlantic mackerel. Comparison of the abundances at age from 2020 to 2023 estimated from the SAM assessments (solid lines) and forecasts (dotted lines) carried out at WGWIDE2022 (blue) and WGWIDE2023 (red).


Figure 8.10.4. NE Atlantic mackerel. Comparison of model parameters and their uncertainty for the 2023 WGWIDE and the 2022 WGWIDE assessment

type

- Exploitable Stock Biomass
- Spawning Stock Biomass
- Total Stock Biomass
source
$\triangle$ SAM
- STF

Figure 8.10.5. NE Atlantic mackerel. Estimated (triangles) and forecasted (dots) trends in exploitable, spawning and total stock biomass from the 2023 WGWIDE assessment and forecast.

advice tac unilatquota catch


Figure 8.11.1. NE Atlantic mackerel. Top: comparison of the ICES advice, the agreed TAC, the sum of the declared quotas and total catch. Bottom: calculated percentage of TAC over Advice (top left), Catch over Sum of unilateral quota (top right), catch over Advice (bottom left) and Catch over TAC (bottom right).

## 9 Red gurnard in Northeast Atlantic

## Chelidonichthys cuculus in subareas 3-8 (gur.27.3-8)

### 9.1 General biology

The main biological features known for red gurnard (Aspitrigla (Chelidonichthys) cuculus) are described in the stock annex. This species is widely distributed in the North-east Atlantic from South Norway and North of the British Isles to Mauritania, on grounds between 20 and 250 m . This benthic species is abundant in the Channel (7de), the shelf West of Brittany (7h, 8a), and west of Scotland (6a), living on gravel or coarse sand. In the Channel, the size at first maturity is $\sim 25 \mathrm{~cm}$ at 3 years old (Dorel, 1986).

### 9.2 Stock identity and possible assessments areas

A compilation of datasets from bottom-trawl surveys undertaken within the project 'Atlas of the marine fishes of the northern European shelf' has produced a distribution map of red gurnard. Higher occurrences of red gurnard with patchy distribution have been observed along the Western approaches from the Shetlands Islands to the Celtic Seas and the Channel.

A continuous distribution of fish crossing the Channel and the area West of Brittany does not suggest a separation of the divisions 7 d from 7 e and 7 h . Therefore, a split of the population between these Ecoregions does not seem appropriate. Divergent trends in survey abundances have been observed within the assessment area, with a sustained spike in abundance in Division 6a in the early 2010's which is not seen in surveys covering Subareas 7-8. Further investigations, such as morphometric studies, tagging and genetic population studies, would be needed to progress on stock boundaries, however SIMWG has advised that for now, there is insufficient evidence to carry out assessments on smaller spatial units.

### 9.3 Management regulations

Currently no technical measures are specifically applied to red gurnard or other gurnard species. The exploitation of red gurnard is submitted to the general regulation in the areas where they are caught. There is no minimum landing size set.

### 9.4 Fisheries data

Red gurnard is mainly landed as by-catch by demersal trawlers in mixed fisheries, predominantly in Divisions 7d, 7e and 7h (Tables 9.1-2). High discard rates and lack of resolution at a species level make interpretation of spatial trends in catches in other areas problematic.

### 9.4.1 Historical landings

Official landings of red gurnard reported to ICES are presented in Tables 9.1 and 9.2. Before 1977, red gurnard was not specifically reported. Landings of gurnards are still not always reported at a species level, but rather as mixed gurnards (GUX). Use of this code is not consistent across countries reporting landings, and some report mixed landings of gurnards under the code GUU - which is unfortunately the FAO species code for tub gurnard (Chelidonichthys lucerna).

A questionnaire was circulated to WGCATCH to gather information on how landings of gurnards are assigned to species. For those countries who responded, only Portugal has presented information on how the reporting of landings at a species level is achieved. Other countries accept the species code as declared at the point of landing, without further validation. This makes interpretation of the records of official landings, and trends within them difficult. Landings of gurnards (red, grey, tub and mixed) are shown in Figure 9.1.

International landings have fluctuated between 2941-5049 tonnes between 2006-2022. Landings in the most recent year (2022) were 2971 tonnes - the second lowest on record. France is the main contributor of 'red gurnard' landings, with around $75 \%$ of landings from Divisions 7d-h (Celtic Sea/English Channel). In the North Sea red gurnard landings are variable, but roughly evenly distributed between Divisions 4a, b and c. Landings from the west of Scotland and Ireland, and the Irish Sea (Divisions 6a-b, 7a-c, 7j) and Bay of Biscay (Subarea 8) have been consistently low.

### 9.4.2 Discards

Discard data for red gurnard has been provided for 2015-2022 through Intercatch (Table 9.3). For those countries which provided data, discard rates are variable but high (Table 9.4). Given the uncertainty associated with landings data, these figures should be treated with caution.

### 9.5 Survey data

Information on gurnard abundance are available in DATRAS for a number of surveys. Those covering the core area of the stock as determined by WKWEST (ICES, 2021) are the Scottish West Coast Groundfish Survey (SCOWCGFS and SWC-IBTS), Irish Groundfish Survey (IE-IGFS), English Channel Beam Trawl Survey (BTS), French Q4 surveys EVHOE in the Celtic Sea and Bay of Biscay and FR-CGFS in Division 7d, and additionally recent data from NS-IBTS Q1 and Q3 (from 2020 onwards). Each of these surveys covers a specific area of red gurnard distribution and no individual survey covers the entire stock area. Lengths at age are available from FR-CGFS-Q4 and (for some years) from IE-IGFS.

## SCO-WCGFS and SWC-IBTS (Q1 and Q4)

Before 1996, red gurnard was scarce on the west of Scotland. The CPUE trended strongly upwards after 1997, reaching a peak in 2013, before declining to around the series average in recent years. The values for 2021 were low. There was, however, a slight increase in 2022-2023 (with no survey in Q1 in 2022, Figures 9.2-3).

## CGFS (Q4)

Over the time-series 1988-2011, CPUE has fluctuated, peaked in 1994, reached a low in 2011, but is above long term mean since 2016. Values in 2022 were down (Figure 9.4).

## EVHOE (Q4)

Over the period 1997-2022, the CPUE has fluctuated. It has been on an increasing trend since 2018, and 2022 is the second highest value in the series(Figure 9.4). Age reading of red gurnards caught during the EVHOE survey has been carried out in 2006 and routinely since 2008. They indicate that the individuals caught are mainly of age 1 and 2.

## IE-IGFS (Q4)

The CPUE of red gurnard in the IE-IGFS series has varied around the series mean without trend between 2003 and 2022. Values in 2022 were slightly up (Figure 9.5).

CPUE in this relatively short series has fluctuated without apparent trend since 2006. Values in 2022 were down (Figure 9.5).

### 9.6 Biological sampling

Number-at-length information was provided by French and Portuguese landings and discards. There remains a lack of regular sampling for red gurnard in commercial landings and discarding to provide series of length or age compositions usable for a preliminary analytical assessment.

### 9.7 Biological parameters and other research

There is no update of growth parameters and available parameters from several authors are summarized in the Stock Annex. They vary widely. Available length-weight relationships are also shown in Stock Annex. Natural mortality has not been estimated in the areas studied at this Working Group. Accurate estimates of landings are still lacking for this species.

### 9.8 Assessment

Having explored the trends in available survey data, the delta-lognormal assessment method developed during WKWEST (ICES, 2021) was applied. This approach extracts the estimates of year effect from the log-normal part of the model (there is no temporal term in the binomial part), together with their associated standard error, and standardises the series relative to its mean value, to provide an index of biomass across the multiple surveys. Goodness of fit metrics of the model remain acceptable (Figures 9.6-7) and the log-normal part of the model has an adjusted $\mathrm{r}^{2}$ value of 0.22 .

After a period of relative stability, the biomass indicator declined in 2018-2019, before recovering strongly in 2020 (Figure 9.8). It declined in the two following years but remains above the biomass limit reference level of 0.81 .

The influence of COVID-19 related disruption to surveys in the Channel during 2020 has not been investigated for this stock.

### 9.9 Data requirements

Gurnards are still not always reported by species, but rather as mixed gurnards. National approaches to validating the species composition of mixed gurnard landings are undocumented, other than for Portuguese landings. This makes interpretation of the records of official landings difficult. An international approach to the collection of data on species composition of gurnard landings is required to support the provision of advice for this stock.

### 9.10 References

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Table 9.1. Red gurnard in Subareas 3-8. Official landings by country in tonnes.


* Preliminary data
** InterCatch data
$+<0.5$ t

Table 9.1. Red gurnard in Subareas 3-8. Official landings by area in tonnes.

| Year | 3 a | 4a | 4b | 4c | 5b | 6a | 6b | 7a | 7b | 7 c | 7d | 7 e | 7 f | 7g | 7h | 7j | 7k | 8a | 8b | 8c | 8d | 8 e | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2006 | 0 | 13 | 83 | 64 | 0 | 32 | 1 | 11 | 9 | 12 | 1101 | 2804 | 229 | 16 | 446 | 5 | + | 153 | 60 | 1 | 5 | 0 | 5047 |
| 2007 | 0 | 12 | 120 | 55 | 2 | 21 | 0 | 7 | 7 | 15 | 1229 | 2674 | 246 | 15 | 437 | 4 | + | 139 | 59 | 3 | 2 | 0 | 5049 |
| 2008 | 0 | 34 | 64 | 54 | + | 28 | 3 | 5 | 7 | 16 | 1236 | 2451 | 249 | 9 | 408 | 5 | + | 66 | 24 | 3 | 1 | 0 | 4664 |
| 2009 | 0 | 58 | 59 | 92 | 0 | 94 | 2 | 4 | 8 | 6 | 1293 | 1557 | 112 | 22 | 510 | 7 | 0 | 98 | 40 | 1 | 3 | 0 | 3966 |
| 2010 | 0 | 79 | 63 | 86 | 0 | 101 | 46 | 13 | 8 | 10 | 1531 | 1608 | 132 | 23 | 433 | 9 | 0 | 100 | 33 | 0 | 2 | 0 | 4278 |
| 2011 | 0 | 66 | 29 | 51 | 0 | 69 | 54 | 13 | 5 | 6 | 1295 | 1753 | 124 | 20 | 372 | 9 | 0 | 112 | 46 | 1 | 3 | + | 4029 |
| 2012 | 0 | 83 | 71 | 78 | 0 | 51 | 7 | 8 | 2 | 5 | 1244 | 1441 | 145 | 53 | 294 | 2 | + | 83 | 50 | 8 | 1 | + | 3629 |
| 2013 | 0 | 88 | 109 | 60 | 0 | 47 | 0 | 10 | 2 | 6 | 1193 | 1692 | 170 | 58 | 477 | 2 | 0 | 79 | 72 | 532 | 1 | 0 | 4598 |
| 2014 | 0 | 102 | 52 | 68 | 0 | 47 | 3 | 7 | 1 | 2 | 1294 | 1642 | 115 | 19 | 1069 | 1 | 0 | 82 | 75 | 363 | 3 | + | 4946 |
| 2015 | 0 | 133 | 102 | 53 | 0 | 58 | 1 | 4 | 3 | 1 | 790 | 1553 | 87 | 6 | 703 | 1 | + | 95 | 70 | 81 | 2 | + | 3744 |
| 2016 | + | 112 | 83 | 117 | + | 76 | 1 | 11 | 3 | 1 | 906 | 1270 | 114 | 16 | 608 | 1 | + | 87 | 63 | 56 | 1 | 0 | 3526 |
| 2017 | 0 | 53 | 44 | 90 | 0 | 27 | 1 | 14 | 1 | + | 874 | 1424 | 83 | 38 | 473 | 3 | 0 | 78 | 48 | 59 | 1 | 0 | 3311 |
| 2018 | 0 | 109 | 40 | 113 | + | 43 | 0 | 7 | + | + | 903 | 1785 | 164 | 28 | 631 | 4 | + | 80 | 43 | 62 | 2 | 0 | 4014 |
| 2019 | 0 | 128 | 19 | 75 | + | 84 | + | 12 | 1 | + | 959 | 1516 | 75 | 24 | 477 | 5 | + | 73 | 38 | 65 | + | 0 | 3551 |
| 2020 | 0 | 57 | 13 | 65 | 2 | 51 | 4 | 12 | 1 | 4 | 685 | 1504 | 90 | 19 | 424 | 4 | + | 69 | 51 | 88 | 1 | 0 | 3142 |
| 2021 | 0 | 58 | 18 | 55 | 0 | 112 | 4 | 4 | 2 | + | 592 | 1390 | 46 | 15 | 471 | 4 | 0 | 61 | 40 | 62 | + | 0 | 2941 |

Table 9.2. Red gurnard in Subareas 3-8. Discards in tonnes by country, 2015-2022.

| Country | 2015 | 2016 | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| France | 1323 | 2249 | 2232 | 770 | 3132 | 292 | 623 | 1690 |
| Ireland | 10 | 147 | 93 | 251 | 180 | 76 | 56 | 83 |
| Spain | 0 | 286 | 272 | 189 | 122 | 161 | 128 | 123 |
| UK (ENG) | 74 | 30 | 0 | 207 | 506 | 110 | 708 | 2562 |
| UK (SCO) | 649 | 411 | 198 | 512 | 331 | 117 | 0 | 40 |
| Total | 2056 | 3123 | 2795 | 1929 | 4270 | 757 | 1515 | 4498 |

Table 9.3. Red gurnard in Subareas 3-8. Discarding of Red gurnard in the Northeast Atlantic, as a percentage of catch, by country, 2017-2022.

|  | Discard rate (\%) |  |  |  |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Country | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 |
| France | 48 | 21 | 56 | 11 | 22 | 42 |
| Ireland | 91 | 95 | 95 | 88 | 87 | 92 |
| Spain | 72 | 68 | 78 | 91 | 80 | 83 |
| UK (England) | NA | NA | 67 | 51 | 83 | 95 |
| UK (Scotland) | 68 | 92 | 60 | 45 | $N A$ | 22 |



Figure 9.1. Red gurnard in Subareas 3-8. Official landings of grey, red, mixed and tub gurnards from the subareas in 2006-2020.


Figure 9.2. Red Gurnard in Subareas 3-8. Trends in mean abundance ( $\mathbf{k g} / \mathrm{hr}$ ) in the Q1 Scottish IBTS (1985-2010) and Q1 Scottish West Coast Groundfish Survey (2011-2023, excluding 2022)


Figure 9.3. Red gurnard in Subareas 3-8. Trends in mean abundance ( $\mathbf{k g} / \mathrm{h}$ ) in the Q4 Scottish IBTS (1990-2009) and Q4 Scottish West Coast Groundfish Survey (2011-2022).


Figure 9.4. Red gurnard in Subareas 3-8. Trends in mean abundance (kg/h) in the EVHOE (top) and French Channel Groundfish Survey (bottom).


Figure 9.5. Red Gurnard in Subareas 3-8. Trends in mean abundance ( $\mathrm{kg} / \mathrm{h}$ ) in the Irish Groundfish Survey (top) and English Channel Beam Trawl Survey (bottom).


Figure 9.6. Red gurnard in Subareas 3-8. Measures of goodness of fit of the binomial part of the assessment model.


Figure 9.7. Red gurnard in Subareas 3-8. Measures of goodness of fit of the lognormal part of the assessment model.


Figure 9.6. Red gurnard in Subareas 3-8. Results of the assessment model. Shading corresponds to 2 standard errors around the estimate. The dashed line represents MSY $B_{\text {trigger }}(0.81)$.

# 10 Striped red mullet in the North Sea, Bay of Biscay, southern Celtic Seas, and Atlantic Iberian waters 

Mullus surmuletus in subareas 6 and 8, and divisions 7.a-c, 7.e-k, and 9.a (mur.27.67a-ce-k89a)

### 10.1 General biology

Striped red mullet (Mullus surmuletus) is a predominantly benthic species found along the coasts of Europe, southern Norway, and northern Scotland (northern Atlantic, Baltic Sea, North Sea, and the English Channel), up to the Northern part of West Africa, in the Mediterranean Basin, and in the Black Sea (Mahe et al., 2005). Young fish are distributed in lower salinity coastal areas, while adults have a more offshore distribution.

Adult red mullets feed on small crustaceans, annelid worms, and molluscs, using their chin barbels to detect prey and search the mud. Consequently, striped red mullets are typically found on sandy, gravelly and shelly sediments where they can excavate sediment with their barbels and dislodge the small invertebrates. The main natural predators of striped red mullets are sea basses, pollacks, barracudas, monkfish, congers, and sharks (Caill-Milly et al., 2017).
Sexual maturity is reached at the beginning of the second year for males, followed by a marked decrease in growth rates, and at the end of the second or beginning of the third year for females which therefore continue their rapid growth a little longer (Déniel, 1991). In the English Channel, this species matures at approximately 16 cm (Mahe et al., 2005), while in the Bay of Biscay, the sizes of first sexual maturity are given by Dorel (1986) as males 16 cm , females 18 cm and a length at which $50 \%$ of the individuals are mature (the distinction between the two sexes is not mentioned) of 22 cm . Along the continental shelf off the Island of Majorca, males and females mature at 15 and 16 cm , respectively corresponding to 1 year of age (Reñones et al., 1995).

Spawning occurs in the spring and early summer -May to June according to Desbrosses (1933)with a spawning peak in June in the northern Bay of Biscay (N'Da and Déniel, 1993). Eggs and larvae average 2.8 mm and are pelagic (Sabatés et al., 2015). The hatching takes place after three days at $18^{\circ} \mathrm{C}$ and after eight days at a temperature of $9^{\circ} \mathrm{C}$ (Quéro and Vayne, 1997) After metamorphosis juveniles become first demersal and then benthic. At the age of one month, they measure about 5 cm and weigh 0.9 to 1.6 g . They show rapid growth during their first four months of life between July and October. Increases in length and mass are about 7 cm and 25 g on average during this period ( $\mathrm{N}^{\prime} \mathrm{Da}$ and Déniel, 2005). The growth rate declines sharply in October due to the cooling of water and the scarcity of trophic resources in the environment. These conditions contribute to the initiation of migration of red mullets to greater depths offshore. Until age two, there is no significant difference in size between males and females; they then measure $20-23 \mathrm{~cm}$. Sexual dimorphism is observed from the age of first maturity due to growth rates that will then differ between the two sexes. From age three, females exceed males in length by 4 cm on average and 7 cm beyond 5 years ( $\mathrm{N}^{\prime}$ Da et al., 2006).
The maximum reported age of the striped red mullet is 11 years (Quéro and Vayne, 1997; ICES, 2012), while the maximum length given is 44.5 cm in the Bay of Biscay (Dorel, 1986) and 40 cm elsewhere (Whitehead et al., 1984; Fischer et al., 1987). The maximum reported mass is 1 kg (Muus and Nielsen, 1999).

### 10.2 Management regulations

Before 2002, France enforced a minimum landing size of 16 cm . Since 2013 a minimal size requirement has been established to 15 cm (France, 2013). There is no TAC for this stock.

### 10.3 Stock ID and possible management areas

In 2004 and 2005, a study using fish geometrical morphometry was carried out in the Eastern English Channel and the Bay of Biscay. It pointed out a morphological difference in striped red mullets between those from the Eastern English Channel and those from the Bay of Biscay (Mahe et al., 2014). Benzinou et al. (2013) conducted stock identification studies based on otolith and fish shape in European waters and showed that striped red mullet can be geographically divided into three zones:

- The Bay of Biscay (Northern Bay of Biscay - NBB, and Southern Bay of Biscay - SBB)
- A mixing zone composed of the Celtic Sea and the Western English Channel (CS + WEC)
- A northern zone composed of the Eastern English Channel and the North Sea (EEC + NS)

The distinction between the putative Biscay and Western Channel/Celtic Sea populations is supported by the distribution of landings at a statistical rectangle level (Figure 10.1). This assessment treats these putative components as one population. At present there are no management measures in place, however this structuring should be considered if manamgement measures are implemented.

### 10.4 Fisheries data

Official landings have been recorded since 1975 and after early increases they have declined in recent years. France and Spain accounts for most removals (Table 10.1) and landings are mainly taken from Subarea 7 and 8 (Table 10.2). Main landings are taken in area 8 by France and Spain (Figure 10.2). Official fishing effort is only reported by 5 countries and seems stable in the last 5 years (Figure 10.3). The striped red mullet is one species among the set of benthic and demersal species targeted by the French fleet, and is mainly caught by bottom trawlers with a mesh size of $70-99 \mathrm{~mm}$. In the Western English Channel striped red mullet is also caught by gillnets. Danish seine appeared in 2008 due to some trawlers converting to use seine gears.

The average characteristics of vessels in French fleets that caught red mullet from 2000 to 2015 are 41.1 GRT, 191.1 kW engine power, 12.9 m length and 22 years of service. Net vessels comprise the smallest units ( $85 \%$ are less than 12 m long), while $52 \%$ of bottom trawlers are less than 15 m ; the seiners are by far the largest and the oldest vessels (Caill-Milly et al., 2017).

The French activity on this species differs between the area composed by West Scotland/Celtic sea (including West Channel) and the area comprising the Bay of Biscay. In the first one, landings are mainly taken by bottom trawlers, followed by gillnet. In the second one, they are mainly done by bottom trawls, seine and nets. French activity in the Atlantic Iberian waters remains limited. The Spanish activity is located in the north (8.a,b) and the south (8.c) of the Bay of Biscay.

The length distribution of the catches has been available discontinuously in Intercatch since 2012 (Figure 10.4). The pattern shown in this information is relatively stable over time.

Discards represented between $3 \%$ and $18 \%$ of the total catch in 2014-22 (Table 10.3). Since 2018, the discard rates have reported reported below $5 \%$, but in 2022 the discard rate reaches $14 \%$. However, there are concerns about how these discards have been estimated due to some
countries' lack of discards data. From the data provided to Intercatch in 2022, discards are composed of individuals measuring less than 20 cm (Figure 10.4).

### 10.5 Survey data

In DATRAS, five surveys report length measurements of red mullet in the stock area since 1993. Survey data is available in DATRAS during 1997-2021 for the French EVHOE survey, covering the Bay of Biscay and the Celtic Sea, during 2001-2016 for the northern Spanish groundfish survey (SP-NSGFS), and from 2002 onwards for the Portuguese groundfish survey (PT-IBTS), covering the Portuguese coast (Figure 10.5). For this working group in the WGWIDE data call framework, indices from two surveys are provided: French EVHOE indices (survey map in Figure 10.6) and Northern Spanish Shelf surveys (survey map in Figure 10.7).

Two indices from the EVHOE survey are variable around the series mean between 1987-2011 before falling to a lower level thereafter and reaching average values during the last three years (Figure 10.8). Abundance and biomass indices in the Spain NSGFS are highly variable around the mean before 2017, then decreasing to a low level since 2017 (Figure 10.9). The mean stratified length distribution from Spain NSGFS is a bit higher in 2022 than in the previous three years but still very low compared with the last 30 years (Figure 10.10).

### 10.6 Biological sampling

In the Bay of Biscay, sexual maturity and length measures were taken in 2009 by AZTI. Since 2004, data (age, length, sexual maturity) are usually collected by France for the Eastern English Channel and the southern North Sea. France started to collect data for 8a,b at the end of 2007. In 2007-2008, the striped red mullet otolith exchange aimed to optimize age estimation between countries. In 2011, an Otolith Exchange Scheme was carried out, the second exercise for the Striped red mullet. Four readers of this exchange interpreted an image collection from the Bay of Biscay, the Spanish and the Mediterranean coasts (Spain and Italy). A set of Mullus surmuletus otoliths ( $\mathrm{N}=75$ ) from the Bay of Biscay presented the highest percentage of agreement $(82 \%)$. On 75 otoliths, 34 were read with $100 \%$ agreement ( $45 \%$ ) and thus a CV of $0 \%$. The modal age of these fishes was comprised between 0 and 3 years (Mahe et al., 2012).

### 10.7 Current research programs

Two research projects are currently investigating (1) the evolution of striped red mullet abundance indices from fishery-dependent data and (2) the temporal evolution of the size and age at maturity for this species in the Bay of Biscay. The first research project (ACOST) extends the analysis presented in
Caill-Milly et al. (2017) and Caill-Milly et al. (2019) and computes 5 abundances indices from 2005 to 2022 based on the landings per unit effort for 5 french fleets. The second project (MATO) updates the maturity data for the species in the Bay of Biscay thanks to a monthly longitudinal study over one reproduction cycle between 2021 and 2022. The final results will be published in 2023, and the references will be added to the next report.

### 10.8 Analysis of stock trends/assessment

An age-structured analytical stock assessment has yet to be developed due to the quality of the data available in the ICES database and the limited size of the available time series.

### 10.8.1 Data requirements

Regular sampling of biological parameters of striped red mullet catches must be continued under DCF. A benchmark to update the available information on this species and to assess the quality of the data presently available in Intercatch and DATRAS would be beneficial to the description of the ongoing fishery, as this species is supposed to expand its distribution area with the warming of the North Atlantic waters (Cheung et al., 2012). Recent studies show a link between the environmental parameters and the catch rates in the fishery for this species (Leitão, 2023).

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Table 10.1: Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8 and 9a. Official landings by country in tonnes.

| Year | Belgium | France | Guernsey | Ireland | Jersey | Netherlands | Portugal | Spain | UK | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2,006 | 33 | 1,947 | 8 | 16 | 1 | 115 | 10 | 387 | 170 | 2,688 |
| 2,007 | 43 | 1,941 | 9 | 23 | 1 | 148 | 222 | 398 | 194 | 2,978 |
| 2,008 | 26 | 1,394 | 9 | 22 | 0 | 165 | 169 | 394 | 165 | 2,345 |
| 2,009 | 20 | 1,562 | 5 | 16 | 0 | 110 | 199 | 520 | 134 | 2,567 |
| 2,010 | 20 | 1,743 | 5 | 8 | 0 | 128 | 276 | 479 | 133 | 2,793 |
| 2,011 | 21 | 1,740 | 0 | 8 | 0 | 130 | 245 | 508 | 155 | 2,806 |
| 2,012 | 37 | 1,342 | 0 | 7 | 1 | 125 | 217 | 332 | 122 | 2,183 |
| 2,013 | 28 | 932 | 5 | 4 | 0 | 50 | 187 | 246 | 71 | 1,522 |
| 2,014 | 12 | 926 | 5 | 2 | 0 | 2 | 221 | 265 | 53 | 1,487 |
| 2,015 | 23 | 1,215 | 5 | 3 | 0 | 111 | 282 | 248 | 102 | 1,989 |
| 2,016 | 28 | 1,179 | 0 | 4 | 0 | 69 | 204 | 194 | 83 | 1,761 |
| 2,017 | 36 | 997 | 0 | 10 | 0 | 13 | 154 | 327 | 64 | 1,601 |
| 2,018 | 37 | 896 | 0 | 0 | 0 | 95 | 122 | 321 | 67 | 1,538 |
| 2,019 | 30 | 1,358 | 0 | 12 | 0 | 91 | 159 | 267 | 55 | 1,973 |
| 2,020 | 50 | 965 | 0 | 6 | 0 | 82 | 109 | 261 | 89 | 1,562 |
| 2,021 | 53 | 836 | 0 | 18 | 0 | 54 | 117 | 274 | 93 | 1,445 |
| 2,022 | 43 | 1,057 | 0 | 20 | 1 | 18 | 136 | 336 | 81 | 1,691 |

Table 10.2: Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8 and 9a. Official landings by area in tonnes.

| Year | 6.a | 7.a | 7.6 | 7.c | 7.e | 7.f | 7.8 | 7.h | 7.j | 7.k | 8.a | 8.6 | $8 . c$ | 8.d | 9.a | $8 . e$ | 6.6 | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2,006 | 0 | 1 | 1 | 0 | 869 | 50 | 24 | 103 | 11 | 0 | 1,023 | 468 | 71 | 28 | 39 | 0 | 0 | 2,688 |
| 2,007 | 1 | 1 | 1 | 1 | 1,047 | 54 | 22 | 104 | 24 | 0 | 861 | 473 | 90 | 32 | 267 | 0 | 0 | 2,978 |
| 2,008 | 0 | 1 | 1 | 0 | 880 | 46 | 16 | 72 | 26 | 0 | 639 | 246 | 86 | 35 | 296 | 0 | 0 | 2,345 |
| 2,009 | 2 | 1 | 2 | 2 | 592 | 25 | 9 | 74 | 35 | 0 | 879 | 460 | 156 | 88 | 243 | 0 | 0 | 2,567 |
| 2,010 | 2 | 1 | 3 | 2 | 642 | 26 | 10 | 59 | 32 | 1 | 1,033 | 467 | 146 | 38 | 331 | 0 | 0 | 2,793 |
| 2,011 | 1 | 1 | 0 | 0 | 665 | 20 | 10 | 55 | 11 | 0 | 970 | 513 | 214 | 35 | 310 | 0 | 1 | 2,806 |
| 2,012 | 0 | 0 | 0 | 0 | 493 | 23 | 7 | 34 | 9 | 0 | 696 | 387 | 200 | 53 | 280 | 0 | 0 | 2,183 |
| 2,013 | 0 | 0 | 1 | 0 | 232 | 23 | 7 | 36 | 4 | 0 | 473 | 328 | 166 | 12 | 241 | 0 | 0 | 1,522 |
| 2,014 | 1 | 0 | 0 | 0 | 192 | 15 | 3 | 40 | 3 | 0 | 523 | 240 | 151 | 23 | 297 | 0 | 0 | 1,487 |
| 2,015 | 0 | 0 | 1 | 0 | 595 | 10 | 2 | 36 | 2 | 0 | 506 | 327 | 126 | 15 | 369 | 0 | 0 | 1,989 |
| 2,016 | 0 | 0 | 2 | 0 | 417 | 21 | 7 | 35 | 5 | 0 | 548 | 311 | 117 | 21 | 277 | 0 | 0 | 1,761 |
| 2,017 | 0 | 0 | 1 | 0 | 277 | 27 | 21 | 37 | 3 | 0 | 514 | 324 | 160 | 5 | 231 | 0 | 0 | 1,601 |
| 2,018 | 0 | 0 | 0 | 0 | 361 | 26 | 7 | 39 | 1 | 0 | 453 | 276 | 144 | 2 | 226 | 0 | 0 | 1,538 |
| 2,019 | 0 | 1 | 1 | 0 | 377 | 23 | 20 | 35 | 1 | 0 | 770 | 388 | 123 | 4 | 229 | 0 | 0 | 1,973 |
| 2,020 | 0 | 2 | 1 | 0 | 386 | 43 | 18 | 40 | 4 | 0 | 502 | 265 | 128 | 3 | 170 | 0 | 0 | 1,562 |
| 2,021 | 0 | 1 | 0 | 0 | 302 | 52 | 30 | 54 | 3 | 0 | 416 | 281 | 114 | 2 | 188 | 0 | 0 | 1,445 |
| 2,022 | 0 | 1 | 0 | 0 | 232 | 39 | 42 | 58 | 2 | 0 | 698 | 284 | 110 | 1 | 224 | 0 | 0 | 1,691 |

Figure 10.2: Striped red mullet in Subareas and Divisions 6, 7a-c, e-f, 8 and 9a. Official landings by area and country in tonnes.

Stock mur.27.67a-ce-k89a


Figure 10.3: Striped red mullet in Subareas and Divisions 6, 7a-c, e-f, 8 and 9a. Official fishing effort reported


Table 10.3: Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8 and 9a. Official discards by country in tonnes. Total is presented with the total discards rates in \%

| Year | UK | France | Belgium | Portugal | Spain | Ireland | Netherlands | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2,013 | 0 |  |  |  |  |  |  | 0 (0\%) |
| 2,014 |  | 98 |  |  |  |  |  | 98 (6.2\%) |
| 2,015 | 77 | 115 |  |  |  |  |  | 192 (8.8\%) |
| 2,016 | 171 | 213 | 1 | 0 | 8 |  |  | $\begin{aligned} & 394 \\ & (18.3 \%) \end{aligned}$ |
| 2,017 | 11 | 74 | 2 | 0 | 0 | 0 |  | 87 (5.1\%) |
| 2,018 | 14 | 35 | 3 | 0 | 2 | 0 |  | 53 (3.3\%) |
| 2,019 | 29 | 67 | 3 |  | 1 | 0 |  | 100 (4.8\%) |
| 2,020 | 39 | 28 | 4 |  | 1 | 9 | 0 | 82 (5\%) |
| 2,021 | 9 | 49 | 4 |  | 0 | 6 | 0 | 67 (4.5\%) |
| 2,022 | 109 | 162 | 4 |  | 0 | 0 | 0 | 275 (14\%) |

Figure 10.1: Striped red mullet in Subareas and Divisions 6, 7a-c, e-f, 8 and 9a. Landings by statistical rectangle for BEL, FRA, IRE, PT, UK (E\&W), UK (SCO) from 2014 to 2021 (Fishery Dependent Information database 2023).


Figure 10.4: Striped red mullet in Subareas and Divisions 6, 7a-c, e-f, 8 and 9a. Length distribution from 2014 to 2021 from Intercatch (D: Discards, L: Landings)


Figure 10.5: Percentage of hauls with striped red mullet length measurement, by year and survey. Black cross means no data.


Figure 10.7: NSGFS survey station map


Figure 10.11: EVHOE survey station map


Figure 10.12: Abundance and biomass indices (black line, with the grey ribbon for the standard deviation) striped red mullet in Subareas and Divisions 6, 7a-c, e-f, 8 and 9a., estimated from the EVHOE campaign, 1987-2022. The blue line is a loess regression underlining the trend in the series.

## EVHOE indices



Figure 10.13: Abundance and biomass indices (black line, with the grey ribbon for the standard error) striped red mullet in Subareas and Divisions 6, 7a-c, e-f, 8 and 9a., estimated from the NSGFS campaign, 1983-2022. The blue line is a loess regression underlining the trend in the series.

NSGFS indices


Figure 10.14: Main stratified length distributions of the striped red mullet in Subareas and Divisions 6, 7a-c, e-f, 8 and 9a., estimated from the NSGFS campaign, 1983-2022.


## Annex 1: List of participants

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## Annex 2: Resolutions

This resolution was approved on the resolution Forum in June 2023
2022/2/FRSG20 The Working Group on Widely Distributed Stocks (WGWIDE), chaired by Erling Kåre Stenevik, Norway, will meet 23-29 August 2023 in ICES HQ, Copenhagen, Denmark to:
a) Address generic ToRs for Regional and Species Working Groups.

The assessments will be carried out on the basis of the stock annex. The assessments must be available for audit on the first day of the meeting.

Material and data relevant to the meeting must be available to the group no later than 14 days prior to the starting date.

WGWIDE will report by 4 September for the attention of ACOM.

Only experts appointed by national Delegates or appointed in consultation with the national Delegates of the expert's country can attend this Expert Group

## Generic ToRs for Regional and Species Working Groups

The following ToRs apply to: AFWG, HAWG, NWWG, NIPAG, WGWIDE, WGBAST, WGBFAS, WGNSSK, WGCSE, WGDEEP, WGBIE, WGEEL, WGEF, WGHANSA and WGNAS.

## The working group should focus on:

a) Consider and comment on Ecosystem and Fisheries Overviews with a focus on:
i) identifying and correcting mistakes and errors (both in the text, tables, and figures);
ii) proposing concrete evidence-based input that is considered essential to the advice but is currently underdeveloped or missing (with references and Data Profiling Tool entries, as appropriate).

The input will feed into the annual updates of the overviews. Delivery of contributions other than those outlined above is also welcomed but will be utilized during the revision process (around every 5 years).
b) Conduct an assessment on the stock(s) to be addressed in 2023 using the method (assessment, forecast or trends indicators) as described in the stock annex; complete and document an audit of the calculations and results; and produce a brief report of the work carried out regarding the stock, providing summaries of the following where relevant:
i) Input data and examination of data quality; in the event of missing or inconsistent survey or catch information refer to the ACOM document for dealing with missing data and the linked template that formulates how deviations from the stock annex are to be reported;
ii) Where misreporting of catches is significant, provide qualitative and where possible quantitative information and describe the methods used to obtain the information;
iii) For relevant stocks (i.e. all stocks with catches in the NEAFC Regulatory Area), estimate the percentage of the total catch that has been taken in the NEAFC Regulatory Area in 2022;
iv) For category 3 and 4 stocks requiring new advice in 2023, implement the methods recommended by WKLIFE X (e.g. SPiCT, rfb, chr, rb rules) to replace the former 2 over 3 advice rule ( 2 over 5 for elasmobranchs). MSY reference points or proxies for the category 3 and 4 stocks (ICES technical guidance for harvest control rules and stock assessments for stocks in categories 2 and 3 );
v) Evaluate spawning-stock biomass, total-stock biomass, fishing mortality, and catches (projected landings and discards) using the method described in the stock annex:

1) For category 1 and 2 stocks, in addition to the other relevant model diagnostics, the recommendations and decision tree formulated by WKFORBIAS (see Annex 2) should be considered as guidance to determine whether an assessment remains sufficiently robust for providing advice.
2) If the assessment is deemed no longer suitable as basis for advice, provide advice using an appropriate Category 2-5 approach as described in ICES technical guidance for harvest control rules and stock assessments for stocks in categories 2 and 3 or ICES.
3) If the assessment has been moved to a Category 2-5 approach in the past year consider what is necessary to move back to a Category 1 and develop proposal for the appropriate benchmark process.
vi) Catch scenarios for the year(s) beyond the terminal year of the data for the stocks for which ICES has been requested to provide advice on fishing opportunities;
vii) Historical and analytical performance of the assessment and catch options with a succinct description of associated quality issues. For the analytical performance of
category 1 and 2 age-structured assessments, report the mean Mohn's rho (assessment retrospective bias analysis) values for time-series of recruitment, spawning-stock biomass, and fishing mortality rate. The WG report should include a plot of this retrospective analysis. The values should be calculated in accordance with the "Guidance for completing ToR viii) of the Generic ToRs for Regional and Species Working Groups - Retrospective bias in assessment" and reported using the ICES application for this purpose.
c) Produce a first draft of the advice on the stocks under consideration according to ACOM guidelines.
d) Review progress on benchmark issues and processes of relevance to the Expert Group:
i) update the benchmark issues lists for the individual stocks in SID;
ii) review progress on benchmark issues and identify potential benchmarks to be initiated in 2024 for conclusion in 2025;
iii) determine the prioritization score for benchmarks proposed for 2024-2025;
iv) as necessary, document generic issues to be addressed by the Benchmark Oversight Group (BOG).
e) Prepare the data calls for the next year's update assessment and for planned data evaluation workshops.
f) Identify research needs of relevance to the work of the Expert Group.
g) Review and update information regarding operational issues and research priorities on the Fisheries Resources Steering Group SharePoint site.
h) If not completed previously, complete the audit spreadsheet 'Monitor and alert for changes in ecosystem/fisheries productivity' for the new assessments and data used for the stocks. Also note in the benchmark report how productivity, species interactions, habitat and distributional changes, including those related to climate change, could be considered in the advice.
i) Deliver conservation status advice in accordance with the Technical guidelines on conservation status advice. The advice is only to be given when conservation aspects were identified and where clear, demonstrable management action can be recommended for any non-catch anthropogenic pressure. It can also be used to highlight clear demonstrable sensitivity to climate change. The qualification required to show clear, demonstratable management action is high. Avoid generic statements that are of no specific application to management.
j) Update SAG and SID with final assessment input and output.

Information on the stocks to be considered by each Expert Group is available here.

## Annex 3: Stock annex edits

| Stock ID | Stock name | Update | Link |
| :---: | :---: | :---: | :---: |
| boc.27.6-8 | Boarfish (Capros aper) in subareas 68 (Celtic Seas, English Channel, and Bay of Biscay) |  |  |
| gur.27.3-8 | Red gurnard (Chelidonichthys cuculus) in subareas 3-8 (Northeast Atlantic) |  |  |
| her.27.1-24a514a | Herring (Clupea harengus) in subareas $1,2,5$ and divisions 4 .a and 14.a, Norwegian spring-spawning herring (the Northeast Atlantic and Arctic Ocean) |  |  |
| hom.27.2a4a5b6a7a-ce-k8 | Horse mackerel (Trachurus trachurus) in Subarea 8 and divisions 2.a, 4.a, 5.b, 6.a, 7.a-c,e-k (the Northeast Atlantic) |  |  |
| hom.27.3a4bc7d | Horse mackerel (Trachurus trachurus) in divisions 3.a, 4.b-c, and 7.d (Skagerrak and Kattegat, southern and central North Sea, eastern English Channel) |  |  |
| mac.27.nea | Mackerel (Scomber scombrus) in subareas 1-8 and 14 and division 9.a (the Northeast Atlantic and adjacent waters) |  |  |
| mur.27.67a-ce-k89a | Striped red mullet (Mullus surmuletus) in subareas 6 and 8 , and divisions 7.a-c, 7.e-k, and 9.a (North Sea, Bay of Biscay, southern Celtic Seas, and Atlantic Iberian waters) |  |  |
| whb.27.1-91214 | Blue whiting (Micromesistius poutassou) in subareas 1-9, 12, and 14 (Northeast Atlantic and adjacent waters) | WGWIDE $2023$ | https://doi.org/10.17895/ices.pub. 24218679 |

## Annex 4: Audit reports

Format for audits (to be drawn up by expert groups and not review groups)
Review of ICES Scientific Report for WGWIDE meeting August 23-29 ${ }^{\text {th }}$, 2023.
Reviewers: Richard Nash, Anna H. Ólafsdóttir, and Joseph Watson
Expert group Chair: Erling K. Stenevik
Secretariat representative: David Miller
Audience to write for: advice drafting group, ACOM, and next year's expert group

## General

Recommendations, general remarks for expert groups, etc. (use bullet points and subheadings if needed)
This is an update assessment with advice provided in 2023 for 2024 and 2025.

## For single-stock summary sheet advice

Stock: boc.27.6-8.

Short description of the assessment as follows (examples in grey text):

1) Assessment type: update/SPALY
2) Assessment: was accepted. Category 3 with biennial advice
3) Forecast: not presented
4) Assessment model: Bayesian state space surplus production model fitted using catch data, 6 delta-lognormal estimated IBTS survey indices, and 1 acoustic survey estimate.
5) Consistency: This updated assessment is consistent with the assessment carried out in 2021. Minor difference in IBTS survey indices in recent years compared to 2021 assessment input data, changes to DATRAS.
6) Stock status: Fishing pressure on the stock is below FMSY proxy, and the stock size index is above Itrigger.
7) Management Plan: management plan agreed in 2016. In 2023, advice based on MSY approach instead of precautionary approach as per ICES guidelines (2022; method 2.1, rfb rule) for category 3 stocks.

## General comments:

- Clarify whether MSY or Precautionary and the use of the Management plan.

Technical comments:

- Advice sheet has minor errors in reference list and rounding of numbers.
- Assessment data file for length in the catch cannot be found on sharepoint, hence calculations for fproxy not checked.
- The data reported in the Advice sheet Table 8 are not in the report - this information should be in the report to allow cross validation.
- Report text reference list needs updating, many reference included which are not referred to. In text appears to be referred to wrong figure, see comments on report file.


## Conclusions

(Single tables or figures can be added in the text, longer texts should be added as annexes.)

Template for audit of assessments made by EG members<br>Text in italics is explanatory - to be deleted from final report

## Audit of Red Gurnard stock assessment

Date: 01.09.2023
Auditor: Joseph Watson and Are Salthaug

## General

Assessment of this stock is not possible due to a lack of reliable catch data. Red gurnard is mainly landed as by-catch by demersal trawlers in mixed fisheries, predominantly in divisions 7d, 7e and 7h. High discard rates and lack of resolution at a species level make interpretation of spatial trends in catches in other areas problematic.

Landings by country and divisions are available from 2006 to 2022 and discard data has been provided for 2015 - 2022. Information on red gurnard abundance is available from 6 surveys. A combined biomass index built on the time series from these using a delta-lognormal model.

For single stock summary sheet advice:

1) Assessment type: delta-lognormal assessment (from WKWEST)
2) Assessment: trend analyses
3) Forecast: not presented
4) Assessment model: surveys indices combined using a delta-lognormal model in an index of biomass to evaluate stock trend
5) Data issues: general lack of catch data reported at species level
6) Consistency: undefined
7) Stock status: undefined.
8) Management Plan: there is no management plan.

## General comments

The section of red gurnard is well structured and documented. The section includes a description regarding the lack of reporting data at species level and also the method used on the computation of a biomass index for this stock.

## Technical comments

## Conclusions

The combined biomass index appears to be correctly computed. There is no assessment for this stock.

Template for audit of assessments made by EG members
Text in italics is explanatory - to be deleted from final report

## Checklist for audit process

## General aspects

- Has the EG answered those TORs relevant to providing advice?
- Is the assessment according to the stock annex description?
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary?
- Have the data been used as specified in the stock annex?
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex?
- Is there any major reason to deviate from the standard procedure for this stock?
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice?


# Audit of Western Horse Mackerel data and assessment 

Date: 03/09/2023
Auditor: Leif Nøttestad, Sólvá Káradóttir Eliasen, Steve Mackinson

## General

Westem horse mackerel is assessed as a Category 1 stock. A Stock Synthesis (SS3) model is nun to determine the state of the stock in relation to reference points for westem horse mackerel.

For single stock summary sheet advice:

1) Assessment type: update
2) Assessment analytical.
3) Forecast: presented
4) Assessment model: SS3 model with commercial catches (length and age data) and three survey indices: Triennial egg survey index (1992-2019); IBTS recruitment index; PELACUS acoustic biomass.
5) Data issues: No data issues.
6) Consistency: The view of the WG was that the assessment should be accepted.
7) Stock status: Fishing pressure on the stock is above $\mathrm{F}_{\text {MSY }}$ but below $\mathrm{F}_{\mathrm{pa}}$ and $\mathrm{F}_{\text {lim; }}$; spawning-stock size is below MSY $\mathrm{B}_{\text {tigger, }}, \mathrm{B}_{\mathrm{pa}}$ and $\mathrm{B}_{\text {lim }}$.
8) Management Plan: No management plan

## General comments

The assessment and forecast have been available for review. Input and output data were correct.

## Technical comments

There are four tables in the tables section to which there, in the text section, are no references (Tables 7.2.4.3-7.2.4.6).
Advice, Table 8, "Total western stock" column is not identical to what is in the input sheets and not identical to what is shown in advice sheet, Figure 1, "SSB panel"
With regards to the WGWIDE report: It's difficult to make audit when the report is still not finished and the audit quality is not as what it would have been if the text was finished.

## Conclusions

The assessment has been performed correctly.

## Checklist for audit process

## General aspects

- Has the EG answered those TORs relevant to providing advice? Yes
- Is the assessment according to the stock annex description? Yes, but it needs to be updated
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary? Yes, no management plan
- Have the data been used as specified in the stock annex? Yes
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex?
Yes
- Is there any major reason to deviate from the standard procedure for this stock? No
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice?
Yes, but partly No. A zero catch advice on WHOM for 2024, as done for the first time for 2023, raises some fundamental questions and practical concerns that need to be solved. In the advice sheet for 2024 there are three new important issues raised this year, which was not included in last years advice sheet for WHOM:

1) A comprehensive genetic research project on stock identification of horse mackerel has being carried out, and initial results suggest that the boundaries of the stocks require revision.
2) A scientific monitoring quota should be considered to support the stock assessment and the advice in future years.
3) Mixed fisheries considerations might be relevant due to horse mackerel being taken as by-catch in some areas.

The first issue influence the quality of the assessment, and the other two are issues relevant for the advice. All these issues need to be dealt with quantitatively during the upcoming benchmark of WHOM in 2024. A sentence stating that this will be dealt with during the benchmark in 2024, will be useful to add to the advice sheet at the end of section issues relevant for the advice.

## Template for audit of assessments made by EG members

Audit of North Seas Horse mackerel stock (hom.27.3a4bc7d)
Date: 29/08/2023-04/09/2023
Auditor: Aril Slotte, Eydna í Homrum

## General

Biannual advice for 2024 and 2025

For single stock summary sheet advice:
Stock: North Sea Horse mackerel
Assessment type: SALY - Survey trends based assessment
Assessment: Survey index and length-based analysis with F/Fmsy proxy, rfb-rule approach
Forecast: There is no forecast in the applied DLS approach
Assessment model:
Consistency:
Stock status: no reference points for stock size have been defined for this stock.
Management plan: There is no management plan for horse mackerel in this area. Advice basis is MSY-approach.

## General comments

The text is well written and gives the reader a good overview of the stock, input data and methodology behind assessment and advice. The recent development in genetics suggest that all horse mackerel caught in 4a belong to the Western stock, with additional mixture between Western and North sea in 7d, which likely bias this stock assessment significantly. This is issue is, however, well presented in the report and will also have focus the upcoming benchmark.

For a general reader, given the suspected mixtures, some comparisons on the biology between Western and North Sea stocks would have been interesting, such as weight and length at age, length/age at maturity, to better grasp the differences in life history strategies.

I find chapter 6.3 the Biology chapter to be lacking valuable information. It is difficult dot understand that a stock that has been monitored over so many years have no information on maturity ogives, maturation at age/length. It is also strange that the Western stock has old fish that are almost absent in the North Sea stock. In a biological sense this is hard to grasp, and it is not described or discussed anywhere in the text. Is this linked to sampling, aging, high exploitation, high natural mortality, or simply uncertainties in stock structure, annual migrations etc??
PS! I noticed in tables 6.3.3-4, that the weight/length at ages $7-9$ in 4 a were significantly larger than in all the other ICES areas, so much that it potentially may be errors, which can be looked further into.

## Technical comments

Advised catch value matches catch scenarios
Advice years are 2024 and 2025 - correct
Units in plots are same as 2021 -advice. Titles are correct/meaningful. Legends are correct/meaningful and caption is descriptive (however, $\mathrm{I}_{\text {tigger }}$ is not described in caption ...). Values in plots match summary table in end of advice.

## Template for audit of assessments made by EG members

Catch and SSB are given in tonnes. A short advice change description is given with catch scenarios
Basis of advice (MSY-approach with proxy MSY reference point) is described consistently throughout the advice sheet.
Basis of assessment is last benchmarked in 2017 - this is consistent in text and tables. DLSapproach has, however, changed and ICES advice is now based on MSY-approach as opposed to Precautionary approach in last advice (2021).
History of advice is consistent with last year's advice and advice for 2024 matches headline advice.
The references for ICES framework for category 3 stocks appear to be missing in the reference list...
(From quality of the assessment: 'The advice framework changed from the previous ICES framework for category 3 stocks (ICES, 2012), to the new framework-for DLS using the rfb rule; method 2.1 (ICES, 2022). ')

## Conclusions

No inconsistency was seen, with exception of what is mentioned above in Tables 6.3.3-4 and the reference list.

## Template for audit of assessments made by EG members

1 Audit of the NEA Mackerel assessment (mac.27.nea)
Date: 01/09/2023
Auditor: Patricia, Jan Arge, Niels

General
The mackerel assessment is carried out using the SAM model and runs locally at the machine of the stock assessor but is uploaded to the sharepoint under Data. This audit focuses on input data, assessment, forecast and draft advice document.

ICES no longer consider the NEA mackerel stock to consist of three spawning components, i.e. the westerm, the southern, and the North Sea components, as in previous assessments (see section 8.11 Manazement Considerations). This has led to removal of all reference to components in the report and advice sheet.

## For single stock summary sheet advice:

1) Assessment type: updated/SALY
2) Assessment: analytical
3) Forecast: presented; derived directly from the outputs of the SAM model Appropriate settings according to Stock Annex.
4) Assessment model: SAM with 3 survey fleet and tagging data
5) Data issues: Russian data was made available through Norway, no further data issues
6) Consistency: The retrospective bias ( 10 years considered), where the F has consistently been overestimated and SSB underestimated, is still present in older years but has become less apparent in recent years. The reason to this behaviour is vears but has become less apparent in recent years. The reason of the stock due to the $\frac{\text { only partly understood. }}{\text { addition of new tagging data }}$
Stock status: Above biomass trigger and above Fmsy
7) Stock status: Above biomass trigger and above Fmsy
8) Management Plan: none agreed. ICES based its advice on the MSY approach.

## General comments

The assessment is performed in line with the stock annex; there were some minor issues with input data on the egg survey data which were resolved during the working group. The stock would benefit from moving to TAF and sourcing more directly its input data from established/referenced sources. At this stage, tracing back the input data for the auditors was difficult and relied completely on the provision of data from the stock assessor. In this situation, the auditors could only check if the assessor entered the provided data into the assessment correctly, but not if the source of the data up to the input data was consistent.

The report is very detailed and very well written, allowing a comprehensive understanding of the particularities of the input data used and on the model configurations. The advice sheet is well documented.

## Technical comments

Explorations on alternative model settings by scientists other than the stock assessor were performed during the WG but the process was rather untransparent and adding these analyses to the report (not available yet) may not be appropriate

## Conclusions

The assessment was carried out in line with the SA and catch and survey data was in line with the data used in the assessment.

Commented [JanArge 1]: We discussed the systematic upward revision of the stock with every successive assessment at WGWIDE,
but were not able to explain this behavicur satisfactorily We have eq underestimated the SSB by almost one third back in 2014 in relation to today (see my suggested text ab ove Thus Wonder it the added tagging data have caused this
retrospective behaviour the last 10 or $s$ y years? Or have I misunderstood this statement?

## Template for audit of assessments made by EG members

## General aspect <br> Checklist for audit process

- Has the EG answered those TORs relevant to providing advice? Yes
- Is the assessment according to the stock annex description? Yes
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary? There is no active agreed management plan by the coastal states and unilateral quota setting exceeding the ICES advice has been the case for the past decade.
- Have the data been used as specified in the stock annex? Yes
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex? Yes
- Is there any major reason to deviate from the standard procedure for this stock? No
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice? Yes


## Format for audits

Review of ICES Scientific Report, (WGWIDE) (2023) (30/08/2023)
Reviewers: losu Paradinas and Åge Høines

Expert group Chair:
Secretariat representative

Audience to write for: advice drafting group, ACOM, and next year's expert group
General

Stock: Striped red mullet in Subareas and Divisions 6, 7a-c, e-k, 8, and 9a

Short description of the assessment as follows (examples in grey text):

1) Assessment type: update
2) Assessment: no assessment due to sparsity and uncertainty in historical data
3) Forecast: not presented
4) Assessment model: none
5) Consistency: undefined
6) Stock status: undefined
7) Management plan: there is no agreed management plan

General comments
Figure 1 still missing.

Technical comments
Conclusions

## Template for audit of assessments made by EG members

## 1 Audit of the Norwegian Spring spawning herring

 (her.27.1-24a514a)Date: 29 August 2023
Auditor: Alessandro Orio, Richard Nash, Claus R. Sparrevohn

## General

Norwegian Spring Spawning Herring is assessed as a Category 1 stock. A XSAM model is run to determine the state of the stock in relation to reference points for Norwegian Spring Spawning Herring.

For single stock summary sheet advice:

1) Assessment type: update/SPALY
2) Assessment: analytical
) Forecast: presented
3) Assessment model: XSAM with 3 survey fleets. The recruitment survey was not undertaken in 2021 nor 2022.
4) Data issues: No data issues.
5) Consistency: The assessment is consistent with last year and the decrease in biomass was also predicted in 2022 although it was a little more pronounced in the 2023 assessment.
6) Stock status: Fishing pressure on the stock is above $\mathrm{F}_{\text {MSY }}$ and $\mathrm{F}_{\mathrm{pa}}$ but below $\mathrm{F}_{\text {lim }}$. The stock is above MSY $\mathrm{B}_{\text {tigger }}$ in 2023 but is predicted to fall below MSY $\mathrm{B}_{\text {tnigger }}$ in 2024 ( $1^{\text {t }}$ of January).
7) Management Plan: There is an agreed management plan which has been evaluated by ICES and found precautionary.

## General comments

The assessment and forecast were available for review. However, the assessment is a XSAM model which at present only can run on one designated MAC computer. Therefore, it is, at present, not possible to re-run the assessment and the forecast but just check input and configuration. A SAM version of the assessment (with the same settings and properties as the XSAM) is developed and shows - except for some very minor details - the same result. The same is the case for the forecast. Therefore, the EG highly recommended moving from using XSAM to the SAM platform to avoid the reliance on using software which is not maintained through the ongoing R updates and to increase reproducibility and transparency. Input and output data were correct. The report was also available for the audit. A few inconsistencies were found in the advice sheet and report and are outlined in the technical comments.

## Technical comment

Advice sheet:

- Suggest changing the formulation "Fishing pressure on the stock is above $\mathrm{F}_{\text {MSY }}$ and between $F_{p a}$ and $F_{\text {lim" }}$ to "Fishing pressure on the stock is above $F_{\text {MSY }}$ and $F_{p a}$ but below $\mathrm{F}_{\text {lim }}$ ".
- In table 6 year 2024 it says "Follow the management strategy, $\mathrm{F}_{\text {met }}=0.123$ and $\mathrm{B}_{\text {mgt }}=3.184$ million tonnes". $\mathrm{F}_{\text {myt }}$ should be changed to 0,124 as the $\mathrm{F}_{\text {met }}$ according to the LTMP is 0,123614 .
- Some comments added into the advice sheet by Coby - suggested responses given.

Commented [AO1]: The sentence in the advice sheed is the
tandard sentence that is in the ICES guidelines so I dont think we tandaran sent
can change it

## Template for audit of assessments made by EG members

- The order of the Figures in chapter 4.5 .1 is not following the order in which they are presented in the text. This is only minor but could be easily fixed.


## Conclusions

The assessment has been performed correctly

## Template for audit of assessments made by EG members

## Checklist for audit process

General aspect

- Has the EG answered those TORs relevant to providing advice? Yes.
- Is the assessment according to the stock annex description? Yes
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary? Yes. The $\mathbf{F}_{\text {mgt }}$ is below FMSY and the HCR different.
- Have the data been used as specified in the stock annex? Yes
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex? Yes
- Is there any major reason to deviate from the standard procedure for this stock? No

Does the update assessment give a valid basis for advice? Yes

## Template for audit of assessments made by EG members

1 Audit of the Blue whiting assessment (whb.27.1-91214)
Date: 31/08/2023
Auditor: Afra Egan, Gersom Costas, Sigurvin Bjarnason

## General

The blue whiting assessment is carried out using the SAM model and is available on Stockassessment.org (WHB-2023) and on SharePoint. This audit focuses on input data, assessment, forecast and draft advice document.

For single stock summary sheet advice:
Assessment type: update/SALY
Assessment: analytical
3) Forecast: presented; derived directly from the outputs of the SAM model. Appropriate settings according to Stock Annex
4) Assessment model: SAM with 1 survey index
5) Data issues: The final data for 2022 are presented in the report. Data for 2023 are preliminary but are included in the model. Data sources are described in the stock annex.
6) Consistency: The assessment is consistent with last years with an upward trend in SSB. There is a revision upward in recruitment at age 1 in 2022 due to a high survey index value and high catch at age of the 2021 year class. The WG accepted the assessment.
7) Stock status: The fishing pressure on the stock is above $\mathrm{F}_{\text {msy, }}$, $\mathrm{F}_{\text {MGT }}$ and Fpa (but below $\mathrm{F}_{\text {lim }}$ ). Spawning-stock size is above MSY $\mathrm{B}_{\text {trigger, }}, \mathrm{B}_{\mathrm{pa}}$, and $\mathrm{Blim}_{\text {lim }}$
8) Management Plan: Agreed by the Coastal States in October 2016 after evaluation of the management strategy by ICES. The long-term management strategy was found to be consistent with the precautionary approach. According to the strategy F is set at $\mathrm{F}_{\text {MSY }}$ when SSB is forecast to be above or equal to $\mathrm{B}_{\text {trigger, }}$ which is the case for 2024 .

## General comments

The input data and assessment are documented as described in the stock annex and the report sections are well ordered.

## Technical comments

In The stock annex Fpa is 0.53 . This should be updated to 0.32 as per the advice sheet. The technical basis should be updated also.

## Conclusions

The assessment has been performed correctly.

## Template for audit of assessments made by EG members

## Checklist for audit process

## General aspects

- Has the EG answered those TORs relevant to providing advice? Yes.
- Is the assessment according to the stock annex description? Yes (in some cases the SA will need minor updates) Yes
- If a management plan is used as the basis of the advice, has been agreed to by the relevant parties and has the plan been evaluated by ICES to be precautionary? Yes
- Have the data been used as specified in the stock annex? Yes
- Has the assessment, recruitment and forecast model been applied as specified in the stock annex? Yes
- Is there any major reason to deviate from the standard procedure for this stock? No
- Does the update assessment give a valid basis for advice? If not, suggested what other basis should be sought for the advice? Yes


## Annex 5: Working documents

## Working Document to

ICES Working Group on Widely Distributed Stocks (WGWIDE, No. 1)
ICES HQ, Copenhagen, Denmark, (hybrid meeting) 23. - 29. August 2023

## Preliminary cruise report from the International Ecosystem Summer Survey in the Nordic Seas (IESSNS) $1^{\text {st }}$ July $-3^{\text {rd }}$ August 2023



Leif Nøttestad, Hector Peña, Age Høines, Kjell Rong Utne, Susanne Tonheim, Stine Karlson, Are Salthaug Institute of Marine Research, Bergen, Norway

Anna Heiða Ólafsdóttir, Thassya Christina dos Santos Schmidt, James Kennedy
Marine and Freshwater Research Institute, Hafnarfjörður, Iceland
Eydna í Homrum, Leon Smith
Faroe Marine Research Institute, Tórshavn, Faroe Islands

## Teunis Jansen,

Greenland Institute of Natural Resources, Nuuk, Greenland
Kai Wieland
National Institute of Aquatic Resources, Denmark
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The International Ecosystem Summer Survey in the Nordic Seas (IESSNS) was performed within approximately 5 weeks from July $1^{\text {st }}$ to August $3^{\text {rd }}$ in 2023 using five vessels from Norway (2), Iceland (1), Faroe Islands (1) and Denmark (1). The main objective is to provide annual age-segregated abundance index, with an uncertainty estimate, for northeast Atlantic mackerel (Scomber scombrus). The index is used as a tuning series in stock assessment according to conclusions from the 2017 and 2019 ICES mackerel benchmarks. A standardised pelagic swept area trawl method is used to obtain the abundance index and to study the spatial distribution of mackerel in relation to other abundant pelagic fish stocks and to environmental factors in the Nordic Seas, as has been done annually since 2010. Another aim is to construct a new time series for blue whiting (Micromesistius poutassou) abundance index and for Norwegian springspawning herring (NSSH) (Clupea harengus) abundance index. This is obtained by utilizing standardized acoustic methods to estimate their abundance in combination with biological trawling on acoustic registrations. The time series for blue whiting and NSSH now consists of eight years (2016-2023).

The total swept-area mackerel index in 2023 was 4.30 million tonnes in biomass and 10.67 billion in numbers, a decline by $42 \%$ for biomass and $39 \%$ for abundance compared to 2022. In 2023, the most abundant year classes were 2020, 2019, respectively. The cohort internal consistency improved compared to last year, particularly for ages 4-7 years. The catch curves showed clear year effects, and that mackerel of ages 1,2 and to some extent also age 3, are not completely recruited to the survey. Most of the surveyed mackerel are still distributed in the Norwegian Sea. However, they were more easterly and northeasterly distributed compared to 2022. The distribution of mackerel in the Norwegian Sea retracted compared to the last decade, particularly withdrawal from the northernmost part was observed. The zero-line was reached for the whole survey area, north of latitude $60^{\circ} \mathrm{N}$.
Other fish species were also monitored such as lumpfish (Cyclopterus lumpus), capelin (Mallotus villosus), polar cod (Boreogadus saida), and Atlantic salmon (Salmo salar). Lumpfish were caught at $76 \%$ of surface trawl stations distributed across the surveyed area from southwestern part of Iceland, central part of North Sea to southwestern-western part of the Svalbard. Both size and abundance were greater north of latitude $72^{\circ} \mathrm{N}$ compared to southern areas. Capelin were caught in the surface trawl on 29 stations along the cold fronts: north of Iceland, north- and northwest of Jan Mayen, northwest of Bear Island and west of Svalbard. There were more trawl stations with catches of capelin in the west and north of Jan Mayen than previous years The polar cod were caught in larger areas in the north and northeast of Iceland compared to the timeseries. A total of 62 North Atlantic salmon were caught in 24 stations both in coastal and offshore areas from $62^{\circ} \mathrm{N}$ to $74^{\circ} \mathrm{N}$ in the upper 30 m of the water column. The salmon ranged from 0.084 kg to 2.7 kg in weight, dominated by post-smolt and 1 sea-winter individuals. We caught from 1 to 12 salmon during individual surface trawl hauls. The length of the salmon ranged from 21 cm to 82 cm , with the highest fraction between 21 cm and 29 cm .
Satellite measurements of sea surface temperature (SST) in the Northeast Atlantic in July 2023 show that the northern regions of the Nordic Seas were slightly warmer than the average, while the East Greenland Current was cooler than the long-term average. The SST in the Irminger Sea and Iceland Basin were slightly warmer than the average.
The average zooplankton biomass increased in the Norwegian Sea and in Icelandic waters compared to 2022. Zooplankton showed patchy distribution throughout the area.

In the present preliminary report, no results of herring and blue whiting measurements are presented. A final survey report including these two species will be published in the fall of 2023.

During approximately five weeks of survey in 2023 ( $1^{\text {st }}$ of July to $3^{\text {rd }}$ of August), five vessels; the M/V "Eros" and M/V "Vendla" from Norway, "Jákup Sverri" operating from Faroe Islands, the R/V "Árni Friðriksson" from Iceland and M/V "Ceton", operating in the North Sea by Danish scientists, participated in the International Ecosystem Summer Survey in the Nordic Seas (IESSNS).
The major aim of the coordinated IESSNS was to collect data on abundance, distribution, migration, and ecology of Northeast Atlantic (NEA) mackerel (Scomber scombrus) during its summer feeding migration phase in the Nordic Seas and surrounding coastal and offshore waters. The resulting abundance index will be used in the stock assessment of NEA mackerel at the annual meeting of ICES working group of widely distributed stocks (WGWIDE). The IESSNS mackerel index time series goes back to 2010. Since 2016, systematic acoustic abundance estimation of both Norwegian spring-spawning herring (Clupea harengus) and blue whiting (Micromesistius poutassou) have also been conducted. This is considered as potential input for stock assessment since the time series are sufficiently long. Furthermore, the IESSNS is a pelagic ecosystem survey collecting data on physical oceanography, plankton, and other fish species such as lumpfish, polar cod, and Atlantic salmon. Opportunistic whale observations were also recorded from Norway and Faroe Islands. The wide geographical coverage, standardization of methods, sampling on many trophic levels and international cooperation around this survey facilitates research on the pelagic ecosystem in the Nordic Seas, see e.g. Nøttestad et al. (2016), Jansen et al. (2016), Bachiller et al. (2018), Olafsdottir et al. (2019), Nikolioudakis et al. (2019), dos Santos Schmidt et al. (2023).
The methods have evolved over time since the survey was initiated by Norway in the Norwegian Sea in the beginning of the 1990s. The main elements of international standardization were conducted in 2010. Smaller improvements have been implemented since 2010. Faroe Islands and Iceland have participated in the joint mackerel-ecosystem survey since 2009. Greenland since 2013 and Denmark from 2018. Greenland did not participate in 2021 and 2023 but participated with their new research vessel R/V "Tarajoq" in 2022.
The North Sea was included in the survey area for the sixth time in 2023, following the recommendations of WGWIDE. This was done by scientists from DTU Aqua, Denmark. The commercial fishing vessels "Ceton S205" was used. No problems applying the IESSNS methods were encountered. Area coverage, however, was restricted to the northern part of the North Sea at water depths deeper than 50 m (see Appendix 1 for comparison with the 2018-2023 results).

## Material and methods

Coordination of the IESSNS 2023 was done during the WGIPS 2023 virtual meeting in January 2023, and by correspondence in December 2022 and during spring and summer 2023. The participating vessels together with their effective survey periods are listed in Table 1.
Overall, the weather conditions were calmer and less windy than usual for the two Norwegian vessels during the entire survey, thus providing very good survey progress as well as favourable conditions for the acoustic recordings and pelagic trawling. The Icelandic vessel, operating in Icelandic waters, experienced in general calm weather for duration of the survey with no survey delay, however three WP2-net sampling were skipped due to high winds. For the Faroese vessel, the survey was not hampered by weather; however technical issues with the trawl (a repair on land was needed) reduced the survey time with approximately two days in addition to skipped trawl stations in southwestern survey area. The chartered vessel Ceton had good weather conditions throughout the survey.
During the IESSNS, the special designed pelagic trawl, Multpelt 832, has been applied by all participating vessels since 2012. This trawl is a product of cooperation between participating institutes in designing and constructing a standardized sampling trawl for the IESSNS. The work was led by trawl gear scientist John Willy Valdemarsen, Institute of Marine Research (IMR), Bergen, Norway (Valdemarsen et al. 2014). The design of the trawl was finalized during meetings of fishing gear experts and skippers at meetings in

January and May 2011. Further discussions on modifications in standardization between the rigging and operation of Multpelt 832 was done during a trawl expert meeting in Copenhagen 17-18 August 2012, in parallel with the post-cruise meeting for the joint ecosystem survey, and then at the WKNAMMM workshop and tank experiments on a prototype (1:32) of the Multpelt 832 pelagic trawl, conducted as a sequence of trials in Hirtshals, Denmark from 26 to 28 February 2013 (ICES 2013a). The swept area methodology was presented and discussed during the WGISDAA workshop in Dublin, Ireland in May 2013 (ICES 2013b). The standardization and quantification of catchability from the Multpelt 832 pelagic trawl was further discussed during the mackerel benchmark in Copenhagen in February 2014. Recommendations and requests coming out of the mackerel benchmark in February 2014, were considered and implemented during the IESSNS survey in July-August 2014 and in the surveys thereafter. Furthermore, recommendations and requests resulting from the mackerel benchmark in January-February 2017 (ICES 2017), were carefully considered and implemented during the IESSNS survey in July-August 2017. In 2018, the Faroese and Icelandic vessels employed new, redesigned cod-ends with the capacity to hold 50 tonnes. This was done to avoid the cod-end from bursting during hauling of large catches as occurred at three stations in the 2017 IESSNS.

Table 1. Survey effort by each of the five vessels during the IESSNS 2023. The number of predetermined ("fixed") trawl stations being part of the swept-area stations for mackerel in the IESSNS are shown after the total number of trawl stations.

| Vessel | Effective survey <br> period | Length of cruise <br> track (nmi) | Total trawl stations/ <br> Fixed stations | CTD stations | Plankton stations |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Árni Friŏriksson | $3-21 / 7$ | 3250 | $43 / 38$ | 38 | 35 |
| Jákup Sverri | $1-16 / 7$ | 2825 | $31 / 27$ | 27 | 27 |
| Ceton | $4-13 / 7$ | 1987 | $36 / 39$ | 36 | - |
| Vendla | $4 / 7-3 / 8$ | 4077 | $66 / 57$ | 57 | 57 |
| Eros | $4 / 7-3 / 8$ | 3349 | $64 / 57$ | 57 | 57 |
| Total | $1 / 7-3 / 8$ | 15488 | $240 / 218$ | 215 | 176 |

### 3.1 Hydrography and Zooplankton

The hydrographical and plankton stations by all vessels combined are shown in Figure 1. Eros, Vendla, Árni Friðriksson and Jákup Sverri were all equipped with a SEABIRD CTD sensor and Árni Friðriksson and Jákup Sverri moreover also had a water rosette. Ceton used a Seabird SeaCat offline CTD. The CTD-sensors were used for recording temperature, salinity, and pressure (depth) from the surface down to 500 m , or to the bottom when at shallower depths.

Zooplankton was sampled with a WP2-net on 4 vessels, excluding Ceton which operates in the North Sea. Mesh sizes were $180 \mu \mathrm{~m}$ (Eros and Vendla) and $200 \mu \mathrm{~m}$ (Árni Friðriksson and Jákup Sverri). The net was hauled vertically from a depth of 200 m (or bottom depth at shallower stations) to the surface at a speed of $0.5 \mathrm{~m} / \mathrm{s}$. All samples were split in two, one half preserved for species identification and enumeration, and the other half dried and weighed. The zooplankton was sorted into three size categories ( $\mu \mathrm{m}$ ) , >2000, 10002000, 180/200-1000, on the Norwegian and Faroese vessels; and two size fractions ( $\mu \mathrm{m}$ ), > 1000 and 2001000, on the Icelandic vessel. Detailed description of the zooplankton and CTD sampling is provided in the survey manual (ICES 2014a).
Three planned WP2-plankton samples were not taken due to bad weather. The number of stations taken by the different vessels is provided in Table 1.

### 3.2 Trawl sampling

All vessels used the standardized Multpelt 832 pelagic trawl (ICES 2013a; Valdemarsen et al. 2014; Nettestad et al. 2016) for trawling, both for fixed surface stations and for trawling at greater depths to
confirm acoustic registrations and to target blue whiting registrations identified by echograms. Standardization of trawl deployment was emphasised during the survey as in previous years (ICES 2013a; ICES 2014b; ICES 2017). Sensors on the trawl doors, headrope and ground rope of the Multpelt 832 trawl recorded data, and allowed live monitoring, of effective trawl width (actually door spread) and trawl depth. The properties of the Multpelt 832 trawl and rigging on each vessel is reported in Table 2.
Trawl catch was sorted to the highest taxonomical level possible, usually to species for fish, and total weight per species recorded. The processing of trawl catch varied between nations. The Norwegian vessels sorted the whole catch to species but the Faroese vessel sub-sampled the catch before sorting if catches were more than 500 kg . Sub-sample size ranged from 90 kg (if it was clean catch of either herring or mackerel) to 200 kg (if it was a mixture of herring and mackerel); however, other species were mostly sorted out of the full catch. On the Icelandic vessel, the whole catch was sorted to species for all species except mackerel and herring when the catch is large ( $>1000 \mathrm{~kg}$ ) and mostly a mix of the two before mentioned species. Then approximately $10 \%$ of the mixed herring and mackerel catch is sorted to species.

The biological sampling protocol for trawl catch varied between nations in number of specimens sampled per station (Table 3).
Results from the survey expansion southward into the North Sea are analyzed separately from the traditional survey grounds north of latitude $60^{\circ} \mathrm{N}$ as per stipulations from the 2017 mackerel benchmark meeting (ICES 2017). However, data collected with the IESSNS methodology from the Skagerrak and the northern and western part of the North Sea are now available for six years (2018-2023).

Table 2. Trawl settings and operation details during the international mackerel survey in the Nordic Seas from $1^{\text {st }}$ July to $3^{\text {rd }}$ August 2023. The column for influence indicates observed differences between vessels likely to influence performance. Influence is categorized as 0 (no influence) and + (some influence).

| Properties | Árni Friouriksson | Vendla | Ceton | Jákup Sverri | Eros | Influence |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Trawl producer | Îsfell new trawl in 2023 | Egersund Trawl AS | Egersund Trawl AS | Vónin (2018) | Egersund Trawl AS | 0 |
| Warp in front of doors | Dynex-34 mm | Dynex -34mm | Dynex | Dynex - 38 mm | Dynex-34 mm | + |
| Warp length during towing | 350 | 350 | 270-320 | 350 (350-360) | 350 | 0 |
| Difference in warp length port/starb. (m) | 16 | 2-10 | 10 | 0-10 | 5-10 | 0 |
| Weight at the lower wing ends (kg) | $2 \times 400 \mathrm{~kg}$ | 2×400 | $2 \times 400$ | $2 \times 400$ | $2 \times 400$ | 0 |
| Setback (m) | 14 | 6 | 6 | 6 | 6 | + |
| Type of trawl door | Jupiter | Seaflex $7.5 \mathrm{~m}^{2}$ adjustable hatches | Thybron type 15 | Twister | Seaflex $7.5 \mathrm{~m}^{2}$ adjustable hatches | 0 |
| Weight of trawl door (kg) | 2200 | 1700 | 1970 | 1650 | 1700 | + |
| Area trawl door $\left(\mathrm{m}^{2}\right)$ | 6 | 7.5 with $25 \%$ hatches (effective 6.5) | 7 | 4.5 | 7 with $50 \%$ hatches (effective 6.5) | + |
| Towing speed (knots) mean (min-max) | 5.1 (4.6-5.6) | 4.6 (4.1-5.5) | 5.1 (4.5-5.8) | 4.7 (4-5.4) | 4.7 (4.1-5.725) | + |


| Trawl height ( m ) mean (min-max) | 31(23-37) | 25-32 | 30 (24-38) | 32.1 (25-51) | 25-32 | + |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Door distance <br> (m) mean (min-max) | 107 (110-130) | 121.8(118-126) | 125 (118-133) | 114 (102-122) | 135 (113-140) | + |
| Trawl width $(\mathrm{m})^{*}$ | 68.4 | 63.8 | 69.7 | 64.1 | 67.5 | + |
| Turn radius (degrees) | 5-10 | 5-12 SB turn | 5-10 | $5 \mathrm{BB} / \mathrm{SB}$ turn | 5-8 SB tum | + |
| Fish lock front of cod-end | Yes | Yes | Yes | Yes | Yes | + |
| Trawl door depth (port, starboard, m) (min-max) | 7-25, 7-21 | 6-22, 8-23 | 6-26, 6-26 | 5-17, 5-17 | 6-18, 8-20 | + |
| Headline depth (m) | 0 | 0 | 0 | 0 | 0 | + |
| Float arrangements on the headline | Kite +1 buoy on each wingtip | Kite with fender buoy +2 buoys on each wingtip | Kite with fender buoy +2 buoys on each wingtip | Kite with +1 buoys and kite on each wingtip | Kite +2 buoy on each wingtips | + |
| Weighing of catch | All weighed | All weighted | All weighed | All weighed | All weighted | + |

* calculated from door distance (Table 6)

Table 3. Protocol of biological sampling during the IESSNS 2023. Numbers denote the maximum number of individuals sampled for each species for the different determinations.

|  | Species | Faroes | Iceland | Norway | Denmark |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Length measurements | Mackere! | 200/100* | 150 | 100 | $\geq 125$ |
|  | Herring | 200/100* | 200 | 100 | 75 |
|  | Blue whiting | 200/100* | 100 | 100 | 75 |
|  | Lumpfish | all | all | all | all |
|  | Salmon | AII (1) | al! | all | - |
|  | Capelin | - | 50 | 25-30 |  |
|  | Other fish sp. | 20-50 | 50 | 25 | As appropriate |
| Weight, sex and | Mackere! | 15-25 | 50 | 25 | *** |
| maturity determination | Herring | 25-50 | 50 | 25 | 0 |
|  | Blue whiting | 15-50 | 50 | 25 | 0 |
|  | Lumpfish | 0-6 | $1^{\wedge}$ | 25 | 0 |
|  | Salmon | All | 0 | 25 | 0 |
|  | Capelin | - | 50 |  |  |
|  | Other fish sp. | 0-20 | 0 | 0 | 0 |
| Otoliths/scales collected | Mackere! | 15-25 | 25 | 25 | *** |
|  | Herring | 25-50 | 25 | 25 | 0 |
|  | Blue whiting | 15-50 | 50 | 25 | 0 |
|  | Lumpfish | 0 | $1^{\wedge}$ | 0 | 0 |
|  | Salmon | - | 0 | 0 | 0 |
|  | Capelin |  | 50 |  |  |
|  | Other fish sp. | 0 | 0 | 0 | 0 |
| Fat content | Mackere! | 0 | 10 | 0 | 0 |
|  | Herring | 0 | $10^{* *}$ | 0 | 0 |
|  | Blue whiting | 0 | 10 | 0 | 0 |
| Stomach sampling | Mackere! | 5 | 10 | 10 | 0 |
|  | Herring | 5 | $10^{* *}$ | 10 | 0 |
|  | Blue whiting | 5 | 10 | 10 | 0 |
|  | Other fish sp. | 0 | 0 | 10 | 0 |
| Tissue for genotyping | Mackere! | 0 | 0 | 0 | 0 |
|  | Herring | 0 | 0 | 25 | 0 |

*Length measurements / weighed individuals
**Sampled at every third station
${ }^{* * *}$ Up to one fish per cm-group $<25 \mathrm{~cm}$, two fish $25-30 \mathrm{~cm}$ and three fish $>30 \mathrm{~cm}$ from each station was weighed and aged.
${ }^{\wedge}$ All live lumpfish were tagged and released, only otoliths taken from fish which were dead when brought aboard.
This year's survey was quite well synchronized in time and was conducted over a relatively short period ( 33 days) given the large spatial coverage of around 2.4 million $\mathrm{km}^{2}$ (Figure 1). This was in line with recommendations put forward in 2016, that the survey period should be around four weeks with mid-point around $20^{\text {th }}$ of July. The main argument for this time-period was to make the IESSNS survey as synoptic as possible in space and time, and at the same time be able to finalize data and report for inclusion in the assessment for the same year.

## Underwater camera observations during trawling

M/V "Eros" and M/V "Vendla" employed an underwater video camera (GoPro HD Hero 5 Black Edition, www.gopro.com) to observe mackerel aggregation, swimming behaviour and possible escapement from the cod end and through meshes. The camera was put in a waterproof box which tolerated pressure down to approximately 100 m depth. No light source was employed with cameras; hence, recordings were limited to day light hours. Some recordings were also taken during night-time when there was midnight sun and good underwater visibility. Video recordings were collected at 74 trawl stations. The camera was attached on the trawl in the transition between 200 mm and 400 mm meshes.

### 3.3 Marine mammals

Opportunistic observations of marine mammals were conducted by scientific personnel and crew members from the bridge between $4^{\text {th }}$ July and $2^{\text {nd }}$ August 2023 onboard M/V "Eros" and M/V "Vendla", and onboard Jákup Sverri ( $30^{\text {th }}$ of June $-17^{\text {th }}$ of July) opportunistic observations were done from the bridge by crew members.

### 3.4Lumpfish tagging

Lumpfish caught during the survey by vessels R/V "Árni Frioriksson", M/V "Eros", M/V "Vendla" were tagged with Peterson disc tags and released. When the catch was brought aboard, any lumpfish caught were transferred to a tank with flow-through sea water. After the catch of other species had been processed, all live lumpfish larger than $\sim 15-20 \mathrm{~cm}$ were tagged. The tags consisted of a plastic disc secured with a titanium pin which was inserted through the rear of the dorsal hump. Contact details of Biopol (www.biopol.is) were printed on the tag. The fish were returned to the tank until all fish were tagged. The fish were then released, and the time of release was noted which was used to determine the latitude and longitude of the release location.

### 3.5Acoustics

## Multifrequency echosounder

The acoustic equipment onboard Vendla and Eros were calibrated $3^{\text {rd }}$ July 2023 for 18, 38, 70, 120 and 200 kHz . Árni Friðriksson was calibrated $4^{\text {th }}$ of May 2023 for frequencies 18, 38, 70, 120 and 200 kHz . Jákup Sverri was calibrated on $23^{\text {d }}$ March 2023 for 18, 38, 120, 200 and 333 kHz . Ceton did not conduct any acoustic data collection because no calibrated equipment was available, and acoustics are done in the same area and period of the year during the ICES coordinated North Sea herring acoustic survey (HERAS). All the other vessels used standard hydro-acoustic calibration procedure for each operating frequency (Foote 1987). CTD measurements were taken in order to get the correct sound velocity as input to the echosounder calibration settings.
Acoustic recordings were scrutinized to herring and blue whiting on daily basis using the post-processing software (LSSS, see Table 4 for details of the acoustic settings by vessel). Species were identified and partitioned using catch information, characteristic of the recordings, and frequency between integration on 38 kHz and on other frequencies by a scientist experienced in viewing echograms.
To estimate the abundance from the allocated NASC-values the following target strengths (TS) relationships were used.

Blue whiting: $\mathrm{TS}=20 \log (\mathrm{~L})-65.2 \mathrm{~dB}$ (rev. acc. ICES CM 2012/SSGESST:01)
Herring: $\mathrm{TS}=20 \log (\mathrm{~L})-71.9 \mathrm{~dB}$ (Foote, 1987)

Table 4. Acoustic instruments and settings for the primary frequency ( 38 kHz ) during IESSNS 2023.

|  | R/V Árni <br> Friơrriksson | M/V Vendia | R/V Jákup Sverri | M/V Eros |
| :---: | :---: | :---: | :---: | :---: |
| Echo sounder | Simrad EK80 | Simrad EK60 | Simrad EK80 | Simrad EK80 |
| Frequency ( kHz ) | 18,38,70, 120, 200 | 18,38,70,120, 200 | $\begin{aligned} & 18,38,70,120,200 \\ & 333 \end{aligned}$ | $\begin{aligned} & 18,38,70,120,200 \\ & 333 \end{aligned}$ |
| Primary transducer | ES38-7 | ES38B | ES38-7 | ES38B |
| Transducer installation | Drop keel | Drop keel | Drop keel | Drop keel |
| Transducer depth (m) | 9.6 | 8 | 6-9 | 6 |
| Upper integration limit (m) | 15 | 15 | 12 | 15 |
| Absorption coeff. ( $\mathrm{dB} / \mathrm{km}$ ) | 9.8 | 9.9 | 10.3 | 9.3 |
| Pulse length (ms) | 1.024 | 1.024 | 1.024 | 1.024 |
| Band width (kHz) | 2.425 | 2.43 | 3.06 | 2.43 |
| Transmitter power (W) | 2000 | 2000 | 2000 | 2000 |
| Angle sensitivity ( dB ) | 18 | 21.90 | 21.9 | 21.9 |
| 2-way beam angle ( dB ) | -20.30 | -20.70 | -20.4 | -20.7 |
| TS Transducer gain (dB) | 27.02 | 25.22 | 26.94 | 25.22 |
| Sa correction (dB) | 0.02 | -0.73 | -0.13 | -0.72 |
| 3 dB beam width alongship: | 6.43 | 6.88 | 6.47 | 6.85 |
| 3 dB beam width athw. ship: | 6.43 | 6.76 | 6.54 | 6.79 |
| Maximum range (m) | 500 | 500 | 500 | 500 |
| Post processing software | LSSS v.2.14.1 | LSSS 2.12.0 | LSSS 2.14.1 | LSSS 2.12.0 |

M/V Ceton: No acoustic data collection because other survey in the same area in June/July (HERAS).

## Multibeam sonar

Both M/V Eros and M/V Vendla were equipped with the Simrad fisheries sonar. Medium frequency CS90 sonar (frequency range: $70-90 \mathrm{kHz}$ ) on $\mathrm{M} / \mathrm{V}$ Eros and low frequency ST90 sonar (frequency range: 14-24 kHz ) on $\mathrm{M} / \mathrm{V}$ Vendla with a scientific output incorporated which allow the storing of the beam data for post-processing. Acoustic multibeam sonar data was stored continuously onboard Eros and Vendla for the entire survey.

## Cruise tracks

The five participating vessels followed predetermined survey lines with predetermined surface trawl stations (Figure 1). Calculations of the mackerel index are based on swept area approach with the survey area split into 10 strata, of which 6 are permanent ( $1,2,3,7,10$ and 13) and four dynamic (4, 5, 6 and 9 ) (Figure 2). Distance between predetermined surface trawl stations is constant within stratum but variable
between strata and ranged from 40 to 70 nmi . The survey design using different strata is done to allow the calculation of abundance indices with uncertainty estimates, both overall and from each stratum in the software program StoX (see Salthaug et al. 2017). Temporal survey progression by vessel along the cruise tracks in July-August 2023 is shown in Figure 3. The cruising speed was between 10-11 knots if the weather permitted, otherwise the cruising speed was adapted to the weather situation.


Figure $1 \mathbf{a}$. Fixed predetermined trawl stations and additional deep hauls included in the IESSNS from July 1 st to August 3rd 2023. At each station a 30 min surface trawl haul was performed.


Figure 1 b . Fixed predetermined hydrographic stations (CTD and WP2) included in the IESSNS from July $1^{\text {st }}$ to August $3^{\text {rd }}$ 2023. CTD station ( $0-500 \mathrm{~m}$ ) and WP2 plankton net samples ( $0-200 \mathrm{~m}$ depth).


Figure 2. Permanent and dynamic strata used in StoX for IESSNS 2023. The survey area is split into 10 strata, of which 6 are permanent ( $1,2,3,7,10$ and 13) and four dynamic ( $4,5,6$ and 9 ). The former stratum 8 (along the Norwegian coast) was merged into adjacent strata 1 and 7. Stratum 10 (northern Greenland waters) and 11 (southern Greenland waters) were not surveyed in 2023 and are not displayed. The former stratum 12 (offshore south of Iceland) is not used any longer, since the southern boundaries of strata 5 and 6 have been converted to dynamic boundaries. For original strata boundaries see WGIPS manual (ICES 2014a). In 2023, stratum 2 was split in two strata, 2 and 14, as two predetermined surface trawl stations were not sampled on the western end of the $2^{\text {nd }}$ transect from the south, see Figure 1a. Due to large variability in mackerel density within in stratum 2, the area around the skipped predetermined stations was defined as a separated stratum to reflect the mackerel density in the area. This was done to prevent inflation on mackerel abundance in the stratum 2 due to under sampling in a low-density part of stratum 2.


Figure 3. Temporal survey progression by vessel along the cruise tracks during IESSNS 2023: Blue represents effective survey start ( 1 'st of July) progressing to red representing a five-week span (survey ended $3^{\text {rd }}$ of August). As Ceton did not submit acoustics, they have been represented by station positions.

### 3.6 StoX

The recorded acoustic and biological data were analysed using the StoX software package which has been used for some years now for WGIPS coordinated surveys. A description of StoX can be found in Johnsen et al. (2019) and here: www.imr.no/forskning/prosjekter/stox. Mackerel swept-area abundance index, excluding the North Sea, was calculated using StoX version 3.6.1. The herring and blue whiting acoustic abundance indices were calculated using StoX version 3.6.2.

### 3.7 Swept area index and biomass estimation

This year the input data for the swept area calculations were taken from the ICES database. Up until 2020 the input data were extracted from the PGNAPES database.
The swept area age segregated index is calculated separately for each stratum (see stratum definition in Figure 2). Individual stratum estimates are added together to get the total estimate for the whole survey area which is approximately defined by the area between $60^{\circ} \mathrm{N}$ and $77^{\circ} \mathrm{N}$ and $40^{\circ} \mathrm{W}$ and $20^{\circ} \mathrm{E}$ in 2023. An additional run is made, including the North Sea. The density of mackerel on a trawl station is calculated by dividing the total number caught by the assumed area swept by the trawl. The area swept is calculated by multiplying the towed distance by the horizontal opening of the trawl. The horizontal opening of the trawl is vessel specific, and the average value across all hauls is calculated based on door spread (Table 5 and Table 6). An estimate of total number of mackerel in a stratum is obtained by taking the average density based on the trawl stations in the stratum and multiplying this with the area of the stratum.

Table 5. Descriptive statistics for trawl door spread, vertical trawl opening and tow speed for each vessel during IESSNS 2023 at predetermined surface trawl stations. Number of trawl stations used in calculations is also reported. Horizontal trawl opening was calculated using average vessel values for trawl door spread and tow speed (details in Table 6).

|  | Jákup Sverri | RV Árni <br> Fridriksson | Eros | Vendla | Ceton |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Trawl doors horizontal spread (m) |  |  |  |  |  |
| Number of stations | 27 | 29 | 57 | 57 | 36 |
| Mean | 113 | 122 | 122 | 112 | 125.2 |
| max | 120 | 130 | 136 | 120 | 132.7 |
| min | 102 | 110 | 115 | 100 | 117.7 |
| st. dev. | 4.2 | 4.6 | 4.8 | 4.0 | 3.8 |
| Vertical trawl opening (m) |  |  |  |  |  |
| Number of stations | 27 | 36 | 57 | 57 | 36 |
| Mean | 32 | 30.8 | 28 | 29 | 30.4 |
| max | 49 | 23.2 | 33 | 32.0 | 37.7 |
| min | 25 | 37.2 | 25 | 24 | 23.6 |
| st. dev. | 5.9 | 2.9 | 2.9 | 4.3 | 3.2 |
| Horizontal trawl opening (m) |  |  |  |  |  |
| Mean | 63.7 | 68.4 | 71.7 | 65.1 | 69.7 |
| Speed (over ground, nmi) |  |  |  |  |  |
| Number of stations | 27 | 38 | 57 | 57 | 36 |
| Mean | 4.7 | 5.1 | 4.5 | 4.6 | 5.1 |
| max | 5.5 | 5.6 | 5.2 | 5.3 | 5.8 |
| min | 3.4 | 4.6 | 4.2 | 4.2 | 4.5 |
| st. dev. | 0.5 | 0.2 | 0.5 | 0.3 | 0.3 |

Horizontal trawl opening was calculated using average vessel values for trawl door spread and tow speed (Table 6). The estimates in the formulae were based on flume tank simulations in 2013 (Hirtshals, Denmark) where formulas were developed from the horizontal trawl opening as a function of door spread, for two towing speeds, 4.5 and 5 knots:

Towing speed 4.5 knots: Horizontal opening (m) $=0.441$ * Door spread (m) + 13.094
Towing speed 5.0 knots: Horizontal opening (m) $=0.3959$ * Door spread ( m ) +20.094

Table 6. Horizontal trawl opening as a function of trawl door spread and towing speed. Relationship based on simulations of horizontal opening of the Multpelt 832 trawl towed at 4.5 and 5 knots, representing the speed range in the 2014 survey, for various door spread. See text for details. In 2017, the towing speed range was extended from 5.0 to 5.2, in 2020 the door spread was extended to 122 m and in 2022 the towing speed range was extended down to 4.3 knots and up to 5.5 knotsThe door spread was furthermore extended to 135 m in 2023. See also Appendix 4.

| Door spread(m) | Towing speed (knots) |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 4.3 | 4.4 | 4.5 | 4.6 | 4.7 | 4.8 | 4.9 | 5 | 5.1 | 5.2 | 5.3 | 5.4 | 5.5 |
| 100 | 56.5 | 56.9 | 57.2 | 57.7 | 58.2 | 58.7 | 59.2 | 59.7 | 60.2 | 60.7 | 61.2 | 61.7 | 62.2 |
| 101 | 56.9 | 57.3 | 57.6 | 58.1 | 58.6 | 59.1 | 59.6 | 60.1 | 60.6 | 61.1 | 61.5 | 62.0 | 62.5 |
| 102 | 57.3 | 57.7 | 58.1 | 58.6 | 59.0 | 59.5 | 60.0 | 60.5 | 60.9 | 61.4 | 61.9 | 62.4 | 62.9 |
| 103 | 57.7 | 58.1 | 58.5 | 59.0 | 59.5 | 59.9 | 60.4 | 60.9 | 61.3 | 61.8 | 62.3 | 62.8 | 63.2 |
| 104 | 58.1 | 58.5 | 59.0 | 59.4 | 59.9 | 60.3 | 60.8 | 61.3 | 61.7 | 62.2 | 62.7 | 63.1 | 63.6 |
| 105 | 58.6 | 59.0 | 59.4 | 59.9 | 60.3 | 60.8 | 61.2 | 61.7 | 62.1 | 62.6 | 63.0 | 63.5 | 63.9 |
| 106 | 59.0 | 59.4 | 59.8 | 60.3 | 60.7 | 61.2 | 61.6 | 62.1 | 62.5 | 62.9 | 63.4 | 63.8 | 64.3 |
| 107 | 59.5 | 59.9 | 60.3 | 60.7 | 61.2 | 61.6 | 62.0 | 62.5 | 62.9 | 63.3 | 63.8 | 64.2 | 64.6 |
| 108 | 59.9 | 60.3 | 60.7 | 61.1 | 61.6 | 62.0 | 62.4 | 62.9 | 63.3 | 63.7 | 64.1 | 64.6 | 65.0 |
| 109 | 60.4 | 60.8 | 61.2 | 61.6 | 62.0 | 62.4 | 62.8 | 63.2 | 63.7 | 64.1 | 64.5 | 64.9 | 65.3 |
| 110 | 60.9 | 61.2 | 61.6 | 62.0 | 62.4 | 62.8 | 63.2 | 63.6 | 64.1 | 64.5 | 64.9 | 65.3 | 65.6 |
| 111 | 61.3 | 61.7 | 62.0 | 62.4 | 62.8 | 63.2 | 63.6 | 64.0 | 64.4 | 64.8 | 65.2 | 65.6 | 66.0 |
| 112 | 61.8 | 62.1 | 62.5 | 62.9 | 63.3 | 63.7 | 64.0 | 64.4 | 64.8 | 65.2 | 65.6 | 66.0 | 66.3 |
| 113 | 62.2 | 62.6 | 62.9 | 63.3 | 63.7 | 64.1 | 64.4 | 64.8 | 65.2 | 65.6 | 66.0 | 66.3 | 66.7 |
| 114 | 62.7 | 63.0 | 63.4 | 63.7 | 64.1 | 64.5 | 64.9 | 65.2 | 65.6 | 66.0 | 66.3 | 66.7 | 67.0 |
| 115 | 63.1 | 63.5 | 63.8 | 64.2 | 64.5 | 64.9 | 65.3 | 65.6 | 66.0 | 66.3 | 66.7 | 67.0 | 67.3 |
| 116 | 63.6 | 63.9 | 64.3 | 64.6 | 65.0 | 65.3 | 65.7 | 66.0 | 66.4 | 66.7 | 67.0 | 67.4 | 67.7 |
| 117 | 64.0 | 64.4 | 64.7 | 65.0 | 65.4 | 65.7 | 66.1 | 66.4 | 66.8 | 67.1 | 67.4 | 67.7 | 68.0 |
| 118 | 64.5 | 64.8 | 65.1 | 65.5 | 65.8 | 66.1 | 66.5 | 66.8 | 67.2 | 67.5 | 67.8 | 68.1 | 68.4 |
| 119 | 64.9 | 65.3 | 65.6 | 65.9 | 66.2 | 66.6 | 66.9 | 67.2 | 67.6 | 67.9 | 68.1 | 68.4 | 68.7 |
| 120 | 65.4 | 65.7 | 66.0 | 66.3 | 66.6 | 67.0 | 67.3 | 67.6 | 67.9 | 68.2 | 68.5 | 68.8 | 69.1 |
| 121 | 65.8 | 66.1 | 66.5 | 66.8 | 67.1 | 67.4 | 67.7 | 68.0 | 68.3 | 68.6 | 68.9 | 69.1 | 69.4 |
| 122 | 66.3 | 66.6 | 66.9 | 67.2 | 67.5 | 67.8 | 68.1 | 68.4 | 68.7 | 69.0 | 69.2 | 69.5 | 69.8 |
| 123 | 66.7 | 67.0 | 67.3 | 67.6 | 67.9 | 68.2 | 68.5 | 68.8 | 69.1 | 69.3 | 69.6 | 69.9 | 70.1 |
| 124 | 67.2 | 67.5 | 67.8 | 68.0 | 68.3 | 68.6 | 68.9 | 69.2 | 69.5 | 69.7 | 70.0 | 70.2 | 70.4 |
| 125 | 67.6 | 67.9 | 68.2 | 68.5 | 68.8 | 69.0 | 69.3 | 69.6 | 69.8 | 70.1 | 70.3 | 70.6 | 70.8 |
| 126 | 68.1 | 68.4 | 68.7 | 68.9 | 69.2 | 69.5 | 69.7 | 70.0 | 70.2 | 70.5 | 70.7 | 70.9 | 71.1 |
| 127 | 68.6 | 68.8 | 69.1 | 69.4 | 69.6 | 69.9 | 70.1 | 70.4 | 70.6 | 70.9 | 71.1 | 71.3 | 71.5 |
| 128 | 69.0 | 69.3 | 69.5 | 69.8 | 70.0 | 70.3 | 70.5 | 70.8 | 71.0 | 71.2 | 71.4 | 71.6 | 71.8 |
| 129 | 69.5 | 69.7 | 70.0 | 70.2 | 70.5 | 70.7 | 71.0 | 71.2 | 71.4 | 71.6 | 71.8 | 72.0 | 72.1 |
| 130 | 69.9 | 70.2 | 70.4 | 70.7 | 70.9 | 71.1 | 71.4 | 71.6 | 71.8 | 72.0 | 72.2 | 72.3 | 72.5 |
| 131 | 70.4 | 70.6 | 70.9 | 71.1 | 71.3 | 71.6 | 71.8 | 72.0 | 72.2 | 72.3 | 72.5 | 72.7 | 72.8 |
| 132 | 70.8 | 71.1 | 71.3 | 71.5 | 71.8 | 72.0 | 72.2 | 72.4 | 72.5 | 72.7 | 72.9 | 73.0 | 73.1 |
| 133 | 71.3 | 71.5 | 71.7 | 72.0 | 72.2 | 72.4 | 72.6 | 72.7 | 72.9 | 73.1 | 73.2 | 73.3 | 73.4 |
| 134 | 71.7 | 71.9 | 72.2 | 72.4 | 72.6 | 72.8 | 72.9 | 73.1 | 73.3 | 73.4 | 73.5 | 73.6 | 73.7 |
| 135 | 72.1 | 72.4 | 72.6 | 72.8 | 73.0 | 73.1 | 73.3 | 73.5 | 73.6 | 73.7 | 73.8 | 73.9 | 74.0 |

### 4.1 Hydrography

Satellite measurements (NOAA OISST) of sea surface temperature (SST) in the central areas in the Northeast Atlantic in July 2023 were slightly warmer than the long-term average for July 1990-2009 based on SST plots (Figure 4a) and SST anomaly plots (Figure 4b). The northern regions of the Nordic Seas were slightly warmer than the average while the East Greenland Current was cooler than the long-term average. The SST in the Irminger Sea and Iceland Basin were slightly warmer than the average.
It should be mentioned that the NOAA SST are sensitive to the weather conditions (i.e. wind and cloudiness) prior to and during the observations and do therefore not necessarily reflect the oceanographic condition of the water masses in the areas, as seen when comparing detailed in situ features of SSTs between years (Figures 4a,b-5). However, since the anomaly is based on the average for the whole month of July, it should give representative results of the surface temperature.
The temperature distribution at 10, 50, 100 and 400 m depths is shown in Figure 5. At 10 m depth, the temperatures ranged from less than $1^{\circ} \mathrm{C}$ in the Greenland Sea to $16^{\circ} \mathrm{C}$ in the North Sea. At all depths there is a clear signal from the cold East Icelandic Current which carries cold and fresh water into the central and south-eastern part of the Norwegian Sea. Along the Norwegian Shelf and in the southernmost areas, the water masses are dominated by warmer waters of Atlantic origin. The CTD measurements at 10 m depths showed that north of Jan-Mayen the $8^{\circ} \mathrm{C}$ isotherm was found more easterly than last year. South of JanMayen the $8^{\circ} \mathrm{C}$ isotherm was found more westerly than last year and was closely aligned to the Jan-Mayen Ridge.

July - average SST


## July SST anomaly



Anomaly + - 0.25


Figure 4. Annual sea surface temperature ( a ; top panel) and its anomaly (b; lower panel; -4 to $+4^{\circ} \mathrm{C}$ ) in Northeast Atlantic for the month of July from 2010 to 2023 showing warm and cold conditions in comparison to the average for July 1990-2009. Based on monthly averages of daily Optimum Interpolation Sea Surface Temperature (Ver. 2.1 NOAA OISST, AVHRR-only, Banzon et al. 2016, https://www.ncei.noaa.gov/products/optimum-interpolation-sst ).


Figure 5. Interpolated temperature $\left({ }^{\circ} \mathrm{C}\right)$ at 10, 50, 100 and 400 m depth in Nordic Seas and the North Sea in July-August 2023.500 m and 2000 m depth contours are shown in light grey.

### 4.2 Zooplankton

The zooplankton biomass varied between areas with a patchy distribution throughout the area (Figure 6a). In the Norwegian Sea areas, the average zooplankton biomass was around $8 \mathrm{~g} / \mathrm{m}^{2}$, which is higher than the last two years (Figure 6b).
The time-series of zooplankton biomass was averaged by three subareas: Greenland region (not covered in 2023), Iceland region, and the Norwegian Sea region is shown in Figure 6b (see definitions in legend). In the Icelandic region and the Norwegian Sea the level was higher than in 2022. The biomass index in the Norwegian Sea varied less compared to the other two indices, and in 2023 it was comparable to 2019-2020 (Figure 6b). The lower variability over time in the Norwegian Sea might in part be explained by the more homogeneous oceanographic conditions in the area defined as Norwegian Sea.

These plankton indices should be treated with some caution as it is only a snapshot of the standing stock biomass, not of the actual production in the area, which complicates spatio-temporal comparisons.


Figure 6a. Interpolated zooplankton biomass ( $\mathrm{g} \mathrm{dw} / \mathrm{m} 2,0-200 \mathrm{~m}$ ) in Nordic Seas in July-August 2023.500 m and 2000 m depth contours are shown in light grey.


Figure 6b. Zooplankton biomass indices ( $\mathrm{g} \mathrm{dw} / \mathrm{m}^{2}, 0-200 \mathrm{~m}$ ). Time-series (2010-2023) of mean zooplankton biomass for three subareas within the survey range: Norwegian Sea (between $14^{\circ} \mathrm{W}-17^{\circ} \mathrm{E}$ \& north of $61^{\circ} \mathrm{N}$ ), Icelandic waters $\left(14^{\circ} \mathrm{W}-30^{\circ} \mathrm{W}\right)$ and Greenlandic waters (2014-2022, west of $30^{\circ} \mathrm{W}$ ).

### 4.3 Mackerel

The total swept-area mackerel index in 2023 was 4.30 million tonnes in biomass and 10.67 billion in numbers, a decrease of $42 \%$ for biomass and $39 \%$ for abundance compared to 2022 . The survey coverage area (excl. the North Sea, 0.28 million $\mathrm{km}^{2}$ ) was 2.36 million $\mathrm{km}^{2}$ in 2023 , which is $19 \%$ smaller compared to 2022. No extreme catches were taken this year, the highest catch was 5.7 tonnes. This reduces the uncertainty of the index in the biomass, $\mathrm{CV}=0.12$ in 2023 compared to $\mathrm{CV}=0.25$ in 2022.
Most of the surveyed mackerel still appears to be in the Norwegian Sea. However, they were more easterly and northeasterly distributed compared to 2022. The zero-line was reached for the whole survey area, north of latitude $60^{\circ} \mathrm{N}$.


Figure 7. Mackerel catch rates by Multpelt 832 pelagic trawl haul at predetermined surface trawl stations (circle areas represent catch rates in $\mathrm{kg} / \mathrm{km}^{2}$ ) overlaid on mean catch rates per standardized rectangles $\left(2^{\circ}\right.$ lat. $x 4^{\circ}$ lon.) in Nordic Seas in July-August 2023.


Figure 8. Annual distribution of mackerel proxied by the absolute distribution of mean mackerel catch rates per standardized rectangles ( $2^{\circ}$ lat. x $4^{\circ}$ lon.), from Multpelt 832 pelagic trawl hauls at predetermined surface trawl stations in Nordic Seas in June-August 2010-2023. Colour scale goes from white $(=0)$ to red $(=$ maximum value for the highest year).


Figure 9. Annual distribution of mackerel proxied by the relative distribution of mean mackerel catch rates per standardized rectangles ( $2^{\circ}$ lat. $\times 4^{\circ}$ lon.), from Multpelt 832 pelagic trawl hauls at predetermined surface trawl stations stations in Nordic Seas in June-August 2010-2023. Colour scale goes from white ( $=0$ ) to red (= maximum value for the given year).


Figure 10. Average weight of mackerel at predetermined surface trawl stations during IESSNS 2023.

The mackerel weight varied between 37 to 858 g with an average of 439 g . The length of mackerel caught in the pelagic trawl hauls onboard the five vessels varied from 17.5 to 45.5 cm , with an average of 34.2 cm . In total we measured 17464 mackerel. Mackerel length distribution followed the same overall pattern as previous years both in the Norwegian Sea, with increasing size towards the distribution boundaries in the north and the north-west, and in the western area with increasing size westward (Figure 10). The spatial distribution and overlap between the major pelagic fish species (mackerel, herring, blue whiting) in 2023 according to surface trawl catches is shown in Figure 11.


Figure 11. Distribution and spatial overlap between mackerel, herring, and blue whiting, at all surface trawl stations during IESSNS 2023. Vessel tracks are shown as continuous lines and predetermined surface trawl stations with no catch of the three species is displayed as + .

## Swept area analyses from standardized pelagic trawling with Multpelt 832

The swept area estimates of mackerel biomass from the 2023 IESSNS were based on abundance of mackerel per stratum (see strata definition in Figure 2) and calculated in StoX version 3.6.1. Mackerel abundance index in 2023 was $39 \%$ lower than in 2022, and $46 \%$ lower index than the average for the last 5 years (Table 7a; Figure 12) and the biomass index was $42 \%$ lower than in 2022, and $50 \%$ lower than the average for the last 5 years (Table 7c). Mackerel estimates of abundance, biomass and mean weight by age and length are displayed in Table 7d. There is no pattern in changing size-at-age between years (Table 7b). In 2023, the two most abundant year-classes were 2020 (age 3), 2019 (age 4), respectively (Figure 13). The 2020-year class contributed with $33 \%$, followed by 2019-year class with $16 \%$. Mackerel of age 1,2 and to some extent also age 3 are not completely recruited to the survey (Figure 15), because the main part of the nursery area was
further south than the survey area. Therefore, information on recruitment is uncertain. Variance in age index estimation is provided in Figure 14.
The overall internal consistency improved slightly compared to last year (Figure 16). There is a good to strong internal consistency for the younger ages (1-5 years) and older ages ( $9-14$ years) with $r$ between 0.71 and 0.89 . The internal consistency is more variable between age 5 to 9 , but improved with the addition of the data from this year, confirming the relations between 5,6 and 7 , and adding a new contrasting data point to the relation between 7 and 8 . More information on the relation between 7 and 8 (which have been the weakest link in the internal consistency since the beginning of the survey) is expected in IESSNS 2024 as the weak 2018 year class gets surveyed at age 8 .
Mackerel index calculations from the catch in the North Sea (Figure 2) were excluded from the index calculations presented in the current chapter to facilitate comparison to previous years and because the 2017 mackerel benchmark stipulated that trawl stations south of latitude $60^{\circ} \mathrm{N}$ be excluded from index calculations (ICES 2017). Results from the mackerel index calculations for the North Sea are presented in Appendix 1.
The indices used for NEA mackerel stock assessment in WGIWIDE are the number-at-age indices for age 3 to 11 year (Table 7a).


Figure 12. Estimated total stock biomass (upper panel) and total stock numbers (lower panel) of mackerel from StoX for the years 2007 and from 2010 to 2023. The red dots are baseline estimates, the black dots are mean of 1000 bootstrap replicates while the error bars represent $90 \%$ confidence intervals based on the bootstrap. Note, in 2011 the northern part of the Norwegian was not surveyed, hence the index for that year is not representative of mackerel stock size. See IESSNS 2011 cruise report for details.


Figure 13. Mackerel age distribution in numbers (\%) and in biomass (\%) from IESSNS 2023.


Figure 14. Number by age for mackerel in 2023. Plot of abundance ( $5 \%$ percentile, mean, $95 \%$ percentile) and relative standard error (CV) obtained by bootstrapping with 1000 replicates using the StoX software.

Table 7. a-d) StoX baseline (point estimate) time series of the IESSNS showing (a) age-disaggregated abundance indices of mackerel (billions), (b) mean weight (grams) per age, (c) estimated biomass at age (million tonnes) in 2007 and from 2010 to 2023, and (d) estimates of abundance, biomass and mean weight by age and length

| a) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year $\backslash$ Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | $14(+)$ | Tot $N$ |
| 2007 | 1.33 | 1.86 | 0.90 | 0.24 | 1.00 | 0.16 | 0.06 | 0.04 | 0.03 | 0.01 | 0.01 | 0.00 | 0.01 | 0.00 | 5.65 |
| 2010 | 0.03 | 2.80 | 1.52 | 4.02 | 3.06 | 1.35 | 0.53 | 0.39 | 0.20 | 0.05 | 0.03 | 0.02 | 0.01 | 0.01 | 13.99 |
| 2011 | 0.21 | 0.26 | 0.87 | 1.11 | 1.64 | 1.22 | 0.57 | 0.28 | 0.12 | 0.07 | 0.06 | 0.02 | 0.01 | 0.00 | $6.42^{*}$ |
| 2012 | 0.50 | 4.99 | 1.22 | 2.11 | 1.82 | 2.42 | 1.64 | 0.65 | 0.34 | 0.12 | 0.07 | 0.02 | 0.01 | 0.01 | 15.91 |
| 2013 | 0.06 | 7.78 | 8.99 | 2.14 | 2.91 | 2.87 | 2.68 | 1.27 | 0.45 | 0.19 | 0.16 | 0.04 | 0.01 | 0.02 | 29.57 |
| 2014 | 0.01 | 0.58 | 7.80 | 5.14 | 2.61 | 2.62 | 2.67 | 1.69 | 0.74 | 0.36 | 0.09 | 0.05 | 0.02 | 0.00 | 24.37 |
| 2015 | 1.20 | 0.83 | 2.41 | 5.77 | 4.56 | 1.94 | 1.83 | 1.04 | 0.62 | 0.32 | 0.08 | 0.07 | 0.04 | 0.02 | 20.72 |
| 2016 | $<0.01$ | 4.98 | 1.37 | 2.64 | 5.24 | 4.37 | 1.89 | 1.66 | 1.11 | 0.75 | 0.45 | 0.20 | 0.07 | 0.07 | 24.81 |
| 2017 | 0.86 | 0.12 | 3.56 | 1.95 | 3.32 | 4.68 | 4.65 | 1.75 | 1.94 | 0.63 | 0.51 | 0.12 | 0.08 | 0.04 | 24.22 |
| 2018 | 2.18 | 2.50 | 0.50 | 2.38 | 1.20 | 1.41 | 2.33 | 1.79 | 1.05 | 0.50 | 0.56 | 0.29 | 0.14 | 0.09 | 16.92 |
| 2019 | 0.08 | 1.35 | 3.81 | 1.21 | 2.92 | 2.86 | 1.95 | 3.91 | 3.82 | 1.50 | 1.25 | 0.58 | 0.59 | 0.57 | 26.4 |
| 2020 | 0.04 | 1.10 | 1.43 | 3.36 | 2.13 | 2.53 | 2.53 | 2.03 | 2.90 | 3.84 | 1.50 | 1.18 | 0.92 | 0.98 | 26.47 |
| 2021 | 0.09 | 2.13 | 0.71 | 1.22 | 1.53 | 0.37 | 1.29 | 0.81 | 1.05 | 0.97 | 0.93 | 0.46 | 0.34 | 0.33 | 12.22 |
| 2022 | 0.02 | 3.91 | 2.36 | 0.94 | 1.31 | 1.04 | 0.60 | 0.96 | 1.00 | 1.86 | 1.61 | 0.90 | 0.56 | 0.45 | 17.51 |
| 2023 | 0.21 | 0.70 | 3.54 | 1.70 | 0.55 | 0.46 | 0.79 | 0.32 | 0.48 | 0.39 | 0.45 | 0.44 | 0.34 | 0.30 | 10.67 |


| b) |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year\Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| 2007 | 133 | 233 | 323 | 390 | 472 | 532 | 536 | 585 | 591 | 640 | 727 | 656 | 685 |
| 2010 | 133 | 212 | 290 | 353 | 388 | 438 | 512 | 527 | 548 | 580 | 645 | 683 | 665 |


| 2011 | 133 | 278 | 318 | 371 | 412 | 440 | 502 | 537 | 564 | 541 | 570 | 632 | 622 |
| ---: | ---: | ---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| 2012 | 112 | 188 | 286 | 347 | 397 | 414 | 437 | 458 | 488 | 523 | 514 | 615 | 509 |
| 2013 | 96 | 184 | 259 | 326 | 374 | 399 | 428 | 445 | 486 | 523 | 499 | 547 | 677 |
| 2014 | 228 | 275 | 288 | 335 | 402 | 433 | 459 | 477 | 488 | 533 | 603 | 544 | 537 |
| 2015 | 128 | 290 | 333 | 342 | 386 | 449 | 463 | 479 | 488 | 505 | 559 | 568 | 583 |
| 2016 | 95 | 231 | 324 | 360 | 371 | 394 | 440 | 458 | 479 | 488 | 494 | 523 | 511 |
| 2017 | 86 | 292 | 330 | 373 | 431 | 437 | 462 | 487 | 536 | 534 | 542 | 574 | 589 |
| 2018 | 67 | 229 | 330 | 390 | 420 | 449 | 458 | 477 | 486 | 515 | 534 | 543 | 575 |
| 2019 | 153 | 212 | 325 | 352 | 428 | 440 | 472 | 477 | 490 | 511 | 524 | 564 | 545 |
| 2020 | 99 | 213 | 315 | 369 | 394 | 468 | 483 | 507 | 520 | 529 | 539 | 567 | 575 |
| 2021 | 140 | 253 | 357 | 377 | 409 | 451 | 467 | 487 | 497 | 505 | 516 | 523 | 544 |
| 2022 | 125 | 263 | 330 | 408 | 438 | 431 | 462 | 508 | 525 | 519 | 531 | 531 | 549 |
| 2023 | 128 | 269 | 347 | 371 | 416 | 435 | 462 | 484 | 506 | 526 | 517 | 533 | 557 |


| c) | 1 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year\Age | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | $14(+)$ | Tot B |
| 2007 | 0.18 | 0.43 | 0.29 | 0.09 | 0.47 | 0.09 | 0.03 | 0.02 | 0.02 | 0.01 | 0.01 | 0.00 | 0.01 | 0.00 | 1.64 |
| 2010 | 0.00 | 0.59 | 0.44 | 1.42 | 1.19 | 0.59 | 0.27 | 0.20 | 0.11 | 0.03 | 0.02 | 0.01 | 0.01 | 0.00 | 4.89 |
| 2011 | 0.03 | 0.07 | 0.28 | 0.41 | 0.67 | 0.54 | 0.29 | 0.15 | 0.07 | 0.04 | 0.03 | 0.01 | 0.01 | 0.00 | $2.69^{*}$ |
| 2012 | 0.06 | 0.94 | 0.35 | 0.73 | 0.72 | 1.00 | 0.72 | 0.30 | 0.17 | 0.06 | 0.03 | 0.01 | 0.00 | 0.00 | 5.09 |
| 2013 | 0.01 | 1.43 | 2.32 | 0.70 | 1.09 | 1.15 | 1.15 | 0.56 | 0.22 | 0.10 | 0.08 | 0.02 | 0.01 | 0.01 | 8.85 |
| 2014 | 0.00 | 0.16 | 2.24 | 1.72 | 1.05 | 1.14 | 1.23 | 0.80 | 0.36 | 0.19 | 0.05 | 0.03 | 0.01 | 0.00 | 8.98 |
| 2015 | 0.15 | 0.24 | 0.80 | 1.97 | 1.76 | 0.87 | 0.85 | 0.50 | 0.30 | 0.16 | 0.04 | 0.04 | 0.02 | 0.01 | 7.72 |
| 2016 | $<0.01$ | 1.15 | 0.45 | 0.95 | 1.95 | 1.72 | 0.83 | 0.76 | 0.53 | 0.37 | 0.22 | 0.10 | 0.04 | 0.04 | 9.11 |
| 2017 | 0.07 | 0.03 | 1.18 | 0.73 | 1.43 | 2.04 | 2.15 | 0.86 | 1.04 | 0.33 | 0.28 | 0.07 | 0.05 | 0.03 | 10.29 |
| 2018 | 0.15 | 0.57 | 0.16 | 0.93 | 0.50 | 0.63 | 1.07 | 0.85 | 0.51 | 0.26 | 0.30 | 0.16 | 0.08 | 0.05 | 6.22 |
| 2019 | 0.01 | 0.29 | 1.24 | 0.43 | 1.25 | 1.26 | 0.92 | 1.86 | 1.87 | 0.77 | 0.65 | 0.33 | 0.32 | 0.32 | 11.52 |
| 2020 | $<0.01$ | 0.23 | 0.45 | 1.24 | 0.84 | 1.18 | 1.22 | 1.03 | 1.51 | 2.03 | 0.81 | 0.67 | 0.53 | 0.58 | 12.33 |
| 2021 | 0.01 | 0.54 | 0.25 | 0.46 | 0.62 | 0.17 | 0.60 | 0.39 | 0.52 | 0.49 | 0.48 | 0.24 | 0.18 | 0.19 | 5.15 |
| 2022 | 0.00 | 1.03 | 0.78 | 0.39 | 0.57 | 0.45 | 0.28 | 0.49 | 0.52 | 0.97 | 0.85 | 0.48 | 0.31 | 0.26 | 7.37 |
| 2023 | 0.03 | 0.19 | 1.23 | 0.63 | 0.23 | 0.20 | 0.36 | 0.16 | 0.24 | 0.20 | 0.23 | 0.24 | 0.19 | 0.17 | 4.30 |

[^11]| $\begin{array}{\|l} \hline \text { d) } \\ \text { Length } \\ \text { (cm) } \end{array}$ | 2022 | 2021 | 2020 | 2019 | 5 2018 | Age 6 2017 | in years 7 7 2016 | (year clas 8 8 2015 | 5) $\begin{array}{r}9 \\ 2014\end{array}$ | 10 2013 | 11 2012 | 12 2011 | 13+ | NA | $\left.\right\|_{\left(10^{\wedge} 6\right)} ^{\text {Number }}$ | Biomass $\left(10^{\wedge} 6 \mathrm{~kg}\right)$ | $\begin{aligned} & \begin{array}{l} \text { Mean } \\ \text { weight } \\ \text { (g) } \end{array} \\ & \hline \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 17-18 |  |  |  |  |  |  |  |  |  |  |  |  |  | <1 | <1 | <1 | 37.5 |
| 18-19 |  |  |  |  |  |  |  |  |  |  |  |  |  | 1 | 1 | <1 | 47.0 |
| 19.20 | 3 |  |  |  |  |  |  |  |  |  |  |  |  |  | 3 | <1 | 50.9 |
| 20-21 | 7 |  |  |  |  |  |  |  |  |  |  |  |  |  | 7 | <1 | 64.1 |
| 21-22 | 9 |  |  |  |  |  |  |  |  |  |  |  |  |  | 9 | <1 | 75.8 |
| $22-23$ | 9 |  |  |  |  |  |  |  |  |  |  |  |  |  | 9 | 1 | 91.7 |
| 23-24 | 46 |  |  |  |  |  |  |  |  |  |  |  |  |  | 46 | 1 | 102.5 |
| 24.25 | 44 |  | 2 |  |  |  |  |  |  |  |  |  |  |  | 46 | 5 | 111.6 |
| 25-26 | 31 |  |  |  |  |  |  |  |  |  |  |  |  |  | 31 | 5 | 124.0 |
| 27-28 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 16 |  | 145.5 |
| 28.29 |  | 26 |  |  |  |  |  |  |  |  |  |  |  |  | 26 | 3 | 193.7 |
| 29-30 | 11 | 96 |  |  |  |  |  |  |  |  |  |  |  |  | 108 | 5 | 219.8 |
| 30-31 |  | 185 | 18 |  |  |  |  |  |  |  |  |  |  |  | 202 | 24 | 247.5 |
| 31-32 |  | 194 | 56 |  |  |  |  |  |  |  |  |  |  |  | 250 | 50 | 276.2 |
| 32-33 | 3 | 75 | 597 | 84 |  |  |  |  |  |  |  |  |  |  | 760 | 69 | 311.4 |
| 33-34 |  | 100 | 1307 | 216 | 19 |  |  |  |  |  |  |  |  |  | 1645 | 237 | 337.9 |
| 34.35 |  | 20 | 1122 | 611 | 26 | 15 | 17 |  |  |  |  |  |  |  | 1812 | 556 | 360.5 |
| 35-36 | 3 | 1 | 346 | 537 | 153 | 65 | 25 |  |  |  |  |  |  |  | 1135 | 653 | 389.1 |
| 36-37 |  |  | 87 | 189 | 217 | 149 | 172 | 29 |  |  |  |  |  |  | 858 | 441 | 414.5 |
| 37-38 |  |  | 3 | 62 | 99 | 155 | 325 | 97 | 68 | 41 | 16 | 29 | 16 |  | 912 | 356 | 456.0 |
| 38-39 |  |  | 0 | 4 | 19 | 60 | 172 | 131 | 184 | 97 | 170 | 131 | 96 |  | 1063 | 416 | 491.3 |
| 39.40 |  |  | 2 |  | 16 | 12 | 63 | 43 | 124 | 166 | 158 | 131 | 161 |  | 877 | 522 | 532.6 |
| 40-41 |  | 0 |  |  |  | 1 | 13 | 20 | 88 | 64 | 78 | 95 | 185 |  | 545 | 467 | 560.8 |
| 41-42 |  |  |  |  |  |  |  |  | 6 | 12 | 17 | 37 | 129 |  | 201 | 306 | 600.8 |
| 42-43 |  |  |  |  |  |  |  |  |  |  | 8 | 9 | 40 |  | 65 | 121 | 639.4 |
| 43.44 |  |  |  |  |  |  |  |  |  |  |  | 3 | 11 |  | 14 | 42 | 675.8 |
| 44.45 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | 2 | 10 | 742.5 |
| 45-46 |  |  |  |  |  |  |  |  |  |  |  |  |  | <1 | $<1$ | 1 | 821.5 |
| NA |  |  |  |  |  |  |  |  |  |  |  |  |  | $<1$ | $<1$ | <1 | 0.0 |
| TSN(mill) | 211.3 | 696.6 | 3539.6 | 1703.2 | 548.7 | 460.3 | 786.8 | 321.3 | 482.9 | 387.5 | 450.2 | 441.0 | 640.3 | 1.1 | 10670.7 | 4298 |  |
| TSB(1000 t) | 27.0 | 187.4 | 1227.6 | 632.4 | 228.7 | 200.2 | 363.5 | 155.7 | 244.6 | 203.9 | 233.0 | 235.2 | 358.8 | 0.4 | 4298 |  |  |
| Mean length(cm) | 24.5 | 30.8 | 33.4 | 34.4 | 35.9 | 36.5 | 37.1 | 37.8 | 38.5 | 38.8 | 38.8 | 39.0 |  |  |  |  |  |
| Mean weight(g) | 128 | 269 | 347 | 371 | 416 | 435 | 462 | 484 | 506 | 526 | 517 | 533 |  |  |  |  |  |

Table 8. Bootstrap estimates from StoX (based on 1000 replicates) of mackerel in 2023. Numbers by age and total number (TSN) are in millions and total biomass (TSB) in million tons.

| Age | $\begin{array}{r} \text { 5th } \\ \text { percentile } \end{array}$ | Median | $\begin{array}{r} 95 \text { th } \\ \text { percentile } \end{array}$ | Mean | SD | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 45.6 | 192.5 | 395.4 | 200.2 | 107.7 | 0.54 |
| 2 | 439.9 | 672.8 | 943.7 | 677.5 | 152.4 | 0.22 |
| 3 | 2449.2 | 3526.3 | 4768.9 | 3544.2 | 706.9 | 0.20 |
| 4 | 1240.4 | 1715.4 | 2267.2 | 1728.9 | 311.6 | 0.18 |
| 5 | 350.0 | 521.1 | 721.1 | 529.5 | 112.1 | 0.21 |
| 6 | 337.7 | 438.1 | 548.6 | 439.8 | 66.1 | 0.15 |
| 7 | 633.0 | 818.3 | 1034.3 | 826.4 | 122.7 | 0.15 |
| 8 | 240.7 | 340.6 | 429.3 | 338.6 | 58.9 | 0.17 |
| 9 | 378.2 | 480.0 | 591.7 | 481.4 | 65.9 | 0.14 |
| 10 | 278.4 | 366.8 | 469.4 | 368.7 | 57.4 | 0.16 |
| 11 | 311.9 | 423.1 | 548.0 | 426.7 | 72.6 | 0.17 |
| 12 | 355.3 | 460.4 | 567.8 | 462.4 | 66.2 | 0.14 |
| 13 | 235.7 | 326.2 | 428.1 | 328.4 | 57.8 | 0.18 |
| 14 | 77.3 | 115.1 | 154.6 | 115.9 | 23.8 | 0.21 |
| 15 | 59.8 | 97.5 | 141.4 | 99.0 | 25.2 | 0.25 |
| 16 | 40.2 | 70.6 | 102.8 | 70.3 | 19.5 | 0.28 |
| 17 | 9.1 | 20.8 | 34.8 | 21.1 | 7.9 | 0.37 |
| 18 | 1.6 | 6.4 | 14.2 | 6.8 | 3.8 | 0.56 |
| 19 | 0.2 | 1.3 | 7.8 | 2.2 | 2.7 | 1.22 |
| 20 | 0.0 | 0.3 | 0.9 | 0.3 | 0.3 | 1.00 |
| TSN | 8590 | 10680 | 13081 | 10680 | 1344 | 0.13 |
| TSB | 3.50 | 4.28 | 5.21 | 4.30 | 0.51 | 0.12 |



Figure 15. Catch curves for the years 2010; 2012-2023. Each cohort of mackerel is marked by a uniquely coloured line that connects the estimates indicated by the respective ages.


Log10 (index+1)
Figure 16. Internal consistency of the of mackerel density index from 2012 to 2023. Ages indicated by white numbers in grey diagonal cells. Statistically significant positive correlations ( $p<0.05$ ) are indicated by regression lines and red cells in upper left half. Correlation coefficients ( $r$ ) are given in the lower right half.

The swept area method assumes that potential distribution of mackerel outside the survey area - both vertically and horizontally - is a constant percentage of the total biomass. In some years, this assumption may be violated, e.g. mackerel may be distributed below the footrope of the trawl or if the proportion of mackerel outside the survey coverage varies among years. In order to improve the precision of the swept area estimate it would be beneficial to extend the survey coverage further south, such that it covers the southwestern waters south of $60^{\circ} \mathrm{N}$, e.g. UK waters.
The standard swept area method using the average horizontal trawl opening by each participating vessel (ranging 63.7-71.7 m; Table 5), assuming that a constant fraction of the mackerel inside the horizontal trawl opening are caught. Further, that if mackerel is distributed below the depth of the trawl (footrope), this fraction is assumed constant from year to year.

As in previous years, there was overlap in the spatio-temporal distribution of mackerel and herring (Figure 11). This overlap occurred mostly between mackerel and Norwegian spring-spawning herring (NSSH) in the western, north-western and north-eastern part of the Norwegian Sea

### 4.4 Norwegian spring-spawning herring

In the present preliminary report, no results of herring measurements are presented. A final survey report including these two species will be published in the fall of 2023. Figures $17-19$, and tables $9-11$ will be added in the final report.

### 4.5 Blue whiting

In the present preliminary report, no results of blue whiting measurements are presented. A final survey report including these two species will be published in the fall of 2023. Figures $20-22$, and tables 12 and 13 will be added in the final report.

### 4.6 Other species

Lumpfish (Cyclopterus lumpus)
Lumpfish was caught in $69 \%$ of trawl stations across the five vessels (Figure 23) and where lumpfish was caught, $70 \%$ of the catches were $\leq 10 \mathrm{~kg}$. Lumpfish was distributed across the entire survey area, from west of Iceland to the Barents Sea in the northeast, and into the North Sea in the southern part of the covered area. Abundance was greatest north of $71^{\circ} \mathrm{N}$, with lower densities in the central Norwegian Sea and mostly absent directly south of Iceland, and south and southwest of the North Sea. The zero line was not hit to the northeast, northwest and west of the survey so it is likely that the distribution of lumpfish extends beyond the survey coverage.

The length of lumpfish caught varied from 4 to 51 cm with a bimodal distribution with the left peak (5-20 cm ) likely corresponding to 1 -group lumpfish and the right peak consisting of a mixture of age groups (Figure 24). Only a small number of fish were sexed (115 of 1985) but for fish $\geq 20 \mathrm{~cm}$ in which sex was determined, the males ( $\mathrm{n}=15$ ) were $22-29 \mathrm{~cm}$ in length. The females ( $\mathrm{n}=100$ ) ranged in length from 21 to 46 cm . Generally, the mean length and mean weight of the lumpfish was highest in Faroese waters, and around Iceland and along the shelf edges of Norway and lowest in the central and northern Norwegian Sea.

A total of 374 fish ( 126 by R/V "Árni Friðriksson", 149 by M/V "Eros" and 99 by M/V Vendla) between 11 and 49 cm were tagged during the survey (Figure 25).


Figure 23. Lumpfish catches at surface trawl stations during IESSNS 2023.


Figure 24. Length distribution of a) all lumpfish caught during the survey and b) length distribution of fish in which sex was determined.


Figure 25. Number tagged, and release location, of lumpfish. Insert shows the length distribution of the tagged fish.

## Salmon (Salmo salar)

A total of 62 North Atlantic salmon were caught in 38 stations both in coastal and offshore areas from $62^{\circ} \mathrm{N}$ to $74^{\circ} \mathrm{N}$ in the upper 30 m of the water column during IESSNS 2023 (Figure 26). The salmon ranged from 0.084 kg to 2.7 kg in weight, dominated by post-smolt and 1 sea-winter individuals. We caught from 1 to 12 salmon during individual surface trawl hauls. The length of the salmon ranged from 20 cm to 82 cm , with the highest fraction between 20 cm and 29 cm .


Figure 26. Catches of salmon at surface trawl stations during IESSNS 2023.

## Capelin (Mallotus villosus)

Capelin was caught in the surface trawl on 29 stations along the cold fronts: North of Iceland, northnorthwest of Jan Mayen, northwest of Bear Island and west of Svalbard (Figure 27a). Both juvenile and adult capelin were caught during the survey. The average length ranged from 6.9 to 18.8 cm and average weight ranged from 1,2 to 32.6 g in the trawl hauls. There were more pelagic trawl stations with catches of capelin in the western and northern part of the Jan Mayen.

## Polar cod (Boreogadus saida)

Polar cod was caught in the surface trawl on 11 stations north and northeast of Iceland (Figure 27b). The catch weight per station ranged from 10 g to 5 kg . The polar cod ranged in total length from 9 cm to 20 cm and in total weight from 5 g to 36 g . Mean length was 13.2 cm (standard deviation $=1.5, \mathrm{n}=225$ ) and mean weight was 15.3 g (standard deviation $=5.4, \mathrm{n}=224$ ). Polar cod was caught in larger area and in greater abundance in 2023 compared to all previous years of the IESSNS survey, hence it was added to the report chapter on other species.


Figure 27a. Presence of capelin in surface trawl stations during IESSNS 2023.


Figure 27b. Presence of polar cod in surface trawl stations during IESSNS 2023.

### 4.7 Marine Mammals

Opportunistic whale observations were done by M/V "Eros" and M/V "Vendla" from Norway in addition to R/V "Jákup Sverri" from Faroe Islands from 1 ${ }^{\text {st }}$ July to $3^{\text {rd }}$ August 2023 (Figure 28). Overall, 1078 marine mammals of 10 different species were observed, which was an increase from an overall 711 marine mammals observed in 2022.
The species that were observed included blue whales (Balaenoptera musculus), fin whales (Balaenoptera physalus), minke whales (Balaenoptera acutorostrata), humpback whales (Megaptera novaeangliae), Northern bottlenose whales (Hyperoodon ampullatus), pilot whales (Globicephala sp.), killer whales (Orcinus orca), sperm whales (Physeter macrocephalus), sei whales (Baleanoptera borealis), white sided dolphins (Lagenorhynchus acutus) white beaked dolphins (Lagenorhynchus albirostris). Basking sharks (Cetorhinus maximus) were also observed during the survey. The dominant number of marine mammal observations were found around along the continental shelf west and north of Jan Mayen and in the southwestern and western areas of Svalbard. We observed higher number of marine mammals in the central part of the Norwegian Sea in July 2023 compared with last year. Altogether eight blue whales were observed in the western and northern areas of Jan Mayen. They appeared either solitary or in groups of two individuals, and was most probably feeding on large swarms of amphipods in cold water. Fin whales ( $n=82$, group size $=1-20$ (average group size $=2.5$ )) and humpback whales ( $\mathrm{n}=44$, group size $=1-50$ (average group size $=2.4$ )) dominated among the large whale species. They were distributed from $64^{\circ} \mathrm{N}$ to $78.30^{\circ} \mathrm{N}$ and from $25^{\circ} \mathrm{E}$ to $15^{\circ} \mathrm{W}$ and they had hotspot southwest and west of Svalbard as well as west and northwest of Jan Mayen. Few sperm whales (n $=8$, group size $=1-2($ average group size $=1.3))$ where observed. Killer whales $(\mathrm{n}=56$, group size $=1-10$ (average groups size $=6.2$ )) dominated in the southern, north-eastern part of the Norwegian Sea, partly overlapping and presumably feeding on NEA mackerel in the upper water masses. Pilot whales ( $\mathrm{n}=112$, group size $=5-100($ average groups size $=37)$ ) where mostly observed in Faroese waters during IESSNS 2023. Five sei whale and 46 northern bottlenose whale were observed in Faroese waters, whereas three basking sharks were observed in Faroese waters and west of Lofoten. White beaked dolphins ( $\mathrm{n}=123$, group size $=1-50($ average group size $=9.5)$ ) were present in the northern part of the Norwegian Sea. Minke whales $(\mathrm{n}=22$, group size $=1-4$ (average group size $=1.4)$ ) were distributed over large areas from western coast of Norway to west of Svalbard, and from $60^{\circ} \mathrm{N}$ to $77^{\circ} \mathrm{N}$, including overlapping and likely feeding on NSS herring in the upper $10-40 \mathrm{~m}$ of the water column. There is available a publication summarizing the main results on marine mammals from the IESSNS surveys from 2013 to 2018, with major focus on hot spot areas of fin whales and humpback whales from 2013 to 2018 (Leviknes et al. 2021)

## IESSNS 2023

## Marine mammals

* Blue whale
- Common bottlenose dolphin
- Fin whale
- Humpback whale

A Killer whale

- Minke whale
- Northern bottlenose whale
- Pilot whale
+ Sei whale
- Sperm whale
$\times$ Unidentified
+ Unidentified dolphin
is White-beaked dolphin
- FO
- NO (Eros)
- NO (Vendla)

Group size (individuals)
○ 1
2-10
$\bigcirc>10$


Figure 28. Overview of all marine mammals sighted during IESSNS 2023.

| The group suggested the following recommendation from WGIPS | To whom |
| :--- | :--- |
| The surveys conducted by Denmark in 2018-2022 have clearly demonstrated that the <br> IESSNS methodology works also for the northern North Sea (i.e. north and west from |  |
| Doggerbank) and the Skagerrak area deeper than 50 m . The survey provides essential <br> fishery-independent information on the stock during its feeding migration in summer <br> and WGIPS recommends that the Danish survey should continue as a regular annual <br> survey. | WGWIDE, RCG <br> NANSEA |


| Action points | Responsible |
| :--- | :--- |
| Criteria and guidelines should be established for discarding substandard trawl sta- <br> tions using live monitoring of headline, footrope and trawl door vertical depth, and <br> horizontal distance between trawl doors. For predetermined surface trawl station, dis- <br> carded hauls should be repeated until performance is satisfactory. | All |
| Explicit guideline for incomplete trawl hauls is to repeat the station or exclude it from |  |
| future analysis. It is not acceptable to visually estimate mackerel catch, it must be |  |
| hauled onboard and weighed. If predetermined trawl hauls are not satisfactory ac- |  |
| cording to criteria the station will be excluded from mackerel index calculations, i.e. |  |
| treated as if it does not exist, but not as a zero mackerel catch station. |  |$\quad$| We encourage registrations of opportunistic marine mammal observations. |
| :--- |
| We should consider calculating the zooplankton index from annually gridded field <br> polygons to extract area-mean time-series. WGINOR is currently working on Norwe- <br> gian Sea polygons, and further work on this issue will start when their work is final- <br> ized. |
| In 2023 the IESSNS survey in the North Sea has been conducted for six consecutive <br> years (2018-2023). It is recommended that a comprehensive report is written about the <br> major results from the NEA mackerel time series from the IESSNS surveys in the <br> North Sea, where an update of the internal consistency between years in the survey for <br> selected age groups is also evaluated. This report should be made available for consid- <br> eration in the next benchmark. A major aim will be to at some stage evaluate and con- <br> sider the possibility to include and implement the IESSNS survey in the North Sea as <br> an abundance index used in ICES for NEA mackerel. |
| Country representatives for the IESSNS survey should rewrite the respective sections <br> (e.g. trawl performance, trawl station data collection) in the survey manual according <br> to the new format by mid-September 2023. |
| and All |

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## Appendix 1

Denmark joined the IESSNS in 2018 for the first time extending the original survey area into the North Sea. The commercial fishing vessels "Ceton S205" was used. No problems applying the IESSNS methods were encountered. Area coverage, however, was restricted to the northern part of the North Sea at water depths larger 50 m . No plankton samples were taken, and no acoustic data were recorded because this is covered by the HERAS survey in June/July in this area.

Based on the experiences made in the previous years, new limits for the stratum in the North Sea were defined in 2022 (Fig. 2, stratum 13). The northern limit for the North Sea and the Skagerrak were defined as $60^{\circ} \mathrm{N}$ and $59^{\circ} \mathrm{N}$, respectively. The western geographical limit in the North Sea was set to $1^{\circ} 30^{\prime} \mathrm{W}$ in the north and $2^{\circ} 30^{\prime} \mathrm{W}$ further south following the UK coastline where the Inner Moray Firth and the Firth of Forth were excluded because mackerel was not recorded there and a high abundance of 0 -group gadoids, sandeel and other species makes a quantitative analysis of the catches very time consuming. The easter limit in the Skagerrak was set to $11^{\circ} \mathrm{E}$, and the southern limit in the North Sea was approximated by the 50 m isobath, which is about the shallowest depth limit for a safe setting of the Multpelt 832 trawl.
In 2023, 36 stations were taken (PT and CTD). Average mackerel catch amounted to $2362 \mathrm{~kg} / \mathrm{km}^{2}$, which was considerably higher than in the previous year ( $2022: 1689 \mathrm{~kg} / \mathrm{km}^{2}$ ) and is the second highest in the time series (2021: $2429 \mathrm{~kg} / \mathrm{km}^{2}$ 2020: $1318 \mathrm{~kg} / \mathrm{km}^{2}, 2019: 1009 \mathrm{~kg} / \mathrm{km}^{2}, 2018: 1743 \mathrm{~kg} / \mathrm{km}^{2}$ ) (Fig. A1-1). The length and age composition indicate a relative high amount of small $(<25 \mathrm{~cm})$ individuals and the abundance of older ( $\geq$ age 3 ) mackerel was higher than in the previous years (Fig. A1-2).

The StoX (version 3.6.1) baseline estimates of mackerel biomass and abundance in the North Sea for 2023 were 650371 tonnes and 3.3 billion individuals (Table A1-1) which is a $27 \%$ higher biomass and a $40 \%$ higher abundance than last year. The biomass and abundance estimates are based on the stratum limits as shown in Fig. 2 (stratum 13). The area of this polygon is $285781 \mathrm{~km}^{2}$.
Catches curves indicate that all ages including age 1 and 2 are well represented in the survey data, and the 2022-year class is the highest at age 1 in the time series (Fig. A1-3).

The internal consistency plots (Fig. A1-4), however, do not show any significant correlations. This is likely due to the low number of observations which are so far available. Furthermore, interannual variations in the migration of the cohorts in and out of the North Sea may have an effect as well.

| $\begin{aligned} & \text { Length } \\ & (\mathrm{cm}) \\ & \hline \end{aligned}$ | Age in years/ Year class |  |  |  |  |  |  |  |  |  |  |  |  | Number (10^6) | Biomass (ton) | $\begin{gathered} \text { Mean } \\ \text { weight } \\ \text { (g) } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |  |  |  |
|  | 2022 | 2021 | 2020 | 2019 | 2018 | 2017 | 2016 | 2015 | 2014 | 2013 | 2012 | 2011 | 2010 |  |  |  |
| 17-18 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 18-19 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 19-20 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 20-21 | 10.7 |  |  |  |  |  |  |  |  |  |  |  |  | 10.7 | 800 |  |
| 21-22 | 151.2 |  |  |  |  |  |  |  |  |  |  |  |  | 151.2 | 12122 |  |
| 22-23 | 605.8 |  |  |  |  |  |  |  |  |  |  |  |  | 605.8 | 55116 | 9 |
| 23-24 | 572.8 |  |  |  |  |  |  |  |  |  |  |  |  | 572.8 | 56187 |  |
| 24-25 | 258.9 |  |  |  |  |  |  |  |  |  |  |  |  | 258.9 | 29431 | 114 |
| 25-26 | 77.8 |  |  |  |  |  |  |  |  |  |  |  |  | 77.8 | 9934 | 128 |
| 26-27 | 50.3 |  |  |  |  |  |  |  |  |  |  |  |  | 50.3 | 7418 | 148 |
| 27-28 | 33.3 | 4.6 |  |  |  |  |  |  |  |  |  |  |  | 38.0 | 6594 | 174 |
| 28-29 | 58.8 | 34.0 |  |  |  |  |  |  |  |  |  |  |  | 92.8 | 18114 | 195 |
| 29-30 | 71.8 | 49.7 |  |  |  |  |  |  |  |  |  |  |  | 121.5 | 26646 | 219 |
| 30-31 | 55.9 | 93.6 | 0.4 |  |  |  |  |  |  |  |  |  |  | 149.9 | 36760 | 245 |
| 31-32 | 18.5 | 195.4 | 18.3 | 0.7 |  |  |  |  |  |  |  |  |  | 232.9 | 61810 | 265 |
| 32-33 |  | 256.6 | 93.9 | 4.7 | 3.0 |  |  |  |  |  |  |  |  | 358.2 | 107294 | 299 |
| 33-34 |  | 75.9 | 159.1 | 53.6 |  |  |  |  |  |  |  |  |  | 288.5 | 92445 | 320 |
| 34-35 |  | 0.6 | 69.5 | 53.6 | 25.5 | 2.7 |  |  |  |  |  |  |  | 151.9 | 52262 | 344 |
| 35-36 |  |  | 7.2 | 13.2 | 19.4 | 18.2 | 13.6 | 1.4 |  |  |  |  |  | 73.1 | 26473 | 36 |
| 36-37 |  |  | 0.8 | 1.5 | 19.7 | 21.2 | 4.3 | 2.1 |  |  |  |  |  | 49.6 | 19162 | 386 |
| 37-38 |  |  |  |  | 2.5 | 10.8 | 6.7 | 6.2 | 1.3 | 0.1 |  |  |  | 27.6 | 11670 | 423 |
| 38.39 |  |  |  |  | 1.6 | 0.4 | 6.5 | 12.7 | 0.7 | 0.7 | 0.2 | 0.2 |  | 23.0 | 10843 | 472 |
| 39-40 |  |  |  |  | 0.8 |  | 0.8 | 3.1 | 3.8 | 0.3 | 0.3 | 0.1 |  | 9.1 | 4421 | 485 |
| 40-41 |  |  |  |  |  |  | 0.8 | 1.4 |  | 0.1 | 0.7 | 0.3 | 0.3 | 3.6 | 2015 | 560 |
| 41-42 |  |  |  |  |  |  |  |  | 0.1 | 0.1 | 0.6 |  |  | 0.8 | 486 | 57 |
| 42-43 |  |  |  |  |  |  |  |  |  |  |  |  |  | 0.0 | 0 |  |
| 43-44 |  |  |  |  |  |  |  |  | 1.1 |  |  |  |  | 1.1 | 915 | 85 |
| 44.45 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| TSN (mill) | 1965.8 | 710.3 | 349.1 | 127.4 | 72.5 | 53.3 | 32.6 | 26.9 | 7.0 | 1.4 | 1.7 | 0.5 | 0.3 | 3,349 | 648919 |  |
| TSB (ton) | 222558 | 194721 | 110530 | 42839 | 26509 | 20636 | 13029 | 12420 | 3690 | 658 | 911 | 265 | 153 |  |  |  |
| Mean length (cm) | 23.5 | 31.1 | 32.9 | 33.6 | 35.0 | 35.8 | 36.4 | 37.7 | 39.2 | 38.6 | 40.0 | 39.2 | 40.0 |  |  |  |
| Mean weight ( 8 ) | 113 | 274 | 317 | 336 | 366 | 387 | 399 | 462 | 530 | 476 | 522 | 494 | 476 |  |  |  |



Fig. A1-1. Biomass density (mean and standard error) of mackerel in the North Sea 2018 to 2023.


Fig. A1-2. Comparison of Iength and age distribution of mackereI in the North Sea 2018 to 2023.

## IESSNS - North Sea



Fig. A1-3. Catch curves for mackerel year classes 2013 to 2022 in the North Sea (lines represents cohorts, numbers denote ages)


Fig. A1-4. Internal consistency of mackerel density indices ages 1 to 9 for the North Sea from 2018 to 2023 (numbers in symbols indicate $2000^{\prime}$ er year classes).

## Appendix 2

The mackerel index is calculated on all valid surface stations. That means, that invalid and potential extra surface stations and deeper stations need to be excluded. Below is the exclusion list used when calculating the mackerel abundance index for IESSNS 2023.

| Vessel | Country | Horizontal trawl <br> opening $(\mathrm{m})$ | Exclusion list |  |
| :--- | :--- | :--- | :--- | :--- |
|  |  |  | Cruise | Stations |
| Vendla | Norway | 65.1 | 2023203003 | $53,59,69,76,77,80,89$, <br> $94,100,105,110$ |
| Eros | Norway | 71.7 | 202204002 | $12,19,42,47,51,57,58$, <br> $62,64,65,70,72$ |
| R/V Árni Friðriksson | Iceland | 68.4 | A8-2023 | $280,281,315,318,321$ |
| R/V Jákup Sverri | Faroe Islands | 64.1 | 2334 | $21,28,37,45^{*}$ |
| Ceton | Denmark | 69.7 | IESSNS2023 | 9,30 |

* Observe that in PGNAPES and the national database station numbers are 4-digit numbers preceded by 2230 (c.g. '22300005')


Figure A2-1. IESSNS 2023. Surface trawl stations included (filled dark blue rectangle) and excluded (filled light blue rectangle) in calculations of mackerel age segregated index used in the assessment. Strata boundary also displayed (grey solid lines).

## Appendix 3

Horizontal trawl opening of the Multpelt 832 trawl is a function of trawl door spread and tow speed (Table 6 in the 2022 report). The estimates in table 6 are originally based on flume tank simulations in 2013 (Hirtshals, Denmark) where two formulas were empirically derived for two towing speeds, 4.5 and 5 knots:

| Towing speed 4.5 knots: | Horizontal opening $(\mathrm{m})=0.441$ * Door spread $(\mathrm{m})+13.094$ |
| :--- | :--- |
| Towing speed 5.0 knots: | Horizontal opening $(\mathrm{m})=0.3959$ * Door spread $(\mathrm{m})+20.094$ |

In 2017, the towing speed range was increased to 5.2 knots, i.e. an extrapolation of the trawl opening as a function of door spread and speed was performed. In 2022 the towing speed range was further extended down to 4.3 knots and up to 5.5 knots, using a kriging gridding method, see figure A4-1. In 2023, the trawl opening was extended to 135 m (Table 6).


Figure A3-1. Table 6 in the report shown as a plot.

Working document 01, WGWIDE 2023

## PFA self-sampling report for WGWIDE 2023

Niels Hintzen, 11/07/2023 14:06:54
PFA report 2023_07

## Executive summary

The Pelagic Freezer-trawler Association (PFA) is an association that has nine member companies that together operate 19 (in 2022) freezer trawlers in six European countries (www.pelagicfish.eu). In 2015, the PFA has initiated a self-sampling program that expands the ongoing monitoring programs on board of pelagic freezer-trawlers aimed at assessing the quality of fish. The expansion in the selfsampling program consists of recording of haul information, recording the species compositions by haul and regularly taking length measurements from the catch. The self-sampling is carried out by the vessel quality managers on board of the vessels, who have a long experience in assessing the quality of fish, and by the skippers/officers with respect to the haul information. The scientific coordination of the self-sampling program is carried out by Niels Hintzen (PFA chief science officer) with support of Lina de Nijs and Floor Quirijns (contractor). The self-sampling program has been incrementally implemented in the fishery and by 2018 all vessels in the PFA fleet participated in the selfsampling.

This report for WGWIDE presents an overview of the results of the Pelagic Freezer-Trawler Association (PFA) self-sampling program for the fisheries for widely-distributed pelagic stocks: Northeast Atlantic mackerel, Blue whiting, Horse mackerel and Atlanto-scandian herring (herring caught north of 62 degrees). The selection of hauls to be included in the analyses was based on first summing all catches by vessel, trip, species and week. For each vessel-trip-species-week combination, the proportion of the species in the catch were calculated. The following filter criteria have applied to the weekly data:

- for horse mackerel: latitude $>45$, proportion in the catch $>10 \%$, weekly catch $>10$ tonnes
- for mackerel : latitude >45, proportion in the catch > 10\%, weekly catch > 10 tonnes
- for blue whiting : latitude $>50$, proportion in the catch $>10 \%$, weekly catch $>10$ tonnes
- for herring : division = 27.2.a, proportion in the catch > 10\%, weekly catch > 10 tonnes

Trips from 2017 up to 02/06/2023 have been processed for this overview. Pelagic fisheries within the Pelagic Freezer-trawler Association are carried out by vessels from different countries. Overall, around $48 \%$ of the catch volume of trips in this overview were taken by Dutch trawlers, $19 \%$ German trawlers, $13 \%$ UK trawlers and $19 \%$ other countries. Blue whiting constitutes the majority of the catch in those trips (57\%), followed by mackerel (25\%) and horse mackerel (10\%). Atlanto-scandian herring
only constitutes around $1 \%$ of the volume in the PFA widely distributed fishery. Note that the North Sea herring fishery is not included in this overview.

The Mackerel fishery takes place from October through to March of the subsequent year. Bycatches of mackerel may also occur during other fisheries,e.g. for horse mackerel or herring. Overall, the selfsampling activities for the mackerel fisheries during the years 2017-2023 (up to 02/06/2023) covered 507 fishing trips with 6977 hauls, a total catch of 416958 tonnes and 112296 individual length measurements. The main fishing areas are ICES division 27.4.a and division 27.6.a. Compared to the previous years, mackerel in the catch in 2022 have been relatively large with a median length of 36.3 cm . The median weight at 400 grams was among average compared to preceding years.

The Western horse mackerel fishery takes place from October through to March of the subsequent year. Overall, the self-sampling activities for the Western horse mackerel fisheries during the years 2017-2023 (up to 02/06/2023) covered 244 fishing trips with 3096 hauls, a total catch of 116548 tonnes and 121815 individual length measurements. The main fishing areas are ICES division 27.6.a, division 27.7.b and division 27.7.j. Western horse mackerel have a wide range in the length distributions in the catch. Median lengths have fluctuated between 22.2 and 30.0 cm .

The North Sea horse mackerel fishery takes place from October through to January of the subsequent year. Overall, the self-sampling activities for the North Sea horse mackerel fisheries during the years 2017-2023 (up to 02/06/2023) covered 120 fishing trips with 969 hauls, a total catch of 46562 tonnes and 38981 individual length measurements. The main fishing areas is ICES division 27.7.d with some minor catches in 27.4.c. Catches in division 27.4.a have been counted as Western Horse mackerel. North Sea horse mackerel have a narrow range in the length distributions in the catch. Median lengths have fluctuated between 21.5 and 23.9 cm .

The blue whiting fishery takes place from February through to May although some minor fisheries for blue whiting may remain over the other months. Overall, the self-sampling activities for the horse mackerel fisheries during the years 2017-2023 (up to 02/06/2023) covered 356 fishing trips with 9426 hauls, a total catch of 945762 tonnes and 554926 individual length measurements. The main fishing areas are ICES division 27.6.a, division 27.7.c and division 27.7.k. Compared to the previous years, blue whiting in the catches during 2020-2022 have been relatively large with a median length of 27.5 cm compared to 24.2-27.1 in the preceding years.

The fishery for Atlanto-scandian herring (ASH) is a relatively smaller fishery for PFA and takes place mostly in October. Overall, the self-sampling activities for the Atlanto-scandian herring fisheries during the years 2017-2023 (up to 02/06/2023) covered 27 fishing trips with 249 hauls, a total catch of 16374 tonnes and 4634 individual length measurements. Only the herring fishery in ICES division 27.2.a is considered for ASH. Note that there are herring catches in other divisions within the selected trips. These are trips where North Sea herring has been fished with some bycatches of mackerel for
example. Atlanto-scandian herring have a relatively narrow range in the length distributions in the catch. Median lengths have been between 30.8 and 35.2 cm

In this 2023 self-sampling report, a standardized CPUE calculation has been included again for for most of the stocks. The standardized CPUE is based on a GLM model with a negative binomial distribution. The response variable is the catch by week and vessel, with an offset of the log effort (number of fishing days per week) and explanatory variables year, GT category, month, division and depth category. An assumed technical efficiency increase of $2.5 \%$ per year has been included in the fitting of the model (Rousseau et al 2019) Due to changes in the fishery, e.g. low TAC / closure of the Western horse mackerel fishery, CPUEs are also low as most catches are reflective of bycatch rather than targeted fishery catches.

## 1 Introduction

The Pelagic Freezer-trawler Association (PFA) is an association that has nine member companies that together operate 18 freezer trawlers (in 2022) in six European countries (www.pelagicfish.eu). In 2015, the PFA has initiated a self-sampling program that expands the ongoing monitoring programs on board of pelagic freezer-trawlers by the specialized crew of the vessels. The primary objective of that monitoring program is to assess the quality of fish. The expansion in the self-sampling program consists of recording of haul information, recording the species compositions per haul and regularly taking random length-samples from the catch. The self-sampling is carried out by the vessel quality managers on board of the vessels, who have a long experience in assessing the quality of fish, and by the skippers/officers with respect to the haul information. The scientific coordination of the selfsampling program is carried out by Niels Hintzen (PFA chief science officer) with support of Lina de Nijs and Floor Quirijns (contractor).

## 2 Overview of self-sampling methodology

The PFA self-sampling program has been implemented incrementally on many vessels that belong to the members of the PFA. The self-sampling program is designed in such a way that it follows as closely as possible the working practices on board of the different vessels and that it delivers relevant information for documenting the performance of the fishery and to assist stock assessments of the stocks involved. The following main elements can be distinguished in the self-sampling protocol:

- haul information (date, time, position, weather conditions, environmental conditions, gear attributed, estimated catch, optionally: species composition)
- batch information (total catch per batch=production unit, including variables like species, average size, average weight, fat content, gonads $\mathrm{y} / \mathrm{n}$ and stomach fill)
- linking batch and haul information (essentially a key of how much of a batch is caught in which of the hauls)
- length information (length frequency measurements, either by batch or by haul)

The self-sampling information is collected using standardized Excel worksheets. Each participating vessel will send in the information collected during a trip by the end of the trip. The data will be checked and added to the database by Lina de Nijs, who will also generate standardized trip reports (using RMarkdown) which will be sent back to the vessel within one or two days. The compiled data for all vessels is being used for specific purposes, e.g. reporting to expert groups, addressing specific fishery or biological questions and supporting detailed biological studies. The PFA publishes an annual report on the self-sampling program.

A major feature of the PFA self-sampling program is that it is tuned to the capacity of the vessel-crew to collect certain kinds of data. Depending on the number of crew and the space available on the vessel, certain types of measurements can or cannot be carried out. That is why the program is essentially tuned to each vessel separately. And that is also the reason that the totals presented in this report can be somewhat different dependent on which variable is used. For example the estimate of total catch is different from the sum of the catch per species because not all vessels have supplied data on the species composition of the catch.

In order to supply relevant information to WGWIDE, the PFA self-sampling data has been filtered using the following approach. First, all catches per vessel, trip and species have been summed by week. For each vessel-trip-species-week combination, the proportion of the species in the catch were calculated. Then the following filter criteria have applied to the weekly data:

- for horse mackerel: latitude > 45, proportion in the catch > $10 \%$, catch > 10 tonnes
- for mackerel : latitude $>45$, proportion in the catch $>10 \%$, catch $>10$ tonnes
- for blue whiting : latitude $>50$, proportion in the catch $>10 \%$, catch $>10$ tonnes
- for herring : division = 27.2.a, proportion in the catch $>10 \%$, catch $>10$ tonnes

For this report, data have been processed for 2017-2023 (up to 02/06/2023).

## 3 Results

### 3.1 General

An overview of all the self-sampled trips for mac, hom, whb, her_ash in 27.2.a, 27.4.a, 27.6.a, 27.7.b, 27.7.j, 27.7.h, 27.4.c, 27.7.d, 27.7.c, 27.7.k, 27.5.b, 27.8.d. The percentage non-target species is defined as the catch of non-pelagic species relative to the catch of pelagic species.

| year | nvessels | ntrips | ndays | nhauls | catch | catch/day | nontarget | nlength | nbio |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |  |
| 2017 | 12 | 62 | 842 | 1,783 | 178,162 | 212 | 0.258 | 91,881 | 0 |  |
| 2018 | 16 | 86 | 1,220 | 2,679 | 253,474 | 208 | 0.228 | 171,600 | 641 |  |
| 2019 | 16 | 97 | 1,233 | 2,668 | 225,059 | 183 | 0.298 | 126,772 | 1,055 |  |
| 2020 | 17 | 113 | 1,434 | 3,051 | 306,172 | 214 | 0.358 | 167,340 | 2,379 |  |
| 2021 | 19 | 119 | 1,401 | 2,881 | 282,898 | 202 | 0.528 | 140,339 | 1,433 |  |
| 2022 | 18 | 114 | 1,259 | 2,736 | 237,438 | 189 | 0.698 | 96,762 | 4,150 |  |
| $2023 \star$ | 18 | 68 | 785 | 1,934 | 189,832 | 242 | 0.189 | 120,862 | 662 |  |
| (al1) |  | 659 | 8,174 | 17,732 | $1,673,036$ |  |  | 915,557 | 10,320 |  |

Table 3.1.1: PFA fisheries for widely distributed species Self-sampling Summary of number of vessels, trips, days, hauls, catch (tonnes), catch per day and number of fish measured. * denotes incomplete year

Catch and number of self-sampled hauls by year and division

| division | 2017 | 2018 |  | 2019 | 2020 | 2021 | 2022 | 2 2023* | * all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 27.6.a | 75,493 | 126,130 |  | 116,241 | 125,729 | 113,522 | 61,529 | 9 83,551 | 1 702,195 | 42.08 |
| 27.4.a | 23,842 | 36,129 |  | 39,494 | 63,061 | 61,135 | 64,700 | 0 11,022 | 2 299,383 | 17.98 |
| 27.7.c | 29,371 | 30,524 |  | 26,772 | 44,548 | 28,885 | 20,391 | 1 9,471 | 1 189,962 | 11.48 |
| 27.7.k | 96 | 7,646 |  | 2,036 | 11,339 | 16,684 | 29,327 | 7 77,358 | 8 144,486 | 8.68 |
| 27.7.j | 663 | 3,648 |  | 8,635 | 16,322 | 14,976 | 17,193 | 3 2,061 | 63,499 | 3.88 |
| 27.7.d | 8,404 | 9,853 |  | 10,373 | 10,763 | 9,934 | 5,463 | 3 4,563 | 63 59,353 | 3.5\% |
| 27.2.a | 20,469 | 18,096 |  | 4,607 | 10,000 | 2,595 | 1,028 |  | 0 56,795 | 3.48 |
| 27.7.b | 8,605 | 5,324 |  | 10,530 | 11,649 | 13,205 | 6,007 | 7145 | 5 55,464 | 3.3\% |
| 27.5.b | 8,061 | 7,933 |  | 3,925 | 10,277 | 8,689 | 9,970 | $0 \quad 70$ | 0 48,925 | 2.98 |
| 27.8.d | 275 | 237 |  | 173 | 889 | 802 | 17,913 | 3876 | 6 21,165 | 1.3\% |
| 27.7.h | 1,330 | 6,571 |  | 1,236 | 111 | 9,012 | 1,503 |  | $0 \quad 19,761$ | 1.2\% |
| 27.4.c | 1,555 | 1,385 |  | 1,036 | 1,483 | 3,460 | 2,414 | 4715 | 512,048 | 0.7\% |
| (a11) | 178,162 | 253,474 |  | 225,059 | 306,172 | 282,898 | 237,438 | 8 189,832 | 2 1,673,036 | 100.0\% |
| division | 2017 | 2018 | 2019 | 92020 | 2021 | 2022 | 2023* | all | perc |  |
| 27.6.a | 668 | 1,267 1, | 1,281 | 1 1,209 | 966 | 740 | 948 | 7.079 | 39.9\% |  |
| 27.4.a | 191 | 374 | 436 | $6 \quad 548$ | 560 | 676 | 111 | 2,896 | 16.38 |  |
| 27.7.c | 255 | 243 | 252 | 2328 | 255 | 154 | 74 | 1,561 | $8.8 \%$ |  |
| 27.7.k | 3 | 59 | 17 | 795 | 131 | 244 | 659 | 1,208 | 6.88 |  |
| 27.7.d | 153 | 187 | 187 | 7187 | 206 | 157 | 103 | 1,180 | 6.78 |  |
| 27.7.j | 17 | 60 | 137 | 7208 | 289 | 320 | 14 | 1,045 | 5.98 |  |
| 27.7.b | 139 | 88 | 175 | 5 207 | 202 | ${ }^{88}$ | 2 | 901 | 5.18 |  |
| 27.2.a | 237 | 207 | 86 | 6142 | 24 | 24 | 0 | 720 | 4.18 |  |
| 27.5.b | 66 | 82 | 38 | 887 | 54 | 70 | 1 | 398 | 2.28 |  |
| 27.7.h | 30 | 94 | 24 | 4 6 | 144 | 50 | 0 | 348 | 2.08 |  |
| 27.4.c | 22 | 16 | 25 | $5 \quad 21$ | 55 | 54 | 14 | 207 | 1.28 |  |
| 27.8.d | 2 | 2 | 10 | 013 | 7 | 159 | 8 | 201 | 1.18 |  |
| (a11) | 1,783 | 2,679 2 | 2,668 | 3,051 | 2,893 | 2,736 | 1,934 1 | 17,744 10 | 00.08 |  |

Table 3.1.2: PFA fisheries for widely distributed species Self-sampling Summary of catch (top) and number of hauls (bottom) per year and division. * denotes incomplete year

Catch and number of self-sampled hauls by year and month

| month | 2017 | 2018 |  | 2019 | 2020 | 2021 | 2022 | 2 2023* | * all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Jan | 28,644 | 25,647 |  | 35,499 | 37,485 | 51,537 | 40,516 | 6 46,091 | 1 265,420 | 15.9\% |
| Feb | 19,390 | 32,600 |  | 32,829 | 28,300 | 31,967 | 44,851 | 1 58,858 | 248,795 | 14.98 |
| Mar | 29,642 | 32,910 |  | 27,992 | 48,658 | 36,936 | 40,093 | 3 46,958 | 263,188 | 15.78 |
| Apr | 28,510 | 58,665 |  | 28,857 | 66,042 | 29,472 | 25,878 | 8 26,738 | 264,163 | $15.8 \%$ |
| May | 12,367 | 30,227 |  | 21,332 | 29,189 | 14,466 | 8,521 | 1 10,748 | 126,850 | 7.68 |
| Jun | 0 | 6,866 |  | 1,498 | 4,219 | 2,467 |  | $0 \quad 438$ | - 15,488 | 0.98 |
| Jul | 665 | 791 |  | 6,185 | 1,566 | 12,330 | 1,174 |  | 0 22,711 | 1.4\% |
| Aug | 6,545 | 4,551 |  | 3,844 | 4,234 | 4,779 | 3,467 | 70 | 027,420 | 1.68 |
| sep | 9,898 | 8,334 |  | 7,775 | 12,586 | 9,134 | 12,297 | 70 | 0 60,024 | 3.64 |
| $\infty$ ct | 17,478 | 22,975 |  | 25,570 | 27,648 | 39,924 | 27,883 |  | 0 161,478 | 9.78 |
| Nov | 21,875 | 20,385 |  | 22,225 | 27,061 | 30,033 | 28,894 |  | 0 150,472 | $9.0 \%$ |
| Dec | 3,148 | 9,522 |  | 11,453 | 19,184 | 19,853 | 3,866 |  | 0 67,027 | 4.08 |
| (a11) | 178,162 | 253,474 |  | 225,059 | 306,172 | 282,898 | 237,438 | 8 189,832 | 2 1,673,036 | 100.0\% |
| month | 2017 | 2018 | 2019 | 9 2020 | 2021 | 2022 | 2023* | all | perc |  |
| Jan | 311 | 309 | 452 | 2355 | 568 | 474 | 469 | 2,938 16 | 16.6\% |  |
| Feb | 207 | 325 | 362 | 287 | 344 | 435 | 525 | 2,485 1 | 14.0\% |  |
| Mar | 227 | 299 | 314 | 4422 | 333 | 389 | 464 | 2,448 13 | 13.88 |  |
| Apr | 201 | 494 | 289 | 9574 | 240 | 359 | 306 | 2,463 13, | 13.98 |  |
| May | 145 | 372 | 250 | 312 | 167 | 144 | 164 | 1,554 | 8.88 |  |
| Jun | 0 | 77 | 23 | 397 | 42 | 0 | 6 | 245 | 1.48 |  |
| Jul | 12 | 10 | 75 | 15 26 | 113 | 19 | 0 | 255 | 1.48 |  |
| Aug | 58 | 39 | 41 | 153 | 33 | 30 | 0 | 254 | 1.4\% |  |
| sep | 130 | 145 | 149 | 9154 | 187 | 177 | 0 | 942 | 5.3* |  |
| oct | 198 | 232 | 306 | 6 296 | 398 | 370 | 0 | 1,800 10, | 10.18 |  |
| Nov | 269 | 291 | 318 | 831 | 305 | 306 | 0 | 1,820 10 | 10.38 |  |
| Dec | 25 | 86 | 89 | 9144 | 163 | 33 | 0 | 540 | 3.08 |  |
| (all) | 1,783 | 2,679 2, | 2,668 | 3,051 | 2,893 | 2,736 | 1,934 | 17,744 10 | 00.08 |  |

Table 3.1.3: PFA fisheries for widely distributed species Self-sampling summary of catch (top) and number of hauls (bottom) per year and month.

Catch and number of self-sampled hauls by year and country (flag)

| flag | 2017 | 2018 |  | 2019 | 2020 | 2021 | 2022 | 2023* | all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| DEU | 27,500 | 55,468 |  | 40,385 | 69,764 | 54,075 | 35,977 | 7 43,857 | 7 327,027 | 19.58 |
| ER | 0 | 11,981 |  | 19,356 | 14,739 | 12,257 | 14,124 | 4 11,586 | 84,043 | 5.0\% |
| LIT | 0 |  | 0 | 1,414 | 13,744 | 23,150 | 10,566 | 6 14,915 | 5 63,788 | 3.8\% |
| NL | 114,844 | 139,403 |  | 107,049 | 117,284 | 124,972 | 126,082 | 2 83,262 | 812,896 | 48.68 |
| POL | 0 | 15,966 |  | 28,044 | 54,615 | 29,675 | 16,202 | 2 18,975 | 5 163,478 | 9.83 |
| UK | 35,818 | 30,657 |  | 28,811 | 36,026 | 35,341 | 34,487 | 7 17,237 | 7 218,376 | 13.18 |
| NA | 0 |  | 0 | 0 | 0 | 3,428 |  | 0 | 0 3,428 | 0.28 |
| (al1) | 179,162 | 253,474 |  | 225,059 | 306,172 | 282,898 | 237,438 | 8 189,832 | 2 1,673,036 | 100.08 |
| $f 1 \mathrm{ag}$ | 2017 | 2018 | 2019 | 92020 | 2021 | 2022 | 2023* | all | perc |  |
| DEU | 276 | 637 | 456 | $6 \quad 633$ | 463 | 376 | 446 | 3,287 | 18.58 |  |
| ER | 0 | 237 | 357 | 7245 | 205 | 234 | 186 | 1,464 | 8.38 |  |
| LIT | 0 | 0 | 34 | $34 \quad 142$ | 165 | 69 | 104 | 514 | 2.98 |  |
| NL | 1,177 | 1,403 1, | 1,318 | 8 1,374 | 1,392 | 1,489 | 842 | 8,995 | 50.78 |  |
| POL | 0 | 111 | 189 | 9323 | 187 | 146 | 169 | 1,125 | 6.3\% |  |
| UK | 330 | 291 | 314 | 4334 | 394 | 422 | 187 | 2,272 | 12.88 |  |
| NA | 0 | 0 |  | 0 | 75 | 0 | 0 | 75 | 0.48 |  |
| (all) | 1,783 | 2,679 2, | 2,668 | 58 3,051 | 2,881 | 2,736 | 1,934 1 | 17,732 10 | 00.08 |  |

Table 3.1.4: PFA fisheries for widely distributed species Self-sampling summary of catch (top) and number of hauls (bottom) per year and month.

## Catch by species and year

| $\begin{aligned} & \text { species } \\ & \text { perc } \end{aligned}$ | english_name | scientific_name | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023 | all |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| whb | blue whiting | Micromesistius poutassou | 79,108 | 154,733 | 113,434 | 174,971 | 149,988 | 122,877 | 150,653 | 945,763 |
| 56.58 \% ${ }^{\text {5 }}$ |  |  |  |  |  |  |  |  |  |  |
| mac | mackerel | Scomber scombrus | 63,279 | 55,989 | 54,005 | 84,855 | 69,094 | 63,097 | 26,639 | 416,958 |
| 24.98 |  |  |  |  |  |  |  |  |  |  |
| hom | horse mackerel | Trachurus trachurus | 20,874 | 28,501 | 31,565 | 25,061 | 34,130 | 20,381 | 2,599 | 163,112 |
| 9.7\% |  |  |  |  |  |  |  |  |  |  |
| her | herring | Clupea harengus | 6,870 | 7,851 | 17,286 | 9,154 | 19,912 | 18,615 | 3,189 | 82,876 |
| 5.08 |  |  |  |  |  |  |  |  |  |  |
| arg | argentines | Argentina spp | 2,596 | 4,097 | 4,566 | 7,036 | 5,457 | 9,595 | 5,449 | 38,794 |
| 2.38 |  |  |  |  |  |  |  |  |  |  |
| her_ash | NA | NA | 4,913 | 1,367 | 3,373 | 3,563 | 2,379 | 778 | 0 | 16,375 |
| $1.08^{-1}$ |  |  |  |  |  |  |  |  |  |  |
| boc | boarfish | Capros aper | 245 | 153 | 288 | 603 | 846 | 744 | 28 | 2,907 |
| 0.28 |  |  |  |  |  |  |  |  |  |  |
| pil | pilchard | Sardina pilchardus | 61 | 371 | 155 | 32 | 325 | 449 | 970 | 2,364 |
| $0.18{ }^{18}$ |  |  |  |  |  |  |  |  |  |  |
| hke | hake | Merluccius merluccius | 107 | 270 | 197 | 181 | 240 | 329 | 131 | 1,456 |
| 0.18 |  |  |  |  |  |  |  |  |  |  |
| spr | sprat | Sprattus sprattus | 0 | 0 | 0 | 415 | 138 | 0 | 0 | 553 |
| 0.0\% |  |  |  |  |  |  |  |  |  |  |
| had | haddock | Melanogrammus aeglefinus | 5 | 15 | 46 | 42 | 66 | 121 | 47 | 342 |
| 0.08 er |  |  |  |  |  |  |  |  |  |  |
| sqr | squid | Loligo vulgaris | 0 | 8 | 8 | 26 | 133 | 68 | 8 | 251 |
| $0.0 \%$ |  |  |  |  |  |  |  |  |  |  |
| brb | black seabream | Spondyliosoma cantharus | 2 | 22 | 3 | 83 | 5 | 6 | 86 | 209 |
| 0.08 |  |  |  |  |  |  |  |  |  |  |
| med | NA | Ceratoscopelus maderensis | 0 | 0 | 32 | 28 | 20 | 105 | 8 | 194 |
| 0.08 |  |  |  |  |  |  |  |  |  |  |
| whg | whiting | Merlangius merlangus | 0 | 24 | 31 | 31 | 30 | 63 | 2 | 181 |
| 0.08 |  |  |  |  |  |  |  |  |  |  |
| oth | NA | NA | 101 | 74 | 70 | 91 | 134 | 210 | 22 | 703 |
|  |  |  |  |  |  |  |  |  |  |  |
| (a11) | (a11) | (all) | 178,162 | 253,474 | 225,059 | 306,172 | 282,898 | 237,438 | 189,832 | 1,673,036 |
| 100.0\% |  |  |  |  |  |  |  |  |  |  |

Table 3.1.5: PFA fisheries for widely distributed species Self-sampling Summary of total catch (tonnes) by species. OTH refers to all other species that are not the main target species

## Haul positions

An overview of all self-sampled hauls in the PFA fisheries for widely distributed species.


Figure 3.1.1: PFA fisheries for widely distributed species Self-sampling haul positions. $N$ indicates the number of hauls.

## Catches for the main target species

Summed catches (tonnes) of the main target species aggregated in rectangles.


Figure 3.1.2: PFA fisheries for widely distributed species Self-sampling catch per species and per rectangle. $N$ indicates the number of hauls. Catch refers to the total catch per year.

## Catch rates (catch/day) for the main target species



Figure 3.1.3: Average catch per day, per species and per rectangle. $N$ indicates the number of hauls; avg refers to the average catch per day.

Average surface temperature by quarter and by rectangle.


Figure 3.1.4: PFA fisheries for widely distributed species Average surface temperature (C) by year and quarter. $N$ indicates the number of hauls. Avg refers to the average temperature.

## Average fishing depth



Figure 3.1.5: PFA fisheries for widely distributed species Average fishing depth ( $m$ ) by year and quarter. $N$ indicates the number of hauls. Avg refers to the average fishing depth.

## Average wind force.



Figure 3.1.6: PFA fisheries for widely distributed species Average windforce (Bft) by year and quarter. $N$ indicates the number of hauls. Avg refers to the average windforce.
3.2 Northeast Atlantic mackerel (MAC, Scomber scombrus)

Northeast Atlantic mackerel self-sampling summary.

| species | year | nvessels | ntrips | ndays | nhauls | catch | catch/day | nlength | nbio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2017 | 11 | 48 | 386 | 690 | 62,715 | 162 | 11,731 | 0 |
| mac | 2018 | 16 | 56 | 501 | 841 | 55,186 | 110 | 13,789 | 32 |
| mac | 2019 | 15 | 72 | 615 | 1,105 | 53,525 | 87 | 18,037 | 476 |
| mac | 2020 | 17 | 84 | 712 | 1,258 | 83,876 | 118 | 31,555 | 64 |
| mac | 2021 | 18 | 78 | 606 | 1,054 | 68,466 | 113 | 11,297 | 684 |
| mac | 2022 | 14 | 74 | 568 | 1,031 | 62,847 | 111 | 15,347 | 3,874 |
| mac | 2023 | 14 | 26 | 243 | 468 | 26,279 | 108 | 5,752 | 129 |
| (all) | (all) |  | 438 | 3,631 | 6,447 | 412,894 |  | 107,508 | 5,841 |

Table 3.2.1: Northeast Atlantic mackerel. Self-sampling summary with the number of days, hauls, trips, vessels, catch (tonnes), catch rate (ton/day), number of fish measured, number of biological observations.

Northeast Atlantic mackerel. Catch by division

| species | division | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | $2023 *$ | all | perc |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |  |

Table 3.2.2: Northeast Atlantic mackerel. Self-sampling summary with the catch (tonnes) by year and division

Northeast Atlantic mackerel. Catch by month

| species | month | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023* | all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Jan | 18,550 | 11,546 | 18,715 | 20,750 | 14,806 | 13,050 | 20,294 | 117,712 | 28.5\% |
| mac | Feb | 8,199 | 7,297 | 11,862 | 19,376 | 5,678 | 6,925 | 5,188 | 64,525 | 15.68 |
| mac | Mar | 4,469 | 1,292 | 4,374 | 5,114 | 2,840 | 6,613 | 13 | 24,716 | 6.08 |
| mac | Apr | 955 | 1,226 | 1,326 | 604 | 366 | 98 | 731 | 5,306 | 1.3\% |
| mac | May | 288 | 192 | 489 | 1,239 | 97 | 71 | 51 | 2,428 | 0.68 |
| mac | Jun | 0 | 60 | 96 | 173 | 35 | 0 | 3 | 368 | 0.13 |
| mac | Jul | 89 | 0 | 262 | 83 | 907 | 55 | 0 | 1,396 | 0.3\% |
| mac | Aug | 237 | 59 | 431 | 296 | 360 | 389 | 0 | 1,772 | 0.48 |
| mac | sep | 9,096 | 4,779 | 3,039 | 6,284 | 2,624 | 3,406 | 0 | 29,228 | 7.18 |
| mac | $\infty$ ct | 7,866 | 19,437 | 11,457 | 20,161 | 30,743 | 17,733 | 0 | 107,398 | 26.08 |
| mac | Nov | 11,595 | 8,934 | 1,473 | 9,461 | 10,009 | 14,231 | 0 | 55,704 | 13.58 |
| mac | Dec | 1,370 | 363 | 0 | 334 | 0 | 275 | 0 | 2,342 | 0.68 |
| (all) | (a11) | 62,715 | 55,186 | 53,525 | 83,876 | 68,466 | 62,847 | 26,279 | 412,894 | 100.0\% |

Table 3.2.3: Northeast Atlantic mackerel. Self-sampling summary with the catch (tonnes) by year and month

## Northeast Atlantic mackerel. Catch by country

```
species flag 2017 2018 2019 2020 2021 
    mac DEU 6,934 9,760 8,735
```

| mac | FR | 0 | 8,096 | 8,962 | 6,375 | 7,086 | 6,226 | 3,003 | 39,748 | 9.68 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| mac | LIT | 0 | 0 | 0 | 827 | 6,876 | 0 | 0 | 7,704 | 1.98 |
| mac | NL | 29,171 | 12,670 | 14,885 | 27,424 | 20,674 | 23,163 | 4,464 | 132,452 | 32.18 |
| mac | POL | 0 | 4,051 | 3,601 | 5,502 | 1,771 | 0 | 140 | 15,066 | 3,68 |
| mac | UK | 26,610 | 20,608 | 17,341 | 20,952 | 19,704 | 21,197 | 7,496 | 133,708 | 3248 |
| mac | NA | 0 | 0 | 0 | 0 | 2,049 | 0 | 0 | 2,049 | 0.58 |
| (all) | (all) | 62,715 | 55,186 | 53,525 | 83,876 | 68,466 | 62,847 | 26,279 | 412,894 | 10008 |

Table 3.2.4: Northeast Atlantic mackerel. Self-sampling summary with the catch (tonnes) by year and country

Northeast Atlantic mackerel. Catch by rectangle


Figure 3.2.1: Northeast Atlantic mackerel. Catch per per rectangle. $N$ indicates the number of hauls; Catch refers to the total catch per year.

Northeast Atlantic mackerel. Catchrate (ton/day) by rectangle


Figure 3.2.2: Northeast Atlantic mackerel. Catchrate (ton/day) per rectangle. $N$ indicates the number of hauls; Avg refers to the average catchrate per rect.

Northeast Atlantic mackerel. Spatio-temporal evolution of catch by month and rectangle


Figure 3.2.3: Northeast Atlantic mackerel. Spatio-temporal evolution of the catches per rectangle and month. $N$ indicates the number of hauls; $C$ refers to the total catch by year and month.

## Northeast Atlantic mackerel. Catch proportion at depth



Figure 3.2.4: Northeast Atlantic mackerel. Catch proportion at depth. $N$ indicates the number of hauls.

Northeast Atlantic mackerel. Length distributions of the catch


Figure 3.2.5: Northeast Atlantic mackerel. Length distributions by year (top) and by year and division (bottom). Nobs refers to the number of observations; median denotes the median length.

Northeast Atlantic mackerel. Length distributions as proportions by (large) rectangle


Figure 3.2.6: Northeast Atlantic mackerel. Length distributions as proportions by large rectangle. Ind. refers to the number of length measurements

Northeast Atlantic mackerel. Average length, weight and fat content by year and month


Figure 3.2.7: Northeast Atlantic mackerel. Average length, average weight, and average fat content. Nobs indicates the number of measurements, median indicates the median values

## Northeast Atlantic mackerel (MAC). Standardized CPUE

Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with $\log$ (days) as offset. It is assumed that a $2.5 \%$ annual efficiency increase takes place (Rousseau et al 2019).


Figure 3.2.8: Northeast Atlantic mackerel. Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with log(days) as offset

### 3.3 Western horse mackerel (HOM, Trachurus trachurus)

Western horse mackerel self-sampling summary.

| species | year | nvessels | ntrips | ndays | nhauls | catch | catch/day | nlength | nbio |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |  |

Table 3.3.1: Western horse mackerel. Self-sampling summary with the number of days, hauls, trips, vessels, catch (tonnes), catch rate (ton/day), number of fish measured, number of biological observations.

Western horse mackerel. Catch by division

| species | division | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | $2023 *$ | all | perc |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |  |

Table 3.3.2: Western horse mackerel. Self-sampling summary with the catch (tonnes) by year and division

Western horse mackerel. Catch by month

| species | month | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023* | all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| hom | Jan | 6,666 | 10,627 | 9,610 | 7,017 | 4,894 | 9,511 | 14 | 48,339 | 41.98 |
| hom | Feb | 3,052 | 5,392 | 3,257 | 4,774 | 6,634 | 909 | 18 | 24,036 | 20.88 |
| hom | Mar | 212 | 3,027 | 1,284 | 1,237 | 245 | 413 | 0 | 6,418 | 5.68 |
| hom | Apr | 0 | 31 | 45 | 0 | 6 | 0 | 144 | 226 | 0.2\% |
| hom | May | 156 | 7 | 42 | 529 | 2 | 0 | 0 | 735 | $0.6 \%$ |
| hom | Jun | 0 | 227 | 1,357 | 642 | 0 | 0 | 0 | 2,226 | 1.9 |
| hom | Jul | 112 | 15 | 5,342 | 420 | 5,809 | 274 | 0 | 11,971 | 10.48 |
| hom | Aug | 0 | 0 | ${ }^{8}$ | 0 | 1,005 | 13 | 0 | 1,026 | 0.98 |
| hom | Sep | 0 | 429 | 335 | 0 | 4,300 | 60 | 0 | 5,125 | 4.48 |
| hom | $\infty$ ct | 15 | 126 | 259 | 1 | 831 | 3,813 | 0 | 5,046 | 4.48 |
| hom | Nov | 1,262 | 1,410 | 2,483 | 1,713 | 2,629 | 298 | 0 | 9,796 | 8.58 |
| hom | Dec | 103 | 120 | 0 | 0 | 221 | 0 | 0 | 444 | 0.48 |
| (a11) | (a11) | 11,578 | 21,412 | 24,022 | 16,334 | 26,576 | 15,291 | 176 | 115,388 | 100.08 |

Table 3.3.3: Western horse mackerel. Self-sampling summary with the catch (tonnes) by year and month

## Western horse mackerel. Catch by country

| species | flag | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023* | all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| hom | DEU | 1,803 | 4,069 | 2,602 | 977 | 4,155 | 725 | 5 | 14,337 | 12.4 |

$$
\begin{array}{rrrrrrrrrrr}
\text { hom } & \text { FR } & 0 & 622 & 864 & 1,370 & 788 & 1,406 & 13 & 5,062 & 4.48 \\
\text { hom } & \text { NL } & 9,239 & 14,617 & 18,011 & 11,535 & 18,234 & 12,698 & 149 & 84,483 & 73.28 \\
\text { hom } & \text { PoL } & 0 & 0 & 4 & 1,005 & 1,210 & 0 & 0 & 2,219 & 1.98 \\
\text { hom } & \text { UK } & 535 & 2,104 & 2,541 & 1,447 & 2,014 & 462 & 9 & 9,113 & 7.98 \\
\text { hom } & \text { NA } & 0 & 0 & 0 & 0 & 175 & 0 & 0 & 175 & 0.28 \\
\text { (all) } & \text { (all) } & 11,578 & 21,412 & 24,022 & 16,334 & 26,576 & 15,291 & 176 & 115,388 & 100.08
\end{array}
$$

Table 3.3.4: Western horse mackerel. Self-sampling summary with the catch (tonnes) by year and country

Western horse mackerel. Catch by rectangle


Figure 3.3.1: Western horse mackerel. Catch per per rectangle. Nindicates the number of hauls; Catch refers to the total catch per year.

Western horse mackerel. Catchrate (ton/day) by rectangle


Figure 3.3.2: Western horse mackerel. Catchrate (ton/day) per rectangle. $N$ indicates the number of hauls; Avg refers to the average catchrate per rect.

Western horse mackerel. Spatio-temporal evolution of catch by month and rectangle


Figure 3.3.3: Western horse mackerel. Spatio-temporal evolution of the catches per rectangle and month. $N$ indicates the number of hauls; C refers to the total catch by year and month.

Western horse mackerel. Catch proportion at depth


Figure 3.3.4: Western horse mackerel. Catch proportion at depth. $N$ indicates the number of hauls.

Western horse mackerel. Length distributions of the catch


Figure 3.3.5: Western horse mackerel. Length distributions by year (top) and by year and division (bottom). Nobs refers to the number of observations; median denotes the median length.

Western horse mackerel. Length distributions as proportions by (large) rectangle


Figure 3.3.6: Western horse mackerel. Length distributions as proportions by large rectangle. Ind. refers to the number of length measurements

Western horse mackerel. Average length, weight and fat content by year and month


Figure 3.3.7: Western horse mackerel. Average length, average weight, and average fat content. Nobs indicates the number of measurements, median indicates the median values

## Western horse mackerel (HOM). Standardized CPUE

Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with $\log$ (days) as offset. It is assumed that a $2.5 \%$ annual efficiency increase takes place (Rousseau et al 2019).


Figure 3.3.8: Western horse mackerel. Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with log(days) as offset

### 3.4 North Sea horse mackerel (HOM, Trachurus trachurus)

North Sea horse mackerel self-sampling summary.

| species | year | nvessels | ntrips | ndays | nhauls | catch | catch/day | nlength | nbio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| hom | 2017 | 6 | 14 | 81 | 156 | 8,568 | 106 | 1,013 | 0 |
| hom | 2018 | 5 | 13 | 80 | 146 | 7,079 | 88 | 4,349 | 0 |
| hom | 2019 | 8 | 14 | 78 | 143 | 7,417 | 95 | 9,454 | 0 |
| hom | 2020 | 7 | 21 | 94 | 150 | 8,726 | 93 | 10,685 | 829 |
| hom | 2021 | 8 | 22 | 94 | 153 | 7,259 | 77 | 6,320 | 0 |
| hom | 2022 | 7 | 23 | 88 | 144 | 5,090 | 58 | 4,756 | 0 |
| hom | 2023 | 5 | 13 | 47 | 77 | 2,423 | 52 | 2,404 | 0 |
| (a11) | (a11) |  | 120 | 562 | 969 | 46,563 |  | 38,981 | 829 |

Table 3.4.1: North Sea horse mackerel. Self-sampling summary with the number of days, hauls, trips, vessels, catch (tonnes), catch rate (ton/day), number of fish measured, number of biological observations.

North Sea horse mackerel. Catch by division

| species | division | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023* | all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| hom | 27.4.c | 1,371 | 853 | 369 | 898 | 1,149 | 1,558 | 715 | 6,913 | 14.8 ¢ |

$$
\begin{array}{rrrrrrrrrrr}
\text { hom } & \text { 27.7.d } & 7,198 & 6,226 & 7,048 & 7,829 & 6,111 & 3,531 & 1,708 & 39,650 & 85.28 \\
\text { (a11) } & \text { (all) } & 8,568 & 7,079 & 7,417 & 8,726 & 7,259 & 5,090 & 2,423 & 46,563 & 100.08
\end{array}
$$

Table 3.4.2: North Sea horse mackerel. Self-sampling summary with the catch (tonnes) by year and division

North Sea horse mackerel. Catch by month

| species | month | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023* | all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| hom | Jan | 2,362 | 892 | 1,382 | 2 | 1,013 | 538 | 1,294 | 7,484 | 16.18 |
| hom | Feb | 0 | 310 | 0 | 0 | 97 | 376 | 849 | 1,632 | 3.58 |
| hom | Mar | 0 | 0 | 0 | 0 | 0 | 0 | 280 | 280 | 0.69 |
| hom | Jun | 0 | 0 | 0 | 6 | 25 | 0 | 0 | 31 | 0.18 |
| hom | sep | 135 | 1,471 | 2,009 | 3,860 | 422 | 1,953 | 0 | 9,850 | .23 |
| hom | oct | 4,490 | 1,391 | 1,967 | 1,834 | 2,349 | 1,398 | 0 | 13,429 | .88 |
| hom | Nov | 1,581 | 2,018 | 1,110 | 1,463 | 1,218 | 485 | 0 | 7,876 | 16.98 |
| hom | Dec | 0 | 998 | 9 | 1,561 | 2,134 | \% | 0 | 5,982 | 12.8 |
| (all) | (a11) | 8,568 | 7,079 | 7,417 | 8,726 | 7,259 | 5,090 | 2,423 | 46,563 | 100.0 |

Table 3.4.3: North Sea horse mackerel. Self-sampling summary with the catch (tonnes) by year and month

North Sea horse mackerel. Catch by country

| species | flag | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023* | all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| hom | deu | 0 | 1,378 | 958 | 0 | 0 | 12 | 0 | 2,348 | 5.08 |
| hom | FR | 0 | 422 | 400 | 238 | 202 | 0 | 0 | 1,262 | 2.78 |
| hom | LIT | 0 | 0 | 1,373 | 0 | 0 | 0 | 0 | 1,373 | 2.98 |
| hom | NL | 4,887 | 1,578 | 1,682 | 4,167 | 2,356 | 2,103 | 90 | 17,675 | 8.0\% |
| hom | UK | 3,682 | 3,701 | 3,004 | 4,322 | 3,674 | 2,975 | 1,519 | 22,876 | 49.18 |
| hom | NA | 0 | 0 | 0 | 0 | 1,028 | 0 | 0 | 1,028 | 2.23 |
| (all) | (a11) | 8,568 | 7,079 | 7,417 | 8,726 | 7,259 | 5,090 | 2,423 | 46,563 | 100.0 |

Table 3.4.4: North Sea horse mackerel. Self-sampling summary with the catch (tonnes) by year and country


Figure 3.4.1: North Sea horse mackerel. Catch per per rectangle. $N$ indicates the number of hauls; Catch refers to the total catch per year.

North Sea horse mackerel. Catchrate (ton/day) by rectangle


Figure 3.4.2: North Sea horse mackerel. Catchrate (ton/day) per rectangle. N indicates the number of hauls; Avg refers to the average catchrate per rect.


Figure 3.4.3: North Sea horse mackerel. Spatio-temporal evolution of the catches per rectangle and month. $N$ indicates the number of hauls; C refers to the total catch by year and month.

North Sea horse mackerel. Catch proportion at depth


Figure 3.4.4: North Sea horse mackerel. Catch proportion at depth. $N$ indicates the number of hauls.

North Sea horse mackerel. Length distributions of the catch


Figure 3.4.5: North Sea horse mackerel. Length distributions by year (top) and by year and division (bottom). Nobs refers to the number of observations; median denotes the median length.

North Sea horse mackerel. Length distributions as proportions by (large) rectangle


Figure 3.4.6: North Sea horse mackerel. Length distributions as proportions by large rectangle. Ind. refers to the number of length measurements

North Sea horse mackerel. Average length, weight and fat content by year and month


Figure 3.4.7: North Sea horse mackerel. Average length, average weight, and average fat content. Nobs indicates the number of measurements, median indicates the median values

## North Sea horse mackerel (HOM). Standardized CPUE

Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with $\log$ (days) as offset. It is assumed that a $2.5 \%$ annual efficiency increase takes place (Rousseau et al 2019).


Figure 3.4.8: North Sea horse mackerel. Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with $\log$ (days) as offset
3.5 Blue whiting (WHB, Micromesistius pouttasseu)

Blue whiting self-sampling summary.

| species | year | nvessels | ntrips | ndays | nhauls | catch | catch/day | nlength | nbio |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| whb | 2017 | 8 | 32 | 348 | 767 | 79,056 | 227 | 64,477 | 0 |
| whb | 2018 | 14 | 45 | 577 | 1,440 | 152,791 | 265 | 119,752 | 0 |
| whb | 2019 | 15 | 51 | 507 | 1,176 | 112,619 | 222 | 56,187 | 0 |
| whb | 2020 | 13 | 61 | 738 | 1,692 | 174,836 | 237 | 90,496 | 78 |
| whb | 2021 | 17 | 68 | 635 | 1,356 | 143,665 | 226 | 61,743 | 0 |
| whb | 2022 | 15 | 48 | 579 | 1,312 | 121,762 | 210 | 46,958 | 0 |
| whb | 2023 | 15 | 48 | 567 | 1,432 | 150,495 | 265 | 101,318 | 533 |
| (all) | (a11) |  | 353 | 3,951 | 9,175 | 935,225 |  | 540,932 | 711 |

Table 3.5.1: Blue whiting. Self-sampling summary with the number of days, hauls, trips, vessels, catch (tonnes), catch rate (ton/day), number of fish measured, number of biological observations.

Blue whiting. Catch by division

| species | division | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023* | all | perc |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| whb | 27.2.a | 2,550 | 11,907 | 998 | 5,718 | 190 | 125 | 0 | 21,487 | 2.3\% |
| whb | 27.4.a | 727 | 2,947 | 2,087 | 5,567 | 2,995 | 1 | 47 | 14,370 | 1.5\% |
| whb | 27.5.b | 7,960 | 7,928 | 3,905 | 10,220 | 8,665 | 9,798 | 65 | 48,540 | 5.28 |
| whb | 27.6.a | 39,085 | 91,738 | 75,707 | 97,232 | 84,794 | 41,129 | 60,795 | 490,481 | 52.48 |


| whb | $27.7 . c$ | 28,731 | 30,504 | 26,587 | 44,309 | 28,613 | 20,360 | 9,471 | 188,575 | 20.28 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| whb | $27.7 . j$ | 4 | 122 | 1,127 | 159 | 1,062 | 3,112 | 1,883 | 7,469 | 0.88 |
| whb | $27.7 . \mathrm{k}$ | 0 | 7,646 | 2,036 | 11,307 | 16,684 | 29,327 | 77,358 | 144,357 | 15.48 |
| whb | $27.8 . \mathrm{d}$ | 0 | 0 | 171 | 324 | 663 | 17,912 | 876 | 19,945 | 2.18 |
| 〈all〉 | $\langle$ all〉 | 79,056 | 152,791 | 112,619 | 174,836 | 143,665 | 121,762 | 150,495 | 935,225 | 100.08 |

Table 3．5．2：Blue whiting．Self－sampling summary with the catch（tonnes）by year and division Blue whiting．Catch by month

| species | month | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | 2023＊ | all | per |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| whb | Jan | 185 | 957 | 4，287 | 9，527 | 29，603 | 14，664 | 21，746 | 80，969 | 8. |
| whb | Feb | 8，027 | 19，108 | 17，504 | 4，051 | 18，915 | 35，879 | 52，004 | 155，488 | 16. |
| whb | Mar | 24，683 | 26，954 | 21，414 | 41，462 | 30，137 | 32，928 | 45，949 | 223，527 | 23. |
| whb | Apr | 27，316 | 55，654 | 26，391 | 61，978 | 25，153 | 19，539 | 24，163 | 240，194 | 25. |
| whb | May | 9，394 | 26，728 | 17，281 | 24，321 | 10，661 | 4，020 | 6，632 | 99，038 | 10. |
| whb | Jun | － | 5，094 | 13 | 872 | 461 | 0 | － | 6，440 |  |
| whb | Jul | 0 | 0 | 102 | 59 | 703 | 75 | 0 | 939 | 0. |
| whb | Aug | 1，265 | 4，219 | 337 | 2，043 | 207 | 1 | 0 | 8，072 | 0. |
| whb | Sep | 538 | 414 | 246 | 1，327 | 128 | 26 | － | 2，678 | 0. |
| whb | $\infty$ ct | 39 | 92 | 1，483 | 2，417 | 5 | 27 | 0 | 4，063 |  |
| whb | Nov | 5，935 | 6，619 | 13，930 | 10，934 | 11，373 | 11，421 | 0 | 60，212 |  |
| whb | Dec | 1，675 | 6，952 | 9，632 | 15，845 | 16，318 | 3，182 | 0 | 53，603 |  |

Table 3．5．3：Blue whiting．Self－sampling summary with the catch（tonnes）by year and month Blue whiting．Catch by country

| species | flag | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | $2023 *$ | all | perc |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  |  |  |  |  |  |  |  |  |  |  |

Table 3．5．4：Blue whiting．Self－sampling summary with the catch（tonnes）by year and country

Blue whiting. Catch by rectangle


Figure 3.5.1: Blue whiting. Catch per per rectangle. $N$ indicates the number of hauls; Catch refers to the total catch per year.


Figure 3.5.2: Blue whiting. Catchrate (ton/day) per rectangle. $N$ indicates the number of hauls; Avg refers to the average catchrate per rect.

Blue whiting. Spatio-temporal evolution of catch by month and rectangle


Figure 3.5.3: Blue whiting. Spatio-temporal evolution of the catches per rectangle and month. N indicates the number of hauls; C refers to the total catch by year and month.

Blue whiting. Catch proportion at depth


Figure 3.5.4: Blue whiting. Catch proportion at depth. $N$ indicates the number of hauls.

Blue whiting. Length distributions of the catch


Figure 3.5.5: Blue whiting. Length distributions by year (top) and by year and division (bottom). Nobs refers to the number of observations; median denotes the median length.

Blue whiting. Length distributions as proportions by (large) rectangle


Figure 3.5.6: Blue whiting. Length distributions as proportions by large rectangle. Ind. refers to the number of length measurements

Blue whiting. Average length, weight and fat content by year and month


Figure 3.5.7: Blue whiting. Average length, average weight, and average fat content. Nobs indicates the number of measurements, median indicates the median values

## Blue whiting (WHB). Standardized CPUE

Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with $\log$ (days) as offset. It is assumed that a $2.5 \%$ annual efficiency increase takes place (Rousseau et al 2019).


Figure 3.5.8: Blue whiting. Standardized CPUE (ton/day) from GLM model with factors year, month, GT, division and depth with $\log ($ days) as offset

### 3.6 Atlanto-scandian herring (HER_ASH, Clupea harengus)

Atlanto-scandian herring self-sampling summary.

| species | year | nvessels | ntrips | ndays | nhauls | catch | catch/day | nlength |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| her_ash | 2017 | 4 | 7 | 31 | 58 | 4,913 | 158 | 678 |
| her_ash | 2018 | 2 | 3 | 18 | 27 | 1,367 | 76 | 4 |
| her_ash | 2019 | 3 | 4 | 23 | 59 | 3,373 | 147 | 838 |
| her_ash | 2020 | 6 | 8 | 44 | 69 | 3,563 | 81 | 989 |
| her_ash | 2021 | 3 | 3 | 10 | 16 | 2,379 | 238 | 1,469 |
| her_ash | 2022 | 2 | 2 | 11 | 20 | 778 | 71 | 657 |
| (all) | (a11) |  | 27 | 137 | 249 | 16,375 |  | 4,635 |

Table 3.6.1: Atlanto-scandian herring. Self-sampling summary with the number of days, hauls, trips, vessels, catch (tonnes), catch rate (ton/day), number of fish measured, number of biological observations.

Atlanto-scandian herring. Catch by division

| species | division | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | all | perc |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| her_ash | 27.2.a | 4,913 | 1,367 | 3,373 | 3,563 | 2,379 | 778 | 16,375 | 100.08 |
| (all) | (all) | 4,913 | 1,367 | 3,373 | 3,563 | 2,379 | 778 | 16,375 | 100.08 |

Table 3.6.2: Atlanto-scandian herring. Self-sampling summary with the catch (tonnes) by year and division

Atlanto-scandian herring. Catch by month

| species | month | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | all | perc |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| - |  |  |  |  |  |  |  |  |  |
| her_ash | May | 0 | 0 | 0 | 26 | 0 | 0 | 26 | 0.28 |
| her_ash | Aug | 118 | 52 | 0 | 61 | 0 | 0 | 232 | 1.48 |
| her_ash | Sep | 7 | 405 | 362 | 53 | 0 | 0 | 827 | 5.18 |
| her_ash | cot | 4,788 | 910 | 2,184 | 2,480 | 1,659 | 0 | 12,021 | 73.48 |
| her_ash | Nov | 0 | 0 | 828 | 942 | 721 | 778 | 3,269 | 20.08 |
| (all) | (all) | 4,913 | 1,367 | 3,373 | 3,563 | 2,379 | 778 | 16,375 | 100.08 |

Table 3.6.3: Atlanto-scandian herring. Self-sampling summary with the catch (tonnes) by year and month

## Atlanto-scandian herring. Catch by country

| species | flag | 2017 | 2018 | 2019 | 2020 | 2021 | 2022 | all | perc |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| - |  |  |  |  |  |  |  |  |  |

Table 3.6.4: Atlanto-scandian herring. Self-sampling summary with the catch (tonnes) by year and country

## Atlanto-scandian herring. Catch by rectangle



Figure 3.6.1: Atlanto-scandian herring. Catch per per rectangle. $N$ indicates the number of hauls; Catch refers to the total catch per year.

Atlanto-scandian herring. Catchrate (ton/day) by rectangle


Figure 3.6.2: Atlanto-scandian herring. Catchrate (ton/day) per rectangle. $N$ indicates the number of hauls; Avg refers to the average catchrate per rect.

## Atlanto-scandian herring. Spatio-temporal evolution of catch by month and rectangle



Figure 3.6.3: Atlanto-scandian herring. Spatio-temporal evolution of the catches per rectangle and month. $N$ indicates the number of hauls; $C$ refers to the total catch by year and month.

Atlanto-scandian herring. Catch proportion at depth


Figure 3.6.4: Atlanto-scandian herring. Catch proportion at depth. $N$ indicates the number of hauls.

## Atlanto-scandian herring. Length distributions of the catch



Figure 3.6.5: Atlanto-scandian herring. Length distributions by year (top) and by year and division (bottom). Nobs refers to the number of observations; median denotes the median length.

## Atlanto-scandian herring. Length distributions as proportions by (large) rectangle



Figure 3.6.6: Atlanto-scandian herring. Length distributions as proportions by large rectangle. Ind. refers to the number of length measurements

## Atlanto-scandian herring. Average length, weight and fat content by year and month



Figure 3.6.7: Atlanto-scandian herring. Average length, average weight, and average fat content. Nobs indicates the number of measurements, median indicates the median values

## 4 Discussion and conclusions

The PFA self-sampling program has been carried out for the nineth year in a row (2015-2023). Here, results have been presented for the years 2017-2023 in terms of meta-information on the sampling (number of vessels, trips, days and length measurements per area and/or season), in terms of the spatio-temporal distribution of catches and the length and weight compositions by area and/or season.

The definition of what constitutes the 'widely distributed fishery' has been approached by selecting all combination of vessel-trip-weeks where hauls were taken in a certain area and where the catch composition consisted of a minimum percentage of certain species (blue whiting, mackerel, horse mackerel, Atlanto-scandian herring) and a minimum weekly catch of 10 tons. Although for herring we aimed to select only trips for Atlanto-scandian herring (in division 27.2.a) some trips with North Sea herring have been included because they were combined with some fishing for mackerel.

The CPUE of Mackerel shows a relative stable picture of the stock in the recent years with a small downward trend from 2020 to 2022. Some smaller fish has been caught in the early part of 2023 that has been absent in previous years. The fishery has been similar in distribution compared to other years except that almost no catches have been taken in the NEAFC area (banana area) in 2021-2022. Fish were less fat in 2022 compared to the years before with an average of $18.8 \%$ of fat compared to 19.1-22\% in recent years.

The CPUE of Blue whiting shows a relatively stable pattern with a small decline from 2021 to 2022. The fishery had a relatively wide spectrum of fish sizes being landed while the distribution of the fishery was more focused on areas 6.a and 7.c. 7.k. and 7.b and not so much on the NEAFC area or area 2.a. The fish was relatively heavy at around 137 grams.

The CPUE of Western horse mackerel shows a declining trend since 2018 and due to bycatch quota only in 2023 the CPUE drops even further. The size range of fish caught was rather tight in 2022 compared to preceding years and the fishery caught most of the horse mackerel further south compared to previoius years in which more substantial catches were also made in area 6.a. The fish had lower fatcontent compared to 2020-2021 but were not that dissimilar from 2018-2019. Length distribution and weight was comparable to the preceding years.

The CPUE of North Sea horse mackerel shows, similar to the Western horse mackerel CPUE, a declining trend since 2018. The fishery had larger catches in 4.c. compared to previous years and mean size of the fish was lower compared to other years. The averge weight however was quite a bit larger than the previous years which may indicate that fish were fatter than in previous years. There were too few fat measurements however to support any statement on this.

Given the low number of hauls of Atlanto-Scandian herring, we did not perform CPUE analyses for this fishery. The catch distribution of this fishery varies substantially from year to year with catches
in 2022 all located in the NEAFC area while in previous years also closer to the Norwegian coast catches had been made.

## 5 Acknowledgements

The skippers, officers and the quality managers of many of the PFA vessels are putting in a lot of effort to make the PFA the self-sampling work. Without their efforts, there would be no self-sampling.

## 6 References and publications

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Rousseau, Y., R. A. Watson, J. L. Blanchard and E. A. Fulton (2019). "Evolution of global marine fishing fleets and the response of fished resources." Proceedings of the National Academy of Sciences 116(25): 12238-12243

## 7 More information

Please contact Niels Hintzen (nhintzen@pelagicfish.eu) if you would have any questions on the PFA self-sampling program or the specific results presented here.

## 8 Northeast Atlantic mackerel: detailed tables

Northeast Atlantic mackerel Sampling overview

| species | year | quarter | area | division | catch | sampleweight | nsamples | count | catchnumber |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2022 | 1 | 27 | 27.4.a | 11982 | 1209 | 82 | 3659 | 24087 |
| mac | 2022 | 1 | 27 | 27.6.a | 9707 | 502 | 42 | 1570 | 15056 |
| mac | 2022 | 1 | 27 | 27.7.b | 4535 | 443 | 33 | 1447 | 10193 |
| mac | 2022 | 1 | 27 | 27.7.j | 363 | 42 | 16 | 138 | 416 |
| mac | 2022 | 2 | 27 | 27.4.a | 23 | NA | NA | NA | NA |
| mac | 2022 | 2 | 27 | 27.6.a | 146 | 12 | 8 | 32 | 158 |
| mac | 2022 | 3 | 27 | 27.2.a | 120 | NA | NA | NA | NA |
| mac | 2022 | 3 | 27 | 27.4.a | 3721 | 181 | 32 | 584 | 5562 |
| mac | 2022 | 3 | 27 | 27.7.b | 3 | 5 | 1 | 12 | 10 |
| mac | 2022 | 3 | 27 | 27.7.j | 3 | 4 | 4 | 14 | 103 |
| mac | 2022 | 4 | 27 | 27.2.a | 0 | NA | NA | NA | NA |
| mac | 2022 | 4 | 27 | 27.4.a | 32238 | 1739 | 195 | 7891 | 59952 |
| mac | 2022 | 4 | 27 | 27.7.j | 0 | NA | NA | NA | NA |
| mac | 2023 | 1 | 27 | 27.4.a | 9331 | 709 | 41 | 2270 | 12373 |
| mac | 2023 | 1 | 27 | 27.6.a | 16150 | 999 | 68 | 3168 | 18582 |
| mac | 2023 | 1 | 27 | 27.7.b | 13 | NA | NA | NA | NA |
| mac | 2023 | 2 | 27 | 27.4.a | 9 | 0 | 1 | 57 | 142 |
| mac | 2023 | 2 | 27 | 27.6.a | 767 | 94 | 38 | 256 | 1533 |
| mac | 2023 | 2 | 27 | 27.7.j | 8 | 0 | 1 | 1 | 18 |

Northeast Atlantic mackerel Length frequencies 2022

| species | stockname | stockcode | area | year | quarter division | length | count | catchnumber | prop |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 26 | 10 | 36464 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2022 | 1 27.4.a | 27 | 54 | 346705 | 0.0144 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | $127.4 . a$ | 28 | 49 | 340570 | 0.0141 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2022 | 1 27.4.a | 29 | 64 | 412899 | 0.0171 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 127.4.a | 30 | 68 | 412400 | 0.0171 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . \mathrm{a}$ | 31 | 82 | 489344 | 0.0203 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . \mathrm{a}$ | 32 | 94 | 569007 | 0.0236 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 33 | 109 | 526606 | 0.0219 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | $127.4 . \mathrm{a}$ | 34 | 181 | 1047462 | 0.0435 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | $127.4 . a$ | 35 | 269 | 1703142 | 0.0707 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2022 | $127.4 . a$ | 36 | 450 | 2723107 | 0.1131 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 37 | 648 | 4162473 | 0.1728 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . \mathrm{a}$ | 38 | 689 | 4592406 | 0.1907 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . \mathrm{a}$ | 39 | 508 | 3774323 | 0.1567 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 40 | 266 | 1885261 | 0.0783 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 41 | 92 | 756625 | 0.0314 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | 1 27.4.a | 42 | 21 | 243746 | 0.0101 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 43 |  | 7439 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2022 | $127.4 . a$ | 44 |  | 57399 | 0.0024 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 17 |  | 4150 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 18 |  | 4150 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 19 |  | 4150 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 23 |  | 15424 | 0.0010 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | 1 27.6.a | 24 |  | 30313 | 0.0020 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | 127.6 .4 | 25 |  | 34891 | 0.0023 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | 127.6 .9 | 26 | 12 | 85246 | 0.0057 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2022 | 127.6 .9 | 27 | 14 | 137644 | 0.0091 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 28 | 24 | 273784 | 0.0182 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . \mathrm{a}$ | 29 | 22 | 150308 | 0.0100 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 30 | 45 | 398239 | 0.0265 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | 1 27.6.a | 31 | 64 | 554722 | 0.0368 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 32 | 84 | 782862 | 0.0520 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 33 | 120 | 1156090 | 0.0768 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2022 | $127.6 . a$ | 34 | 95 | 882994 | 0.0586 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | $127.6 . a$ | 35 | 115 | 1091725 | 0.0725 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2022 | $127.6 . a$ | 36 | 105 | 1096835 | 0.0728 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 37 | 209 | 2050525 | 0.1362 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 38 | 274 | 2565070 | 0.1704 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | 1 27.6.a | 39 | 214 | 2171710 | 0.1442 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2022 | 127.6 .8 | 40 | 117 | 1193579 | 0.0793 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | $127.6 . a$ | 41 | 29 | 282289 | 0.0187 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | $127.6 . a$ | 42 |  | 89517 | 0.0059 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.7 . \mathrm{b}$ | 20 |  | 34527 | 0.0034 |



|  |  |  |
| :---: | :---: | :---: |
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| Nor | Atlantic mackerel |  |
| ortheast | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northee | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northe | Atlantic mackerel |  |
| Nort | Atlantic mack |  |
| Northeast A | At1 |  |
| Northeast $A$ |  |  |
| No |  |  |
| Northeast $A$ | Atlantic mackerel |  |
| Northeast $A$ | Atlantic mackerel |  |
| Northeast A | Atlantic mackerel |  |
| Northeast A | lan |  |
| Northea | Atlantic mackere |  |
| Northeas | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| rtheas | Atlantic mackerel |  |
| rthea | Atlantic mackerel |  |
| Northe | Atlantic mackerel |  |
| Nor | Atlantic mackerel |  |
| No | Atlantic mack |  |
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| Nor | Atlantic mackerel |  |
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| Northeast | clantic mack |  |
| Northeas | Atlantic mack |  |
| Northeas | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
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| Northeast | lantic mack | nac. |
| Northeast | , |  |
| ortheast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast A | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Nort | Atlantic macker |  |
| Nort | Atlantic macker |  |
| Nort | Atlantic macke |  |
| Northeast | Atlantic macke |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel | ac. |
| Nort | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Nort | Atlantic mackerel |  |
| Northee | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast A | Atlantic mackerel |  |
| Northeast $A$ | Atlantic mackere |  |
| Northeast $A$ | Atlantic mackere |  |
| Northeast A | Atlantic mackerel |  |
| Northeast A | Atlantic mackere |  |
| Northeast A | Atlantic mackere |  |
| North | tlan |  |
| Nort | lantio macker |  |
|  |  |  |




|  | 57545 | 0.005 |
| :---: | :---: | :---: |
| 7 | 80563 | 0.0079 |
| 7 | 80563 | 0.0079 |
| 4 | 46036 | 0.0045 |
| 5 | 57545 | 0.0056 |
| 4 | 46036 | 0.0045 |
| 4 | 35245 | 0.0035 |
| 3 | 31860 | 0.0031 |
| 7 | 84735 | 0.008 |
| 25 | 216377 | 0.0212 |
| 24 | 196195 | 0.019 |
| 31 | 296552 | 0.0291 |
| 45 | 315556 | 0.0310 |
| 72 | 507485 | 0.0498 |
| 162 | 1101891 | 0.1081 |
| 203 | 1318078 | 0.1293 |
| 201 | 1323574 | 0.1298 |
| 270 | 1866023 | 0.1831 |
| 211 | 1395078 | 0.1369 |
| 110 | 751821 | 0.0738 |
| 34 | 283483 | 0.0278 |
| 6 | 46318 | 0.0045 |
| 3 | 9300 | 0.0009 |
|  | 10778 | 0.0011 |
| 2 | 6604 | 0.0159 |
|  | 3302 | 0.0079 |
| 2 | 10388 | 0.0250 |
| 2 | 6163 | 0.0148 |
|  | 34010 | 0.0817 |
| 15 | 55985 | 0.1345 |
| 10 | 47240 | 0.1135 |
| 10 | 44617 | 0.1072 |
| 13 | 48956 | 0.1176 |
| 15 | 44410 | 0.1067 |
| 16 | 30610 | 0.0736 |
| 17 | 21822 | 0.0524 |
| 11 | 26273 | 0.0631 |
| 8 | 18422 | 0.0443 |
|  | 12616 | 0.0303 |
| 3 | 4754 | 0.011 |
|  | 20848 | 0.1319 |
|  | 31786 | 0.2011 |
|  | 28235 | 0.1786 |
|  | 27219 | 0.1722 |
|  | 28043 | 0.17 |
|  | 14125 | 0.0894 |
|  | 2102 | 0.0133 |
|  | 2102 | 0.0133 |
|  | 3602 | 0.0228 |
| 5 | 18037 | 0.0032 |
| 15 | 53765 | 0.0097 |
| 10 | 36074 | 0.0065 |
| 20 | 72148 | 0.0130 |
| 15 | 132662 | 0.0239 |
| 89 | 985972 | 0.1773 |
| 117 | 1255600 | 0.2257 |
| 97 | 977413 | 0.1757 |
| 66 | 687811 | 0.1237 |
| 44 | 386556 | 0.0695 |
| 23 | 210254 | 0.0378 |
| 25 | 218654 | 0.0393 |
| 29 | 302428 | 0.054 |
| 22 | 180212 | 0.032 |
| 6 | 31873 | 0.0057 |
| 1 | 12791 | 0.0023 |
| 1 | 873 | 0.0833 |
| 5 | 4369 | 0.4168 |
| 1 | 873 | 0.0833 |
| 1 | 873 | 0.0833 |
| 2 | 1747 | 0.1667 |
| 1 | 873 | 0.0833 |
|  | 873 | 0.0833 |
|  | 4751 | 0.0457 |
|  | 17221 | 0.1656 |
|  | 12469 | 0.1199 |
|  | 9293 | 0.0894 |
| 3 | 14255 | 0.1371 |
| 2 | 9503 | 0.0914 |
|  | 13494 | 0.12 |


| mac | Northeast | Atlantic mackerel | mac. 27. nea | 27 | 2022 | $327.7 . j$ | 39 |  | 4751 | 0.0457 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 24 | 1 | 1806 | 0.0000 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 25 | 1 | 16414 | 0.0003 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 26 | 7 | 41869 | 0.0007 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 27 | 18 | 82129 | 0.0014 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 28 | 29 | 173550 | 0.0029 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 29 | 53 | 361704 | 0.0060 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 30 | 207 | 1585267 | 0.0264 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 31 | 625 | 4705556 | 0.0785 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 32 | 849 | 6064281 | 0.1012 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 33 | 703 | 4940919 | 0.0824 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 34 | 583 | 4329125 | 0.0722 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 35 | 666 | 5068623 | 0.0845 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 36 | 814 | 7045002 | 0.1175 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 37 | 977 | 7928056 | 0.1322 |
| mac | Northeast | Atlantic mackerel | mac. 27 .nea | 27 | 2022 | $427.4 . a$ | 38 | 1030 | 7884558 | 0.1315 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 39 | 772 | 5641257 | 0.0941 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 40 | 398 | 2789539 | 0.0465 |
| mac | Northeast | Atlantic mackerel | mac. 27. nea | 27 | 2022 | 4 27.4.a | 41 | 133 | 1066549 | 0.0178 |
| mac | Northeast | Atlantic mackerel | mac. 27. nea | 27 | 2022 | 4 27.4.a | 42 | 19 | 185216 | 0.0031 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 43 |  | 36034 | 0.0006 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .8 | 44 | 1 | 3324 | 0.0001 |
| mac | Northeast | Atlantic mackerel | mac. 27 .nea | 27 | 2022 | 4 27.4.a | 45 |  | 1929 | 0.0000 |

Northeast Atlantic mackerel Length frequencies 2023

| species | stockname | stockcode | area | year | quarter division | length | count | catchnumber | prop |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 20 |  | 7095 | 0.0006 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 21 | 8 | 9143 | 0.0007 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 22 | 6 | 6710 | 0.0005 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 23 | 3 | 3606 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 24 |  | 14339 | 0.0012 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 25 | 19 | 38546 | 0.0031 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 26 | 23 | 46971 | 0.0038 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | 1 27.4.a | 27 | 34 | 76906 | 0.0062 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | 1 27.4.a | 28 | 43 | 164557 | 0.0133 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 29 | 66 | 340411 | 0.0275 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 30 | 141 | 824473 | 0.0666 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 31 | 157 | 854314 | 0.0690 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 32 | 142 | 750761 | 0.0607 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 33 | 148 | 771996 | 0.0624 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | 1 27.4.a | 34 | 151 | 834645 | 0.0675 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 35 | 203 | 1251271 | 0.1011 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 36 | 284 | 1741581 | 0.1407 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 37 | 344 | 1974576 | 0.1596 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 38 | 249 | 1460466 | 0.1180 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 39 | 142 | 742746 | 0.0600 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | 127.4.a | 40 | 63 | 348290 | 0.0281 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.4 . a$ | 41 | 15 | 89649 | 0.0072 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.4 . a$ | 42 | 2 | 18736 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.4 . a$ | 43 |  | 2073 | 0.0002 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.6 .9 | 16 | 0 | 985 | 0.0001 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 17 |  | 7268 | 0.0004 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 18 |  | 27792 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 19 |  | 19368 | 0.0010 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 20 | 9 | 47732 | 0.0026 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 21 | 3 | 17037 | 0.0009 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.6 . a$ | 22 | 0 | 1663 | 0.0001 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 25 | 0 | 3719 | 0.0002 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 26 |  | 27399 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 27 | , | 46455 | 0.0025 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 28 | 28 | 146393 | 0.0079 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.6 . a$ | 29 | 95 | 525911 | 0.0283 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.6 . a$ | 30 | 132 | 733950 | 0.0395 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.6 . a$ | 31 | 192 | 1070533 | 0.0576 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | 127.6 .9 | 32 | 212 | 1274502 | 0.0686 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 33 | 203 | 1197143 | 0.0644 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 34 | 232 | 1343748 | 0.0723 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 35 | 299 | 1643786 | 0.0885 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 36 | 424 | 2467270 | 0.1328 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.6 . \mathrm{a}$ | 37 | 497 | 2878717 | 0.1549 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 38 | 395 | 2400487 | 0.1292 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.6 . a$ | 39 | 245 | 1535283 | 0.0826 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.6 . a$ | 40 | 123 | 795265 | 0.0428 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 41 | 39 | 271188 | 0.0146 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2023 | $127.6 . \mathrm{a}$ | 42 |  | 93608 | 0.0050 |


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| $\begin{aligned} & \ddot{9} \\ & \stackrel{1}{\dot{N}} \\ & \underset{\sim}{1} \end{aligned}$ | デ む̇ৰ 亡্ৰ <br>  |
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## 9 Western horse mackerel: detailed tables

Western horse mackerel Sampling overview

| species | year | quarter | area | division | catch | sampleweight | nsamples | count | catchnumber |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2022 | 1 | 27 | 27.4.a | 11982 | 1209 | 82 | 3659 | 24087 |
| mac | 2022 | 1 | 27 | 27.6.a | 9707 | 502 | 42 | 1570 | 15056 |
| mac | 2022 | 1 | 27 | 27.7.b | 4535 | 443 | 33 | 1447 | 10193 |
| mac | 2022 | 1 | 27 | 27.7.j | 363 | 42 | 16 | 138 | 416 |
| mac | 2022 | 2 | 27 | 27.4.a | 23 | NA | NA | NA | NA |
| mac | 2022 | 2 | 27 | 27.6.a | 146 | 12 | 8 | 32 | 158 |
| mac | 2022 | 3 | 27 | 27.2.a | 120 | NA | NA | NA | NA |
| mac | 2022 | 3 | 27 | 27.4.a | 3721 | 181 | 32 | 584 | 5562 |
| mac | 2022 | 3 | 27 | 27.7.b | 3 | 5 | 1 | 12 | 10 |
| mac | 2022 | 3 | 27 | 27.7.j | 3 | 4 | 4 | 14 | 103 |
| mac | 2022 | 4 | 27 | 27.2.a | 0 | NA | NA | NA | NA |
| mac | 2022 | 4 | 27 | 27.4.a | 32238 | 1739 | 195 | 7891 | 59952 |
| mac | 2022 | 4 | 27 | 27.7.j | 0 | NA | NA | NA | NA |
| mac | 2023 | 1 | 27 | 27.4.a | 9331 | 709 | 41 | 2270 | 12373 |
| mac | 2023 | 1 | 27 | 27.6.a | 16150 | 999 | 68 | 3168 | 18582 |
| mac | 2023 | 1 | 27 | 27.7.b | 13 | NA | NA | NA | NA |
| mac | 2023 | 2 | 27 | 27.4.a | 9 | 0 | 1 | 57 | 142 |
| mac | 2023 | 2 | 27 | 27.6.a | 767 | 94 | 38 | 256 | 1533 |
| mac | 2023 | 2 | 27 | 27.7.j | 8 | 0 | 1 | 1 | 18 |

Western horse mackerel Length frequencies 2022

| species | stockname | stockcode | area | year | quarter division | length | count | catchnumber | prop |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 26 | 10 | 36464 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 127.4.a | 27 | 54 | 346705 | 0.0144 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 28 | 49 | 340570 | 0.0141 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 127.4.a | 29 | 64 | 412899 | 0.0171 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 30 | 68 | 412400 | 0.0171 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 31 | 82 | 489344 | 0.0203 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 32 | 94 | 569007 | 0.0236 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . \mathrm{a}$ | 33 | 109 | 526606 | 0.0219 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 34 | 181 | 1047462 | 0.0435 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2022 | 1 27.4.a | 35 | 269 | 1703142 | 0.0707 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 36 | 450 | 2723107 | 0.1131 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 37 | 648 | 4162473 | 0.1728 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 38 | 689 | 4592406 | 0.1907 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 127.4.a | 39 | 508 | 3774323 | 0.1567 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 40 | 266 | 1885261 | 0.0783 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 41 | 92 | 756625 | 0.0314 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 42 | 21 | 243746 | 0.0101 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 43 |  | 7439 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 127.4.a | 44 | 3 | 57399 | 0.0024 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 17 |  | 4150 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 18 |  | 4150 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 19 |  | 4150 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 23 |  | 15424 | 0.0010 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 24 |  | 30313 | 0.0020 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 25 | 8 | 34891 | 0.0023 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 26 | 12 | 85246 | 0.0057 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 127.6 .8 | 27 | 14 | 137644 | 0.0091 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 28 | 24 | 273784 | 0.0182 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 29 | 22 | 150308 | 0.0100 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 30 | 45 | 398239 | 0.0265 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 31 | 64 | 554722 | 0.0368 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 32 | 84 | 782862 | 0.0520 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 33 | 120 | 1156090 | 0.0768 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 34 | 95 | 882994 | 0.0586 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 35 | 115 | 1091725 | 0.0725 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . \mathrm{a}$ | 36 | 105 | 1096835 | 0.0728 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . \mathrm{a}$ | 37 | 209 | 2050525 | 0.1362 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 38 | 274 | 2565070 | 0.1704 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 39 | 214 | 2171710 | 0.1442 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 40 | 117 | 1193579 | 0.0793 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 41 | 29 | 282289 | 0.0187 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 42 |  | 89517 | 0.0059 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.7 . \mathrm{b}$ | 20 | 3 | 34527 | 0.0034 |



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| :---: | :---: | :---: |
|  |  |  |
| Nor | Atlantic mackerel |  |
| ortheast | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northee | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northe | Atlantic mackerel |  |
| Nort | Atlantic mack |  |
| Northeast A | At1 |  |
| Northeast $A$ |  |  |
| No |  |  |
| Northeast $A$ | Atlantic mackerel |  |
| Northeast $A$ | Atlantic mackerel |  |
| Northeast A | Atlantic mackerel |  |
| Northeast A | lan |  |
| Northea | Atlantic mackere |  |
| Northeas | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| rtheas | Atlantic mackerel |  |
| rthea | Atlantic mackerel |  |
| Northe | Atlantic mackerel |  |
| Nor | Atlantic mackerel |  |
| No | Atlantic mack |  |
| Northeast A | Atlan |  |
| Nor | Atlantic mackerel |  |
| Nor | Atlantic mackerel |  |
| Nor | Atlantic mackerel |  |
| Nort | ack |  |
| Northeast | clantic mack |  |
| Northeas | Atlantic mack |  |
| Northeas | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
|  |  |  |
| Northea | Atlantic mack |  |
| Nort | At1 |  |
| Northeast A | Atlantic mackerel |  |
| Northeast $A$ | Atlantic mackerel |  |
| or | Atlantic mackerel |  |
| Northeast A | la |  |
| Northeast | lantic mack | nac. |
| Northeast | , |  |
| ortheast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast A | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Nort | Atlantic macker |  |
| Nort | Atlantic macker |  |
| Nort | Atlantic macke |  |
| Northeast | Atlantic macke |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel | ac. |
| Nort | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Nort | Atlantic mackerel |  |
| Northee | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast A | Atlantic mackerel |  |
| Northeast $A$ | Atlantic mackere |  |
| Northeast $A$ | Atlantic mackere |  |
| Northeast A | Atlantic mackerel |  |
| Northeast A | Atlantic mackere |  |
| Northeast A | Atlantic mackere |  |
| North | tlan |  |
| Nort | lantio macker |  |
|  |  |  |




|  | 57545 | 0.0056 |
| :---: | :---: | :---: |
| 7 | 80563 | 0.0079 |
| 7 | 80563 | 0.0079 |
| 4 | 46036 | 0.0045 |
| 5 | 57545 | 0.0056 |
| 4 | 46036 | 0.0045 |
| 4 | 35245 | 0.0035 |
| 3 | 31860 | 0.0031 |
| 7 | 84735 | 0.0083 |
| 25 | 216377 | 0.0212 |
| 24 | 196195 | 0.0192 |
| 31 | 296552 | 0.0291 |
| 45 | 315556 | 0.0310 |
| 72 | 507485 | 0.0498 |
| 162 | 1101891 | 0.1081 |
| 203 | 1318078 | 0.1293 |
| 201 | 1323574 | 0.1298 |
| 270 | 1866023 | 0.1831 |
| 211 | 1395078 | 0.1369 |
| 110 | 751821 | 0.0738 |
| 34 | 283483 | 0.0278 |
| 6 | 46318 | 0.0045 |
| 3 | 9300 | 0.0009 |
| 1 | 10778 | 0.0011 |
| 2 | 6604 | 0.0159 |
| 1 | 3302 | 0.0079 |
| 2 | 10388 | 0.0250 |
| 2 | 6163 | 0.0148 |
| 7 | 34010 | 0.0817 |
| 15 | 55985 | 0.1345 |
| 10 | 47240 | 0.1135 |
| 10 | 44617 | 0.1072 |
| 13 | 48956 | 0.1176 |
| 15 | 44410 | 0.1067 |
| 16 | 30610 | 0.0736 |
| 17 | 21822 | 0.0524 |
| 11 | 26273 | 0.0631 |
| 8 | 18422 | 0.0443 |
| 6 | 12616 | 0.0303 |
| 5 | 4754 | 0.0114 |
|  | 20848 | 0.1319 |
| 5 | 31786 | 0.2011 |
| 5 | 28235 | 0.1786 |
| 6 | 27219 | 0.1722 |
| 6 | 28043 | 0.1774 |
| 3 | 14125 | 0.0894 |
|  | 2102 | 0.0133 |
|  | 2102 | 0.0133 |
|  | 3602 | 0.0228 |
| 5 | 18037 | 0.0032 |
| 15 | 53765 | 0.0097 |
| 10 | 36074 | 0.0065 |
| 20 | 72148 | 0.0130 |
| 15 | 132662 | 0.0239 |
| 89 | 985972 | 0.1773 |
| 117 | 1255600 | 0.2257 |
| 97 | 977413 | 0.1757 |
| 66 | 687811 | 0.1237 |
| 44 | 386556 | 0.0695 |
| 23 | 210254 | 0.0378 |
| 25 | 218654 | 0.0393 |
| 29 | 302428 | 0.0544 |
| 22 | 180212 | 0.0324 |
| 6 | 31873 | 0.0057 |
| 1 | 12791 | 0.0023 |
| 1 | 873 | 0.0833 |
| 5 | 4369 | 0.4168 |
| 1 | 873 | 0.0833 |
| 1 | 873 | 0.0833 |
| 2 | 1747 | 0.1667 |
| 1 | 873 | 0.0833 |
| 1 | 873 | 0.0833 |
| 1 | 4751 | 0.0457 |
| 2 | 17221 | 0.1656 |
| 1 | 12469 | 0.1199 |
| 1 | 9293 | 0.0894 |
| 3 | 14255 | 0.1371 |
| 2 | 9503 | 0.0914 |
|  | 13494 | 0.1298 |


| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $327.7 . j$ | 39 |  | 4751 | 0.0457 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 24 | 1 | 1806 | 0.0000 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 25 | 1 | 16414 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 26 | 7 | 41869 | 0.0007 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .9 | 27 | 18 | 82129 | 0.0014 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 28 | 29 | 173550 | 0.0029 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 29 | 53 | 361704 | 0.0060 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 30 | 207 | 1585267 | 0.0264 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 31 | 625 | 4705556 | 0.0785 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 32 | 849 | 6064281 | 0.1012 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 33 | 703 | 4940919 | 0.0824 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .8 | 34 | 583 | 4329125 | 0.0722 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .9 | 35 | 666 | 5068623 | 0.0845 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 36 | 814 | 7045002 | 0.1175 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .9 | 37 | 977 | 7928056 | 0.1322 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 38 | 1030 | 7884558 | 0.1315 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 39 | 772 | 5641257 | 0.0941 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 40 | 398 | 2789539 | 0.0465 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 41 | 133 | 1066549 | 0.0178 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .9 | 42 | 19 | 185216 | 0.0031 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 43 | 4 | 36034 | 0.0006 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .8 | 44 | 1 | 3324 | 0.0001 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 45 |  | 1929 | 0.0000 |

Western horse mackerel Length frequencies 2023

| species | stockname | stockcode | area | year | quarter division | length | count | catchnumber | prop |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 20 |  | 7095 | 0.0006 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 21 |  | 9143 | 0.0007 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 22 |  | 6710 | 0.0005 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 23 |  | 3606 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 24 |  | 14339 | 0.0012 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 25 | 19 | 38546 | 0.0031 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 26 | 23 | 46971 | 0.0038 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 27 | 34 | 76906 | 0.0062 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 28 | 43 | 164557 | 0.0133 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 29 | 66 | 340411 | 0.0275 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 30 | 141 | 824473 | 0.0666 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 31 | 157 | 854314 | 0.0690 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 32 | 142 | 750761 | 0.0607 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 33 | 148 | 771996 | 0.0624 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 34 | 151 | 834645 | 0.0675 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 35 | 203 | 1251271 | 0.1011 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 36 | 284 | 1741581 | 0.1407 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 37 | 344 | 1974576 | 0.1596 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 38 | 249 | 1460466 | 0.1180 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 39 | 142 | 742746 | 0.0600 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 40 | 63 | 348290 | 0.0281 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 41 | 15 | 89649 | 0.0072 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 42 |  | 18736 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 43 |  | 2073 | 0.0002 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.6 .9 | 16 |  | 985 | 0.0001 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . \mathrm{a}$ | 17 |  | 7268 | 0.0004 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 18 |  | 27792 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 19 |  | 19368 | 0.0010 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 20 |  | 47732 | 0.0026 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 21 | 3 | 17037 | 0.0009 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 22 |  | 1663 | 0.0001 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 25 |  | 3719 | 0.0002 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 26 |  | 27399 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 27 |  | 46455 | 0.0025 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 28 | 28 | 146393 | 0.0079 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 29 | 95 | 525911 | 0.0283 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 30 | 132 | 733950 | 0.0395 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 31 | 192 | 1070533 | 0.0576 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.6 .9 | 32 | 212 | 1274502 | 0.0686 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 33 | 203 | 1197143 | 0.0644 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 34 | 232 | 1343748 | 0.0723 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 35 | 299 | 1643786 | 0.0885 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 36 | 424 | 2467270 | 0.1328 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 37 | 497 | 2878717 | 0.1549 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 38 | 395 | 2400487 | 0.1292 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 39 | 245 | 1535283 | 0.0826 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 40 | 123 | 795265 | 0.0428 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 41 | 39 | 271188 | 0.0146 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 42 |  | 93608 | 0.0050 |


| $\begin{aligned} & \text { M} \\ & \stackrel{\rightharpoonup}{0} \\ & \dot{\circ} \end{aligned}$ |  <br>  |
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| $\begin{aligned} & \ddot{9} \\ & \stackrel{1}{\dot{N}} \\ & \underset{\sim}{1} \end{aligned}$ | デ む̇ৰ 亡্ৰ <br>  |
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## 10 North Sea horse mackerel: detailed tables

North Sea horse mackerel Sampling overview

| species | year | quarter | area | division | catch | sampleweight | nsamples | count | catchnumber |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2022 | 1 | 27 | 27.4.a | 11982 | 1209 | 82 | 3659 | 24087 |
| mac | 2022 | 1 | 27 | 27.6.a | 9707 | 502 | 42 | 1570 | 15056 |
| mac | 2022 | 1 | 27 | 27.7.b | 4535 | 443 | 33 | 1447 | 10193 |
| mac | 2022 | 1 | 27 | 27.7.j | 363 | 42 | 16 | 138 | 416 |
| mac | 2022 | 2 | 27 | 27.4.a | 23 | NA | NA | NA | NA |
| mac | 2022 | 2 | 27 | 27.6.a | 146 | 12 | 8 | 32 | 158 |
| mac | 2022 | 3 | 27 | 27.2.a | 120 | NA | NA | NA | NA |
| mac | 2022 | 3 | 27 | 27.4.a | 3721 | 181 | 32 | 584 | 5562 |
| mac | 2022 | 3 | 27 | 27.7.b | 3 | 5 | 1 | 12 | 10 |
| mac | 2022 | 3 | 27 | 27.7.j | 3 | 4 | 4 | 14 | 103 |
| mac | 2022 | 4 | 27 | 27.2.a | 0 | NA | NA | NA | NA |
| mac | 2022 | 4 | 27 | 27.4.a | 32238 | 1739 | 195 | 7891 | 59952 |
| mac | 2022 | 4 | 27 | 27.7.j | 0 | NA | NA | NA | NA |
| mac | 2023 | 1 | 27 | 27.4.a | 9331 | 709 | 41 | 2270 | 12373 |
| mac | 2023 | 1 |  | 27.6.a | 16150 | 999 | 68 | 3168 | 18582 |
| mac | 2023 | 1 | 27 | 27.7.b | 13 | NA | NA | NA | NA |
| mac | 2023 | 2 | 27 | 27.4.a | 9 | 0 | 1 | 57 | 142 |
| mac | 2023 | 2 | 27 | 27.6.a | 767 | 94 | 38 | 256 | 1533 |
| mac | 2023 | 2 | 27 | 27.7.j | 8 | 0 | 1 | 1 | 18 |

North Sea horse mackerel Length frequencies 2022



|  |  |  |
| :---: | :---: | :---: |
|  |  |  |
| Nor | Atlantic mackerel |  |
| ortheast | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northee | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northe | Atlantic mackerel |  |
| Nort | Atlantic mack |  |
| Northeast A | At1 |  |
| Northeast $A$ |  |  |
| No |  |  |
| Northeast $A$ | Atlantic mackerel |  |
| Northeast $A$ | Atlantic mackerel |  |
| Northeast A | Atlantic mackerel |  |
| Northeast A | lan |  |
| Northea | Atlantic mackere |  |
| Northeas | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| rtheas | Atlantic mackerel |  |
| rthea | Atlantic mackerel |  |
| Northe | Atlantic mackerel |  |
| Nor | Atlantic mackerel |  |
| No | Atlantic mack |  |
| Northeast A | Atlan |  |
| Nor | Atlantic mackerel |  |
| Nor | Atlantic mackerel |  |
| Nor | Atlantic mackerel |  |
| Nort | ack |  |
| Northeast | clantic mack |  |
| Northeas | Atlantic mack |  |
| Northeas | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
|  |  |  |
| Northea | Atlantic mack |  |
| Nort | At1 |  |
| Northeast A | Atlantic mackerel |  |
| Northeast $A$ | Atlantic mackerel |  |
| or | Atlantic mackerel |  |
| Northeast A | la |  |
| Northeast | lantic mack | nac. |
| Northeast | , |  |
| ortheast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast A | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Nort | Atlantic macker |  |
| Nort | Atlantic macker |  |
| Nort | Atlantic macke |  |
| Northeast | Atlantic macke |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel | ac. |
| Nort | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Nort | Atlantic mackerel |  |
| Northee | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast A | Atlantic mackerel |  |
| Northeast $A$ | Atlantic mackere |  |
| Northeast $A$ | Atlantic mackere |  |
| Northeast A | Atlantic mackerel |  |
| Northeast A | Atlantic mackere |  |
| Northeast A | Atlantic mackere |  |
| North | tlan |  |
| Nort | lantio macker |  |
|  |  |  |




| 5 | 57545 | 0.0056 |
| :---: | :---: | :---: |
| 7 | 80563 | 0.0079 |
| 7 | 80563 | 0.0079 |
| 4 | 46036 | 0.0045 |
| 5 | 57545 | 0.0056 |
| 4 | 46036 | 0.0045 |
| 4 | 35245 | 0.0035 |
| 3 | 31860 | 0.0031 |
| 7 | 84735 | 0.0083 |
| 25 | 216377 | 0.0212 |
| 24 | 196195 | 0.0192 |
| 31 | 296552 | 0.0291 |
| 45 | 315556 | 0.0310 |
| 72 | 507485 | 0.0498 |
| 162 | 1101891 | 0.1081 |
| 203 | 1318078 | 0.1293 |
| 201 | 1323574 | 0.1298 |
| 270 | 1866023 | 0.1831 |
| 211 | 1395078 | 0.1369 |
| 110 | 751821 | 0.0738 |
| 34 | 283483 | 0.0278 |
| 6 | 46318 | 0.0045 |
| 3 | 9300 | 0.0009 |
| 1 | 10778 | 0.0011 |
| 2 | 6604 | 0.0159 |
| 1 | 3302 | 0.0079 |
| 2 | 10388 | 0.0250 |
| 2 | 6163 | 0.0148 |
| 7 | 34010 | 0.0817 |
| 15 | 55985 | 0.1345 |
| 10 | 47240 | 0.1135 |
| 10 | 44617 | 0.1072 |
| 13 | 48956 | 0.1176 |
| 15 | 44410 | 0.1067 |
| 16 | 30610 | 0.0736 |
| 17 | 21822 | 0.0524 |
| 11 | 26273 | 0.0631 |
| 8 | 18422 | ${ }^{0.0443}$ |
| 6 | 12616 | 0.0303 |
| 3 | 4754 | 0.0114 |
| 4 | 20848 | 0.1319 |
| 5 | 31786 | 0.2011 |
| 5 | 28235 | 0.1786 |
| 6 | 27219 | 0.1722 |
| 6 | 28043 | 0.1774 |
| 3 | 14125 | 0.0894 |
| 1 | 2102 | 0.0133 |
| 1 | 2102 | 0.0133 |
| 1 | 3602 | 0.0228 |
| 5 | 18037 | 0.0032 |
| 15 | 53765 | 0.0097 |
| 10 | 36074 | 0.0065 |
| 20 | 72148 | 0.0130 |
| 15 | 132662 | 0.0239 |
| 89 | 985972 | 0.1773 |
| 117 | 1255600 | 0.2257 |
| 97 | 977413 | 0.1757 |
| 66 | 687811 | 0.1237 |
| 44 | 386556 | 0.0695 |
| 23 | 210254 | 0.0378 |
| 25 | 218654 | 0.0393 |
| 29 | 302428 | 0.0544 |
| 22 | 180212 | 0.0324 |
| 6 | 31873 | 0.0057 |
| 1 | 12791 | 0.0023 |
| 1 | 873 | 0.0833 |
| 5 | 4369 | 0.4168 |
| 1 | 873 | 0.0833 |
| 1 | 873 | 0.0833 |
| 2 | 1747 | 0.1667 |
| 1 | 873 | 0.0833 |
| 1 | 873 | 0.0833 |
| 1 | 4751 | 0.0457 |
| 2 | 17221 | 0.1656 |
| 1 | 12469 | 0.1199 |
| 1 | 9293 | 0.0894 |
| 3 | 14255 | 0.1371 |
| 2 | 9503 | 0.0914 |
|  | 13494 | 0.1298 |


| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $327.7 . j$ | 39 |  | 4751 | 0.0457 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 24 | 1 | 1806 | 0.0000 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 25 | 1 | 16414 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 26 | 7 | 41869 | 0.0007 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .9 | 27 | 18 | 82129 | 0.0014 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 28 | 29 | 173550 | 0.0029 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 29 | 53 | 361704 | 0.0060 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 30 | 207 | 1585267 | 0.0264 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 31 | 625 | 4705556 | 0.0785 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 32 | 849 | 6064281 | 0.1012 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 33 | 703 | 4940919 | 0.0824 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .8 | 34 | 583 | 4329125 | 0.0722 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .9 | 35 | 666 | 5068623 | 0.0845 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 36 | 814 | 7045002 | 0.1175 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .9 | 37 | 977 | 7928056 | 0.1322 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 38 | 1030 | 7884558 | 0.1315 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 39 | 772 | 5641257 | 0.0941 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 40 | 398 | 2789539 | 0.0465 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 41 | 133 | 1066549 | 0.0178 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .9 | 42 | 19 | 185216 | 0.0031 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 43 | 4 | 36034 | 0.0006 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .8 | 44 | 1 | 3324 | 0.0001 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 45 |  | 1929 | 0.0000 |

North Sea horse mackerel Length frequencies 2023

| species | stockname | stockcode | area | year | quarter division | length | count | catchnumber | prop |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 20 |  | 7095 | 0.0006 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 21 | 8 | 9143 | 0.0007 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4 .8 | 22 | 6 | 6710 | 0.0005 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 23 | 3 | 3606 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4 .8 | 24 | 7 | 14339 | 0.0012 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 25 | 19 | 38546 | 0.0031 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 26 | 23 | 46971 | 0.0038 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 27 | 34 | 76906 | 0.0062 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 28 | 43 | 164557 | 0.0133 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 29 | 66 | 340411 | 0.0275 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 30 | 141 | 824473 | 0.0666 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 31 | 157 | 854314 | 0.0690 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4 .8 | 32 | 142 | 750761 | 0.0607 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 33 | 148 | 771996 | 0.0624 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 34 | 151 | 834645 | 0.0675 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 35 | 203 | 1251271 | 0.1011 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 36 | 284 | 1741581 | 0.1407 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 37 | 344 | 1974576 | 0.1596 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 38 | 249 | 1460466 | 0.1180 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 39 | 142 | 742746 | 0.0600 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 40 | 63 | 348290 | 0.0281 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 41 | 15 | 89649 | 0.0072 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 42 | 2 | 18736 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 43 |  | 2073 | 0.0002 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 16 | 0 | 985 | 0.0001 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 17 | 1 | 7268 | 0.0004 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 18 | 5 | 27792 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 19 | 3 | 19368 | 0.0010 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 20 | 9 | 47732 | 0.0026 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 21 | 3 | 17037 | 0.0009 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 22 | 0 | 1663 | 0.0001 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 25 |  | 3719 | 0.0002 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 26 |  | 27399 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.6 .8 | 27 | 7 | 46455 | 0.0025 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 28 | 28 | 146393 | 0.0079 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 29 | 95 | 525911 | 0.0283 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 30 | 132 | 733950 | 0.0395 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 31 | 192 | 1070533 | 0.0576 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 32 | 212 | 1274502 | 0.0686 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 33 | 203 | 1197143 | 0.0644 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 34 | 232 | 1343748 | 0.0723 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 35 | 299 | 1643786 | 0.0885 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 36 | 424 | 2467270 | 0.1328 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . \mathrm{a}$ | 37 | 497 | 2878717 | 0.1549 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . \mathrm{a}$ | 38 | 395 | 2400487 | 0.1292 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 39 | 245 | 1535283 | 0.0826 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.6 .9 | 40 | 123 | 795265 | 0.0428 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 41 | 39 | 271188 | 0.0146 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 42 | 8 | 93608 | 0.0050 |


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## 11 Blue whiting: detailed tables

## Blue whiting Sampling overview



Blue whiting Length frequencies 2022



|  |  |  |
| :---: | :---: | :---: |
|  |  |  |
| Nor | Atlantic mackerel |  |
| ortheast | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northee | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northe | Atlantic mackerel |  |
| Nort | Atlantic mack |  |
| Northeast A | At1 |  |
| Northeast $A$ |  |  |
| No |  |  |
| Northeast $A$ | Atlantic mackerel |  |
| Northeast $A$ | Atlantic mackerel |  |
| Northeast A | Atlantic mackerel |  |
| Northeast A | lan |  |
| Northea | Atlantic mackere |  |
| Northeas | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| rtheas | Atlantic mackerel |  |
| rthea | Atlantic mackerel |  |
| Northe | Atlantic mackerel |  |
| Nor | Atlantic mackerel |  |
| No | Atlantic mack |  |
| Northeast A | Atlan |  |
| Nor | Atlantic mackerel |  |
| Nor | Atlantic mackerel |  |
| Nor | Atlantic mackerel |  |
| Nort | ack |  |
| Northeast | clantic mack |  |
| Northeas | Atlantic mack |  |
| Northeas | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
|  |  |  |
| Northea | Atlantic mack |  |
| Nort | At1 |  |
| Northeast A | Atlantic mackerel |  |
| Northeast $A$ | Atlantic mackerel |  |
| or | Atlantic mackerel |  |
| Northeast A | la |  |
| Northeast | lantic mack | nac. |
| Northeast | , |  |
| ortheast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeas | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast A | Atlantic mackerel |  |
| Northea | Atlantic mackerel |  |
| Nort | Atlantic macker |  |
| Nort | Atlantic macker |  |
| Nort | Atlantic macke |  |
| Northeast | Atlantic macke |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel | ac. |
| Nort | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Nort | Atlantic mackerel |  |
| Northee | Atlantic mackerel |  |
| Northeast | Atlantic mackerel |  |
| Northeast A | Atlantic mackerel |  |
| Northeast $A$ | Atlantic mackere |  |
| Northeast $A$ | Atlantic mackere |  |
| Northeast A | Atlantic mackerel |  |
| Northeast A | Atlantic mackere |  |
| Northeast A | Atlantic mackere |  |
| North | tlan |  |
| Nort | lantio macker |  |
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|  | 57545 | . 00 |
| :---: | :---: | :---: |
| 7 | 80563 | 0.0079 |
| 7 | 80563 | 0.0079 |
| 4 | 46036 | 0.0045 |
| 5 | 57545 | 0.0056 |
| 4 | 46036 | 0.0045 |
| 4 | 35245 | 0.0035 |
| 3 | 31860 | 0.0031 |
| 7 | 84735 | 0.0083 |
| 25 | 216377 | 0.0212 |
| 24 | 196195 | 0.0192 |
| 31 | 296552 | 0.0291 |
| 45 | 315556 | 0.0310 |
| 72 | 507485 | 0.0498 |
| 162 | 1101891 | 0.1081 |
| 203 | 1318078 | 0.1293 |
| 201 | 1323574 | 0.1298 |
| 270 | 1866023 | 0.1831 |
| 211 | 1395078 | 0.1369 |
| 110 | 751821 | 0.0738 |
| 34 | 283483 | 0.0278 |
| 6 | 46318 | 0.0045 |
| 3 | 9300 | 0.0009 |
| 1 | 10778 | 0.0011 |
| 2 | 6604 | 0.0159 |
| 1 | 3302 | 0.0079 |
| 2 | 10388 | 0.0250 |
| 2 | 6163 | 0.0148 |
| 7 | 34010 | 0.0817 |
| 15 | 55985 | 0.1345 |
| 10 | 47240 | 0.1135 |
| 10 | 44617 | 0.1072 |
| 13 | 48956 | 0.1176 |
| 15 | 44410 | 0.1067 |
| 16 | 30610 | 0.0736 |
| 17 | 21822 | 0.0524 |
| 11 | 26273 | 0.0631 |
| 8 | 18422 | ${ }^{0.0443}$ |
| 6 | 12616 | 0.0303 |
| 3 | 4754 | 0.0114 |
|  | 20848 | 0.1319 |
| 5 | 31786 | 0.2011 |
| 5 | 28235 | 0.1786 |
| 6 | 27219 | 0.1722 |
| 6 | 28043 | 0.1774 |
| 3 | 14125 | 0.0894 |
| 1 | 2102 | 0.0133 |
| 1 | 2102 | 0.0133 |
| 1 | 3602 | 0.0228 |
| 5 | 18037 | 0.0032 |
| 15 | 53765 | 0.0097 |
| 10 | 36074 | 0.0065 |
| 20 | 72148 | 0.0130 |
| 15 | 132662 | 0.0239 |
| 89 | 985972 | 0.1773 |
| 117 | 1255600 | 0.2257 |
| 97 | 977413 | 0.1757 |
| 66 | 687811 | 0.1237 |
| 44 | 386556 | 0.0695 |
| 23 | 210254 | 0.0378 |
| 25 | 218654 | 0.0393 |
| 29 | 302428 | 0.0544 |
| 22 | 180212 | 0.0324 |
| 6 | 31873 | 0.0057 |
| 1 | 12791 | 0.0023 |
| 1 | 873 | 0.0833 |
| 5 | 4369 | 0.4168 |
| 1 | 873 | 0.0833 |
|  | 873 | 0.0833 |
| 2 | 1747 | 0.1667 |
| 1 | 873 | 0.0833 |
| 1 | 873 | 0.0833 |
| 1 | 4751 | 0.0457 |
| 2 | 17221 | 0.1656 |
| 1 | 12469 | 0.1199 |
| 1 | 9293 | 0.0894 |
| 3 | 14255 | 0.1371 |
| 2 | 9503 | 0.0914 |
|  | 13494 | 0.1298 |


| mac | Northeast | Atlantic mackerel | mac. 27. nea | 27 | 2022 | $327.7 . j$ | 39 | 1 | 4751 | 0.0457 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 24 | 1 | 1806 | 0.0000 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 25 |  | 16414 | 0.0003 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 26 | 7 | 41869 | 0.0007 |
| mac | Northeast | Atlantic mackerel | mac. 27. nea | 27 | 2022 | 4 27.4.a | 27 | 18 | 82129 | 0.0014 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 28 | 29 | 173550 | 0.0029 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 29 | 53 | 361704 | 0.0060 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 30 | 207 | 1585267 | 0.0264 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 31 | 625 | 4705556 | 0.0785 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 32 | 849 | 6064281 | 0.1012 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 33 | 703 | 4940919 | 0.0824 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 34 | 583 | 4329125 | 0.0722 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 35 | 666 | 5068623 | 0.0845 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 36 | 814 | 7045002 | 0.1175 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 37 | 977 | 7928056 | 0.1322 |
| mac | Northeast | Atlantic mackerel | mac. 27. nea | 27 | 2022 | 4 27.4.a | 38 | 1030 | 7884558 | 0.1315 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 39 | 772 | 5641257 | 0.0941 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 40 | 398 | 2789539 | 0.0465 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 41 | 133 | 1066549 | 0.0178 |
| mac | Northeast | Atlantic mackerel | mac. $27 . \mathrm{nea}$ | 27 | 2022 | 4 27.4.a | 42 | 19 | 185216 | 0.0031 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 43 | 4 | 36034 | 0.0006 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 44 | 1 | 3324 | 0.0001 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 45 |  | 1929 | 0.0000 |

Blue whiting Length frequencies 2023

| species | stockname | stockcode | area | year | quarter division | length | count | catchnumber | prop |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 20 |  | 7095 | 0.0006 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | 1 27.4.a | 21 |  | 9143 | 0.0007 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 22 |  | 6710 | 0.0005 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 23 |  | 3606 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 24 |  | 14339 | 0.0012 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 25 | 19 | 38546 | 0.0031 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 26 | 23 | 46971 | 0.0038 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 27 | 34 | 76906 | 0.0062 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.4 . a$ | 28 | 43 | 164557 | 0.0133 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 29 | 66 | 340411 | 0.0275 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 30 | 141 | 824473 | 0.0666 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 31 | 157 | 854314 | 0.0690 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4 .8 | 32 | 142 | 750761 | 0.0607 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 33 | 148 | 771996 | 0.0624 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 34 | 151 | 834645 | 0.0675 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 35 | 203 | 1251271 | 0.1011 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | 1 27.4.a | 36 | 284 | 1741581 | 0.1407 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 37 | 344 | 1974576 | 0.1596 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 38 | 249 | 1460466 | 0.1180 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 39 | 142 | 742746 | 0.0600 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.4 . a$ | 40 | 63 | 348290 | 0.0281 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 41 | 15 | 89649 | 0.0072 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 42 |  | 18736 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 43 |  | 2073 | 0.0002 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.6 .9 | 16 |  | 985 | 0.0001 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 17 |  | 7268 | 0.0004 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.6 . a$ | 18 |  | 27792 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.6 . \mathrm{a}$ | 19 | 3 | 19368 | 0.0010 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . \mathrm{a}$ | 20 |  | 47732 | 0.0026 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 21 |  | 17037 | 0.0009 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.6 . a$ | 22 |  | 1663 | 0.0001 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | 127.6 .8 | 25 |  | 3719 | 0.0002 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 26 |  | 27399 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 27 |  | 46455 | 0.0025 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 28 | 28 | 146393 | 0.0079 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.6 . a$ | 29 | 95 | 525911 | 0.0283 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 30 | 132 | 733950 | 0.0395 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 31 | 192 | 1070533 | 0.0576 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 32 | 212 | 1274502 | 0.0686 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 33 | 203 | 1197143 | 0.0644 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . \mathrm{a}$ | 34 | 232 | 1343748 | 0.0723 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 35 | 299 | 1643786 | 0.0885 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2023 | $127.6 . a$ | 36 | 424 | 2467270 | 0.1328 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2023 | $127.6 . a$ | 37 | 497 | 2878717 | 0.1549 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | 1 27.6.a | 38 | 395 | 2400487 | 0.1292 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 39 | 245 | 1535283 | 0.0826 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | $127.6 . a$ | 40 | 123 | 795265 | 0.0428 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 41 | 39 | 271188 | 0.0146 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2023 | $127.6 . a$ | 42 | 8 | 93608 | 0.0050 |


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| $\begin{aligned} & \dot{9} \\ & \dot{\sim} \\ & \dot{\sim} \end{aligned}$ |  <br>  <br>  |
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| $\begin{aligned} & \mathscr{g} \\ & \stackrel{y}{c} \\ & \tilde{M} \\ & \dot{U} \\ & \ddot{E} \end{aligned}$ |  えָ זָ <br>  |
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## 12 Atlanto-scandian herring: detailed tables

Atlanto-scandian herring Sampling overview

| species | year | quarter | area | division | catch | sampleweight | nsamples | count | catchnumber |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | 2022 | 1 | 27 | 27.4.a | 11982 | 1209 | 82 | 3659 | 24087 |
| mac | 2022 | 1 | 27 | 27.6.a | 9707 | 502 | 42 | 1570 | 15056 |
| mac | 2022 | 1 | 27 | 27.7.b | 4535 | 443 | 33 | 1447 | 10193 |
| mac | 2022 | 1 | 27 | 27.7.j | 363 | 42 | 16 | 138 | 416 |
| mac | 2022 | 2 | 27 | 27.4.a | 23 | NA | NA | NA | NA |
| mac | 2022 | 2 | 27 | 27.6.a | 146 | 12 | 8 | 32 | 158 |
| mac | 2022 | 3 | 27 | 27.2.a | 120 | NA | NA | NA | NA |
| mac | 2022 | 3 | 27 | 27.4.a | 3721 | 181 | 32 | 584 | 5562 |
| mac | 2022 | 3 | 27 | 27.7.b | 3 | 5 | 1 | 12 | 10 |
| mac | 2022 | 3 |  | 27.7.j | 3 | 4 | 4 | 14 | 103 |
| mac | 2022 | 4 | 27 | 27.2.a | 0 | NA | NA | NA | NA |
| mac | 2022 | 4 | 27 | 27.4.a | 32238 | 1739 | 195 | 7891 | 59952 |
| mac | 2022 | 4 |  | 27.7.j | 0 | NA | NA | NA | NA |
| mac | 2023 | 1 | 27 | 27.4.a | 9331 | 709 | 41 | 2270 | 12373 |
| mac | 2023 | 1 |  | 27.6.a | 16150 | 999 | 68 | 3168 | 18582 |
| mac | 2023 | 1 |  | 27.7.b | 13 | NA | NA | NA | NA |
| mac | 2023 | 2 |  | 27.4.a | 9 | 0 | 1 | 57 | 142 |
| mac | 2023 | 2 |  | 27.6.a | 767 | 94 | 38 | 256 | 1533 |
| mac | 2023 | 2 |  | 27.7.j | 8 | 0 | 1 | 1 | 18 |

Atlanto-scandian herring Length frequencies 2022

| species | stockname | stockcode | area | year | quarter division | length | count | catchnumber | prop |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 26 | 10 | 36464 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 27 | 54 | 346705 | 0.0144 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 28 | 49 | 340570 | 0.0141 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 127.4.a | 29 | 64 | 412899 | 0.0171 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 30 | 68 | 41240 | 0.0171 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . \mathrm{a}$ | 31 | 82 | 489344 | 0.0203 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 32 | 94 | 569007 | 0.0236 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . \mathrm{a}$ | 33 | 109 | 526606 | 0.0219 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 34 | 181 | 1047462 | 0.0435 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 35 | 269 | 1703142 | 0.0707 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | 1 27.4.a | 36 | 450 | 2723107 | 0.1131 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 37 | 648 | 4162473 | 0.1728 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 38 | 689 | 4592406 | 0.1907 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 127.4.a | 39 | 508 | 3774323 | 0.1567 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 40 | 266 | 1885261 | 0.0783 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 41 | 92 | 756625 | 0.0314 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 42 | 21 | 243746 | 0.0101 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.4 . a$ | 43 | 2 | 7439 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.4.a | 44 | 3 | 57399 | 0.0024 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 17 | 1 | 4150 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 18 | 1 | 4150 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 19 | 1 | 4150 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 23 | 4 | 15424 | 0.0010 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 24 | 5 | 30313 | 0.0020 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 25 | 8 | 34891 | 0.0023 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | 1 27.6.a | 26 | 12 | 85246 | 0.0057 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 27 | 14 | 137644 | 0.0091 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 28 | 24 | 273784 | 0.0182 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 29 | 22 | 150308 | 0.0100 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 30 | 45 | 398239 | 0.0265 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 31 | 64 | 554722 | 0.0368 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 32 | 84 | 782862 | 0.0520 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 33 | 120 | 1156090 | 0.0768 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 34 | 95 | 882994 | 0.0586 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 35 | 115 | 1091725 | 0.0725 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 36 | 105 | 1096835 | 0.0728 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 37 | 209 | 2050525 | 0.1362 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 38 | 274 | 2565070 | 0.1704 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.6.a | 39 | 214 | 2171710 | 0.1442 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 40 | 117 | 1193579 | 0.0793 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2022 | $127.6 . a$ | 41 | 29 | 282289 | 0.0187 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | $127.6 . a$ | 42 | 7 | 89517 | 0.0059 |
| ${ }_{\text {mac }}$ | Northeast Atlantic mackerel | mac.27.nea | 27 | 2022 | 1 27.7.b | 20 | 3 | 34527 | 0.0034 |



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| 7 | 80563 | 0.0079 |
| 7 | 80563 | 0.0079 |
| 4 | 46036 | 0.0045 |
| 5 | 57545 | 0.0056 |
| 4 | 46036 | 0.0045 |
| 4 | 35245 | 0.0035 |
| 3 | 31860 | 0.0031 |
| 7 | 84735 | 0.0083 |
| 25 | 216377 | 0.0212 |
| 24 | 196195 | 0.0192 |
| 31 | 296552 | 0.0291 |
| 45 | 315556 | 0.0310 |
| 72 | 507485 | 0.0498 |
| 162 | 1101891 | 0.1081 |
| 203 | 1318078 | 0.1293 |
| 201 | 1323574 | 0.1298 |
| 270 | 1866023 | 0.1831 |
| 211 | 1395078 | 0.1369 |
| 110 | 751821 | 0.0738 |
| 34 | 283483 | 0.0278 |
| 6 | 46318 | 0.0045 |
| 3 | 9300 | 0.0009 |
| 1 | 10778 | 0.0011 |
| 2 | 6604 | 0.0159 |
| 1 | 3302 | 0.0079 |
| 2 | 10388 | 0.0250 |
| 2 | 6163 | 0.0148 |
| 7 | 34010 | 0.0817 |
| 15 | 55985 | 0.1345 |
| 10 | 47240 | 0.1135 |
| 10 | 44617 | 0.1072 |
| 13 | 48956 | 0.1176 |
| 15 | 44410 | 0.1067 |
| 16 | 30610 | 0.0736 |
| 17 | 21822 | 0.0524 |
| 11 | 26273 | 0.0631 |
| 8 | 18422 | 0.0443 |
| 6 | 12616 | 0.0303 |
| 3 | 4754 | 0.0114 |
|  | 20848 | 0.1319 |
| 5 | 31786 | 0.2011 |
| 5 | 28235 | 0.1786 |
| 6 | 27219 | 0.1722 |
|  | 28043 | 0.1774 |
| 3 | 14125 | 0.0894 |
|  | 2102 | 0.0133 |
|  | 2102 | 0.0133 |
|  | 3602 | 0.0228 |
| 5 | 18037 | 0.0032 |
| 15 | 53765 | 0.0097 |
| 10 | 36074 | 0.0065 |
| 20 | 72148 | 0.0130 |
| 15 | 132662 | 0.0239 |
| 89 | 985972 | 0.1773 |
| 117 | 1255600 | 0.2257 |
| 97 | 977413 | 0.1757 |
| 66 | 687811 | 0.1237 |
| 44 | 386556 | 0.0695 |
| 23 | 210254 | 0.0378 |
| 25 | 218654 | 0.0393 |
| 29 | 302428 | 0.0544 |
| 22 | 180212 | 0.0324 |
|  | 31873 | 0.0057 |
|  | 12791 | 0.0023 |
|  | 873 | 0.0833 |
| 5 | 4369 | 0.4168 |
| 1 | 873 | 0.0833 |
|  | 873 | 0.0833 |
| 2 | 1747 | 0.1667 |
| 1 | 873 | 0.0833 |
| 1 | 873 | 0.0833 |
| 1 | 4751 | 0.0457 |
| 2 | 17221 | 0.1656 |
|  | 12469 | 0.1199 |
| 1 | 9293 | 0.0894 |
| 3 | 14255 | 0.1371 |
| 2 | 9503 | 0.0914 |
|  | 13494 | 0.1298 |


| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $327.7 . j$ | 39 |  | 4751 | 0.0457 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 24 |  | 1806 | 0.0000 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 25 |  | 16414 | 0.0003 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 26 | 7 | 41869 | 0.0007 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 27 | 18 | 82129 | 0.0014 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 28 | 29 | 173550 | 0.0029 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 29 | 53 | 361704 | 0.0060 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 30 | 207 | 1585267 | 0.0264 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 31 | 625 | 4705556 | 0.0785 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 32 | 849 | 6064281 | 0.1012 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . \mathrm{a}$ | 33 | 703 | 4940919 | 0.0824 |
| mac | Northeast | Atlantic mackerel | mac. 27 .nea | 27 | 2022 | 427.4 .8 | 34 | 583 | 4329125 | 0.0722 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .9 | 35 | 666 | 5068623 | 0.0845 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 36 | 814 | 7045002 | 0.1175 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 37 | 977 | 7928056 | 0.1322 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 38 | 1030 | 7884558 | 0.1315 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 39 | 772 | 5641257 | 0.0941 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 40 | 398 | 2789539 | 0.0465 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 427.4 .9 | 41 | 133 | 1066549 | 0.0178 |
| mac | Northeast | Atlantic mackerel | mac. 27 .nea | 27 | 2022 | $427.4 . a$ | 42 | 19 | 185216 | 0.0031 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | $427.4 . a$ | 43 |  | 36034 | 0.0006 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 44 | 1 | 3324 | 0.0001 |
| mac | Northeast | Atlantic mackerel | mac.27.nea | 27 | 2022 | 4 27.4.a | 45 |  | 1929 | 0.0000 |

Atlanto-scandian herring Length frequencies 2023

| species | stockname | stockcode | area | year | quarter division | length | count | catchnumber | prop |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 20 |  | 7095 | 0.0006 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 21 | 8 | 9143 | 0.0007 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 22 | 6 | 6710 | 0.0005 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 23 | 3 | 3606 | 0.0003 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 24 | 7 | 14339 | 0.0012 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . \mathrm{a}$ | 25 | 19 | 38546 | 0.0031 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 26 | 23 | 46971 | 0.0038 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 27 | 34 | 76906 | 0.0062 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 28 | 43 | 164557 | 0.0133 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 29 | 66 | 340411 | 0.0275 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 30 | 141 | 824473 | 0.0666 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 31 | 157 | 854314 | 0.0690 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4 .8 | 32 | 142 | 750761 | 0.0607 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 33 | 148 | 771996 | 0.0624 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 34 | 151 | 834645 | 0.0675 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 35 | 203 | 1251271 | 0.1011 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 36 | 284 | 1741581 | 0.1407 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 37 | 344 | 1974576 | 0.1596 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 38 | 249 | 1460466 | 0.1180 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 39 | 142 | 742746 | 0.0600 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.4 . a$ | 40 | 63 | 348290 | 0.0281 |
| mac | Northeast Atlantic mackerel | mac. 27 .nea | 27 | 2023 | 1 27.4.a | 41 | 15 | 89649 | 0.0072 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.4.a | 42 |  | 18736 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.4.a | 43 |  | 2073 | 0.0002 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 16 |  | 985 | 0.0001 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 17 |  | 7268 | 0.0004 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 18 |  | 27792 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 19 | 3 | 19368 | 0.0010 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 20 | 9 | 47732 | 0.0026 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 21 | 3 | 17037 | 0.0009 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 22 | 0 | 1663 | 0.0001 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 25 |  | 3719 | 0.0002 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 26 |  | 27399 | 0.0015 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 127.6 .8 | 27 | 7 | 46455 | 0.0025 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 28 | 28 | 146393 | 0.0079 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 29 | 95 | 525911 | 0.0283 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 30 | 132 | 733950 | 0.0395 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 31 | 192 | 1070533 | 0.0576 |
| mac | Northeast Atlantic mackerel | mac. 27. nea | 27 | 2023 | 1 27.6.a | 32 | 212 | 1274502 | 0.0686 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 33 | 203 | 1197143 | 0.0644 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 34 | 232 | 1343748 | 0.0723 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 35 | 299 | 1643786 | 0.0885 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 36 | 424 | 2467270 | 0.1328 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . \mathrm{a}$ | 37 | 497 | 2878717 | 0.1549 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 38 | 395 | 2400487 | 0.1292 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 39 | 245 | 1535283 | 0.0826 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 40 | 123 | 795265 | 0.0428 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | $127.6 . a$ | 41 | 39 | 271188 | 0.0146 |
| mac | Northeast Atlantic mackerel | mac.27.nea | 27 | 2023 | 1 27.6.a | 42 |  | 93608 | 0.0050 |


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IESNS post-cruise meeting, Teams 13-15/6 2023

## Working Document to

Working Group on International Pelagic Surveys (WGIPS)
22 - 26 January 2023
and
Working Group on Widely Distributed Stocks (WGWIDE)
23-29 August 2023

# INTERNATIONAL ECOSYSTEM SURVEY IN NORDIC SEA (IESNS) <br> in April - May 2023 

Post-cruise meeting on Teams, 13-15 June 2023

Are Salthaug ${ }^{1}$, Erling Kåre Stenevik ${ }^{1}$, Sindre Vatnehol ${ }^{1}$, Age Høines ${ }^{1}$, Justine Diaz ${ }^{1}$, Lea Hellenbrecht ${ }^{1}$, Kjell Arne Mork ${ }^{1}$, Cecilie Thorsen Broms ${ }^{1}$

RV G.O. Sars

Susan Mærsk Lusseau ${ }^{2}$ and Serdar Sakinan ${ }^{5}$
RV Dana

Sigurvin Bjarnason ${ }^{3}$
RV Árni Fridriksson

Sólvá Káradóttir Eliasen ${ }^{4}$, Leon Smith ${ }^{4}$
RV Jákup Sverri

Fabio Campanella ${ }^{6}$, Richard Humphreys ${ }^{6}$, Samantha Barnett ${ }^{6}$, Nicola Hampton ${ }^{6}$, Gary Burt ${ }^{6}$, Matthew Eade ${ }^{6}$
MS Resolute

[^12]In April-June 2023, four research vessels and one hired commercial vessel participated in the International ecosystem survey in the Nordic Seas (IESNS); R/V Dana, Denmark (joint EU survey by Denmark, Germany, Ireland, The Netherlands and Sweden), R/V Jákup Sverri, Faroe Islands, R/V Árni Friðriksson, Iceland, R/V G.O. Sars, Norway and M/S Resolute, United Kingdom (UK). Like in 2022, the Barents Sea was not surveyed by a Russian research vessel. The aim of the survey was to cover the whole distribution area of the Norwegian Spring-spawning herring with the objective of estimating the total abundance of the herring stock, in addition to collect data on plankton and hydrographical conditions in the area. The survey was initiated by the Faroes, Iceland, Norway and Russia in 1995. Since 1997 the EU has also participated (except 2002 and 2003) and from 2004 onwards the survey has been more integrated into an ecosystem survey.

This report represents analyses of data from this International survey in 2023 that are stored in the PGNAPES database and the ICES acoustic database and supported by national survey reports from some survey participants (Dana: Cruise Report R/V Dana Cruise 03/2023. International Ecosystem survey in the Nordic Seas (IESNS) in 2023, Árni Friðriksson: A62023 Cruise Report, Bjarnason, 2023, Jákup Sverri: Cruise Report 2320).

## Material and methods

Coordination of the survey was done during the WGIPS meeting in January 2023 and by correspondence. Planning of the acoustic transects, hydrographic stations and plankton stations were carried out by using the survey planner function in the r-package Rstox version 1.11 (see https://www.hi.no/en/hi/forskning/projects/stox). The survey planner function generates the survey plan (transect lines) in a cartesian coordinate system and transforms the positions to the geographical coordinate system (longitude, latitude) using the azimuthal equal distance projection, which ensures that distances, and also equal coverage, if the method used is designed with this prerequisite, are preserved in the transformation. Figure 1 shows the planned acoustic transects and hydrographic and plankton stations in each stratum. Only parallel transects were used this year, however, because the transects follow great circles they appear bended in a Mercator projection. The participating vessels together with their effective survey periods are listed in the table below:

IESNS post-cruise meeting, Teams 13-15/6 2023

| Vessel | Institute | Survey period |
| :--- | :--- | :--- |
| Dana | DTU Aqua - National Institute of Natural Resources, <br> Denmark | $28 / 04-24 / 05$ |
| G.O. Sars | Institute of Marine Research, Bergen, Norway | $27 / 04-01 / 06$ |
| Jákup Sverri | Faroe Marine Research Institute, Faroe Islands | $05 / 05-16 / 05$ |
| Ámi Friòriksson | Marine and Freshwater Research Institute, Iceland | $08 / 05-27 / 05$ |
| Resolute | CEFAS, United Kingdom | $24 / 04-06 / 05$ |

Figure 2 shows the cruise tracks, Figure 3 the hydrographic and WPII plankton stations and, Figure 4 Macroplankton trawl and Multinet stations and Figure 5 the pelagic trawl stations. Survey effort by each vessel is detailed in Table 1. Daily contacts were maintained between the vessels during the course of the survey, primarily through electronic mail. The temporal progression of the survey is shown in Figure 6.

In general, the weather conditions did not affect the survey even if there were a few days in the northern part at the end of the survey that were not favourable and trawling, WP2 and Multinet sampling at some stations were prevented. The weather conditions in the first part of the survey, when most of the herring was observed, were unusually good. The survey was based on scientific echosounders using 38 kHz frequency. Transducers were calibrated with the standard sphere calibration (Foote et al., 1987) prior to the survey. Salient acoustic settings are summarized in the text table below.

Acoustic instruments and settings for the primary frequency (boldface).

|  | Dana | G. O. Sars | Arni <br> Friöriksson | Jákup Sverri | Resolute |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Echo sounder | Simrad EK60 | Simrad EK80 | Simrad EK80 | Simrad EK80 | Simrad EK80 |
| Frequency ( kHz ) | 38 | $\begin{aligned} & 38,18,70, \\ & 120,200,333 \end{aligned}$ | $\begin{aligned} & 38,18,70, \\ & 120,200 \end{aligned}$ | $\begin{aligned} & 18,38,70, \\ & 120,200,333 \end{aligned}$ | 38,200 |
| Primary transducer | ES38BP | ES 38-7 | ES38-7 | ES38-7 | ES38-7 |
| Transducer installation | Towed body | Drop keel | Drop keel | Drop keel | Hull-mounted |
| Transducer depth (m) | 4-6 | 6 | 6.9 | 6-9 | 6 |
| Upper integration limit ( m ) | 10 | 15 | 12 | 15 | 10 |
| Absorption coeff. (dB/km) | 8.9 | 10.1 | 10.6 | 10.3 | 10.1 |
| Puise length ( ms ) | 1.024 | 1.024 | 1.024 | 1.024 | 1.024 |
| Band width ( kHz ) | 2.425 | 2.43 | 2.425 | 3.06 | 2.425 |
| Transmitter power (W) | 2000 | 2000 | 2000 | 2000 | 2000 |
| Angle sensitivity (dB) | 21.9 | 21.9 | 18 | 21.9 | 18 |
| 2-way beam angle | -20.5 | -20.7 | $-20.3$ | -20.4 | -20.7 |

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|  | Dana | G. O. Sars | Arni <br> Friđriksson | Jákup Sverri | Resolute |
| :--- | :--- | :--- | :--- | :--- | :--- |
| (dB) |  |  |  |  |  |
| Sv Transducer <br> gain (dB) |  |  |  |  |  |
| Ts Transducer <br> gain (dB) | 25.51 | 26.11 | 27.05 | 26.94 | 26.75 |
| Sacorrection (dB) | -0.60 | -0.04 | 0.01 | -0.13 | -0.07 |
| 3 dB beam width <br> (dg) |  |  |  |  |  |
| alongship: | 6.76 | 6.39 | 6.44 | 6.47 | 6.32 |
| athw. ship: | 6.99 | 6.38 | 6.52 | 6.54 | 6.40 |
| Maximum range <br> (m) | 500 | 500 | 500 | 500 | 500 |
| Post processing <br> software | LSSS | LSSS | LSSS | LSSS | Echoview |

All participants except UK used the same post-processing software (LSSS). The UK data were, however, scrutinized using Echoview. Scrutinization was carried out according to an agreement at the PGNAPES scrutinizing workshop in Bergen in February 2009 (ICES 2009), and "Notes from acoustic Scrutinizing workshop in relation to the IESNS", Reykjavik 3.-5. March 2015 (Annex 4 in ICES 2015). Generally, acoustic recordings were scrutinized on daily basis and species identified and partitioned using catch information, characteristic of the recordings, and frequency between integration on 38 kHz and on other frequencies by a scientist experienced in viewing echograms. Immediately after the 2023 survey an online meeting was held to standardise the scrutiny and to agree on particularly difficult scrutiny situations encountered. All vessels used a large or medium-sized pelagic trawl as the main tool for biological sampling. The salient properties of the trawls, plankton nets and hydrographic equipment are as follows:

|  | Dana | G.O. Sars | Arni <br> Friöriksson | Jákup Sverri | Resolute |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Trawl dimensions |  |  |  |  |  |
| Circumference (m) |  | 496 | 832 | 832 | 972 |
| Vertical opening (m) | 20-30 | 25-30 | 20-35 | 45-55 | 30-50 |
| Mesh size in codend (mm) | 18 | 24 | 20 | 45 | 20 |
| Typical towing speed (kn) | 3.5-4.5 | 3.0-4.5 | 3.1-5.0 | 3.5-4.5 | 3.5-5 |
| Plankton sampling |  |  |  |  |  |
| Sampling net | WP2 | WP2 | WP2 | WP2 | WP2 |
| Standard sampling depth (m) | 200 | 200 | 200 | 200 | 200 |

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|  | Dana | G.O. Sars | Arni <br> Frioriksson | Jákup Sverri | Resolute |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Hydrographic sampling |  |  |  |  |  |
| CTD unit | SBE911 | SBE911 | SBE911 | SBE911 | SATV SD208 |
| Standard sampling depth <br> $(\mathrm{m})$ | 1000 | 1000 | 1000 | 1000 | 250 |

Catches from trawl hauls were sorted and weighed; fish were identified to species level, when possible, and other taxa to higher taxonomic levels. A subsample of herring, blue whiting and mackerel were sexed, aged, and measured for length and weight, and their maturity status was estimated using established methods. An additional sample of fish was measured for length. For the Norwegian, Icelandic and Faroese vessel, a smaller subsample of stomachs was sampled for further analyses on land. As part of ongoing stock identity research, herring genetic samples were collected. Salient biological sampling protocols for trawl catches are listed in the table below.

|  | Species | Dana | G.O. Sars | Arri <br> Friŏriksson | Jákup <br> Sverri | Resolute |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Length measurements | Herring | $200-300$ | 100 | 300 | $100-200$ | 100 |
|  | Blue whiting | $200-300$ | 100 | 50 | $100-200$ | 100 |
|  | Mackerel | $100-200$ | 100 | 50 | $100-200$ | 100 |
| Weighed, sexed | Other fish sp. | 50 | 30 | 30 | $100-150$ | 30 |
| maturity determination |  |  |  |  |  |  |
|  | Herring | 50 | $25-100$ | 100 | $50^{*}$ | 50 |
|  | Blue whiting | 50 | $25-100$ | 50 | $50^{*}$ | 50 |
|  | Mackerel | 50 | $25-100$ | 50 | $50^{*}$ | 50 |
| Otoliths/scales collected | Other fish sp. | 0 | 0 | 0 | $0^{*}$ | 0 |
|  | Herring | 50 | $25-30$ | 100 | $25-50$ | 50 |
|  | Blue whiting | 50 | $25-30$ | 50 | $25-50$ | 50 |
|  | Mackerel | 0 | $25-30$ | 50 | $25-50$ | 50 |
| Stomach sampling | Other fish sp. | 0 | 0 | 0 | 0 | 0 |
|  | Herring | 0 | 10 | 10 | 5 | 0 |
|  | Blue whiting | 0 | 10 | 10 | 5 | 0 |
|  | Mackerel | 0 | 10 | 10 | 5 | 0 |
| Genetic samples | Other fish sp. | 0 | 0 | 0 | 0 | 0 |

* If the catch is sufficiently large 100 individuals are always weighed.

Acoustic data were analysed using the StoX software package (version 3.6.1) which has been used for many years now for WGIPS coordinated surveys. A description of StoX can be found in Johnsen et al. (2019) and here: https://www.hi.no/en/hi/forskning/projects/stox. Estimation of abundance from acoustic surveys with StoX is carried out according to the stratified transect design model developed by Jolly and Hampton (1990). This method requires pre-defined strata, and the survey area was therefore split into 5 strata with predefined acoustic transects (this year only 4 strata, as the Barents Sea was not surveyed).

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Within each stratum, parallel transects with equal distances were used. The distance between transects was based on available survey time, and the starting point of the first transect in each stratum was randomized. This approach allows for robust statistical analyses of uncertainty of the acoustic estimates. The strata and transects used in StoX are shown in Figure 2. Generally, and in accordance with most WGIPS coordinated surveys, all trawl stations within a given stratum with catches of the target species (either blue whiting or herring) were assigned to all transects within the stratum, and the length distributions were weighted equally within the stratum.

The following target strength (TS)-to-fish length (L) relationships were used:
Blue whiting: $\mathrm{TS}=20.0 \log (\mathrm{~L})-65.2 \mathrm{~dB}$ (ICES 2012)
Herring: $\quad \mathrm{TS}=20.0 \log (\mathrm{~L})-71.9 \mathrm{~dB}$ (Foote et al. 1987)

The target strength for herring is the traditionally one used while this target strength for blue whiting was first applied in 2012 (ICES 2012).

The hydrographical and plankton stations by survey are shown in Figure 3. Most vessels collected hydrographical data using a SBE 911 CTD. Maximum sampling depth was 1000 m .

Zooplankton was sampled by WP2 nets on all vessels, according to the standard procedure for the surveys. Mesh sizes were 180 or $200 \mu \mathrm{~m}$. The net was hauled vertically from 200 m to the surface or from the bottom whenever bottom depth was less than 200 m . All samples were split in two and one half was preserved in formalin while the other half was dried and weighed. The samples for dry weight were size fractionated before drying by sieving the samples through $2000 \mu \mathrm{~m}$ and $1000 \mu \mathrm{~m}$ sieves, giving the size fractions $180 / 200-1000 \mu \mathrm{~m}$, $1000-2000 \mu \mathrm{~m}$, and $>2000 \mu \mathrm{~m}$. Data are presented as total mg dry weight per m 2 . For the zooplankton distribution map, all stations are presented. Interpolation was carried out using Bratseth's Successive Correction Method (Bratsheth, 1986). This method was designed specifically for marine data, and it uses bottom depth to calculate the similarity among the interpolation points. More specifically, it uses objective analysis with a Gaussian correlation function where the effective distance between the observations and the nodes of the interpolation grids is defined based on the difference in bottom depths, as follows:

$$
r^{2}=r_{x}^{2}+r_{y}^{2}+\left(\lambda \frac{H_{\mathrm{a}}-H_{\mathrm{o}}}{H_{\mathrm{a}}+H_{\mathrm{o}}}\right)^{2}
$$

where rx and ry is the geographic distance in the zonal and meridional directions, and Ha and Ho are the bottom depths at the analysis and observation points, respectively (Skagseth and Mork, 2012). The analysis was done using an R script based on a MATLAB routine developed by Kjell Arne Mork (Mork et al. 2014). For the time series, stations in the Norwegian Sea delimited to east of $14^{\circ} \mathrm{W}$ and west of $20^{\circ} \mathrm{E}$ have been included. Estimates of the statistical distribution of the zooplankton biomass indices is done by simple bootstrapping by re-sampling with replacement.

## Results and Discussion

## Hydrography

The temperature distributions in the ocean, averaged over selected depth intervals; $0-50 \mathrm{~m}$, $50-200 \mathrm{~m}$, and $200-500 \mathrm{~m}$, are shown in Figures $7 \mathrm{a}-\mathrm{c}$. The temperatures in the surface layer $(0-50 \mathrm{~m})$ ranged from below $0^{\circ} \mathrm{C}$ in the Greenland Sea to $9-10^{\circ} \mathrm{C}$ in the southern part of the Norwegian Sea (Figure 7a). The Arctic front was encountered south of $65^{\circ} \mathrm{N}$ east of Iceland extending eastwards towards about $2^{\circ} \mathrm{W}$ where it turned north-eastwards to $65^{\circ} \mathrm{N}$ and then almost straight northwards. The front sharpened and had a more eastern location with depths. Further to west at about $8^{\circ} \mathrm{W}$, another front runs northward to Jan Mayen, the Jan Mayen Front, that was most distinct in the upper 200 m . The warmer North Atlantic water formed a broad tongue that stretched far northwards along the Norwegian coast with temperatures about $5^{\circ} \mathrm{C}$ to the Bear Island at $74.5^{\circ} \mathrm{N}$ in the surface layer.

Relative to the long-term mean, from 1995 to 2021, the temperatures at $0-50 \mathrm{~m}$ were below the mean at the western and eastern parts of Norwegian Sea while in the central areas, the temperatures were mostly above the mean (Figure 7a). At $50-200 \mathrm{~m}$ depth, the patterns were also fragmented, but the Norwegian Sea was, in general, colder than the long-term mean in the western part and in the Lofoten Basin (Figure 7b). At 200-500 m depth, the patterns were less fragmented and nearly the whole Norwegian Sea was colder than the long-term mean (Figure 7c). The negative anomalies North of the Faroese derive likely from increased influenced of the East Iceland Current compared to the long-term mean.

Two main features of the circulation in the Norwegian Sea, where the herring stock is grazing, are the Norwegian Atlantic Current (NWAC) and the East Icelandic Current (EIC). The NWAC with its offshoots forms the northern limb of the North Atlantic current system and carries relatively warm and salty water from the North Atlantic into the Nordic Seas. The EIC, on the other hand, carries Arctic waters. To a large extent this water derives from the East Greenland Current, but to a varying extent, some of its waters may also have been formed in the Iceland and Greenland Seas. The EIC flows into the southwestern Norwegian Sea where its waters subduct under the Atlantic waters to form an intermediate Arctic layer. While such a layer has long been known in the area north of the Faroes and in the Faroe-Shetland Channel, it is in the last four decades a similar layer has been observed all over the Norwegian Sea. Also, in periods this layer has been less well-defined.

This circulation pattern creates a water mass structure with warm Atlantic Water in the eastern part of the area and more Arctic conditions in the western part. The NWAC is rather narrow in the southern Norwegian Sea, but when meeting the Vøring Plateau off Mid Norway it is deflected westward. The western branch of the NWAC reaches the area of Jan Mayen at about $71^{\circ} \mathrm{N}$. Further northward in the Lofoten Basin the lateral extent of the Atlantic water
gradually narrows again, apparently under topographic influence of the mid-ocean ridge. It has been shown that atmospheric forcing largely controls the distribution of the water masses in the Nordic Seas. Hence, the lateral extent of the NWAC, and consequently the position of the Arctic Front, that separates the warm North Atlantic waters from the cold Arctic waters, is correlated with the large-scale distribution of the atmospheric sea level pressure. The local air-sea heat flux in addition influence the upper layer and it is found that it can explain about half of the year-to-year variability of the ocean heat content in the Norwegian Sea.

## Zooplankton

The zooplankton biomass ( mg dry weight $\mathrm{m}^{-2}$ ) distribution in the upper 200 m in 2023 is shown in Figure 8b). Sampling stations were evenly spread over the area, covering Atlantic water, Arctic water, and the Arctic frontal zone. The highest zooplankton biomasses were found in the Iceland Sea, northeast of Iceland. This was also reflected in the time series. A smaller area with high biomass was also found outside the Norwegian continental shelf, at $66^{\circ} \mathrm{N}$. In the rest of the investigated area there was a relatively even distribution of zooplankton biomass. This was different from the distribution in 2022, where the highest zooplankton biomasses were found in the eastern and southeastern parts.

Figure 9b) shows the zooplankton time series indices for the sampling area (delimited to east of $14^{\circ} \mathrm{W}$ and west of $20^{\circ} \mathrm{E}$ ). To examine regional biomass differences, the area was divided into 4 sub-areas 1) East of Iceland, 2) the Jan Mayen Arctic front, 3) the Lofoten Basin (covering the northern Norwegian Sea, and 4) the Norwegian Basin (covering the southern Norwegian Sea), figure 10 a). The zooplankton biomass index for 2023 was respectively: 13394, 8009,8484 and 9688 mg dry weight $\mathrm{m}^{-2}$. There was an increase in zooplankton biomass in Icelandic water, where the biomass was almost three times higher in 2023 compared to 2022. For the other sub-areas minor changes were observed from last year. The zooplankton biomass indices for the Norwegian Sea in May have been estimated since 1995. All sub-areas had a high biomass period until mid-2000, and a lower period thereafter. The long-term decrease has been most pronounced in the Iceland Sea. In the Lofoten- and Norwegian Basins there has been an increasing trend during the low-biomass period.

The reasons for the changes in zooplankton biomass are not obvious. It is worth noting that the period with lower zooplankton biomass coincides with higher-than-average heat content in the Norwegian Sea (ICES, 2020) and reduced inflow of Arctic water into the southwestern Norwegian Sea (Kristiansen et al., 2019; Skagseth et al., 2022). Timing effects, such as match/mismatch with the phytoplankton bloom, can also affect the zooplankton abundance. The high biomass of pelagic fish feeding on zooplankton has been suggested to be one of the main causes for the reduction in zooplankton biomass. However, carnivorous zooplankton and not pelagic fish may be the main predators of zooplankton in the Norwegian Sea (Skjoldal et al., 2004), and we do not have good data on the development of the carnivorous zooplankton stocks.

## Norwegian spring-spawning herring

Survey coverage in the Norwegian Sea was considered adequate in 2023. The zero-line was believed to be reached for adult NSS herring in most of the areas. It is recommended that the results from IESNS 2023 can be used for assessment purpose. The herring was primarily distributed in the western and northwestern area this year (Figure 10). The 2016-year-class was by far the most abundant year class in the areas where most of the herring biomass was found. It is a commonly observed pattern that the older fish are distributed in the southwest while the younger fish are found closer to the nursery areas in the Barents Sea (Figure 11).

Seven-year-old herring (2016-year class) dominated both in terms of number and biomass (both around $57 \%$ ) on basis of the StoX bootstrap estimates for the Norwegian Sea (Table 2). The point estimate of abundance of the 2016 year-class decreased by $2 \%$ compared to last year's estimate which is much less than the decline between 2021 and 2022 (Figure 12). However, the 2015-2013 year classes decrease with 47-66\% compared to last year's estimates. This indicates that the mortality of older herring has been higher than that of the 2016 year class. Uncertainty estimates for number at age based on bootstrapping within StoX are shown in Figure 13 and Table 2. The relative standard error (CV) is $23 \%$ for the total biomass and $22 \%$ for the total numbers estimate, and the relative standard errors for the dominating age groups is around $24-32 \%$ (Figure 13).

The total estimate of herring in the Norwegian Sea from the 2023 survey was 16.5 billion in number and the biomass was 4.1 million tonnes. The biomass estimate is about $8 \%$ lower than the 2022 survey estimate and the estimated number is about $17 \%$ lower than in 2022 . The biomass estimate decreased significantly from 2009 to 2012 and has since then been rather stable at 4.1 to 5.9 million tonnes with similar confidence interval (Figure 14), with the lowest abundance occurring in 2023.

Since 2015 an increased awareness has been raised around the age reading of herring. It appeared that the age distributions from the different participants some years showed differences and also the older specimens appear to have uncertain ages. An age-reading workshop was held in Bergen 17.-19. April 2023 (WKARNSSH2). This workshop was based on otoliths and scales collected in 2021 and subsequently exchanged between the participating countries. At the publication of this survey report the concluding report from WKARNSSH2 is not yet available, but it will be for next year's survey report.

With respect to age-reading, the comparison between the nations, in this year's survey, show that there were some differences within strata (Figure 15). Particularly, in stratum 2 there were differences between the EU vessel (3-year-olds dominating) and the Norwegian vessel (7-year-olds dominating). This could at least partly be explained by spatial differences in sampling between vessels; the EU vessel have all their samples in the northern part of stratum 2 while the Norwegian vessel have samples both from the northern and the southern part of
the stratum. It is well known that mean age of herring in the IESNS generally decreases with increasing latitude.

Recently, concerns have been raised by the survey groups for the International ecosystem surveys in the Nordic Seas (IESNS and IESSNS) on mixing issues between Norwegian spring-spawning herring and other herring stocks (e.g. Icelandic summer-spawning, Faroese autumn-spawning, Norwegian summer-spawning and North Sea type autumn-spawning herring) occurring in some of the fringe regions in the Norwegian Sea. Until now, fixed cut lines have been used by the survey group to exclude herring of presumed other types than NSS herring, however this simple procedure is thought to introduce some contamination of the stock indices of the target NSS herring. WGIPS noted in their 2019 report that the separation of different herring stock components is an issue in several of the surveys coordinated in WGIPS and the needs for development of standardized stock splitting methods was also noted in the WKSIDAC (ICES 2017).

## Blue whiting

Boostrap estimates of abundance, biomass, mean length and mean weight of blue whiting during IESNS 2023 are shown in Table 3. The estimated biomass was 961 thousand tons $(\mathrm{CV}=0.14)$ which is a $36 \%$ decrease from last year's estimate, and slightly below the average from the period 2008-2022. The estimated total abundance was 12.8 billion ( $\mathrm{CV}=0.15$ ) which is a $57 \%$ decrease from last year's estimate. The stock is dominated by 1-3 years old blue whiting. Uncertainty estimates for numbers at age based on bootstrapping with StoX are shown in Figure 18 and Table 3.

The spatial distribution of blue whiting in 2023 is shown in Figure 16. As usual, most of the fish was registered in the eastern part of the Norwegian Sea. The largest fish was found in the northwestern part of the of the survey area (Figure 17). Comparison of the size and age distributions of blue whiting by stratum and country are shown in Figure 19 and 20, and they seem to be in fairly good agreement.

## Mackerel

Trawl catches of mackerel are shown in Figure 21. Mackerel was present in the southern and eastern part of the Norwegian Sea in the beginning of May. The spatial distribution of catches in 2023 were similar to 2022, i.e. a lower northward extent than in the period 2008-2021. No further quantitative information can be drawn from these data as this survey is not designed to monitor mackerel.

## General recommendations and comments

RECOMMENDATION $\quad$ ADDRESSED TO

1. Continue the methodological research in distinguishing between WGIPS herring and blue whiting in the interpretation of echograms.
2. Implement logging of sonar data to measure the amount of WGIPS herring in the surface blind zone

## Next year's post-cruise meeting

We will aim for next meeting in 16-18 June 2024. The final decision will be made at the next WGIPS meeting.

## Concluding remarks

- The sea temperature in 2023 was generally below the long-term mean (1995-2021) in the Norwegian Sea
- The 2023 indices of meso-zooplankton biomass in the Norwegian Sea and adjoining waters were fairly similar to last year's estimates for all areas except the Icelandic area where the index is much higher compared to the most recent years
- The total biomass estimate of NSSH in herring in the Norwegian Sea was 4.1 million tonnes, which is an $8 \%$ decrease from the 2022 survey estimate. The estimate of total number of NSSH was 16.5 billion, which is $17 \%$ lower than in the 2022 survey. The survey followed the pre-planned protocol and the survey group recommends using the abundance estimates in the analytical assessment.
- The 2016 year class of NSSH dominated in the survey indices both in numbers and biomass (both $57 \%$ ). The abundance of the 2016 year-class decreased by $2 \%$ compared to last year's estimate.
- The biomass of blue whiting measured in the 2022 survey decreased by $36 \%$ from last year's survey and $57 \%$ in terms of numbers. The stock is dominated by the 2020 to 2022 year classes.


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## Tables

Table 1. Survey effort by vessel for the International ecosystem survey in the Nordic Seas in May - June 2023.

| Vessel | Effective <br> survey <br> period | Effective <br> acoustic <br> cruise <br> track <br> (nm) | Trawl <br> stations | Ctd <br> stations | Aged <br> fish <br> (HER) | Length <br> fish <br> (HER) | Plankton <br> stations |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Dana | $28 / 4-19 / 5$ | 2218 | 31 | 38 | 154 | 335 | 37 |
| Jákup Sverri | $6 / 5-14 / 5$ | 1297 | 17 | 22 | 569 | 1849 | 22 |
| Árni Fridriksson | $10 / 5-24 / 5$ | 2700 | 19 | 31 | 1424 | 4608 | 29 |
| G.O. Sars | $27 / 4-01 / 6$ | 3858 | 36 | 78 | 226 | 557 | 67 |
| Resolute | $24 / 4-06 / 5$ | 1345 | 13 | 19 | 145 | 244 | 19 |

Table 2. IESNS 2023 in the Norwegian Sea. Estimates of abundance, mean weight and mean length of Norwegian spring-spawning herring. The estimates are mean of 1000 bootstrap replicates in Stox.


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Table 3. IESNS 2023 in the Norwegian Sea. Estimates of abundance, mean weight and mean length of blue whiting. The estimates are mean of 1000 bootstrap replicates in Stox.


Figures


Figure 1. The pre-planned strata and transects for the IESNS survey in 2023 (red: EU, dark blue: Norway, yellow: Faroes Islands, violet: UK, green: Iceland). Hydrographic stations and plankton stations are shown as blue circles with diamonds. All the transects have numbered waypoints for each 30 nautical mile and at the ends.

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Figure 2. Cruise tracks and strata (with numbers) for the IESNS survey in May 2023.

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Figure 3. IESNS survey in May 2023: location of hydrographic and WPII plankton stations.

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Figure 4. IESNS survey in May 2023: location of Macroplankton/Krill trawl and Multinet stations.


Figure 5. IESNS survey in May 2023: cruise tracks and location of pelagic trawl stations.

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Figure 6. Temporal progression IESNS in April-May 2023.


Figure 7a. Temperature (left) and temperature anomaly (right) averaged over $0-50 \mathrm{~m}$ depth in May 2023. Anomaly is relative to the 1995-2021 mean.


Figure 7b. Same as above but averaged over $50-200 \mathrm{~m}$ depth.

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Figure 7c. Same as above but averaged over 200-500 m depth.
a)

b)


Figure 8. Distribution of zooplankton biomass ( mg dry weight $\mathrm{m}^{-2}$ ) in the upper 200 m in May a) IESNS 2022 and b) IESNS 2023.
a)

b)


Figure 9 a) shows the sub-areas, and b) the indices of zooplankton biomass ( mg dry weight $\mathrm{m}^{-2}$ ) sampled by WP2 in May from 1995-2023.

IESNS post-cruise meeting, Teams 13-15/6 2023


IESNS post-cruise meeting, Teams 13-15/6 2023
(b)


Figure 10. Distribution of Norwegian spring-spawning herring as measured during the IESNS survey in May 2023 in terms of NASC values $\left(\mathrm{m}^{2} / \mathrm{nm}^{2}\right)$ averaged for every 1 nautical mile. The NASC values are represented as both bars (a) and bubbles (b)

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Figure 11. Mean length of Norwegian spring-spawning herring in all hauls in IESNS 2023.


Figure 12. Tracking of the Total Stock Number at age (TSN, in billions) of Norwegian spring-spawning herring for each cohort since 2004 from age 2 to age 8. From 2008, stock is estimated using the StoX software. Prior to 2008, stock was estimated using BEAM.


Figure 13. IESNS 2023. Norwegian spring-spawning herring in the Norwegian Sea: $R$ boxplot of abundance and relative standard error (CV) obtained by bootstrapping with 1000 replicates using the StoX software

IESNS,TSB


Figure 14. Biomass estimates of Norwegian-spring spawning herring in the IESNS survey (Barents Sea, east of $20^{\circ} \mathrm{E}$, is excluded) from 1996 to 2023 as estimated using BEAM (1996-2007; calculated on basis of rectangles) and as estimated with the software StoX (2008-2023; bootstrap means with $90 \%$ confidence interval; calculated on basis of standard stratified transect design).

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Figure 15. Comparison of the age distributions of NSS-herring by stratum and country in IESNS 2023. The strata are shown in Figure 3.

IESNS post-cruise meeting, Teams 13-15/6 2023


IESNS post-cruise meeting, Teams 13-15/6 2023
(b)


Figure 16. Distribution of blue whiting as measured during the IESNS survey in May 2023 in terms of NASC values $\left(\mathrm{m}^{2} / \mathrm{nm}^{2}\right)$ (a) averaged for every 1 nautical mile. The NASC values are represented as both bars (a) and bubbles (b).

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Figure 17. Mean length of blue whiting in all hauls in IESNS 2023. The strata are shown.


Figure 18. IESNS 2023. Blue whiting in the Norwegian Sea: R boxplot of abundance and relative standard error (CV) obtained by bootstrapping with 1000 replicates using the StoX software.

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Figure 19. Comparison of the length distributions of blue whiting by stratum and country in IESNS 2023. The strata are shown in Figure 3

IESNS post-cruise meeting, Teams 13-15/6 2023


Figure 20. Comparison of the age distributions of blue whiting by stratum and country in IESNS 2023. The strata are shown in Figure 3.


Figure 21. Pelagic trawl catches of mackerel in IESNS 2023.


DISTRIBUTION AND ABUNDANCE OF NORWEGIAN SPRINGSPAWNING HERRING DURING THE SPAWNING SEASON IN 2023

Tittel (norsk og engelsk):
Distribution and abundance of Norwegian spring-spawning herring during the spawning season in 2023
Fordeling og mengde av norsk vårgytende sild under gytesesongen i 2023

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I perioden 13. - 26 . Februar 2023 ble gytefeltene til norsk vårgytende sild fra Møre $\left(62^{\circ} 15^{\prime} \mathrm{N}\right)$ til Troms $\left(70^{\circ} 37^{\prime} \mathrm{N}\right)$ dekket akustisk med de kommersielle fartøyene MS Eros og MS Vendla. Den estimerte biomassen var omtrent $18 \%$ høyere, og det estimerte antallet omtrent $7 \%$ høyere sammenlignet med fjorårets tokt. Usikkerheten i årets estimater er imidlertid vesentlig høyere enn i de seks foregående årene, til tross for at gjennomføringen av toktet gikk i henhold til planen. Gytebestanden var dominert av 2016 -årsklassen med $46 \%$ i antall og $42 \%$ i vekt. Mesteparten av gytebestanden befant seg vest av Lofoten i år, mer konsentrert enn tidligere år. Sammenlignet med toktet i fjor var silda kommet kortere i modningsprosessen i âr noe som indikerer senere gyting. Det anbefales å bruke estimatene av relativ mengde fra toktet i 2023 i ICES sin bestandsvurdering av norsk vårgytende sild.

## Sammendrag (engelsk):

During the period 13-26th of February 2023 the spawning grounds of Norwegian spring-spawning herring from Møre ( $62^{\circ} 15^{\prime} \mathrm{N}$ ) to Troms $\left(70^{\circ} 37^{\prime} \mathrm{N}\right.$ ) were covered acoustically by the commercial vessels MS Eros and MS Vendla. The estimated biomass was about $18 \%$ higher, and the estimated total number was about $7 \%$ higher this year compared to the last year's survey. The uncertainty of the estimates in 2023 was much higher than what was observed the last six years, although the survey was conducted according to the plan. The spawning stock was dominated by the 2016 year class; $46 \%$ in numbers and $42 \%$ in biomass. Most of the spawning stock was found west of Lofoten this year, more concentrated than earlier years. Compared with last year's survey the herring was earlier in the spawning process this year which indicates later spawning. The estimates of relative abundance from the survey in 2023 are recommended to be used in this year's ICES stock assessment of Norwegian spring-spawning herring.
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## 1 - Introduction

Acoustic surveys on Norwegian spring-spawning herring during the spawning season has been carried out regularly since 1988, with some breaks (in 1992-1993, 1997, 2001-2004 and 2009-2014). In 2015 the survey was initiated again partly based on the feedback from fishermen and fishermenấ ${ }^{T M}$ s organizations that IMR should conduct more surveys on this commercially important stock. Since then this survey, hereafter termed the NSSH spawning survey, has continued using hired commercial fishing vessels. In the ICES benchmark assessment of NSS herring in 2016 it was decided to use the data from this time series as input to the stock assessment, together with the ecosystem survey in the Norwegian Sea in May and catch data. Thus, the results from the NSSH spawning survey, have significant influence on the ICES catch advice.

The objective of the NSSH spawning survey 2023 was to continue the time series of abundance estimates, both point estimates and uncertainty, for use in the ICES WGWIDE stock assessment. Moreover, other biological information about the surveyed spawning stock of Norwegian spring-spawning herring is presented: spatial distribution of biomass and acoustic densities, spatial patterns in maturity and variations in temperature.

## 2 - Material and methods

## 2.1 - Survey design

During the period 13-26th of February 2023 (same period as in 2017-2022) the spawning grounds from Møre $\left(62^{\circ} 15^{\circ} \mathrm{N}\right)$ to Troms $\left(70^{\circ} 37^{\prime} \mathrm{N}\right)$ were covered acoustically by the commercial fishing vessels MS Eros and MS Vendla. The survey was planned based on information from the previous spawning surveys and the distribution of the herring fishery during the 2022/2023 season and earlier years. Figure 1 shows the distribution of commercial herring catches during the 2022/2023 winter season. The fishery seems to follow the herring migration from the continental slope area north of Andenes in the October into the overwintering areas around Kvænangen in November-January, and out again in January-February to start the spawning migration heading south. During the last nine years in November-January there has also been a fishery that seemed to follow a herring migration from the Norwegian Sea towards Røst and the Træna deep, however, this was not observed in 2023. In the two weeks prior to the start of the survey in 2023 the fishery had not moved south of Andenes, which is unusually far north for this period. Thus, the information from the fishery prior to the survey was sparse this year. Based on the observations from previous years it was decided to plan the survey coverage to take account of herring both migrating from the North along Lofoten and Vesterålen, and herring entering the Røst/Træna deep area from the west (i.e. from Norwegian Sea). As seen in Figure 1, the fishery during the survey in 2023 took place west and southwest of Røst and Lofoten, and north of Andenes, which is further north than usual at this time. The survey design followed a standard stratified design (Jolly and Hampton 1990), where the survey area was stratified before the survey start according to the assumed density structures of herring during the spawning migration (based on previous surveys and fisheries). All strata were covered with a zigzag design since this is the most efficient use of survey effort (Harbitz 2019). The survey planner function in the Rstox_1.11 package in $r$ was used to generate the transects, and this function generates survey tracks with uniform coverage of strata and a random starting position in the start of each stratum. Each straight line in the zigzag track within a stratum was considered as a transect and a primary sampling unit (Simmonds and MacLennan 2005). Transit tracks between strata, i.e. from the end of the zigzag in one stratum to the start of the zigzag in the next stratum, were not used as primary sampling units. Prior to the start of the survey in 2023 all reported NSSH catches had been taken north of Lofoten, hence, the survey coverage (see Aglen 1989) was planned to be low to moderate south of $66^{\circ} \mathrm{N}$ since it was assumed that the fishing fleet followed the front of the herring migrating south and that the abundance of herring south of the fleet therefore was insignificant.

Trawl sampling was planned to be carried out on a regular basis during the survey to confirm the acoustic observations along transects and to be able to give estimates of abundance for different size and age groups. Both vessels used a Multpelt 832 scientific sampling trawl with small meshed ( 20 mm ) inner net in the codend and a slit (so called "splitt") close to the codend to avoid too large catches. The following variables of individual herring were analysed from each station with herring catch: total weight in grams and total length in cm (rounded down to the nearest 0.5 cm ) of up to 100 individuals per sample. In addition, age from scales, sex, maturity stage, stomach fullness and gonad weight in grams were recorded in up to 50 individuals per sample. Some genetic samples and otoliths were also collected to be used in later research projects.

CTD casts (using Seabird 911 systems) were taken by both vessels, spread out haphazardly in the survey area.
Echosounder data from the 38 kHz transducers was, as usual, the basis for measurement of fish density. The software LSSS version 2.14.0 was used for post-processing. Echogram scrutinization was carried out by the cruise leader and the chief instrument officer. Data was partitioned into the following categories: "herring", "other", "capelin" and "air bubbles" (upper 20 meters from the transducer near field).

## 2.2 - Abundance estimation

The acoustic density values were stored by species category in nautical area scattering coefficient (NASC) [m2 n.mi.-2] units (MacLennan et al. 2002) in a database with a horizontal resolution of 0.1 nmi and a vertical resolution of 10 m , referenced to the sea surface. The software Stox version 3.6 (Johnsen et al. 2019) was used to estimate abundance. More details and equations can be found in Salthaug et al. (2021). All trawl stations with herring were used to derive a common length distribution for all transects within the respective strata and all stations had equal weight. Abundance of fish in numbers and biomass were estimated based on 1000 bootstrapping iterations of biotic stations and acoustic transects. All results presented in this report are based on the output of this. The following target strength (TS)-to-fish length ( L ) relationship was used for herring: $T S=20 \log L-71.9$ (Foote 1987).

## 3 - Results and discussion

## 3.1 - Survey coverage

The cruise tracks of the NSSH spawning survey in 2023, together with pelagic trawl stations and CTD stations are shown in Figure 2. As mentioned above, the coverage south of $66^{\circ} \mathrm{N}$ was low to moderate since we expected low abundance in this area, which turned out to be the case (see below). Thus, most of the available survey effort was used to carry out dense coverage of the strata north of $66^{\circ} \mathrm{N}$. The survey coverage (see Aglen 1989) of the first three strata was 5,7 and 9 respectively (starting from south) and 11 in the two strata west of Lofoten and Vesterallen. The northernmost stratum had a survey coverage of 9 . Pelagic trawl hauls were carried out regularly (Fig. 2) in the areas where herring like marks were observed on the echo sounder, to confirm the acoustic observations based on species composition in the catch and to obtain biological samples like size, maturity stage and age of herring. A total of 29 CTD casts were carried out in the surveyed area (Fig. 2). Bad weather prevented some of the planned CTD casts. Nautical area scattering coefficients (NASC) allocated to herring from acoustic transects by each nautical mile are shown in Figure 3. Marks on the echosounders that seemed like herring started to occur near Træna around $66^{\circ} \mathrm{N}$, but these were weak registrations. A trawl haul also confirmed presence of herring in the area. Strong herring registrations were not observed before north of $67^{\circ} \mathrm{N}$, approximately 60 nautical miles southwest of Røst. Herring was observed on the echosounder and in trawl hauls in all strata north of this. The highest concentrations were observed west of Lofoten, both around the shelf edge and on the banks nearer land. The zero-line was established in the north at around $70.3^{\circ} \mathrm{N}$, which is further south than last year. A few capelin marks were observed on the northernmost transects and these were also confirmed by trawl samples.

## 3.2 - Estimates of abundance

The abundance estimates from this survey are viewed as relative, i.e. as indices of abundance, since there are highly uncertain scaling parameters like acoustic target strength and compensation for herring migrating in the opposite direction of the survey. The abundance estimates are shown in Table 1 and 2. In terms of the point estimates, the 2016 year class (age 7) dominated both in numbers ( $46 \%$ ) and biomass ( $42 \%$ ). Compared with the point estimates from last year (see Salthaug et al. 2022) the 2016 year class was reduced by 6-9 \% in numbers (largest reduction for the median). The point estimate (mean) of total stock biomass (TSB) in the survey area was 3.91 million tons which is $18 \%$ higher than last year's estimate. The time series of mean of total stock biomass from the survey is shown in Figure 4. The point estimate of total stock number (TSN) in the survey area was 13.04 billion which is $7 \%$ higher than last year's estimate. This year's estimate of TSB corresponds to the mean of the time series while TSN is below the mean. The relative standard error (CV) of TSB is $40 \%$ and that of TSN is $39 \%$ (Tab. 1 and 2). These estimates of sample uncertainty are much higher compared to those from the last six surveys (7-17\%). The CV per age (Tab.1, Tab. 2 and Fig. 5) shows the normal pattern with high uncertainty for the very young and old ages and lower but still very high uncertainty (around $50 \%$ ) for the most abundant ages in the survey. The uncertainty estimates are mainly driven by the highly patchy observations in the stratum west/southwest of Lofoten which contained most of the estimated abundance (see below). Some very strong/dense schools in this area caused high variation in the values of acoustic density samples (EDSUs), i.e. some values were much higher than the rest. One way to reduce the uncertainty when the fish have such patchy distributions is to increase the sampling effort. However, the degree of coverage in this stratum was already high (11) since we expected high densities of fish here. It was therefore unexpected that the uncertainty estimates become much higher than in the most recent years, and we conclude that this is due to randomness. Some years the survey will "hit" large schools which increases the uncertainty and point estimates compared to years when fewer such large schools are observed. This phenomenon is a well-known problem in abundance surveys, see e.g. Pennington (1996). It is doubtful whether changing the survey design or timing of the survey would have prevented the high uncertainty. Figure 6 shows estimates of number per year class in the four most recent surveys. These numbers are expected to decline between consecutive years due to mortality, for ages that are fully recruited to the survey (seems to be age 5 or 6 ). Such declines are observed for the 2013 and 2016 year classes which are the two strongest
in the survey, however, many of the other weaker year classes show an increase from 2022 to 2023. The latter "unexpected" observation is probably a reflection of the very high uncertainty in the 2023 estimates. As observed in the most recent surveys, the 2016 and 2013 year classes are estimated to be the most abundant also in 2023 which shows that this survey is fairly internally consistent. Mean weight and length from the 2023 spawning survey are shown in Table 3. The point estimates are higher than last year for ages 3-6 which indicates more favorable growth conditions.

## 3.3 - Spatial distribution of the stock and maturity

The relative distribution of the estimated biomass per stratum is shown in Figure 7. Most of the biomass (87\%) was found in the stratum west and southwest of Lofoten and a significant proportion (10\%) was found in the stratum west of Vesterålen. This shows that the spawning stock was very concentrated and far north this year. Figure 8 shows the proportion of different maturation stages at different latitudes. Most of the herring was classified as maturing or ripe. The proportion of maturing herring was highest (i.e. lowest proportion of spawning and ripe) in the northernmost area as usual. Regarding spawning area, the fishery, like last year, indicates that most of the spawning occurred west of Lofoten around the Røst bank this year (Fig. 1). The fishery in 2023 south of $67^{\circ} \mathrm{N}$ has been insignificant and based on this it is likely that the spatial distribution of the herring observed during the spawning survey reflects the spawning area. The proportion of the different spawning stages among mature herring is shown for the period 2017-2023 in Figure 9 (the surveys have been carried out in the second part of February during all these years). The large proportion of maturing herring and low proportion of spawning/ripe herring in 2023 compared to earlier years indicates that the spawning is late this year, however, the distribution in 2019 was very similar so not exceptional late.

## 3.4 - Geographical variation in temperatures experienced by the herring

Temperatures experienced by herring from close to the surface and down to 250 m are shown in Figure 10 for the areas south and north of $67^{\circ} \mathrm{N}$, for the years after 2016 when the survey has been carried out in the same period (latter half of February). The temperatures in 2023 varied from $8.1^{\circ} \mathrm{C}$ at 150 m depth south of $67^{\circ} \mathrm{N}$ to $6.6^{\circ} \mathrm{C}$ at 5 m depth north of 67 N . This range is lower than last year when the variation was large, and the temperature in 2023 was higher at all depths compared to last year. At typical spawning depths of herring at $100-200 \mathrm{~m}$ depth, the temperature conditions were quite similar to those observed during the most recent NSSH spawning surveys.

## 3.5 - Quality of the survey

Both vessels were equipped with multifrequency equipment on a drop keel. The weather conditions were more challenging this year than last, and the vessels had to lie still and wait for better conditions a couple of times, but acoustic data with good quality was recorded and trawling on registrations could be carried out adequately most of the time. Luckily, the weather was very nice in the area where most of the stock was found. Like last year the zero line was clearly established in the north, and we are not aware of any observations that indicates presence of mature NSS herring outside the survey area during the survey this year. To conclude, the acoustic and biological data recorded in 2023 on the NSSH spawning survey were of satisfactory quality and the estimates from the survey are recommended to be used in the stock assessment of Norwegian spring-spawning herring in 2023.

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## 5 - Tables

Table 1. Abundance estimates (million individuals) of Norwegian spring-spawning herring during the spawning survey 13.-26 February 2023, based on 1000 bootstrap replicates in Stox.

| Age | 5th percentile | Median | 95th percentile | Mean | SD | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 55 | 184 | 402 | 205 | 115 | 0.56 |
| 4 | 25 | 170 | 425 | 189 | 132 | 0.70 |
| 5 | 200 | 538 | 1015 | 568 | 255 | 0.45 |
| 6 | 247 | 550 | 1094 | 594 | 269 | 0.45 |
| 7 | 2725 | 5751 | 10287 | 6014 | 2333 | 0.39 |
| 8 | 111 | 327 | 712 | 351 | 181 | 0.52 |
| 9 | 359 | 981 | 1981 | 1057 | 523 | 0.49 |
| 10 | 521 | 1280 | 2379 | 1349 | 585 | 0.43 |
| 11 | 79 | 277 | 622 | 304 | 171 | 0.56 |
| 12 | 160 | 496 | 1039 | 537 | 279 | 0.52 |
| 13 | 152 | 460 | 1025 | 508 | 270 | 0.53 |
| 14 | 142 | 483 | 1151 | 538 | 316 | 0.59 |
| 15 | 97 | 293 | 643 | 324 | 175 | 0.54 |
| 16 | 23 | 97 | 276 | 116 | 79 | 0.68 |
| 17 | 88 | 314 | 766 | 356 | 221 | 0.62 |
| 19 | 0 | 9 | 30 | 10 | 10 | 1.00 |
| TSN | 5725 | 12572 | 21883 | 13037 | 5040 | 0.39 |

Table 2. Biomass estimates (thousand tons) of Norwegian spring-spawning herring during the spawning survey 13.-26. February 2023, based on 1000 bootstrap replicates in Stox.

| Age | 5th percentile | Median | 95th percentile | Mean | SD | CV |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 6 | 23 | 53 | 26 | 15 | 0.58 |
| 4 | 4 | 31 | 83 | 35 | 27 | 0.76 |
| 5 | 42 | 107 | 199 | 113 | 51 | 0.45 |
| 6 | 59 | 133 | 263 | 143 | 64 | 0.45 |
| 7 | 745 | 1591 | 2852 | 1657 | 648 | 0.39 |
| 8 | 34 | 106 | 237 | 114 | 61 | 0.53 |
| 9 | 118 | 323 | 663 | 350 | 175 | 0.50 |
| 10 | 174 | 430 | 802 | 453 | 197 | 0.44 |
| 11 | 28 | 98 | 223 | 109 | 62 | 0.57 |
| 12 | 61 | 187 | 395 | 204 | 107 | 0.52 |
| 13 | 56 | 169 | 376 | 187 | 100 | 0.53 |
| 14 | 53 | 183 | 438 | 204 | 121 | 0.59 |
| 15 | 38 | 112 | 249 | 125 | 68 | 0.54 |
| 16 | 9 | 38 | 109 | 46 | 31 | 0.68 |
| 17 | 34 | 119 | 292 | 134 | 83 | 0.61 |
| 19 | 0 | 4 | 11 | 4 | 4 | 0.99 |
| TSB | 1674 | 3768 | 6660 | 3910 | 1560 | 0.40 |

Table 3. Estimated length and weight of individuals by age group of Norwegian spring-spawning herring during the spawning survey 13.-26. February 2023, based on 1000 bootstrap replicates in Stox

| Age | mean weight (g) | CV(weight) | mean length (cm) | CV(length) |
| :---: | :---: | :---: | :---: | :---: |
| 3 | 121.1 | 0.050 | 26.2 | 0.012 |
| 4 | 166.7 | 0.081 | 28.7 | 0.025 |
| 5 | 193.7 | 0.043 | 29.7 | 0.009 |
| 6 | 236.9 | 0.023 | 31.4 | 0.006 |
| 7 | 269.5 | 0.008 | 32.7 | 0.002 |
| 8 | 318.4 | 0.034 | 34.4 | 0.011 |
| 9 | 328.7 | 0.014 | 35.0 | 0.003 |
| 10 | 332.5 | 0.009 | 35.0 | 0.003 |
| 11 | 352.8 | 0.022 | 35.9 | 0.006 |
| 12 | 379.8 | 0.017 | 36.9 | 0.006 |
| 13 | 365.8 | 0.014 | 36.4 | 0.004 |
| 14 | 381.4 | 0.019 | 37.0 | 0.006 |
| 15 | 385.5 | 0.019 | 37.4 | 0.006 |
| 16 | 390.1 | 0.028 | 37.3 | 0.010 |
| 17 | 385.0 | 0.021 | 37.0 | 0.009 |
| 19 | 390.2 | 0.016 | 37.9 | 0.006 |

## 6 - Figures



Figure 1. Distribution of commercial catches of Norwegian spring-spawning herring from October 2022 until February 2023, based on electronic logbooks. Each point represent one catch, only catches larger than 10 tons are shown


Figure 2. Cruise tracks (mostly acoustic transects), pelagic trawl stations (triangles), and CTD stations (Z) covered by Eros and Vendla on the Norwegian spring-spawning herring spawning survey 13.-26. February 2023. Strata limits area also shown.


Figure 3. Acoustic densities (NASC) of herring recorded during the Norwegian spring-spawning herring spawning survey 13.-26. February 2023. Points represent NASC values per nautical mile. Depth contours are shown for $50 \mathrm{~m}, 100 \mathrm{~m}, 150 \mathrm{~m}, 200 \mathrm{~m}, 500 \mathrm{~m}$, $1000 \mathrm{~m}, 1500 \mathrm{~m}$ and 2000 m .

## SPAWNING SURVEY,TSB



Figure 4. Estimates of total biomass from the Norwegian spring-spawning herring spawning surveys during 1988-2023. The estimates are mean of 1000 bootstrap replicates in Stox and the error bars represent $90 \%$ confidence intervals.


Figure 5. Abundance estimates (left axis) and relative standard error (right axis) by age. Black dots are mean of 1000 boostrap replicates in Stox, error bars represent $90 \%$ confidence intervals and triangles relative standard error (CV).





Figure 7. Relative distribution by stratum of the biomass of herring from the Norwegian spring-spawning herring spawning survey $13 .-$ 26. February 2023 (mean of 1000 bootstrap replicates).


Figure 8. Proportions of different maturity stages from the Norwegian spring-spawning herring spawning survey 13.-26. February 2023. The proportions are based on sampled herring in the following latitude intervals: $<68^{\circ} \mathrm{N}, 68-68.5^{\circ} \mathrm{N}, 68.5-69.3^{\circ} \mathrm{N},>69.3^{\circ} \mathrm{N}$.

 sbatroung survegs in 2017-2023

 spatwoing stoveg in 2017.2023


Farose herring catch samples 2022
Mix between spring spawning and summer/autumn spawning herring - highlighting the importance of doing genetic sampling of herring

Samples taken in winter (Jan 2022 and Dec 2022) in the international zone (Lat $>67 \mathrm{~N}$ and Lon $>1 / 2^{\circ} \mathrm{E}$ ) had $6 \%, 16 \%, 34 \%$ and $46 \%$ of hyaline otoliths, respectively. This indicates a relatively high mix of autumn and spring spawners that occupy the central part of the Norwegian Sea in the winter (Dec-Jan).

Hyaline otoliths are also more prevalent in samples taken close to Iceland.


Thus, the fishery close to Iceland in the second semester of the year, and in the central part of the Norwegian Sea in the early winter (which is the quarter when most of the catch is taken), not only targets NSSH, but also other herring stocks.


[^13]
## Working Document <br> Working Group on International Pelagic Surveys <br> January 2024 <br> Working Group on Widely Distributed Stocks

August 2023


# INTERNATIONAL BLUE WHITING SPAWNING STOCK SURVEY (IBWSS) SPRING 2023 

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Survey coordinator

## Material and methods

## Survey planning and Coordination

Coordination of the survey was initiated at the meeting of the Working Group on International Pelagic Surveys (WGIPS) in January 2023 and continued by correspondence until the start of the survey. During the survey, effort was refined and adjusted by the survey coordinator (Norway) using real time observations. Participating vessels together with their effective survey periods are listed below:

| Vessel | Institute | Survey period |
| :--- | :--- | :---: |
| Celtic Explorer | Marine Institute, Ireland | $25 / 3-04 / 04$ |
| Jákup Sverri | Faroe Marine Research Institute, Faroe Islands | $25 / 3-04 / 04$ |
| Tridens | Wageningen Marine Research, the Netherlands | $25 / 3-31 / 03$ |
| Vendla | Institute of Marine Research, Norway | $19 / 3-01 / 04$ |
| Vizconde de Eza | Spanish Institute of Oceanography, Spain | $26 / 3-02 / 04$ |

Survey design was based on methods described in ICES Manual for International Pelagic Surveys (ICES, 2015). Overall, weather conditions were mixed compared to 2022, with poor weather delaying the start of the survey by 4 days in the south for the Irish and Dutch vessels. Further north, further downtime was experienced by some vessels, accounting for slowed progress. The entire survey was completed in 17 days, within the agreed 21-day target threshold (Figure 4). Area coverage was considered comprehensive in both core and peripheral areas, with all vessels completing the planned routes. Spanish survey effort (Strata 1b) was included into the 2023 estimate as temporal alignment was achieved compared to 2022.

Vessel cruise tracks, trawl positions and survey stratification are shown in Figure 1. CTD and plankton stations are in shown in Figure 2. Communication between vessels occurred daily via email to the coordinator (Norway) exchanging up to date information on blue whiting distribution, echograms, fleet activity and biological information. Tridens keeps a weblog during the survey with echograms, catches and additional information.

## Sampling equipment

All vessels employed a single midwater trawl for biological sampling, the properties of which are given in Table 1. Acoustic equipment for data collection and processing are presented in Table 2. Survey abundance estimates are based on acoustic data collected from calibrated scientific echo sounders using an operating frequency of 38 kHz . All transducers were calibrated using a standardised sphere calibration (Demer et al. 2015) prior, during or directly after the survey. Acoustic settings by vessel are summarised in Table 2.

## Biological sampling

All components of the trawl haul catch were sorted and weighed; fish and other taxa were identified to species level where possible. A summary of biological sampling by vessel is provided in Table 3.

## Hydrographic sampling

Hydrographic sampling (vertical CTD casts) was carried out by each vessel at predetermined locations (Figure 3 and Table 3). Depth was capped at a maximum depth of 1000 m in open water, with the exception of the Faroese and Spanish vessels $(500 \mathrm{~m})$.

## Plankton sampling

Plankton sampling, by way of vertical WP2 casts, was carried out by the RV Jákup Sverri (FO) to a depth of 200 m (Table 3). WP2 casts were also carried out by FV Vendla (NO), with a focus on sampling blue whiting eggs to a depth of 400 m .

## Acoustic data processing

Echogram scrutinisation for blue whiting was carried out by experienced personnel, with the aid of trawl composition information. Post-processing software and procedures are described by vessel below;
On RV Celtic Explorer, acoustic data were backed up every 24 hrs and scrutinised using Echoview (V 13.0) post-processing software for the previous day's work. Data was partitioned into the following categories: blue whiting and mesopelagic fish species. For mesopelagic fish, categorisation was based on criteria agreed at WGIPS 2021 (ICES 2021, Annex 22).
On RV Jákup Sverri, acoustic data were scrutinised every 24 hrs on board using LSSS (2.14.1) post processing software. Data was partitioned into the following categories: plankton, mesopelagics/krill and blue whiting. Partitioning of data into the above categories was based on trawl samples and acoustic characteristics on the echograms.

On RV Tridens, acoustic data were backed up continuously and scrutinised every 24 hrs using the Large Scale Survey System LSSS (2.10.0) post-processing software. Blue whiting was identified and separated from other recordings based on trawl catch information and characteristics of the recordings. Recordings have been assigned to blue whiting and mesopelagic fish species, based on the criteria at WGIPS 2021 (ICES 2021, Annex 22).

On FV Vendla, the acoustic recordings were scrutinized using LSSS (V. 2.14.0) once or twice per day. Data was partitioned into the following categories: plankton ( $<120 \mathrm{~m}$ depth layer), mesopelagic species and blue whiting.

On RV Vizconde de Eza, acoustic data were backed up every 24 hrs and scrutinised after the survey using Echoview (V 9.0) post processing software. Data were partitioned into the following categories: Blue whiting and Mueller's pearlside and boarfish which were identified and separated from other recordings based on trawl catch information and characteristics of the recordings.

Acoustic categorisation and targeted biological sampling of mesopelagic fish species is ongoing and will be further refined during future surveys. Progress updates will be reported through WGIPS.

## Acoustic data analvsis

Acoustic data were analysed using the StoX software package (V3.6.0) and R-StoX packages software package (RStoX Framework 3.6.0, RStoX Base 1.11.0 and RStoX Data 1.8.0). A description of StoX software package is provided by Johnsen et. al. (2019). Estimation of abundance from acoustic surveys using StoX is carried out according to the stratified transect design model developed by Jolly and Hampton (1990). Baseline survey strata, established in 2017, were adjusted based on survey effort and observations in 2023 (Figure 1). Area stratification and transect design are shown in Figure 1 and 4. Within StoX, length and weight data from trawl samples were equally weighted and applied across all transects within a given stratum (Figure 4).
Following the decisions made at the Workshop on implementing a new TS relationship for blue whiting abundance estimates (WKTSBLUES, ICES 2012), the following target strength (TS)-to-fish length (L) relationship (Pedersen et al. 2011) is used:

$$
\mathrm{TS}=20 \log 10(\mathrm{~L})-65.2
$$

In StoX an impute super-individual table is produced where abundance is linked to population parameters including age, length, weight, sex, maturity etc. This table is used to split the total abundance estimate by any combination of population parameters. The StoX project folder for 2023 is available on request.

Estimate of relative sampling error
For the baseline run, StoX estimates the number of individuals by length group which are further grouped into population characteristics such as numbers at age and sex.
A total length distribution is calculated, by transect, using all the trawl stations assigned to the individual transects. Conversion from NASC (by transect) to mean density by length group by stratum uses the calculated length distribution and a standard target strength equation with user defined parameters. Thereafter, the mean density by stratum is estimated by using a standard weighted mean function, where each transect density is weighted by transect distance. The number of individuals by stratum is given as the product of stratum area and area density.
The bootstrap procedure to estimate the coefficient of variance randomly replaces transects and trawl stations within a stratum on each successive run. The output of all runs are stored in a RData-file, which is used to calculate the relative sampling error.

## Results

Stock size
The estimated total stock biomass (TSB) of blue whiting for the 2023 international survey was 2.5 million tonnes, representing an abundance of $29.9 \times 10^{9}$ individuals (Table 4). This is an $8 \%$ decrease in total stock biomass and a $5 \%$ decrease in total stock numbers (TSN) from observations in 2022 (Table 4). The spawning stock biomass (SSB) was estimated at 2.3 million tonnes representing $27.5 \times 10^{9}$ individuals (Table 5). This is a $1 \%$ decrease in the observed spawning stock biomass and a $15 \%$ increase in the spawning stock numbers (SSN) compared to last year. Spawning stock biomass, as determined from biological samples, showed a slight decrease in SSB with a moderate increase in SSN driven by recruitment of fish from the strong 2021-year class to the spawning stock.

## Distribution of blue whiting

In total, $8,571 \mathrm{nmi}$ (nautical miles) of survey transects were completed across six strata, relating to an overall geographical coverage of $147,968 \mathrm{nmi}^{2}$ (Figure 1, Tables $3 \& 7$ ), Area coverage increased by $17 \%$ as compared to 2022. Acoustic sampling (transect miles) saw an increase of $47 \%$ compared to 2022 (from $5,812 \mathrm{nmi}$ to $8,571 \mathrm{nmi}$ ). The increase in acoustic sampling effort and area coverage can be accounted for by the below average effort in 2022, due to the early departure of the RV Celtic Explorer and the addition of Spanish survey effort this year in the south. The westward extension of blue whiting observed in 2023 was notable from previous years and required further survey effort to ensure containment in the west.

The stock was considered contained within core and peripheral abundance areas (Rockall Bank, Porcupine Bank and Porcupine Seabight). The distribution of blue whiting, as observed during the survey, is shown in Figures 5 and 6. The main body of the stock occurred between $54^{\circ}$ to $57^{\circ} \mathrm{N}$ and schools in this stratum were observed to span the Rockall Trough from east to west. Blue whiting were observed in moderate numbers on the western flanks of the Rockall Bank (Stratum 5) for the first time in recent years.

Overall, the distribution of blue whiting was found further west into open water than observed in either 2021 or 2022 , most notably in the Rockall Trough. This considered, the highest concentrations still occurred in the east at the continental shelf break in stratum 3.

The stock was distributed within core strata $1-3$, totalling $77.5 \%$ of TSB and $78.8 \%$ of TSN (Table 4). A second area of abundance, somewhat geographically distinct from the main body of the stock, was observed in the northern strata ( $4 \& 6$ ) accounting for $13.8 \%$ of TSB and $12.8 \%$ of TSN and composed of mainly 2, 3, 4 and 1-year-old fish respectively.

The Rockall Trough (stratum 3) contained the highest abundance overall, accounting for $52 \%$ of TSB and $51 \%$ of TSN (Table 4), within this stratum the highest densities of fish were observed between $55^{\circ} \mathrm{N}$ to $58^{\circ} \mathrm{N}$. Compared to 2022 , this stratum saw a moderate increase in both TSB and TSN of $12 \%$. The Porcupine Bank (Strata 1) saw a large decrease in TSB of $67 \%$ compared to 2022 and a corresponding decrease in TSN of $55 \%$, indicating the bulk of the stock had already migrated northwards. The North Porcupine (Strata 2) saw a decrease of $21 \%$ of TSB and $5 \%$ in TSN. An increase of $12 \%$ TSB and TSN was recorded in Stratum 3 (Rockall Trough). The Rockall Bank (Strata 5) saw a marked increase of $1,162 \%$ TSB (from $15,000 \mathrm{t}$ to $183,000 \mathrm{t}$ ) and an increase of $1,677 \%$ TSN as compared to 2022.

Last year the most notable change was the large increase in biomass observed in northern survey area; south Faroes and Faroe/Shetland Channel (Strata $4 \& 6$, respectively). In 2023, these strata also contained a relatively high proportion of the stock ( $13 \%$ ) compared to $20 \%$ in 2022 (Table 4). The south Faroes strata showed a slight decrease in TSB of 4\% (from 193,000 $\mathfrak{t}$ to $186,000 \mathrm{t}$ ) and $6 \%$ in TSN compared to 2022, whereas the Faroe/Shetland Channel saw a decrease of $29 \%$ in TSB (from $226,000 \mathrm{t}$ to $159,000 \mathrm{t}$ ) and $56 \%$ in TSN. Both strata were dominated by 2- and 3-year-old fish, followed by 4 - and 1-year-old fish, respectively.

## Echograms

The highest $\mathrm{s}_{\mathrm{A}}$ value ( $91,787 \mathrm{~m}^{2} / \mathrm{nmi}^{2}$ - per 1 nmi EDSU) observed during the combined survey was recorded by FV Vendla in the North Porcupine stratum (Figure 7a). The second highest density value, was recorded in the Rockall Trough stratum, by the RV Tridens ( 65 , $955 \mathrm{~m}^{2} / \mathrm{nmi}^{2}$ - per 1 nmi EDSU), Figure 7 b . The third highest value ( $56,888 \mathrm{~m}^{2} / \mathrm{nmi}^{2}-$ per 1 nmi EDSU) was recorded by RV Celtic Explorer on the shelf edge in the Rockall Trough stratum (Figure 7c). Blue whiting observations by RV Jákup Sverri (Figure 7d) on the Wywille Thompson ridge (in the Faroe-Shetland Channel stratum) showing a scattering layer of relatively high concentrations at depths between $400-500 \mathrm{~m}$. The only significant concentrations of Blue Whiting observed by RV Vizconde de Eza were in the western part of the South Porcupine Bank at a depth of 400 m . (Figure 7e).

## Stock composition

Survey samples found fish ages from 1 to 15 years ( $10+$ group) during the survey (Table 5 ).
The main contribution to the spawning stock biomass was composed of the age groups 3, 2, 4 and 5 years, respectively. Combined these age cohorts represent $84 \%$ of TSB. In terms of abundance, 3 -year-olds ( 2020 year-class) were most abundant ( $49 \%$ ), followed by the 2 -yearolds ( $27 \%$ ), 4 -year-olds ( $9 \%$ ) and 5 -year-olds ( $5 \%$ ), respectively (Table 5).
The largest mean length value of blue whiting by strata obtained from catches came from Stratum $5(25.5 \mathrm{~cm})$, Figure 8. The largest mean weight came from Stratum 4 and was 96.3 g (Figure 9).
The bulk of the stock was composed of mixed age classes of 1 to 15 years, and dominated by mature individuals aged 3 and 4 (Figure 11). The aggregations in the northern area were dominated by $2 \& 3$-year-old fish with a relatively high proportion of mature individuals
(Figure 11). The abundance of these two year classes in 2023 were the highest in the time series and above the numbers associated with the 2014 record year class (Table 6, Figure 12).

Immature fish represented $5 \%$ of TSB and $7 \%$ of TSN. Over $87 \%$ of the 2-year old fish were mature contributing to the SSB of the stock (Table 5).
The CV of the total estimate of abundance was 0.16 , which is lower than 2022 (2022 $=0.19$ and $2021=0.14$ ).
The survey time series (2004-2023) of TSN and TSB are presented in Figures 13 and 14 respectively and Table 6.

## Hydrography

A total of 102 CTD casts were undertaken over the course of the survey (Table 1). Horizontal plots of temperature and salinity at depths of $50 \mathrm{~m}, 100 \mathrm{~m}, 200 \mathrm{~m}$ and 500 m as derived from vertical CTD casts are displayed in Figures $15-18$, respectively. A decrease in salinity observed in 2017 persisted through 2018 and 2019, but seems to have reversed again in 2020 with an increasing trend (K.M. Larsen, pers. comm., Faroe Marine Research Institute). Pre2020 , this is thought to have limited the western extent of the blue whiting spawning distribution on the Rockall and Hatton Bank areas in recent years. Observations in 2022 and 2023 are in agreement with a reversing trend (in salinity mainly), with a more western extension of fish into the Rockall Trough than observed in recent years, with blue whiting approaching the eastern slopes of the Rockall Plateau in 2023.

## Mesopelagic fish

Echogram scrutinisation for mesopelagic fish species was conducted by participants during the survey and will be uploaded to the ICES database after further analysis. Due to ongoing complexities regarding representative trawl catches these data are considered as experimental and outputs reported to the ICES database should be treated as such.

## Concluding remarks

## Main results

- Weather conditions were mixed and the survey start was delayed by 4 days in the southern area (Ireland \& Netherlands)
- The total area surveyed and acoustic sampling effort increased from 2022 (17\% and 48\% respectively) and was required to contain the stock in the west and from contributions for the Spanish survey in the south
- In terms of biological sampling effort, the number of trawl stations was reduced due to the mechanical breakdown on the RV CE that prevented any fishing for the duration of the survey. However, RV Tridens undertook trawling on some aggregations identified by the RV CE. The total number of fish aged and measured was comparable to 2022. The stock was considered representatively sampled in number, and across the distribution area.
- The International Blue Whiting Spawning Stock Survey 2023 shows an $8 \%$ decrease in TSB and a corresponding $5 \%$ decrease in TSN when compared to the 2022 estimate for comparable survey effort and coverage.
- In terms of abundance, 3-year-olds (2020 year-class) were most abundant (49\%), followed by the 2 -year-olds ( $27 \%$ ), 4 -year-olds ( $9 \%$ ) and 5 -year-olds ( $5 \%$ ), respectively.
- Immature fish represented $5 \%$ of TSB and $7 \%$ of TSN. Over $87 \%$ of the 2-year old fish were mature contributing to the SSB of the stock.
- The abundance of the two year classes, 2 \& 3-year-old fish, in 2023 were the highest in the time series and above the numbers observed associated with the previous 2014 record year class.
- Estimated uncertainty around the total stock abundance was $\mathrm{CV}=0.16$ (CV=0.19 in 2022).
- The survey was carried out over 17 days, below the 21-day time window target. With core areas representatively sampled by multiple vessels.


## Interpretation of the results

- The group considers the 2023 estimate of abundance as robust. Good stock containment was achieved for both core and peripheral strata.
- Temporal progression was miss aligned due to poor weather delaying the vessels starting in the south leading to a mismatch timing in core regions. However, the group considers this as not ideal but acceptable nonetheless
- The bulk of SSB was distributed from northern Porcupine Bank northwards in the Rockall Trough from $54^{\circ} \mathrm{N}$ to $57^{\circ} \mathrm{N}$.
- Contribution of the 2020-year class remains significant as the largest and strongest year class in the time series, surpassing the previous in 2014. This year class is now considered fully recruited to the spawning stock.


## Recommendations

- The group recommends that coverage in the western Rockall/Hatton Bank (stratum 5) should be carried out based on real time observations. Stock size and distribution in recent years combined with hydrographic conditions would indicate a period of westward expansion that requires monitoring
- To facilitate the process of calculating global biomass the group requires that all data be made available at least 72 hours in advance of the meeting start date and made available through the ICES database.
- Hydrographic and Plankton data along with Log book files formats should still be submitted in the PGNAPES format.
- The group recommends that the process of producing output reporting tables, figures and maps from StoX outputs files (StoX) are standardised in R code for consistency of reporting and replication.
- It is recommended that the effective timing of the survey starting point is maintained to begin around the $20^{\text {th }}$ March in 2024.
- Faroes and Spain collect CTD data to $1,000 \mathrm{~m}$ in line with the other participants


## Achievements

- Good stock containment within the survey area, with comprehensive trawl and biological sampling achieved.
- All survey data were uploaded to the ICES trawl-acoustic database in advance of the post cruise meeting, with the exception of the Spanish data, where corrected data was provided 7 days after the meeting.
- Survey area covered completed within 17 days.


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Table 1. Country and vessel specific details, IBWSS March-April 2023.

|  | Celtic <br> Explorer | Jákup <br> Sverri | Tridens | Vendla | Vizconde <br> de Eza |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Trawl dimensions | 768 | 832 | 860 | 832 | 752 |
| Circumference (m) | 50 | 45 | $30-70$ | 45 | 30 |
| Vertical opening (m) | 20 | 45 | 40 | 40 | 20 |
| Mesh size in codend (mm) | 3.5 | 3.4 | $3.5-4.0$ | $3.5-4.0$ | $3.5-4.0$ |
| Typical towing speed (kts) |  |  |  |  |  |
| Plankton sampling | - | Wlankton | - | plankton |  |
| Sampling net |  | net |  | net |  |
| Standard sampling depth (m) | - |  |  | 400 |  |
| Hydrographic sampling |  |  |  |  |  |
| CTD Unit |  |  |  |  |  |
| SBE911 | SBE911 | SBE911 | SBE25 | SBE25Plus |  |
| Standard sampling depth (m) | 1000 | 500 | 1000 | 1000 | 1000 |

Table 2. Acoustic instruments and settings for the primary acoustic sampling frequency, IBWSS March-April 2023.

|  | Celtic Explorer | Jákup Sverri | Tridens | Vendla | Vizconde de Eza |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Echo sounder | Simrad | Simrad | Simrad | Simrad | Simrad |
| Echo sounder | EK 60 | EK80 | EK 80 | EK 80 | EK 80 |
| Frequency (kHz) | 38, 18, 120 , | 18, 38, 70, | $18, \mathbf{3 8}, 70$ | 18, 38, 70 | 38, 18,70 , |
| Frequency (kHz) | 200 | 120,200, 333 | 120,200, 333 | 18, 38, 70 | 120, 200 |
| Primary transducer | ES 38B | 38-7 | ES 38B | ES 38B | ES 38-7 |
| Transducer installation | Drop keel | Drop keel | Drop keel | Drop keel | Drop keel |
| Transducer depth (m) | 8.8 | 6 | 8 | 8.5 | 5 |
| Upper integration limit (m) | 20 | 15 | 15 | 15 | 23.91 |
| Absorption coeff. (dB/km) | 9.4 | 10 | 9.5 | 9.5 | 9.4 |
| Pulse length (ms) | 1.024 | 1.024 | 1.024 | 1.024 | 1.024 |
| Band width (kHz) | 2.43 | 3.06 | 2.43 | 2.43 | 3.06 |
| Transmitter power (W) | 2000 | 2000 | 2000 | 2000 | 2000 |
| Angle sensitivity (dB) | 21.9 | 21.9 | 21.9 | 21.9 | 21.9 |
| 2-way beam angle ( dB ) | -20.6 | -20.4 | -20.6 | -20.7 | -20.7 |
| Ts Transducer gain (dB) | 25.86 | 26.87 | 27.27 | 25.28 | 26.84 |
| $\mathrm{S}_{\mathrm{A}}$ correction ( dB ) | -0.66 | -0.1377 | -0.01 | -0.66 | -0.246 |
| 3 dB beam width (dg) |  |  |  |  |  |
| alongship: | 6.78 | 6.44 | 6.86 | 6.89 | 6.51 |
| athw. ship: | 6.91 | 6.40 | 6.89 | 6.85 | 6.6 |
| Maximum range (m) | 1000 | 750 | 750 | 750 | 1000 |
| Post processing software | Echoview | LSSS | LSSS | LSSS | Echoview |

Table 3. Survey effort by vessel, IBWSS March-April 2023. RV Tridens carried out 4 trawl hauls on the transects of RV Celtic Explorer.

| Vessel | Effective <br> survey period | Length of <br> cruise track <br> $(\mathrm{nmi})$ | Trawl <br> stations | CTD <br> stations | Mesopelagic <br> sampling | Aged <br> fish | Length- <br> measured <br> fish |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Celtic Explorer^ | $24 / 03-04 / 04$ | 2279 | - | 29 | - | - | - |
| Jákup Sverri | $25 / 03-04 / 04$ | 1688 | 8 | 26 | - | 548 | 1931 |
| Vendla | $19 / 03-01 / 04$ | 2558 | 18 | 31 | - | 527 | 1723 |
| Tridens | $25 / 03-31 / 03$ | 793 | 13 | 16 | 2 | 1200 | 3499 |
| Vizconde de Eza* | $26 / 03-02 / 04$ | 1253 | 3 | - | - | 90 | 645 |
| Total | $24 / 03-04 / 04$ | 8571 | 42 | 102 | 2 | 2365 | 7798 |

$\wedge$ No trawl sampling carried out by Ireland due to mechanical failure of the net Sonde winch *CTD data from the RV Vizconde de Eza not available due to failure of the equipment.

Table 4 Abundance and biomass estimates of blue whiting by strata in 2022 and 2021. IBWSS March-April 2023.

| Strata | Name | 3023 |  |  |  | 2022 |  |  |  | 1)ifference 2023-2022 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\begin{gathered} \mathrm{TSIS} \\ \left(1 \mathcal{B}^{\prime} 1\right) \end{gathered}$ | $\begin{aligned} & \text { TSN } \\ & \left(10^{6}\right) \\ & \hline \end{aligned}$ | \% TS3 | \% TSN | $\begin{array}{r} \text { TS }{ }^{13} \\ \left(10^{3} 1\right) \\ \hline \end{array}$ | $\begin{aligned} & \text { TSN } \\ & \left(10^{6}\right) \\ & \hline \end{aligned}$ | \% TSI3 | \#TSN | TS13 | TSN |
| 1 | Porcupine Bank | 170 | 2:116 | 6.8 | 7.1 | 510 | 4,714 | 18.9 | 15.0 | -67\% | -55\% |
| 2 | -2 Porcupine Bark | 475 | 6,156 | 19.0 | 20.6 | 599 | 6,469 | 221 | 20.6 | -21\% | -5\%, |
| 3 | Rockall Trough | 1,295 | 15,261 | 51.8 | 51.1 | 1,151 | 13,672 | 42.5 | 43.5 | 12\% | 12\% |
| 4 | South l'aroes | 186 | 1,970 | 7.4 | 6.5 | 193 | 2.042 | 7.1 | 6.5 | -4\%\% | -6\% |
| 5 | Ruckall Eankis | 18.3 | 2,083 | 7.3 | 7.0 | 15 | 117 | 0.5 | 0.4 | 1.16.2\% | 1.677\% |
| 6 | Faroeshetland Ch . | 159 | 1,897 | 6.4 | 6.3 | 226 | 4,276 | 83 | 13.6 | -29\% | - $56 \%$ |
| 7 | Porcupinc Scabigh | 33 | 439 | 1.3 | 1.5 | 13 | 151 | 0.5 | 0.5 | 163\%\% | 190\%\% |
|  | Italal | 2,501 | 20,883 | 100 | 100 | 2,707 | 31,442 | 100 | 100 | -8\% | -50\% |

Table 5. Survey stock estimate of blue whiting (determined from StoX baseline output), IBWSS March-April 2023.

| $\begin{aligned} & \text { Length } \\ & (\mathrm{cm}) \\ & \hline \end{aligned}$ | Age in years (year class) |  |  |  |  |  |  |  |  |  | Number$\left(10^{\wedge} 6\right)$ | Biomass$\left(10^{\wedge} 6 \mathrm{~kg}\right)$ | Mean weight (g) | $\begin{array}{\|l\|} \hline \text { Prop } \\ \text { Mature } \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | $10+$ |  |  |  |  |
|  | 2021 | 2020 | 2019 | 2018 | 2017 | 2016 | 2015 | 2014 | 2013 |  |  |  |  |  |
| 14-15 |  |  |  |  |  |  |  |  |  |  | 0 | 0 | 0.0 | 0 |
| 15-16 |  |  |  |  |  |  |  |  |  |  | 0 | 0 | 0.0 | 0 |
| 16-17 | 1 |  |  |  |  |  |  |  |  |  | 1 | 0 | 26.0 | 0 |
| 17-18 | 30 |  |  |  |  |  |  |  |  |  | 30 | 1 | 25.2 | 0 |
| 18-19 | 148 |  |  |  |  |  |  |  |  |  | 148 | 4 | 29.4 | 31 |
| 19-20 | 305 | 48 |  |  |  |  |  |  |  |  | 354 | 13 | 36.3 | 35 |
| 20-21 | 242 | 70 |  |  |  |  |  |  |  |  | 312 | 14 | 43.7 | 23 |
| 21-22 | 148 | 551 | 78 |  |  |  |  |  |  |  | 777 | 39 | 50.3 | 71 |
| 22.23 |  | 2,364 | 218 | 9 |  |  |  |  |  |  | 2,591 | 148 | 57.0 | 91 |
| 23-24 |  | 3,135 | 1,349 | 14 |  |  |  |  |  |  | 4,497 | 285 | 63.3 | 93 |
| 24-25 |  | 1,504 | 3,932 | 253 |  |  |  |  |  |  | 5,690 | 394 | 69.3 | 95 |
| 25-26 |  | 328 | 5,063 | 410 |  |  |  |  |  |  | 5,801 | 444 | 76.5 | 93 |
| 26-27 |  | 110 | 2,627 | 527 | 119 |  |  |  |  |  | 3,383 | 292 | 86.4 | 94 |
| 27-28 |  | 25 | 1,014 | 560 | 169 |  |  |  |  |  | 1,768 | 178 | 100.9 | 96 |
| 28-29 |  |  | 297 | 439 | 155 | 24 | 107 | 20 | 30 |  | 1,071 | 125 | 116.6 | 97 |
| 29-30 |  |  | 127 | 231 | 222 | 140 | 42 | 8 | 39 |  | 810 | 108 | 133.9 | 100 |
| 30-31 |  |  | 43 | 151 | 453 | 190 | 70 | 39 | 10 |  | 955 | 142 | 148.6 | 100 |
| 31.32 |  |  | 12 | 120 | 123 | 82 | 80 | 28 | 91 |  | 536 | 86 | 161.3 | 97 |
| 32-33 |  |  | 12 | 8 | 51 | 172 | 158 | 16 | 159 |  | 575 | 98 | 170.9 | 100 |
| 33-34 |  |  |  | 20 | 46 | 33 | 49 | 83 | 51 |  | 281 | 52 | 184.3 | 99 |
| 34-35 |  |  |  |  | 14 | 28 | 8 | 9 | 94 |  | 153 | 35 | 228.9 | 93 |
| 35-36 |  |  |  |  |  | 19 |  |  | 13 | 29 | 61 | 14 | 234.9 | 100 |
| 36-37 |  |  |  |  |  |  |  |  | 13 | 13 | 25 | 6 | 240.0 | 100 |
| 37.38 |  |  |  |  |  | 8 |  |  | 8 |  | 16 | 6 | 348.5 | 100 |
| 38-39 |  |  |  |  |  | 16 |  |  |  |  | 16 | 5 | 283.5 | 100 |
| 39-40 |  |  |  |  |  |  |  |  |  | 13 | 13 | 4 | 352.0 | 100 |
| 40-41 |  |  |  |  |  |  | 6 |  |  | 13 | 19 | 7 | 379.0 | 100 |
| 41-42 |  |  |  |  |  |  |  |  |  |  | 0 | 0 | 0.0 | 100 |
| 42-43 |  |  |  |  |  |  |  |  |  |  |  | 0 | 0.0 | 100 |
| 43-44 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 44.45 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| TSN(mill) | 873 | 8,135 | 14,771 | 2,744 | 1,352 | 711 | 520 | 202 | 508 | 67 | 29,883 |  |  |  |
| TSB(1000t) | 33.6 | 517.4 | 1,143.0 | 265.9 | 190.3 | 117.8 | 84.8 | 31.1 | 96.5 | 20.6 | 2,501.0 |  |  |  |
| Mean length(cm | 19.4 | 22.9 | 24.9 | 27.0 | 29.2 | 31.0 | 30.5 | 31.3 | 32.0 |  |  |  |  |  |
| Mean weight(g) | 39 | 63 | 77 | 103 | 138 | 172 | 158 | 159 | 183 |  |  |  |  |  |
| \% Mature | 23 | 87 | 96 | 97 | 99 | 97 | 99 | 100 | 99 | 100 |  |  |  |  |
| SSB (1000 t) | 7.6 | 452.5 | 1098.2 | 256.6 | 188.7 | 114.0 | 84.3 | 31.1 | 95.7 | 20.6 | 2,349.3 |  |  |  |
| SSN (mill) | 198 | 7,115 | 14,192 | 2,648 | 1,341 | 688 | 517 | 202 | 503 | 67 | 27,470.4 |  |  |  |

Table 6. Time series of StoX abundance estimates of blue whiting (millions) by age in the IBWSS, 2023. Total biomass in last column (1000 t). Note: * indicates survey excluded or not undertaken.

| Age |  |  |  |  |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | $910+$ | ГSB $(1000 \mathrm{t})$ |  |
| 2004 | 1,097 | 5,538 | 13,062 | 15,134 | 5,119 | 1,086 | 994 | 593 | 164 | 3,505 |  |
| 2005 | 2,129 | 1,413 | 5,601 | 7,780 | 8,500 | 2,925 | 632 | 280 | 129 | 23 | 2,513 |
| 2006 | 2,512 | 2,222 | 10,858 | 11,677 | 4,713 | 2,717 | 923 | 352 | 198 | 31 | 3,512 |
| 2007 | 468 | 706 | 5,241 | 11,244 | 8,437 | 3,155 | 1,110 | 456 | 123 | 58 | 3,274 |
| 2008 | 337 | 523 | 1,451 | 6,642 | 6,722 | 3,869 | 1,715 | 1,028 | 269 | 284 | 2,639 |
| 2009 | 275 | 329 | 360 | 1,292 | 3,739 | 3,457 | 1,636 | 587 | 250 | 162 | 1,599 |
| $2010^{*}$ |  |  |  |  |  |  |  |  |  |  |  |
| 2011 | 312 | 1,361 | 1,135 | 930 | 1,043 | 1,712 | 2,170 | 2,422 | 1,298 | 250 | 1,826 |
| 2012 | 1,141 | 1,818 | 6,464 | 1,022 | 596 | 1,420 | 2,231 | 1,785 | 1,256 | 1,022 | 2,355 |
| 2013 | 586 | 1,346 | 6,183 | 7,197 | 2,933 | 1,280 | 1,306 | 1,396 | 927 | 1,670 | 3,107 |
| 2014 | 4,183 | 1,491 | 5,239 | 8,420 | 10,202 | 2,754 | 772 | 577 | 899 | 1,585 | 3,337 |
| 2015 | 3,255 | 4,565 | 1,888 | 3,630 | 1,792 | 465 | 173 | 108 | 206 | 247 | 1,403 |
| 2016 | 2,745 | 7,893 | 10,164 | 6,274 | 4,687 | 1,539 | 413 | 133 | 235 | 256 | 2,873 |
| 2017 | 275 | 2,180 | 15,939 | 10,196 | 3,621 | 1,711 | 900 | 75 | 66 | 144 | 3,135 |
| 2018 | 836 | 628 | 6,615 | 21,490 | 7,692 | 2,187 | 755 | 188 | 72 | 144 | 4,035 |
| 2019 | 1,129 | 1,169 | 3,468 | 9,590 | 16,979 | 3,434 | 484 | 513 | 99 | 144 | 4,198 |
| $2020^{*}$ |  |  |  |  |  |  |  |  |  |  |  |
| 2021 | 1,948 | 2,095 | 2,545 | 2,275 | 3,914 | 3,197 | 3,379 | 463 | 189 | 114 | 2,357 |
| 2022 | 4,461 | 9,313 | 4,830 | 5,460 | 2,587 | 1,880 | 898 | 1,764 | 71 | 178 | 2,707 |
| 2023 | 873 | 8,135 | 14,771 | 2,744 | 1,352 | 711 | 520 | 202 | 508 | 67 | 2,501 |

Table 7. IBWSS survey effort time series.

| Survey <br> effort | Survey <br> area <br> $\left(\right.$ nmi $\left.^{2}\right)$ | Transect <br> n. miles <br> $(\mathrm{nmi})$ | Trawls | CTDs | Plankton | Measured | Aged |
| :---: | :---: | :---: | ---: | ---: | ---: | ---: | ---: |
| 2004 | 149000 |  | 76 | 196 |  |  |  |
| 2005 | 172000 | 12385 | 111 | 248 | - | 29935 | 4623 |
| 2006 | 170000 | 10393 | 95 | 201 | - | 7211 | 2731 |
| 2007 | 135000 | 6455 | 52 | 92 |  | 5367 | 2037 |
| 2008 | 127000 | 9173 | 68 | 161 | - | 10045 | 3636 |
| 2009 | 133900 | 9798 | 78 | 160 | - | 11460 | 3265 |
| 2010 | 109320 | 9015 | 62 | 174 | - | 8057 | 2617 |
| 2011 | 68851 | 6470 | 52 | 140 | 16 | 3810 | 1794 |
| 2012 | 88746 | 8629 | 69 | 150 | 47 | 8597 | 3194 |
| 2013 | 87895 | 7456 | 44 | 130 | 21 | 7044 | 3004 |
| 2014 | 125319 | 8231 | 52 | 167 | 59 | 7728 | 3292 |
| 2015 | 123840 | 7436 | 48 | 139 | 39 | 8037 | 2423 |
| $2016^{*}$ | 134429 | 6257 | 45 | 110 | 47 | 5390 | 2441 |
| 2017 | 135085 | 6105 | 46 | 100 | 33 | 5269 | 2477 |
| 2018 | 128030 | 7296 | 49 | 101 | 45 | 5315 | 2619 |
| 2019 | 121397 | 7610 | 38 | 118 | 17 | 6228 | 1938 |
| 2021 | 118169 | 7794 | 45 | 102 | 8 | 12019 | 2089 |
| $2022^{\wedge}$ | 126235 | 5812 | 47 | 99 | 57 | 6499 | 2372 |
| 2023 | 147,968 | 8571 | 42 | 102 | 54 | 7798 | 2365 |

[^14]

Figure 1. Strata, cruise tracks and trawl hauls for the individual vessels (country) during the International Blue Whiting Spawning Stock Survey (IBWSS) from March-April 2023. Faroe Islands (RV Jákup Sverri); Ireland (RV Celtic Explorer); Netherlands (RV Tridens); Norway (FV Vendla); Spain (RV Vizconde de Lza).


Figure 2. Vessel cruise tracks with hydrographic CTD stations (z) and WP2 plankton net samples (circles) during the International Blue Whiting Spawning Stock Survey (IBWSS) from March-April 2023.


Figure 3. Temporal progression for the International Blue Whiting Spawning Stock Survey (IBWSS) from March-April 2023


Figure 4. Tagged acoustic transects (green circles) with associated trawl stations containing blue whiting (dark blue squares) used in the StoX abundance estimation. IBWSS March-April 2023.


Figure 5. Acoustic density heat map ( $\mathrm{s}_{\mathrm{A}} \mathrm{m}^{2} / \mathrm{nmi}^{2}$ ) of blue whiting during the International Blue Whiting Spawning Stock Survey (IBWSS) from March-April 2023.


Figure 6. Map of proportional acoustic density $\left(\mathrm{s}_{\wedge} \mathrm{m}^{2} / \mathrm{nmi}^{2}\right)$ of blue whiting by 1 nmi sampling unit. IBWSS March-April 2023.

a) Highest density blue whiting per Inmi log interval $(91,787)$ recorded in during the IBWSS survey in the North Porcupine area (Strata 2) FV Vendla, Norway.

b) Single highest density blue whiting layer ( $\mathrm{s}_{\mathrm{A}}$ value $65955 \mathrm{~m}^{2} / \mathrm{nmi}^{2}$ ) by 1 nmi recorded by the RV Tridens at the shelf edge at $54.56 \mathrm{~N}-10.25 \mathrm{~W}$ (Stratum 3).

c) High density blue whiting layer per 1 nmi $\log$ interval $(56,888)$ observed during the survey recorded by the RV Celtic Explorer on the shelf margin in the southern Rockall Trough (Stratum 3) in 300-400 m.

d) Acoustic registrations ( 38 kHz ) of bluc whiting with the Farocse Jákup Sverrion on 30 March, 2023 on the Wyville Thomson ridge.

e) Blue Whiting concentration recorded by RV Vizconde de Eza in the western shelf edge of South Porcupine Bank on 26 March 2023 (Stratum 1b)

Figure 7. Echograms of interest encountered during the IBWSS, March-April 2023. Vertical banding represents 1 nmi acoustic sampling intervals (EDSU). All echograms presented at 38 kHz .


Figure 8. Combined mean length of blue whiting from trawl catches by vessel, IBWSS in March- April 2023


Figure 9. Combined mean weight of blue whiting from trawl catches, IBWSS March- April 2023


Figure 10. Blue whiting bootstrap abundance (millions) by age (left axis) and associated CVs (right axis) in 2021 (top panel), 2022 (middle panel) and 2023 (lower panel). From StoX.


Figure 11. Length and age distribution (numbers) of blue whiting by survey strata. March-April 2023.


Figure 12. Length and age distribution (numbers) of total stock of blue whiting. March-April 2023.

IBWSS,TSN


Figure 13. Time series of StoX survey indices of blue whiting abundance, 2004-2023, excluding 2010 and 2020.

IBWSS,TSB


Figure 14. Time series of StoX survey indices of blue whiting biomass, 2004-2023, excluding 2010 and 2020.


Figure 15. Horizontal temperature (left panel) and salinity (right panel) at 50 m subsurface as derived from vertical CTD casts. IBWSS March-April 2023.


Figure 16. Horizontal temperature (left panel) and salinity (right panel) at 100 m subsurface as derived from vertical CTD casts. IBWSS March-April 2023.


Figure 17. Horizontal temperature (left panel) and salinity (right panel) at 200 m subsurface as derived from vertical CTD casts. IBWSS March-April 2023.


Figure 18. Horizontal temperature (left panel) and salinity (right panel) at 500 m subsurface as derived from vertical CTD casts. IBWSS March-April 2023.

# Analysis of Mackerel assessment 2023 

Höskuldur Björnsson

2023-08-28

The Muppet model was run as comparison with the SAM assessment conducted at WGWIDE. There are a number of differences between the setup, one important difference is the lack of process error that plays a large role in the official assessment as the data are relatively poor and inconsistent.
The main differences in setup are.

1. Selection in Muppet pattern fixed for 2 periods, 1980-1998 and 1999-2022. Selection pattern in SAM variable. Selection pattern of the models is similar and variability in SAM not large (figure 5).
2. F for ages 9 and older is constant, 7 and older in SAM (figure 5).
3. Tagdata used are from all tagging experiments 2011-2021 and recaptures 2012-2022. Recaptures in the tagging year not included. Tagloss estimated.
4 The tagging data have the tendency to take over the assessment. The number data points are 1320 and grow quite rapidly with more years. In retrospective runs the tagging data have quickly less weight.
4. One of the Muppet runs includes the recruitment index that is estimated poor by the Muppet model (CV estimated 0.75-1.1 in different runs but 0.26 in the sam run).
5. In muppet, $q$ in the Pelagic survey is constrained to be the same for ages 7 and older. In the SAM run it is constrained to be the same for ages 10 and 11 (10 and older). Estimated Q in the survey does not change much after age 7 so this difference is probably not important.
6. In Muppet, CV of residuals in survey and catch is specified as a multiplier on a pattern given but in SAM the CV is constrained to be identical for a large range of age groups.
7. In Muppet $\mathrm{p} M$ and pF proportions of F and M before spawning are constant in time. Those values vary with time in the adopted SAM assessment. Therefore SSB is not comparable.
8. The treatment of catch data before 1998 is different from the adopted asssessment. In the Muppet run catch in numbers by age are used in the traditional way before 1998 but a multiplier on the catches estimated. In different runs the value of the multiplier was estimated as $1.25-1.38$ with standard error of $0.16-0.20$. Interpreting this factor as misreporting means misreporting of $25-40 \%$. Considerable difference in results from Muppet and SAM before 1998 is therefore expected.


Biomass of age 3 and older was calculated for both models to solve the problem of SSB not being comparable, the values of B3+ and SSB are though similar. Looking at the Muppet results (figure 1), the weight of tagging data has most effect on the estimated biomass. The run labelled "SSB+Pelsur+lowwtTags" is based on weight of 0.1 for the tagging data but the number of data points in the tag-recapture file is 1320 ,certainly not 1320 independent data points. 119 observations are used for the RFID tags in SAM and the weight of 0.1 fits reasonably well to get same effective number of data points.
One interesting difference between Muppet and Sam results is that average recruitment at age 3 is much less in SAM (figure 2), still the biomass is similar. It has always been known that number at ages 0 and 1 were questionable, if looked at from the "fixed M world", but that difference should not be seen so strong at age 3 . The difference is really large enough that the results should be investigated to see if there was something wrong with the data in either model. Muppet is just a simple catch at age model that where fish can only be lost by catches and $\mathrm{M}=0.15$. SAM has more ways of loosing or adding fish.
Recruitment at age 0 is remarkably stable in SAM (figure 3). This is to be able to fit the recruitment survey that is estimated good $(C V \approx 0.26)$. Also the the parameter " $\operatorname{logSdLogN}$ " is -1.6372523 for age 0 and -1.5756247 for ages $1-12$, very unconventual as the most of the uncertainty in recruitment is usually taken up by the first parameter.
There is as shown above considerable uncertainty about the stock size of Mackerel. The tagging data indicate that the stock might be decreasing, we are getting too many tags back compared to recent years. The recruitment index leads to too optimistic recruitment estimates but it has too much weight in the adopted runs. Looking at the 4 alternative Muppet runs the advice for 2024 would be 407, 541, 769 and 872 thous tonnes, quite some difference between runs. $F_{m s y}$ from the SAM assessment gives advice of 740 thous tonnes. The Muppet run with weight of tag data closest to what it is in the SAM assessment gives an advice of 769 thous. tonnes based on $F=0.26$. The advice is surprisingly similar taking into account how different stock in numbers by age are (figures $6, ? ?, 3$ and 2 ).


Figure 1: Biomass of age 3 and older for various setup of Muppet and the adopted SAM assessment


Figure 2: Recruitment at age 3 for various setup of Muppet and the adopted SAM assessment


Figure 3: Recruitment at age 0 for various setup of Muppet and the adopted SAM assessment


Figure 4: F4-8 of Muppet and the adopted SAM assessment


Figure 5: Selection pattern in Muppet and SAM. Values from Muppet are labelled as range of years.


Figure 6: Stock in numbers 2023 for various setup of the Muppet model and the adopted SAM assessment


Retrospective pattern
It has been claimed that the Mackerel assessment is always an underestimate. This has probably been true, the reason being that tuning data were extremely scarce. IESNSS started in 2010, 2011 is not used but has been conducted every year since 2012. The run shown here uses age 3-11 in the survey but it would be better also to include the plus group (age 12+).

Relatively much change is seen between 2017 and 2018 but after that the stock is over and underestimated in this model run. F as estimated now was relatively low in the period after IESNSS started leading to slow convergence of estimated $q$.



Looking at observed and predicted survey biomass (figure 7) 2019 and 2020 are outliers. What is happening there? 2022 is the survey with really high CV. Interesting with this picture is that the survey biomass and $B 4+$ are about the same $(q=1)$.

Taking a simple stupid approach from this picture B4+ now should be similar to what it was 2010 and 2012 Average from 2010 and 2012 is 3900 tons but 3700 tons in 2023 so this works ???!!!


Figure 7: Observed and predicted biomass from IESNSS and B4+ from assessment based on the survey. B4+ is shown as age 3 is only half recruited to this survey

## Analysis of blue whiting assessment 2023

Höskuldur Björnsson

2023-08-27

The Muppet model was run as comparison with the SAM assessment conducted at WGWIDE. There are a number of differences between the setup but the results are generally more consistent than for example for the Mackerel.
The main differences in setup are

1. Selection in Muppet pattern fixed for 2 periods, 1980-1996 and 1997-2022. Selection pattern in SAM variable. Selection pattern of the models is similar and variability in SAM not large (figure 7).
2. F for ages 7 and older is constant, 9 and older in SAM (figure 7 ).
3. 2 Versions of Muppet are run, one using only Tac constraint in the assessment year (Muppet)but the other using catch in numbers in the assement year (Muppetw2023)
4. In Muppet, CV of residuals in survey and catch is specified as an estimate multiplier on a pattern given, but in SAM the CV is constrained to be identical for specified range of age groups. The setting of those patterns affects the results but were not investigated much this time. The pattern of survey CV was set using the Muppet model in VPA mode, estimating the survey CV for each age group (can in principle be done by a VPA model)
5. Fishing mortality is modelled by random walk in SAM but not in Muppet. The random walk limits varibility in $F$ and does lead to larger estimated stock when $F$ is increasing and vice versa. The random walk constraint is probably weak in the blue withing stock where F has been quite variable $(\operatorname{logSdLogFsta}=-1.03353539$ and corFlag=2 $)$
All the models show similar trends in the stock unlike what was seen for Mackerel. There is though difference in the estimate of current stocksize with the SAM model giving indicating larger stock than the Muppet models (figures 1 and 3). The Muppet model using catch at age 2023 indicates smaller stock than the one using TAC constraint in 2023.

The models agree on the amount of older fish (figure 2) but considerable difference is seen regarding yearclasses 2020 and 2021 (figure 8).

The Muppet model includes prediction and the TAC for 2024 based on $F=0.32$ is 960 thous tonnes for the model using catch at age 2023 and 1216 thous. tonnes for the model not using catch at age 2023. Comparable number for the adopted model is around 1500 thous. tonnes.
In the Muppet model the first guess of each yearclass is geometric mean (or SSB recruitment function). CV around this mean is estimated and this term can have substantial effect on estimate of yearclasses far from the geometric mean, how much depends on the quality of the survey data. This term is downweighted in the runs here (weight $=0.1$ ) and does not have large effects.

In summary, considerable difference can be seen by different models and settings. Most of the stock is from 2 yearclasses 2020 and 2021 and the weighting by age of survey and and catchdata might have considerable effect on the results.


Figure 1: SSB from Muppet and from the adopted SAM assessment


Figure 2: Biomass of age 4 and older from Muppet and the adopted SAM assessment


Figure 3: Biomass of age 3 and older fromMuppet and the adopted SAM assessment


Figure 4: Recruitment at age 3 from Muppet and the adopted SAM assessment


Figure 5: Recruitment at age 1 from Muppet and the adopted SAM assessment


Figure 6: F4-8 from Muppet and the adopted SAM assessment


Figure 7: Selection pattern from Muppet and SAM. Values from Muppet are labelled as range of years (2 selection periods)


Figure 8: Stock in numbers 2023 from the Muppet model and the adopted SAM assessment


Figure 9: Catch in numbers 2022 from the Muppet model and the adopted SAM assessment


Figure 10: Biomass by age 2023 from the Muppet model and the adopted SAM assessment

# Working Group on Widely Distributed Stocks (WGWIDE) 23-29 August 2023, Copenhagen (Denmark) 

# Identification of new reference fleets regarding striped red mullet from professional fishing data for Subareas 7 and 8. 

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## Context

For striped red mullet (Mullus surmuletus), there is no scientific campaign to provide one or more robust abundance indices for the mur-west stock. One of the objectives of the ACOST project is therefore to construct, and in some cases consolidate, abundance indices based on commercial fishing data. Classically, two approaches can be developed to build such indices:

- the "design-based" approach, based on the selection of representative statistical units (in this case, vessels), from which data are then collected to build an abundance index;
- the "model-based" approach, which, on the other hand, studies the variability of catch data and reworks them to construct such an index.

The design-based approach was initiated during the ROMELIGO project (Léauté et al., 2018). It is based on a data-filtering method that examines the activity of fishing fleets that are potential candidates to be retained to provide an index of abundance through mandatory and optional criteria. These criteria were defined and discussed throughout the project in working groups involving the fishing industry. At the end of the ROMELIGO project, five reference fleets were proposed because their Landings Per Unit Effort (LPUE) were considered to be potentially able to account for the level of abundance of striped red mullet in the Bay of Biscay (ICES Subarea 8) (Léauté et al., 2018 with publication of the method in Caill-Milly et al., 2019). These fleets were as follows:

$$
\begin{aligned}
& \text { - « OTB» fleet - cluster } 1 \text { - mesh size class } 70-79 \mathrm{~mm} \text {; } \\
& \text { - « GNS» fleet - cluster } 2 \text { - mesh size class 50-59 mm - 2nd quarter; } \\
& \text { - « GNS » fleet - cluster } 2 \text { - mesh size class 50-59 mm-3rd quarter; } \\
& \text { - «GNS» fleet - cluster } 2 \text { - mesh size class 60-69 mm-2nd quarter; } \\
& \text { - «GNS» fleet - cluster } 2 \text { - mesh size class over } 90 \mathrm{~mm} \text { - 2nd quarter. }
\end{aligned}
$$

Since the ROMELIGO project, their CPUE has been updated and transmitted to ICES WGWIDE. Of the five original fleets, there are now four reference fleets, as the second one (« GNS » fleet - cluster 2-mesh size class 50-59 mm-2nd quarter) has been withdrawn as its use reached a very low level (around 40 fishing trips in 2018).

Within the framework of the ACOST project and for this ICES Subarea 8, the present work aims to consider the interest of a new fleet using Danish seines. The latter were not considered before because the time series was too short. The interest of an extent of the methodology
to ICES Subarea 7 and/or ICES Subareas 7 and 8 for OTB, GNS and SDN is also viewed to take into account the entire geographical area delimiting the mur-west stock.
NB: the ACOST project also deploys a model-based approach based on the standardization of LPUE using statistical models. It has been used for whiting and pollack, and is currently being adapted for red mullet, to be communicated to WGWIDE at a later date.

## This WD refers only to new results obtained with the design-based approach

## 1. Data and approach

### 1.1. Data origin

The data used relate to commercial fishing catches and are of two types: data regarding landings (SACROIS data) and data regarding discards (OBSMER data). SACROIS data have been available since 2000; they were extracted for the period 2000-2022. For the area corresponding to the spatial delimitation of the mur-west stock, $3,263,982$ fishing trips were related to striped red mullet catch (each line containing the quantity of red mullet landed per fishing trips and per vessel). It therefore represents a quantification of landings per unit effort (LPUE). OBSMER data have been available since 2003; they were extracted for the period 2003-2021. To use the OBSMER data, the following criteria were considered: the availability of a sufficient number of observations, the proportion of discards mean by month and by year from the number of total operations realized and observed, the stability of this proportion over time, and the size composition of the discards. Regarding size, there is no minimum size in force for striped red mullet, but a minimum commercial weight set at 40 grams, equivalent to individuals of around 18 cm . For the area corresponding to the spatial delimitation of the mur-west stock, 6,504 fishing operations were observed entirety (landing part, discard part, all species). Their breakdown by Subarea and by gear is provided in Table 1.

Table 1: Number of fishing operations observed according to the Subareas and gears considered over the 2003-2021 period.

|  | OTB | GNS | SDN |
| :--- | :--- | :--- | :--- |
| Subarea 7 | 3453 | 383 | 353 |
| Subarea 8 | 1001 | 941 | 373 |

Regarding the vessels, their technical characteristics may change over time as a result of modifications (such as remotorization, changes in fishing gear, etc.). In order to provide the most accurate information throughout a vessel's life cycle, the annual characteristics were collected from the Community fishing fleet register (https://webgate.ec.europa.eu/fleeteuropa/search_en).

### 1.2. Selection, cleaning and data preparation

Three gears were selected using their FAO code: "OTB" for the bottom otter trawls (1 vessel), "GNS" for the set gillnets and "SDN" for the Danish seines. As OBSMER data for OTB and GNS in Subarea 8 has already been processed by Léauté et al. (2018), these data were not considered in this report.

### 1.2.1. Data extraction and definition of the study population

Differences in activity for the three gears exist between Subarea 7 and Subarea 8 (seasonality, average vessel length, year of first use of gear, etc.). Analyses by gear are therefore initially undertaken separately for these two Subareas:

- For Subarea 7, the spatial coverage selection corresponds to ICES divisions "27.7.a", "27.7.b", "27.7.c", "27.7.e", "27.7.f", "27.7.g", "27.7.h", "27.7.j" and "27.7.k";
- For Subarea 8, it corresponds to ICES divisions "27.8.a", "27.8.b", "27.8.c", "27.8.d" and "27.8.e".
The reference fleet selection for the construction of abundance indices consists in selecting vessels that will enable us to pinpoint the phenomenon of interest, namely striped red mullet catches. Therefore, vessels were selected in these 2 subareas, according to 2 criteria:
- Criterion 1: those having caught striped red mullet at least once between 2000 and 2020 for each Subarea;
- Criterion 2: only statistical rectangles consistent with the bathymetric distribution of striped red mullet were selected. They correspond to rectangles with bathymetries of less than 300 m and/or an intersection with this depth (except for rectangles 28E8 and 29E8, which have been excluded because they have a surface almost entirely in "27.7.d" and not in "27.7.e").
Vessels were then selected on the basis of activity, namely a realization of at least 24 fishing trips per year with striped red mullet, using trawls, nets or Danish seines respectively. At the end of this first stage, 757 vessels formed the statistical population considered.


### 1.2.2. Data cleaning and final data set preparation

The "cleaning" related only to the landing data. It included detected outliers following expert opinions (atypical values of landings, unrealistic values of gear mesh sizes), data corresponding to vessels with fishing time equal to zero or not filled in and non-available data (fishing sequences without statistical rectangles information). The deletion also affected lines describing fishing sequences whose occurrence per month was less than 2 (as they cannot be used to consider the variability of landings for a given year, month, rectangle and gear) and the sequences for the year 2020 due to the health crisis linked to Covid19 and its impact on fishing activity (this year deemed too atypical to be considered for the fleet selection process). For landings, to conduct the analyses, level indicators (mean, median) and dispersal indicators (interquartile range IQR, variance Var, standard deviation StD, and coefficient of variation CV) of the LPUE were calculated and displayed in columns using the following aggregation: year, month, rectangle, month, vessel registration, vessel length, vessel power, vessel tonnage, gear and gear mesh size (grouped in mesh class).
For discards, the calculated column parameters were the mean, the median and the standard deviation of the catches and the discards. The aggregation used was: year, month, rectangle and gear.

### 1.3. Analytical process of landing data to select a representative fleet

The approach was described by Caill-Milly et al (2019); only the main outlines are given hereafter.
Preliminary stage: a focus on discards is presented first due to the nature of the professional data available (related to landings). A condition for the indicators constructed to reflect abundance at the time of capture, it is first necessary to ensure that the loss of information linked to discards between catch and landing is controlled, or at least acceptable. To this end,
an analysis of OBSMER data was carried out for the pre-selected fleets; note that this consideration of discards is not always made in the literature
Once this stage has been completed, the approach developed comprises four steps (Figure 1):

- a focus on the variability of LPUE and a hierarchization of causes (spatio-temporal variables, vessel and gear characteristics);
- a clustering to obtain a typology of vessels based on the technical characteristics of the vessels that emerge as the main explanatory factors for the variability of LPUE (excluding spatio-temporal variables);
- analysis of average LPUE per cluster. To identify clusters (or fleets) particularly useful for providing an overview of abundance, a multi-criteria approach with mandatory and optional conditions was applied (Table 2); it sets conditions for the level, variability and duration of activity and is applied for each cluster. Then for these conditions, a points system was developed. Points were attributed by expert opinion on a scale of 3 ( 1 - null or weak; 2 medium; 3 - strong);

Table 2: Mandatory and optional conditions applied to identify the cluster of interest

| Mandatory conditions | Optional conditions |
| :--- | :--- |
| - a sufficient number of vessels constituting | - a seasonal signal stable in amplitude and |
| the cluster, set at a minimum of 10 for OTB, | periodicity; |
| GNS and SDN; | - activity north and south of the Bay of Biscay |
| - presence in the subarea for a long period. It | for Subarea 8, in 7e-k for Subarea 7; |
| is considered here that the series must be | - moderate seasonal variability. |
| available for a period longer than the |  |
| maximum longevity known for striped red |  |
| mullet, i.e. 12 years; |  |
| - a sufficient average level of LPUE. The value |  |
| for this criterion varies according to the gear. |  |
| It is set at 5 kg for nets and trawls, and $\mathbf{1 0 ~ k g ~}$ |  |
| for Danish seines, based on an examination of |  |
| LPUE values. |  |

- Once selected, fleets were refined using gear mesh sizes or mesh classes, and possibly seasonal (quarterly) activity considerations.

NB : the analyses carried out to identify new reference fleets were started in 2021; they were therefore conducted over the 2005-2019 period for trawl and net, and over the 2007-2019 or 2008-2019 period for Danish seine (2020 having been excluded due to COVID). Since the start of these analyses, other years have become available: 2021 and 2022. The question of their impact on the results if they had been included in the analyses was therefore raised at the WG Reference Fleet associating the professional structures. Steps 1 to 2 were therefore reworked to take them into account. The results are similar to the previous ones in terms of fleet selection. This does not justify repeating all the graphs and tables presented (because of the time it would take). However, for the fleets selected, the "number of uses" and "average LPUE" indicators were updated using the latest SACROIS database available, which includes 2021 and 2022.


Figure 1: Retained methodology for the identification of reference fleets from the LPUE study (from Caill-Mily, 2023)

### 1.4. Engagement of organizations representing professional fishermen

As with the ROMELIGO project, exchanges with professional representatives have been organized throughout the ACOST project. These took place in Working Groups (WG) "Reference Fleets", attended by scientists, representatives of producers' organizations, fisheries committees and Aglia.... These groups enable scientists to present the data used and the method(s) selected, to discuss the relevance of thresholds and aggregations, to detect possible biases linked to changes in regulations (from local to European level), etc. in the construction of indices. Appendix 1 details the progress of the WGs regarding striped red mullet. The approach adopted can be described as collaborative research. The nature of the collaboration with professional representatives is presented in Table 3, given that the entire project is not yet complete.

Table 3: Detailed participation of the professional stakeholders in the various stages of the ACOST research project by $08 / 31 / 2023$, using the modalities described by Macher et al. (2022)

| Nature of the collaboration | Realization <br> ACOST <br> framework by <br> $\mathbf{3 1 / 0 8 / 2 0 2 3 ~}$ |  |
| :--- | :---: | :--- |
| Stakeholder participation in data analysis |  |  |
| Stakeholder participation in project management |  |  |
| Stakeholder participation in interpreting results | X |  |
| Stakeholder participation in financing project | X |  |
| Stakeholder participation in disseminating results | (X) | Planned but not yet <br> implemented as project in <br> progress |
| Co-construction of research questions | X |  |
| Stakeholder participation in data collection | X | By auto-sampling ... |
| Stakeholder consultation/knowledge gathering | X |  |
| Knowledge transfer from science to stakeholders | (X) | Planned but not yet <br> implemented as project in <br> progress |

In terms of stakeholder engagement ${ }^{1}$, ICES has recently produced a document describing the key principles of engagement in the work of the council and defining the roles of the professional parties as well as those of scientists (ICES, 2023a). In addition, ICES organized a workshop in May 2023 dedicated to the implementation of this stakeholder engagement strategy (WKSTIMP). This provided an opportunity to deepen some definitions and propose actions (ICES, 2023b). With reference to these two documents, and as far as the ACOST project is concerned, the reason for stakeholder involvement can be defined by the term "knowledge production", with stakeholders playing a "contributor" role. The last document also notes the importance of effective communication and learning for successful engagement. At the end of the ACOST project, feedback will be provided to professional structures to identify possible improvements for future projects.

[^15]
## 2. Results for the data-filtering approach

For this WD and for the fleets studied (gear/Subarea combination) but not retained, the reasons for their exclusion are presented but not detailed. For each fleet selected, the presentation of results begins with the vessel typology. This typology is based on the technical characteristics that have the greatest influence on the variability of LPUE. Steps 2 to 4 are therefore presented and detailed. The figures and tables regarding the preliminary stage and step 1 are available in the appendixes 2,3 and 4 . This choice was made because all the stages have already been presented in a WD transmitted and presented to the WGWIDE in 2018 (Caill-Milly et al., 2018) for other fleets.

### 2.1. Applied to vessels using Danish seine (SDN) in Subareas 8 and 7

In previous work, the length of the Danish seine data series was too limited. This is no longer the case, so this document aims to consider the interest of the Danish seine metier in defining, or not, additional reference fleet(s).

## For Subarea 8

Spatio-temporal factors are more important than technical characteristics in accounting for the variability of LPUE (Appendix 2). Once these factors have been removed, technical characteristics of the gear and vessels are highlighted. Regarding vessel characteristics and for SDN Subarea 8, they relate to gauge and length.

Vessel typology using technical characteristics and evolution of average LPUE for each cluster Danish seine (SDN)
Implementation of Agglomerative Hierarchical Clustering (AHC) on the technical characteristics (gauge and vessel length) allows distinguishing two clusters (figure 2; characteristics displayed in table 4).


Figure 2: AHC of vessels using SDN in Subarea 8 according to their technical characteristics - gauge and length (standardized) (red line: visual cut)

Table 4 : Values of technical characteristics per cluster for vessels using SDN in Subarea 8

| Cluster |  | Vessel length (m) |  |  | Gauge (grt) |  |  | Engine power (kW) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Code | Number | Min | Mean | Max | Min | Mean | Max | Min | Mean | Max |
| $\mathbf{1}$ | 3 | 34.0 | 34.0 | 34.0 | 274 | 274 | 274 | 588 | 588 | 588 |
| $\mathbf{2}$ | 17 | 18.0 | 22.0 | 24.9 | 88 | 149 | 227 | 331 | 459 | 626 |
| Total | 20 |  |  |  |  |  |  |  |  |  |

For each candidate cluster, changes in landing average calculated by month and by year are shown in Figure 3. Cluster 1 displays highly variable activity (with a very high peak in 2009) at the start of the period, with a stop from 2014 onwards. Cluster 2 is only interrupted at the end of 2009/beginning of 2010, and shows a relative cyclicality in the LPUE levels.


Figure 3: Average LPUEs of striped red mullet per cluster of vessels using SDN in Subarea 8

Within the area studied in the Bay of Biscay, an additional North (8a) / South (8b) division was applied. This division is made at latitude $46^{\circ} 0^{\prime} \mathrm{N}$. In the northern part of the Bay of Biscay (delimitation $46^{\circ} 0^{\prime} \mathrm{N}$ ) (Figure 4) and for each cluster, the average LPUE is very similar to that observed for the entire Bay.


Figure 4: Average LPUEs of striped red mullet per cluster of vessels using SDN in the northern part of Subarea 8

In the southern Bay of Biscay (Figure 5) and for both clusters, the average LPUE is close to that observed for the whole Bay, but there are differences in levels for both clusters at some periods compared with Figure 3. Moreover, the series are discontinuous for both clusters.


Figure 5: Average LPUEs of striped red mullet per cluster of vessels using SDN in the southern part of Subarea 8

LPUE levels differ according to the technical characteristics of the vessels. Among all the series observed, the northern Bay of Biscay shows the most cyclical and uninterrupted trend in average catch levels for Cluster 2.

Cluster selection with a multicriteria approach
Table 5 summarizes the results following the application of the madatory and optional conditions on the data series.

Table 5: Classification of clusters for SDN in Subarea 8

| Multi-criteria selection method |  |  |  |
| :--- | :--- | :---: | :---: |
| Level of <br> obligation | Vessel typology (technical characteristics) | Cluster 1 | Cluster 2 |
| Mandatory | Sufficient number of vessels (> 10) |  | X |
|  | Long data series (> 12 years) |  | X |
|  | LPUE levels medium to high (> 10 kg/UE) over the period | $\mathrm{X}(133)$ | $\mathrm{X}(77)$ |
| Optional | Stable seasonal signal (both in amplitude and <br> periodicity) during the series | $1 / 3$ | $2 / 3$ |
|  | Activity present in N and S of the Bay of Biscay | $1 / 3$ | $2 / 3$ |
|  | Moderate seasonal variability | $1 / 3$ | $2 / 3$ |
| Notation |  |  |  |
| Proposed ranking |  | $3 / 9$ | $6 / 9$ |

The points system allows us to propose cluster 2 for Danish sein (SDN) in Subarea 8. This cluster 2 is composed of medium vessels with a mean length of 22 m (between 18.0 et 24.9 m ), with a gauge between 88 and 227 grt , and an engine power between 331 and 626 kW .

The question of spatial selection was posed for Cluster 2 . In order not to reduce the number of observations, and because the signals between the northern and southern Bay of Biscay were fairly similar, we chose to keep the whole Subarea 8 for this fleet. The following analyses therefore refer solely to this cluster.

Study of mesh sizes and seasonal variations in LPUE for Danish seine Cluster 2 (SDN)
Mesh sizes were considered according to several criteria: representativeness of the landing levels, continuity of use and a sufficient number of uses. The 70 mm mesh size is by far the most represented class for this cluster over the period (Figure 6).


Figure 6: Number of mesh uses between 2008 and 2019 for Cluster 2 - Danish seine (SDN) - Subarea 8

Ffigure 7 shows the monthly evolution of LPUE for the three main meshes of cluster $2(70 \mathrm{~mm}, 80$ mm and 100 mm ).


Figure 7: Monthly evolution of LPUE for the main mesh sizes used by Cluster 2 - Danish seine (SDN) - Subarea 8

For the criteria considered (Table 6), the 70 mm mesh size for the Danish seine (SDN) - cluster 2 is seen as the most interesting. This result, presented at the WG Reference Fleets, was discussed and it was decided to extend it to the $70-80 \mathrm{~mm}$ class. There are regulatory reasons for this, as some vessels previously working with 70 mm have already been required to use 80 mm in certain months as part of the management of the Bay of Biscay common sole stock ${ }^{2}$. The $70-80 \mathrm{~mm}$ class is therefore preferred, as it will make it possible to anticipate other possible 80 mm obligations for these vessels, and thus facilitate the construction of a long-term series.

Table 6: Mesh size results for Cluster 2 - Danish seine (SDN) - Subarea 8

|  | Proposed mesh size selection(s) |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Métier | Danish seine |  |  |  |  |  |
| Cluster | 2 |  |  |  |  |  |
| Mesh size (mm) | 70 | 75 | 80 | 90 | 100 | 110 |
| Sufficient LPUE level | X | X | X | X | X | X |
| Availability over a long period | X |  | X |  | X |  |
| High number of uses | X |  |  |  |  |  |
| Limited confidence interval | (X) | $\mathrm{X})$ | $(\mathrm{X})$ |  | $(\mathrm{X})$ |  |
| Proposed mesh size | X |  |  |  |  |  |
| Quarterly study | Yes |  |  |  |  |  |

For the 70-80 mm mesh class for SDN cluster 2, LPUE displays strong seasonality over the period, with higher levels in May. The confidence interval (CI) is also widest in May (Figure 8).


Figure 8: Average LPUE (with CI) per month over the period 2010-2019 for Cluster 2 - Danish Seine for mesh size class $70-80 \mathrm{~mm}$ - Subarea 8

Based on these characteristics, working on quarter 2 is proposed for this fleet.

[^16]Number of uses and LPUE levels per year
For cluster 2-SDN with a 70-80 mm gear mesh size, the evolution of its use over time and of the LPUEs for the whole Bay of Biscay are considered.

The number of uses has been rising since 2012. It has reached a level of around 1,000 sequences per year since 2015, with 2021 marking the highest value in the series, with 1,356 sequences. The first LPUEs appear in 2008, with an interruption in 2010 (they only related to one vessel in 2008 and 2009). In 2011, average LPUE is high (around 270 kg/fishing trip). They fall until 2014 (67 kg/fishing trip), then rise between 2014 and 2016, then fall until 2018, before rebounding in 2019 to $210 \mathrm{~kg} /$ fishing trip. The last two years, 2021 and 2022, average around $100 \mathrm{~kg} /$ fishing trip (Figure 9).


Figure 9: Number of uses and associated average LPUE levels for the Bay of Biscay - Danish seine - cluster 2 mesh size class $70-80 \mathrm{~mm}$ - Quarter 2

## For Subarea 7

Implementation of Agglomerative Hierarchical Clustering (AHC) on the technical characteristics accounting for the variability of LPUE (power engine) allows distinguishing two clusters (Figure 10). However, neither of these 2 clusters is made up of a sufficient number of vessels ( 5 and 2 respectively). In addition, the length of the LPUE series for the 2 clusters is insufficient (data not shown). There is therefore no Danish seine reference fleet proposed for Subarea 7.


Figure 10: AHC of vessels using SDN in Subarea 7 according to their technical characteristics - engine power (standardized) (red line: visual cut)

## Summary

The data-filtering method applied on vessels using Danish seine vessels allows to propose a reference fleet for the Bay of Biscay (Subarea 8). This fleet is composed of vessels using SDN with a mesh size of $70-80 \mathrm{~mm}$ in quarter 2 and having an average length of 22.0 m (between 18.0 and 24.9 m ), an average gauge of 149 grt (between 88 and 227 grt ) and an average engine power of 459 kW (between 331 and 626 kW ).

No reference fleet is proposed for SDN in Subarea 7.

### 2.2. Applied to vessels using otter bottom trawls (OTB) in Subarea 7

Vessel typology using technical characteristics and evolution of average LPUE for each cluster - otter bottom trawls (OTB)
Implementation of Agglomerative Hierarchical Clustering (AHC) on the technical characteristics accounting for the variability of LPUE (vessel length, gauge and power engine) allows distinguishing three clusters (figure 11; characteristics displayed in Table 7).


Figure 11: AHC of vessels using OTB in Subarea 7 according to their technical characteristics - vessel length, gauge and engine power (standardized) (red line: visual cut)

Table 7: Values of technical characteristics per cluster for vessels using OTB in Subarea 7

| Cluster |  | Vessel length (m) |  |  | Gauge (grt) |  |  | Engine power (kW) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Code | Number | Min | Mean | Max | Min | Mean | Max | Min | Mean | Max |
| $\mathbf{1}$ | 203 | 9.00 | 12.5 | 17.6 | 4 | 27 | 92 | 71 | 177 | 350 |
| $\mathbf{2}$ | 27 | 28.4 | 33.4 | 38.0 | 179 | 271 | 453 | 459 | 625 | 883 |
| $\mathbf{3}$ | 181 | 17.5 | 22.5 | 26.1 | 83 | 146 | 226 | 305 | 463 | 738 |
| Total | 411 |  |  |  |  |  |  |  |  |  |

For each candidate cluster, changes in landing average calculated by month and by year are shown in Figure 12. The three clusters display highly variable activity. Cluster 1 shows a low level of CPUE, as does Cluster 2, with additional breaks in the data series from 2013 onwards. Cluster 3 shows low activity, generally around the months of April, May and June.


Figure 12: Average LPUEs of striped red mullet per cluster of vessels using OTB in Subarea 7

The catches distribution by statistical rectangle for Cluster 3 over the period 2005-2019 presents a concentration of catches in 7 E , with four rectangles accounting for $67 \%$ of catches (Figure 13). $37 \%$ of catches are even located on the eastern border of the mur-west stock. Under these conditions, the indicator that could be constructed from this fleet does not seem to us to adequately reflect the stock under consideration (mur-west). An interaction with the mur.27.3a47d stock is likely.


Figure 13: Distribution (in percentages) of trawl catches (OTB) - cluster 3 - between 2005 and 2019 in Subarea 7

## Summary

No indicator of striped red mullet abundance from vessels using bottom otter trawls in Subarea 7 is proposed. In addition, the Brexit has repercussions on the activity both in terms of the number of licenses and access to specific statistical rectangles. This fleet is too heavily impacted by regulatory aspects to be retained as a reference fleet.

### 2.3. Applied to vessels using set gillnet (GNS) in Subarea 7

Vessel typology using technical characteristics and evolution of average LPUE for each cluster -set gillnets (GNS)
Implementation of Agglomerative Hierarchical Clustering (AHC) on the technical characteristics accounting for the variability of LPUE (vessel length and gauge) allows distinguishing three clusters (Figure 14; characteristics displayed in Table 8).


Figure 14: AHC of vessels using GNS in Subarea 7 according to their technical characteristics - vessel length and gauge (standardized) (red line: visual cut)

Table 8: Values of technical characteristics per cluster for vessels using GNS in Subarea 7

| Cluster |  | Vessel length (m) |  |  | Gauge (grt) |  |  | Engine power (kW) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Code | Number | Min | Mean | Max | Min | Mean | Max | Min | Mean | Max |
| $\mathbf{1}$ | 103 | 5.2 | 7.8 | 10.2 | 0.5 | 3.8 | 8.6 | 7 | 72 | 177 |
| $\mathbf{2}$ | 21 | 10.0 | 11.2 | 13.1 | 7.3 | 13.4 | 31.0 | 61 | 141 | 275 |
| $\mathbf{3}$ | 4 | 15.4 | 16.8 | 19.4 | 59.4 | 72.6 | 105.0 | 191 | 272 | 336 |
| Total | 128 |  |  |  |  |  |  |  |  |  |

For each candidate cluster, changes in landing average calculated by month and by year are shown in Figure 15. Clusters 1 and 2 display relatively similar activity from 2015 onwards, while at the start of the period, Cluster 1 is marked by occasionally high levels. There is almost no data for cluster 3.


Figure 15: Average LPUEs of striped red mullet per cluster of vessels using GNS in Subarea 7

The catches distribution by statistical rectangle for Clusters 1 and 2 over the period 2005-2019 presents a concentration of catches in 7E at the Brittany extremity, with three rectangles accounting for $90 \%$ of catches (Figure 16). These statistical rectangles extend the statistical rectangles that concentrate a large proportion of the Bay of Biscay's catches.


Figure 16: Distribution (in percentages) of gillnetters (GNS) - clusters 1 and 2 - between 2005 and 2019 in Subarea 7

## Summary

No indicator of striped red mullet abundance from vessels using gillnets in Subarea 7 is proposed. Instead, it might be worthwhile reworking these data by integrating them with those for Subarea 8, in order to build a "gillnet" indicator for Subareas 7-8.

### 2.4. Applied to vessels using set gillnet (GNS) in Subareas 7-8

Vessel typology using technical characteristics and evolution of average LPUE for each cluster -set gillnets (GNS)
Implementation of Agglomerative Hierarchical Clustering (AHC) on the technical characteristics accounting for the variability of LPUE (vessel length, gauge and engine power) allows distinguishing three clusters (figure 17; characteristics displayed in Table 9).


Figure 17: AHC of vessels using GNS in Subareas 7-8 according to their technical characteristics - vessel length, gauge and engine power (standardized) (red line: visual cut)

Table 9: Values of technical characteristics per cluster for vessels using GNS in Subareas 7-8

| Cluster |  | Vessel length (m) |  |  | Gauge (grt) |  |  | Engine power (kW) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Code | Number | Min | Mean | Max | Min | Mean | Max | Min | Mean | Max |
| $\mathbf{1}$ | 253 | 4.2 | 7.7 | 10.2 | 0.4 | 3.7 | 22.3 | 4 | 63 | 124 |
| $\mathbf{2}$ | 168 | 7.5 | 10.8 | 14.8 | 2.0 | 11.4 | 31.0 | 70 | 151 | 525 |
| $\mathbf{3}$ | 50 | 15.1 | 19.3 | 30.3 | 32.4 | 104.4 | 195.0 | 152 | 310 | 558 |
| Total | 471 |  |  |  |  |  |  |  |  |  |

For each candidate cluster, changes in landing average calculated by month and by year are shown in Figure 18. Clusters 1 and 2 show relatively similar activity, with cluster 2 generally showing higher values. Cluster 3 is characterized by breaks in the data series, with high peaks in the first half of the period followed by low levels in the second half. The highest levels are mainly observed around May or June, although this seems less evident in the second half of the period.


Figure 18: Average LPUEs of striped red mullet per cluster of vessels using GNS in Subareas 7-8

In the Subarea 8 (Figure 19) and for each cluster, the average LPUE is very similar to that observed for the entire zone comprising Subareas 7 and 8.


Figure 19: Average LPUEs of striped red mullet per cluster of vessels using GNS in Subarea 8

In the Subarea 7 (Figure 20), and for clusters 1 and 2 overall, seasonality is much less marked than in zone 8 . The beginning of the period is marked by strong variations; the second period seems a little more synchronized in terms of seasonality for clusters 1 and 2 . There is very little data for cluster 3 in this subarea.


Figure 20: Average LPUEs of striped red mullet per cluster of vessels using GNS in Subarea 7

LPUE levels differ according to the technical characteristics of the vessels. The trend in average catch levels is the most cyclical, with no interruption for clusters 1 and 2 for the entire zone comprising Subareas 7 and 8 .

Cluster selection with a multicriteria approach
Ttable 10 summarizes the results following the application of the madatory and optional conditions on the data series.

Table 10: Classification of clusters for GNS in Subareas 7-8

| Multi-criteria selection method |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: |
| Level of obligation | Vessel typology (technical characteristics) | Cluster 1 | Cluster 2 | Cluster 3 |
| Mandatory | Sufficient number of vessels (> 10) | X | X | X |
|  | Long data series (> 12 years) | X | X |  |
|  | LPUE levels medium to high (> $5 \mathrm{~kg} / \mathrm{UE}$ ) over the period | X (19) | X (29) | (28) |
| Optional | Stable seasonal signal (both in amplitude and periodicity) during the series | 1/3 | 2/3 |  |
|  | Activity present in N and S of the zone comprising Subareas 7 and 8 | 2/3 | 2/3 |  |
|  | Moderate seasonal variability | 2/3 | 1/3 |  |
| Note |  | 5/9 | 5/9 |  |
| Intra engin |  | 1 | 1 |  |

The points system ranks clusters 1 and 2 ex aequo. Compared with the average LPUE of cluster 1 in Subarea 7, those of cluster 2 in Subarea 7 seem to show seasonal trends closer to those observed for Subareas 7-8 as a whole. In addition, discussions in the Reference Fleets WG also highlighted the fact that Cluster 1 comprises small, polyvalent vessels (which are therefore likely to use the
line or other gear as well). These factors lead us to choose Cluster 2 for GNS. The vessels in this cluster have an average size of 10.8 m (between 7.5 and 14.8 m ), an average gauge of 11.4 grt (between 2.0 and 31.0 grt ) and an average power of 151 kW (between 70 and 525 kW ). The following analyses therefore refer solely to this cluster.

Study of mesh sizes and seasonal variations in LPUE for Gillnet Cluster 2 (GNS)
Mesh sizes were considered according to several criteria: representativeness of the landing levels, continuity of use and a sufficient number of uses. The mesh size class greater than 90 mm is the most used over the period, although it is mainly the 100 mm mesh size (Figure 21). The other mesh size classes heavily used are $50-59 \mathrm{~mm}$ and $60-69 \mathrm{~mm}$.


Figure 21: Number of mesh uses between 2005 and 2019 for Cluster 2 - Gillenet (GNS) - Subareas 7-8

Figure 22 shows the monthly evolution of LPUE for the three main meshes of cluster 2 ( $50-59$ $\mathrm{mm}, 60-69 \mathrm{~mm}$ and above 90 mm ).


Figure 22: Monthly evolution of LPUE for the main mesh sizes used by Cluster 2 - Gillnet (GNS) - Subareas 7-8

For the criteria considered (Table 11), the three mesh size classes 50-59, 60-69 mm and above 90 mm are seen as interesting for gillnet. However, the presentation of these results to the WG Reference Fleets highlights the need to handle information on the class above 90 mm with care. In this case, there may be a problem with the accuracy of declarations for striped red mullet. For a single gear declared in the logbook (e.g. GNS with 100 mm mesh), some 50 mm mesh nappes targeting striped red mullet may have been used.

Table 11: Mesh size results for Cluster 2 - Gillnet (GNS) - Subareas 7-8

|  | Proposed mesh size selection(s) |  |  |
| :--- | :---: | :---: | :---: |
| Métier | Gillnet |  |  |
| Cluster | 2 |  |  |
| Mesh size (mm) | $50-59$ | $60-69$ | Sup 90 |
| Sufficient LPUE level | X | X | X |
| Availability over a long period | X | X | X |
| High number of uses | X | X | X |
| Limited confidence interval |  | $\mathrm{X})$ | (X) |
| Proposed mesh size | Oui | Oui | Oui |
| Quarterly study | T3 | T 2 | T 2 |

For the $50-59 \mathrm{~mm}$ mesh class for GNS cluster 2, LPUE displays low LPUE in the first quarter then higher. The highest levels are observed in September and October. Confidence intervals are particularly wide for many months, particularly in May (Figure 23).


Figure 23: Average LPUE (with Cl ) per month over the period 2005-2019 for Cluster 2 - Gillnet for mesh size class $50-59 \mathrm{~mm}$ - Subareas 7-8

The 60-69 mm mesh size class displays high LPUE from April to June. Confidence intervals around the mean are widest between April and June (Figure 24).


Figure 24: Average LPUE (with CI) per month over the period 2005-2019 for Cluster 2 - Gillnet for mesh size class $60-69 \mathrm{~mm}$ - Subareas 7-8

Mesh sizes larger than 90 mm have high LPUE in April, May, June and July. Confidence intervals around the mean are widest between May and July (Figure 25).


Figure 25: Average LPUE (with Cl ) per month over the period 2005-2019 for Cluster 2 - Gillnet for mesh size class above 90 mm - Subareas 7-8

For gillnets (GNS), a selection of quarters appears necessary: quarter 3 for the $50-59 \mathrm{~mm}$ mesh class (quarter 2 having particularly wide confidence intervals), quarter 2 for the $60-69 \mathrm{~mm}$ and over 90 mm mesh classes.

Number of uses and LPUE levels per year
For cluster 2 - GNS with a $50-59 \mathrm{~mm}, 60-69 \mathrm{~mm}$ and above 90 mm gear mesh size classes, and for selected quarters, the evolution of their respective uses over time and of the LPUEs for the Subareas 7-8 are considered.

The number of uses of the 50-59 mm mesh class in quarter 3 for GNS cluster 2 remains at a fair level of uses (around 250) up to 2019, then presents a lower level. The level of average LPUE does not display any significant trend over the period (Figure 26).

The number of uses of the 60-69 mm mesh class in quarter 2 for GNS cluster 2 shows a downward trend over the period. The level of average LPUE is marked by low levels in 2014, 2015, 2016 and 2018. The highest levels are observed at the beginning of the period (Figure 27).

The number of uses of the mesh sizes above 90 mm in quarter 2 for GNS cluster 2 also displays a downward trend over the period. The level of average LPUE is again marked by low levels in 2014, 2015, 2016 and 2018. The highest levels are observed at the beginning of the period (Figure 28). It should also be remembered that discussions at the WG Reference Fleets alerted us to the precautions to be taken when using data for this mesh size and for striped red mullet (see p21 above).


Figure 26: Number of uses and associated average LPUE levels for Subareas 7-8 - Gillnet - cluster 2 - mesh size class $50-59 \mathrm{~mm}$ - Quarter 3


Figure 27: Number of uses and associated average LPUE levels for Subareas 7-8 - Gillnet - cluster 2 - mesh size class $60-69 \mathrm{~mm}$ - Quarter 2


Figure 28: Number of uses and associated average LPUE levels for Subareas 7-8-Gillnet - cluster 2 - mesh sizes above 90 mm - Quarter 2

## Summary

The downward trend in the number of uses observed for all the mesh size classes (recently for $50-59 \mathrm{~mm}$, all along the period for $60-69 \mathrm{~mm}$ and above 90 mm ), and the above alert for mesh sizes above 90 mm , lead us to be very careful if using a GNS reference fleet. The first one, sub-cluster GNS - cluster 2 , using mesh sizes $50-59 \mathrm{~mm}$ in quarter 3 , may be used if its number of uses increases again.

## 3. Discussion of the relevance of applying the method to the area comprising Subareas 7 and 8 for Danish seine, trawl and gillnet.

The use of the Danish seine appears too different between Subareas 7 and 8 (vessel length, seasonality, etc.) to be relevant to define a single reference fleet for Subareas 7 and 8. Moreover, in Subarea 8, activity is fairly homogeneously distributed throughout the zone, whereas in Subarea 7, significant activity is found at the eastern limit of the stock's spatial extent.

For vessels using OTB, there is no interest in reworking Subarea 7 data by integrating them with OTB data from Subarea 8 to build a "bottom otter trawl" indicator for Subareas 7-8. The reason for this is that a significant proportion of catches come from the eastern limit of the mur-west stock boundary zone, and interactions with the mur.27.3a47d stock cannot be excluded.

For vessels using GNS, there was an interest in reworking the data by integrating those from Subarea 7 with those from Subarea 8 to build a "gillnet" Subareas $7-8$ indicator. The reason for this is that the catches in Subarea 7 are highly localized, and the statistical rectangles in question extend the statistical rectangles that account for a large proportion of the catch in the Bay of Biscay. This work has led us to propose a subfleet for GNS in Subareas 7-8 even if attention must be paid to its use.

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Appendixes

## Appendix 1

Details of COPIL and GT ACOST reference fleets, with focus on striped red mullet

| Date | Designation | Items presented/discussed | Decisions |
| :---: | :---: | :---: | :---: |
| 11/03/2021 | COPIL 1 | * Scheduled approach for the ACOST project (data, tools, etc.) on striped red mullet <br> * Dedicated WG Reference Fleets announced |  |
| 22/11/2021 | WG Reference Fleets 1 | * Detection of outlier data for trawls, gillnets and Danish seines in Subareas 7 and 8 <br> * OBSMER SDN data in Subareas 7 and 8 <br> * Update of LPUE for OTB and GNS fleets identified during the ROMELIGO project <br> * Application of the "ROMELIGO" method to Danish seines in Subareas 8 and 7, interest in an index for Subareas 7+8? | * For the selected clusters, further analysis of 6 trips identified as outliers for final decision. <br> * Verification of the use of 100 mm mesh in Subarea 8 for Danish seine* <br> * For SDN Subarea 8, look at what the results would be if $70-80 \mathrm{~mm}$ meshes were grouped together |
| 09/06/2022 | COPIL 2 and WG Reference Fleets 2 | * Further data cleaning for the 3 gears <br> * Verification of $\mathbf{1 0 0} \mathbf{m m}$ mesh sizes in Subareas 8 for Dansih seine <br> * Application of the ROMELIGO method in Subarea 7 for trawl and gillnet and interest in building an index for Subareas $7+8$ for trawl and gillnet <br> * First tests of decomposition methods (season, trend) and generalized additive models (GAMs) | * The 6 trips were eliminated from the data set <br> * For the next WG Reference Fleets: <br> - look at what the results would be if the $70-80 \mathrm{~mm}$ meshes in zone 8 were grouped together; - test the method on gillnets in Subareas 7+8 |
| 23/11/2022 | WG Reference Fleets 3 | No red mullet element presented GT dedicated to whiting and pollack |  |
| 02/02/2023 | WG Reference Fleets 4 | No red mullet element presented GT dedicated to whiting and pollack |  |
| 28/03/2023 | COPIL 3 | * Brief review of work presented at the WG Reference Fleets 1 and 2 regarding striped red mullet <br> * Start of the model-based method (first handling of data with presentation of first descriptive analyses) |  |
| 06/06/2023 | WG Reference Fleets 5 | * Presentation of all that has been done on the selection of reference fleets by data-filtering for discussion and validation. <br> * Presentation of the work in progress regarding LPUE | * Rework the section describing collaboration with professional structures, based on the work of C . Macher and the ICES "Stakeholder |


|  |  | standardization by modeling, as <br> part of the Master 2 internship on <br> the subject. The aim was to <br> validate the filters used, answer <br> any methodological questions and <br> explain the work still to be done. <br> In both cases, a document was <br> prepared and sent to all <br> participants prior to the WG. | Engagement Strategy" <br> approach. <br> * Redo the analyses leading to <br> the selection of fleets by <br> integrating 2021 and 2022 and <br> see if this would have changed <br> the results. <br> * Addition of regulatory <br> elements provided by <br> professional structures and <br> useful for analysis and <br> interpretation. <br> * Checking of the vessels <br> having worked with the <br> Danish seine in Subarea 8 (the <br> POs concerned send a list of <br> vessels). |
| :--- | :--- | :--- | :--- |

## Appendix 2

## SDN - preliminary stage and step 1 - Figures and tables

Analysis of discards data

- For Subarea 8

Table 1: Distribution of fishing operations observed by the OBSMER program - SDN - Subarea 8 between 2011 and 2021

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | Total général |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2011 |  |  |  |  | 16 |  |  |  |  |  | 26 |  |  |
| 2012 | 3 | 5 | 2 |  | 5 | 5 | 2 |  | 2 | 2 | 3 | 1 |  |
| 2013 | 2 |  |  | 5 | 8 | 3 | 2 | 2 | 2 | 4 | 2 | 27 | 30 |
| 2014 |  |  |  |  | 5 | 7 | 2 |  | 4 | 5 | 2 | 2 |  |
| 2015 |  |  | 3 | 5 |  | 4 |  | 2 |  | 3 |  |  |  |
| 2016 | 2 |  |  |  | 11 |  | 7 | 7 | 3 | 3 | 5 | 2 |  |
| 2017 |  |  | 7 | 1 | 7 | 3 |  |  | 5 |  | 7 | 3 |  |
| 2018 | 2 |  |  | 4 | 14 |  | 14 |  | 3 | 9 | 6 | 1 | 40 |
| 2019 | 3 |  |  | 15 | 4 | 6 | 2 | 4 |  | 6 | 6 | 3 | 53 |
| 2020 |  | 8 |  |  |  |  |  | 3 |  |  |  |  |  |
| 2021 |  |  |  |  | 6 | 8 |  |  |  |  |  |  | 49 |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| Total général | 12 | 13 | 12 | 30 | 76 | 36 | 29 | 18 | 19 | 32 | 57 | 39 |  |


a)

b)

c)

Figure 1: Annual characteristics of discards observed per fishing operation for SDN - Subarea 8: a) mean and standard error in kg ; b) mean percentage and its standard error; c) cumulative mean percentage for individuals below and above the minimum commercial weight.


Figure 2: Monthly characteristics of discards observed per fishing operation for SDN - Subarea 8: a) mean and standard error in kg ; b) mean percentage and its standard error; c) cumulative mean percentage for individuals below and above the minimum commercial weight.

- For Subarea 7

Table 2: Distribution of fishing operations observed by the OBSMER program - SDN - Subarea 7 between 2011 and 2021

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | Total général |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 |  |  |  |  |  |  |  |  |  |  | 15 |  | 15 |
| 2012 |  |  |  |  |  |  | 37 | 4 |  |  | 27 |  | 68 |
| 2013 |  | 1 | 6 |  |  | 5 | 8 |  |  |  |  |  | 20 |
| 2014 |  |  | 11 | 9 |  |  |  | 1 |  | 18 |  |  | 39 |
| 2015 |  | 9 |  | 19 | 4 |  | 12 |  |  | 6 | 21 |  | 71 |
| 2016 | 25 |  |  | 17 |  |  |  |  |  | 24 |  |  | 66 |
| 2017 | 11 |  |  | 9 |  |  |  |  |  |  | 3 | 12 | 35 |
| 2018 |  |  |  |  |  |  | 2 | 7 |  | 4 |  |  | 13 |
| 2019 |  | 6 |  |  |  | 8 |  |  | 4 |  |  |  | 18 |
| 2020 | Pas de données |  |  |  |  |  |  |  |  |  |  |  |  |
| 2021 |  |  |  | 2 |  |  |  |  | 6 |  |  |  | 8 |
| Total général | 36 | 16 | 17 | 56 | 4 | 13 | 59 | 12 | 10 | 52 | 66 | 12 | 353 |



Figure 3: Annual characteristics of discards observed per fishing operation for SDN - Subarea 7: a) mean and standard error in kg ; b) mean percentage and its standard error; c) cumulative mean percentage for individuals below and above the minimum commercial weight


Figure 4: Monthly characteristics of discards observed per fishing operation for SDN - Subarea 7: a) mean and standard error in kg ; b) mean percentage and its standard error; c) cumulative mean percentage for individuals below and above the minimum commercial weight

Analysis of LPUE data for Subarea 8
Search for linear relationships between variables of interest characterizing LPUE, vessel and gear characteristics, and spatio-temporal factors


Figure 5: Correlation circle for SDN - Subarea 8 using the first two dimensions


Figure 6: Individuals plot on the first factorial plane of the PCA for the "Danish seine" gear (SDN) - Subarea 8


Figure 7: Barycenter plot of years on the PCA principal factorial plane and associated CAH for the "Danish Seine" gear (SDN) - Subarea 8


Figure 8 : Barycenter plot of months on the PCA principal factorial plane and associated CAH for the "Danish Seine" gear (SDN) - Subarea 8


Figure 9: Barycenter plot of statistical rectangles on the PCA principal factorial plane and associated CAH for the "Danish Seine" gear (SDN) - Subarea 8


Figure 10: Barycenter plot of mesh size classes on the PCA principal factorial plane and associated CAH for the "Danish Seine" gear (SDN) - Subarea 8

Search for other relationships between variables of interest characterizing LPUE, vessel and gear characteristics, and spatio-temporal factors


| Moy | $\mathbf{1 4 0}$ | 73 | $\mathbf{7 2}$ | 19 | 101 | $\mathbf{2 1 1}$ | 99 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Med | $\mathbf{8 7}$ | 30 | 23 | 5 | 57 | $\mathbf{1 4 7}$ | 39 |
| N | 211 | 2391 | 203 | 79 | 600 | 34 | 1059 |

Figure 11: Conditional regression tree on log10Moy (standardized LPUE) with technical characteristics (the values in columns correspond to the values of each node) for SDN - Subarea 8

Analysis of LPUE data for Subarea 8
Search for linear relationships between variables of interest characterizing LPUE, vessel and gear characteristics, and spatio-temporal factors


Figure 12: Correlation circle for SDN - Subarea 7 using the first two dimensions


Figure 13: Individuals plot on the first factorial plane of the PCA for the "Danish seine" gear (SDN) - Subarea 7


Figure 14: Barycenter plot of years on the PCA principal factorial plane and associated CAH for the "Danish Seine" gear (SDN) - Subarea 7


Figure 15: Barycenter plot of months on the PCA principal factorial plane and associated CAH for the "Danish Seine" gear (SDN) - Subarea 7


Figure 16: Barycenter plot of statistical rectangles on the PCA principal factorial plane and associated CAH for the "Danish Seine" gear (SDN) - Subarea 7



Figure 17: Barycenter plot of mesh size classes on the PCA principal factorial plane and associated CAH for the "Danish Seine" gear (SDN) - Subarea 7

Search for other relationships between variables of interest characterizing LPUE, vessel and gear characteristics, and spatio-temporal factors


| Moy | 142 | 58 | $\sim 0$ |
| :--- | :---: | :---: | :---: |
| Med | 75 | 11 | $\sim 0$ |
| N | 232 | 238 | 8 |

Figure 18: Conditional regression tree on log10Moy (standardized LPUE) with technical characteristics (the values in columns correspond to the values of each node) for SDN - Subarea 7

## Appendix 3

## OTB in Subarea 7 - preliminary stage and step 1 - Figures and tables

Analysis of discards data
Table 1: Distribution of fishing operations observed by the OBSMER program - OTB - Subarea 7 between 2003 and 2021

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | Total | général |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2003 |  |  |  |  |  | 2 |  | 6 | 14 | 27 | 5 |  |  | 54 |
| 2004 |  |  |  |  | 2 | 2 | 22 | 46 | 25 | 19 | 14 |  | $\square$ | 130 |
| 2005 |  |  |  |  | 3 |  | 25 | 39 | 27 | 33 | 13 | 6 | - | 146 |
| 2006 |  |  | 13 | 11 |  |  | 6 | 5 |  | 7 |  | 4 | - | 46 |
| 2007 |  |  |  |  | 2 |  | 9 | 24 |  | 25 | 11 | 2 |  | 73 |
| 2008 | 3 | 4 |  | 8 | 1 | 12 | 24 | 10 | 6 | 32 | 14 | 22 | - | 136 |
| 2009 | 14 |  | 10 | 59 | 46 | 10 | 11 | 13 | 7 | 40 | 24 | 6 |  | 240 |
| 2010 | 19 | 1 |  | 4 | 13 | 16 | 20 | 31 | 41 | 27 | 5 | 7 |  | 184 |
| 2011 | 2 | 5 | 14 | 19 | 15 | 10 | 26 | 12 | 31 | 22 | 23 | 4 |  | 183 |
| 2012 | 20 | 13 | 12 | 7 | 11 | 7 | 20 | 13 | 34 | 6 | 14 | 3 |  | 160 |
| 2013 |  | 4 | 3 | 4 | 7 | 8 | 15 | 17 | 20 | 17 | 15 | 5 | $\square$ | 105 |
| 2014 | 24 | 8 | 17 | 16 | 13 | 16 | 30 | 19 | 48 | 19 | 23 | 16 |  | 249 |
| 2015 | 10 | 5 | 31 | 20 | 39 | 51 | 43 | 43 | 49 | 46 | 61 | 45 |  | 443 |
| 2016 | 15 | 23 | 15 | 23 | 29 | 35 | 31 | 42 | 41 | 32 | 31 | 44 |  | 361 |
| 2017 | 39 | 33 | 14 | 27 | 13 | 30 | 15 | 23 | 33 | 15 | 11 | 45 |  | 298 |
| 2018 | 7 | 2 | 11 | 3 | 19 | 24 | 11 | 22 | 25 | 37 | 17 | 20 |  | 198 |
| 2019 | 2 | 32 | 26 | 5 | 35 | 15 | 16 | 28 | 55 | 26 | 24 | 19 |  | 283 |
| 2020 | 9 | 16 | 2 |  |  |  | 5 | 4 | 7 |  | 9 | 8 |  | 60 |
| 2021 | 7 | 14 | 9 | 5 | 7 | 24 | 4 | 12 | 22 |  |  |  | $\square$ | 104 |
| Total général | 171 | 160 | 177 | 211 | 255 | 262 | 333 | 409 | 485 | 430 | 304 | 256 |  | 3453 |


a)

b)

c)

Figure 1: Annual characteristics of discards observed per fishing operation for OTB - Subarea 7: a) mean and standard error in kg; b) mean percentage and its standard error; c) cumulative mean percentage for individuals below and above the minimum commercial weight

a)

b)

c)

Figure 2: Monthly characteristics of discards observed per fishing operation for OTB - Subarea 7: a) mean and standard error in kg ; b) mean percentage and its standard error; c) cumulative mean percentage for individuals below and above the minimum commercial weight

Analysis of LPUE data
Search for linear relationships between variables of interest characterizing LPUE, vessel and gear characteristics, and spatio-temporal factors


Figure 3: Correlation circle for OTB - Subarea 7 using the first two dimensions


Figure 4: Individuals plot on the first factorial plane of the PCA for the "Bottom otter trawl" gear (OTB) Subarea 7


Figure 5: Barycenter plot of years on the PCA principal factorial plane and associated CAH for the "Bottom otter trawl " gear (OTB) - Subarea 7


Figure 6: Barycenter plot of months on the PCA principal factorial plane and associated CAH for the "Bottom otter trawl " gear (OTB) - Subarea 7


Figure 7: Barycenter plot of statistical rectangles on the PCA principal factorial plane and associated CAH for the " Bottom otter trawl " gear (OTB) - Subarea 7


Figure 8: Barycenter plot of mesh size classes on the PCA principal factorial plane and associated CAH for the " Bottom otter trawl " gear (OTB) - Subarea 7

Search for other relationships between variables of interest characterizing LPUE, vessel and gear characteristics, and spatio-temporal factors


| Moy | 5.6 | 7.5 | 8.6 | 5.6 | $\mathbf{1 1 . 2}$ | $\mathbf{1 5 . 2}$ | $\mathbf{1 1 . 2}$ | 5.2 | 5.6 | 0.6 | 3.2 | 4.3 | 0.6 | 4.4 | 3.2 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Med | 2.8 | 3.9 | 2.9 | 1.3 | $\mathbf{4 . 7}$ | $\mathbf{6 . 7}$ | $\mathbf{4 . 7}$ | 3.1 | 3.0 | 0.4 | 0.9 | 1.8 | 0.3 | 0.9 | 0.9 |
| $\mathbf{N}$ | 412 | 6118 | 145 | 2661 | $\mathbf{2 5 7 0}$ | $\mathbf{1 4 9 9 6}$ | $\mathbf{4 4 8 7}$ | 4793 | 779 | 13 | 2321 | 13084 | 61 | 56 | 1607 |

Figure 9: Conditional regression tree on $\log 10 \mathrm{Moy}$ (standardized LPUE) with technical characteristics (the values in columns correspond to the values of each node) for OTB - Subarea 7

## Appendix 4

## GNS - preliminary stage and step 1 - Figures and tables

Analysis of discards data
Table 1: Distribution of fishing operations observed by the OBSMER program - GNS - Subarea 7 between 2004 and 2021

|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |  | otal général |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2004 |  |  |  |  |  |  | 9 | 3 |  |  |  | 4 |  | 16 |
| 2007 |  |  |  | 1 |  |  |  |  |  |  |  |  |  | 1 |
| 2008 |  |  |  |  |  |  |  |  | 1 |  |  |  |  | 1 |
| 2009 |  |  |  |  |  |  |  | 1 | 2 | 9 |  |  | $\square$ | 12 |
| 2010 |  | 1 |  | 1 |  |  |  |  |  |  |  |  |  | 2 |
| 2011 |  |  |  |  |  |  |  | 4 | 3 | 4 | 9 | 1 |  | 21 |
| 2012 | 6 | 16 | 8 |  | 1 | 1 | 6 | 12 | 2 | 2 | 10 | 4 |  | 68 |
| 2013 |  |  |  | 2 |  | 2 | 1 | 5 | 2 |  | 2 |  |  | 14 |
| 2014 |  |  |  | 1 |  |  | 2 | 2 | 3 | 2 | 1 |  | - | 11 |
| 2015 | 2 |  |  | 1 |  |  | 8 |  | 15 | 16 | 2 | 10 |  | 54 |
| 2016 | 8 | 4 | 6 | 1 |  |  |  | 3 | 2 | 2 | 1 | 6 |  | 33 |
| 2017 | 8 | 4 | 1 |  | 1 |  |  | 7 | 2 | 2 | 9 |  |  | 34 |
| 2018 | 5 | 4 |  | 3 |  |  | 3 | 6 |  | 3 | 9 | 1 |  | 34 |
| 2019 |  | 12 | 2 | 2 |  |  | 1 |  | 9 | 3 | 2 |  |  | 31 |
| 2020 |  | 2 |  |  |  |  | 2 |  |  |  | 10 | 1 |  | $\square 15$ |
| 2021 | 1 |  | 2 |  |  |  | 8 | 17 | 8 |  |  |  |  | 36 |
| Total général | 30 | 43 | 19 | 12 | 2 | 3 | 40 | 60 | 49 | 43 | 55 | 27 |  | 383 |


a)

b)

c)

Figure 1: Annual characteristics of discards observed per fishing operation for GNS - Subarea 7: a) mean and standard error in kg ; b) mean percentage and its standard error; c) cumulative mean percentage for individuals below and above the minimum commercial weight

a)

b)


Figure 2: Monthly characteristics of discards observed per fishing operation for GNS - Subarea 7: a) mean and standard error in kg ; b) mean percentage and its standard error; c) cumulative mean percentage for individuals below and above the minimum commercial weight

## Analysis of LPUE data for Subarea 7

Search for linear relationships between variables of interest characterizing LPUE, vessel and gear characteristics, and spatio-temporal factors


Figure 3: Correlation circle for GNS - Subarea 7 using the first two dimensions


Figure 4: Individuals plot on the first factorial plane of the PCA for the "Gillnet" gear (GNS) - Subarea 7


Figure 5: Barycenter plot of years on the PCA principal factorial plane and associated CAH for the "Gillnet " gear (GNS) - Subarea 7


Figure 6: Barycenter plot of months on the PCA principal factorial plane and associated CAH for the " Gillnet " gear (GNS) - Subarea 7


Figure 7: Barycenter plot of statistical rectangles on the PCA principal factorial plane and associated CAH for the " Gillnet " gear (GNS) - Subarea 7


Figure 8: Barycenter plot of mesh size classes on the PCA principal factorial plane and associated CAH for the " Gillnet " gear (GNS) - Subarea 7

Search for other relationships between variables of interest characterizing LPUE, vessel and gear characteristics, and spatio-temporal factors


| Moy | 20.2 | 13.6 | 37.0 | 25.7 | 17.6 | 17.8 | 0.3 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Med | 14.0 | 8.9 | 27.9 | 16.0 | 9.4 | 2.6 | 0.2 |
| N | 1031 | 542 | 621 | 602 | 892 | 217 | 15 |

Figure 9: Conditional regression tree on log10Moy (standardized LPUE) with technical characteristics (the values in columns correspond to the values of each node) for GNS - Subarea 7

Analysis of LPUE data for Subareas 7-8
Search for linear relationships between variables of interest characterizing LPUE, vessel and gear characteristics, and spatio-temporal factors


Figure 10: Correlation circle for GNS - Subareas 7-8 using the first two dimensions


Figure 11: Individuals plot on the first factorial plane of the PCA for the "Gillnet" gear (GNS) - Subareas 7-8


Figure 12: Barycenter plot of years on the PCA principal factorial plane and associated CAH for the " Gillnet " gear (GNS) - Subareas 7-8


Figure 13: Barycenter plot of months on the PCA principal factorial plane and associated CAH for the " Gillnet " gear (GNS) - Subareas 7-8


Figure 14: Barycenter plot of statistical rectangles on the PCA principal factorial plane and associated CAH for the " Gillnet " gear (GNS) - Subareas 7-8


Figure 15: Barycenter plot of mesh size classes on the PCA principal factorial plane and associated CAH for the " Gillnet " gear (GNS) - Subareas 7-8

Search for other relationships between variables of interest characterizing LPUE, vessel and gear characteristics, and spatio-temporal factors


| Moy | 19.4 | 31.3 | 11.0 | 20.8 | 36.6 | 24.7 | 154.7 | 22.1 | 15.1 | 10.2 | 36.0 | 4.9 | 36.6 | 2.3 | 17.1 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Med | 13.7 | 24.9 | 7.8 | 14.8 | 23.7 | 11.6 | 118.8 | 1.4 | 8.2 | 5.1 | 15.3 | 2.4 | 23.7 | 0.7 | 1.9 |
| N | 3418 | 166 | 1503 | 2948 | 2200 | 295 | 660 | 39 | 2786 | 825 | 870 | 35 | 13 | 7 | 1448 |

Figure 16: Conditional regression tree on $\log 10$ Moy (standardized LPUE) with technical characteristics (the values in columns correspond to the values of each node) for GNS - Subareas 7-8

# Validation of prediction-variance relation in a state-space fish stock assessment model 

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#### Abstract

It has been proposed by Breivik et al. (2021) that an observation variance can be estimated within an assessment model through the relationship $v=\alpha \mu^{\beta}$, where $\alpha$ and $\beta$ are positive parameters and $\mu$ is the expectation on natural scale. They show that this feature gives a better optimisation for their case studies, but it is not validated against externally estimated variances. One can certainly argue that it does not matter as long as you get a better fit to data, but since survey indices from standard software often come with a variance estimate, it is a good exercise to compare the internal- and the external variance estimates. This is our main objective. For case studies, we use the Norwegian spring-spawning herring and blue whiting assessments. In the former, the external variance estimates are used in the assessment but not for the latter.


Keywords: SAM, stock assessment, prediction-variance

## 1 Introduction

Breivik et al. (2021) proposed that an observation variance can be estimated through the relationship $v=\alpha \mu^{\beta}$ (Taylor, 1961), where $\alpha$ and $\beta$ are positive parameters and $\mu$ is the expectation on natural scale. A key feature is that both $\alpha$ and $\beta$ are estimated within the model, which has the benefits that the parameter estimator uncertainty propagates correctly and it is convenient for the user. We will refer to their method as the internal Taylor approach. The authors test this new feature of the SAM model (Nielsen and Berg, 2014) on both the cod and the haddock assessments and conclude that it improves the fit to data.

This relationship, originally proposed by Taylor (1961), has been shown to be adequate both for fishery independent estimates and for catch at age estimates (Aanes, 2016). An implementation of this relationship in SAM, or any other assessment models, is thus sound, and can give a better observation variance if the assumption of this relationship is valid for the different input data. However, our concern is that this model configuration may be influential for the assessment, and that the estimated variances may not reflect the uncertainty of the computed indices.

Both fishery dependent and fishery independent indices from surveys are typically computed via existing software, where these may also include estimates of the survey variance (Johnsen et al., 2019; Hirst et al., 2004). With a few exceptions, the variance estimates are typically not included in the assessment. The reason for this may be that either the assessment model does not facilitate inclusion of variance estimates, or that the variance has not been estimated. In ICES (2021), there is some survey data that also has variance estimates that can be used to validate the internal Taylor approach. While the approach assumes no changes in the Taylor (1961) relationship over time, the externally estimated index variance takes into account inflated uncertainty in certain years due to the occurrence of special events. Thus, these estimates can be used as confirmation or falsification of the homogeneity assumption across years.

In this paper we investigate firstly, if the internally estimated observation variances are consistent with the external ones, and secondly; if this may have substantial influence on a stock's development for Norwegian spring-spawning herring and blue whiting.

## 2 Method

To evaluate the new model configuration proposed by Breivik et al. (2021) we use the 2021 assessments of two pelagic fish species; the Norwegian spring-spawning herring (Clupea Harengus) and blue whiting (Micromesistius Poutassou) (ICES, 2021). For both species, the surveys indices and variances have been computed using the StoX software (Johnsen et al., 2019) for the fishery independent data. Henceforth, we refer to these as external
variances, as opposed to the internal ones predicted by SAM. The data input and general assessment configurations are described in the subsequent sections for each species. With the exception of model features directly linked to the observation variance, we keep the model configurations the same as from the official assessments. In the technical terms of the current implementation of SAM, this means exclusively adjusting the keyVarObs and predVarLink options of the model configuration.

These two options control the $\alpha_{i a}$ and $\beta_{i a}$ parameters, for fleet $i$ and age group $a$, which give the observation variance $v_{i a y}=\alpha_{i a} \mu_{i a y}^{\beta_{i a}}$, where $y$ denotes the year of the observation. In SAM the observations are log-transformed and the variance of the log-observations is given by

$$
\begin{equation*}
\sigma_{i a y}^{2}=\log \left(\frac{v_{i a y}}{\mu_{i a y}^{2}}+1\right)=\log \left(\alpha_{i a} \mu_{i a y}^{\beta_{i a}-2}+1\right), \tag{1}
\end{equation*}
$$

inserting the assumed relation between $v_{i a y}$ and $\mu_{i a y}$. Note that we have constant variance for a particular fleet $i$ and age group $a$ if $\beta_{i a}=2$. If $\beta_{i a}=1$ the variance increases linearly with $\mu_{i a y}$ on the natural scale and $\beta_{i a}<2$ means $\sigma_{i a y}^{2}$ decreases with increasing $\mu_{i a y}$. As indicated by the notation used, $\alpha_{i a}$ and $\beta_{i a}$ are allowed to vary with fleet and age group, but assumed constant over time. In the implementation of Breivik et al. (2021), it is assumed that $\beta_{i a}>1$ with default value of $\beta_{i a}=2$ corresponding to constant variance.

### 2.1 Herring data and assessment

The assessment of the Norwegian spring-spawning herring has four indices indicating abundance at age; the catch at age (caa), the acoustic-trawl spawning survey in February (NORHERSS), the international acoustic-trawl survey in May (IESNS), and the acoustictrawl recruitment at age 2 index survey in the Barent Sea (RI). The caa is developed by aggregating the estimated catch at age from each nation. The external variance is computed using the ECA-model (Hirst et al., 2004) directly on the Norwegian part of the caa. The approximation by Taylor (1961) is used on the Norwegian data, and this relationship is then used to compute the variance of the combined caa. The external variance of the survey input data is estimated from the StoX software, but is smoothed using the approximation by Taylor (1961). The age structured model uses age groups $2-12+, 3-12^{+}, 2,3-12^{+}$for the caa, NORHERSS, RI and IESNS, respectively. The $12^{+}$ indicates a plus group for age 12 and older. Note that there are more available ages from the indices computation from StoX, but these are the ones used in the assessment. Also note that we use the raw variance output from StoX, not the smoothed that are used in the assessment.

The official assessment model is the XSAM model developed by Aanes (2016) where the variance from all data sources can be included in the model. For this analyses we use the SAM framework (Nielsen and Berg, 2014). The current implementation of the

SAM framework has adopted the general model from XSAM and can produce similar, but not identical, estimates as XSAM. The similarity is still sufficient for the purpose of this exercise.

### 2.2 Blue whiting data and assessment

The ICES assessment of blue whiting (ICES, 2021) uses caa and the international blue whiting spawning stock survey (IBWSS) as input data and the assessment model is implemented in the SAM framework (Nielsen and Berg, 2014). The IBWSS index is estimated using StoX (Johnsen et al., 2019). The external variances are also estimated, but this information is currently not being used in the official assessment model, as opposed to the herring case. For the caa, no estimates of the observation variance for blue whiting have been made. The age structured model uses age groups $1-10^{+}$and $1-8$ for the caa and IBWSS, respectively. Similar to herring, there are more ages available, but these are the ones used in the assessment. The $10^{+}$indicates a plus group for age 10 and older. The assessment from 2021 is available at stockassessment.org.

### 2.3 Objectives

We have two main objectives: 1) Compare the internally estimated relationship between observation variance and mean against the external ones, and 2) compare the differences in model output using the current assessment versus the alternative using the proposed internal Taylor feature. We will not change anything in the configuration not related to this feature. For the alternative assessment using internal Taylor variances, we assume one $\alpha$ and $\beta$-parameter for each data source. That is, $\alpha$ and $\beta$ are assumed constant across age groups, except age 2 for herring caa, to maintain consistency with the external procedure.

The two case studies are interesting in their own ways. In both cases, it will be interesting to compare the externally estimated variances to the internal ones. In the specific herring case, we can study the effects of either specifying the index variances as fixed weights fed into the model or allowing the model to estimate these variances so that they are in accordance with the data and model structure. For the blue whiting assessment, the differences are more related to taking into account the Taylor relationship or not doing so.

## 3 Result

The external variance estimates from the StoX software for both species, all ages, and for the different input data are shown as black dots in Figure 1 with a linear regression attached to each data. The slope of the regression line is shown as $\beta^{\text {ext }}$ in Table 1. The
coloured lines indicate the internally estimated $\log$-variance from $\alpha^{\text {int }}$ and $\beta^{\text {int }}$ for the different age spans. Values for $\beta^{\text {int }}$, with their $95 \%$ confidence intervals, are also shown in Table 1. The model configurations of this run is shown in Appendix A. Note that $\beta^{\text {int }}$ had to be excluded for age 2 for herring for the model to converge, where this is likely due to few data points for those ages. Also, no external variance estimates for caa is available for blue whiting.

Both Figure 1 and Table 1 show a difference between the $\beta^{\text {int }}$ and $\beta^{\text {ext }}$ values, but $\beta^{\text {ext }}$ is within the $95 \%$ confidence interval of $\beta^{\text {int }}$. For herring caa $\beta^{\text {int }}$ is higher than $\beta^{\text {ext }}$, while all other cases the $\beta^{\text {int }}$ is lower.

With the internal Taylor approach, AIC for blue whiting is lower, from 605.3 (22 parameters) to 587.3 (17 parameters), than for the official assessment, while for herring, external weighting changes the input data and not just the model setup, making AIC comparison controversial. The consequence from the different cases are shown in Figure 2, for both estimated spawning stock biomass (SSB), fishing mortality (Fbar) and recruitment. The stock development for both species show an upward revision in the recruitment for the larger cohorts when using the method proposed by Breivik et al. (2021). A larger recruitment of these cohorts will thus increase the SSB and reduce the Fbar. Note however that (almost) all changes are within the confidence intervals of the official assessment estimates.

## 4 Discussion and Conclusion

In many cases, for both fishery-dependent and fishery-independent surveys, the Taylor relationship between mean and variance is observed empirically; but for several surveys, the variance has not been calculated. The approach by Breivik et al. (2021) suggests that this relationship can be estimated by the assessment model itself in any case where the variance is not available. In their case studies, the internal Taylor approach produced a better optimisation in terms of AIC and this feature removed a clear pattern between the one-step-ahead residuals and the prediction, but the predicted variance development was not validated with any external variances. The focus in this work is to compare the variance from the internal Taylor approach with available external variance estimates. The configuration settings used in the cases is thus to evaluate the performance of this configuration, not to find the best configuration for estimating stock dynamics with the purpose of developing a catch advice.

The internal estimates describe a prediction-variance link, but the external procedure do not have the prediction, and must therefore use the observation directly as an approximation. In this regard, the external Taylor approach is an observation-variance link. Assuming that the predictions are proportional to the observations on $\log$-scale (i.e. $\log \mu \approx \log x$, where $x$ is a observation and $\mu$ the corresponding expected value). By
comparing the predictions from the alternative assessments to the observations, we find this assumption to be reasonable (see appendix B).

The $\alpha$ parameters are different between the externally and internally developed variances (Figure 1). The estimated $\alpha$ is not only dependent on the state of the data, but also the configuration of the model. The internal and external $\alpha$ estimates should thus not be compared directly, but it should be noted that this parameter is also important for the weighting of the different data sources. Since $\alpha$ is different, and this difference is not constant between the different data sources, then this indicates that the weight of the data sources has changed. A direct comparison can be made for the $\beta$ parameter. The $\beta$ parameter shows if the development between $\log$ variance and $\log$ expectation is similar when compared to an externally computed $\log$ variance. Our findings show that there is a difference in the $\beta$ parameters, but this difference is within the $95 \%$ confidence intervals of the internally estimated ones (Table 1).

For herring, we experienced some convergence issues when including a $\beta$ estimate for age 2 for RI. Correspondingly, we excluded age 2 for caa due to an issue with infinite standard error for the $\beta$ estimate. This could also be solved by splitting $\alpha$ and $\beta$ further based on age, but we maintain the same $\beta$ for each fleet for comparison reasons.

In Appendix B we have included some supplementary plots of one-step-ahead residuals by year and age for the different fleets, models and species. In Figure B. 2 (herring) and B. 3 (blue whiting) we observed minor differences in the patterns. We have also conducted a simulation study and checked that most of the simulations fall within the confidence intervals of the corresponding models.

In simple terms, $\beta \neq 2$ will cause the $\log$ observation within and across a time series to be weighted depending the relationship shown in eq. 1 . When $\beta \leq 2$, for the cases shown in Table 1, this causes the higher values of $\log$ observation to be weighted more in an assessment model. This effect apply both when the $\beta$ is estimated within or outside the model. The impact this functionality will have on a stock estimation depends on the stock dynamics. For species with a spasmodic recruitment, i.e. sudden large cohorts, the effect will be larger than for species without a spasmodic recruitment, e.g. cod (Breivik et al., 2021). Herring and blue whiting are species with a spasmodic recruitment, and for blue whiting it is shown that the recruitment for the larger cohorts increases when applying this relationship (Figure 2). For herring, the official assessment uses external variance from the data but the recruitment of the larger cohort are still being increased when instead estimating the variance within the assessment model. It should be noted that this increase cannot be explained by a different $\beta$ value alone, as the different configuration settings may cause the weighting of the data sources to vary.

External variances can be included in the assessment if these are available, as the case for the herring assessment. The benefit of external variance is that any temporal effects, i.e. increasing or reducing the survey efficiency, are then also included in the assessment.

The benefit of using the internal Taylor method is that the uncertainty in the $\beta$ parameter is propagated to the entire uncertainty of the assessment, but any temporal effects is ignored. The latter is also true if the Taylor approximation has been used externally. If temporal effects can be ignored, the approach of estimating the variance within the model is appealing and we have found that, if configured correctly, $\beta^{e x t}$ is within the uncertainty of $\beta^{\text {int }}$ and will thus produce similar development of the variance. This approach can thus be implemented in cases where any external variance is not available; but, we will still argue that the external variance should be computed with the purpose to evaluate if the approach proposed by Breivik et al. (2021) is suitable.

| Species | Data | $\beta^{\text {int }}$ | $\beta^{\text {ext }}$ |
| :---: | :---: | :---: | :---: |
| Herring | caa | $1.49 \pm 0.11$ | 1.35 |
| Herring | NORHERSS | $1.30 \pm 0.28$ | 1.56 |
| Herring | IESNS | $1.37 \pm 0.25$ | 1.52 |
| Herring | RI | - | 1.81 |
| Blue Whiting | caa | $1.45 \pm 0.21$ | - |
| Blue Whiting | IBWSS | $1.33 \pm 0.36$ | 1.49 |

Table 1: The internally ( $\beta^{\text {int }}$ with $95 \% \mathrm{CI}$ ) and externally estimated ( $\beta^{\text {ext }}$ ) slope parameters between $\log$ observation and $\log$ variance for the different input data and species.


Figure 1: The relationship between the $\log$ variance and the $\log$ observation. The black dots with a blue regression line show the relationship between externally estimated log variance and $\log$ observation for all available ages, while the lines colored by age group show relationship as outputted from the assessment model. Grey lines indicate the relationship when $\beta=2$, but with different values for $\alpha$.


Figure 2: Stock estimates (SSB, Fbar and recruitment) from the two different cases, the official WGWIDE assessments and the alternative assessments, with herring to the left and blue whiting to the right.

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## Data availability statement

All the code and data necessary to reproduce the results are available at: https://github.com/IMRpelagic/PredLinkVsExternalweights.

## Author Contribution

Both authors contributed equally in all parts of this study.

## A Configurations

The configurations of \$keyVarObs and \$predVarObsLink for the official and alternative assessments for the two species are reported below. The full configuration files can be found in the github repository.

## Herring

Official assessment
\$keyVarObs

| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| -1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| -1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| \$predVarObsLink |  |  |  |  |  |  |  |  |  |  |
| -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| NA | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| -1 | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| NA | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |

Alternative assessment

## \$keyVarObs

| 0 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| -1 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 | 2 |
| 3 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 | -1 |
| -1 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 | 4 |

## \$predVarObsLink

| -1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

$\begin{array}{lllllllllll}\text { NA } & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1 & 1\end{array}$
-1 NA NA NA NA NA NA NA NA NA NA
$\begin{array}{lllllllllll}\text { NA } & 2 & 2 & 2 & 2 & 2 & 2 & 2 & 2 & 2 & 2\end{array}$

## Blue whiting

Official assessment

```
$keyVarObs
    0
    4
$predVarObsLink
    -1
    -1
```

Alternative assessment

| \$keyVarObs |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | -1 | -1 |
| \$predVarObsLink |  |  |  |  |  |  |  |  |  |
| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | NA | NA |

## B Supplementary plots



Figure B.1: Observation versus prediction for the alternative models for herring and blue whiting.


Figure B.2: One-step-ahead residuals by model, survey, age and year for herring. The size of each buble is proportional to the absolute value of the residual while the color indicates if it is positive or negative.


Figure B.3: One-step-ahead residuals by model, survey, age and year for blue whiting. The size of each buble is proportional to the absolute value of the residual while the color indicates whether it is positive or negative.

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# 2022 Mackerel and Horse Mackerel Egg Survey 

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Not to be cited without prior reference to the authors

## Introduction

The mackerel and horse mackerel egg survey is an ICES-coordinated international study in the north east Atlantic conducted during the first half of 2022. This study is a combined plankton and fishery investigation formed by a series of individual surveys which have taken place triennially since the late 1970 s and is coordinated by the ICES Working Group on Mackerel and Horse Mackerel Egg Surveys (WGMEGS).

The main objective is to estimate a biomass index for the Northeast Atlantic mackerel stock and an egg abundance index for the Western horse mackerel stock using the AEPM, and an SSB index for the Southern horse mackerel stock using the DEPM during thiigs series of individual cruises from January to July.. A secondary objective is to estimate the SSB indices for the western and southern areas as well as for the North Sea area using the DEPM. The results have been used in the assessment for NEA mackerel stock since 1977 and for western horse mackerel stock since 1992. The mackerel and horse mackerel egg survey is still the main source of data providing fisheries independent information for these stocks.

The general method is to quantify the freshly spawned eggs in the water column on the spawning grounds. To be able to establish a relationship between eggs and biomass of the spawning stock, the fecundity of the females must also be determined. This is undertaken by sampling ovaries before and during spawning. The potential fecundity is counted from whole mount volumetric subsamples using a dissecting microscope while atresia is counted histologically from slides. Realised fecundity is estimated as potential fecundity minus atresia. The realised fecundity is used in combination with the calculated number of freshly spawned eggs in the water to estimate the spawning stock biomass. (ICES 2019a; ICES 2019b)

To provide reliable estimates of spawned eggs and fecundity an extensive coverage of the spawning area is required both in time and space. The spawning of the southern horse mackerel stock and mackerel starts in late December off the Portuguese coast. Spawning proceeds further north along the continental shelf edge as water temperature increases during late winter and spring. In the past peak spawning of mackerel has normally occurred in April-May in the area of the Sole Banks with an
extension to the Porcupine Bank. Whilst the distribution and timing of peak western horse mackerel spawning has remained fairly stable during recent surveys the same cannot be said for NEA mackerel. Recent surveys in 2010 and 2013 saw peak mackerel spawning in March with 2013 also demonstrating a shift in the geographical centre of spawning further south within the southern Biscay region. Away from these areas mackerel spawning is now observed over a large region of the Northeast Atlantic both on and off the continental shelf, ranging as far west as Hatton Bank and as far north as Iceland and the Faroe Islands as well as the Shetland Islands and the Norwegian coast in the Northeast.

## Survey effort

As a consequence of the long spawning period and the large survey area involved, the mackerel and horse mackerel egg surveys have always relied on broad international participation. In 2022 a total of 16 individual cruises were carried out with a total of 284 survey days in the western and southern areas, with the contribution of Spain (IEO: 42 days at sea, AZTI: 30 days), Scotland ( 59 days), the Netherlands ( 39 days), Ireland ( 28 days), Portugal ( 34 days), Germany ( 23 days), Norway ( 15 days) and the Faroe Islands ( 14 days). The North Sea survey was carried out in the same year as the Atlantic surveys for the first time. Thirty five survey days were devoted to this survey, with England contributing 23 days and Denmark 12 days.

## Survey design

The aim of the triennial egg survey is to determine the annual egg production (AEP). This is calculated using the mean daily egg production rates per pre-defined sampling period for the complete spawning area of the Northeast Atlantic Mackerel and Horse Mackerel Stocks. To achieve this, one plankton haul per each half rectangle (separated by approximately 15 nm ) is conducted on alternating transects covering the complete spawning area. The 2022 egg survey was designed in order to maximise both the spatial and temporal coverage in each of the sampling periods. Given the very large area to be surveyed this design minimises the chances of under/overestimation of the egg production (ICES 2008).

The 2022 survey plan was split into 6 sampling periods (Table 1). Originally Portugal were assigned a Period 1 survey which would extend into Period 2. Due to a delay in the start of their survey it was decided to modify the start date of period 2 in the southern area and include the survey into period 2. No sampling was scheduled to take place in ICES division 9a after Period 2. Sampling of the western area commenced in period 3, and included coverage of the west of Scotland, west of Ireland and Biscay. Surveying in the Cantabrian sea ended at the end of period 5. In periods 6 and 7 the surveys were designed to identify a southern boundary of spawning and to survey all areas north of this boundary.

Maximum deployment of effort in the western area was during periods three, four, five and six. Historically these periods would have coincided with the expected peak spawning of both mackerel and horse mackerel. Recent years have seen mackerel peak spawning taking place during periods 3 and 5 .

Due to the expansion of the spawning area which has been observed since 2007 the emphasis was even more focused on full area coverage and delineation of the spawning boundaries. Cruise leaders had been asked to cover their entire assigned area using alternate transects and then use any remaining time to fill in the missed transects.

Table 1. Participating countries, vessels, areas covered, dates and sampling periods of the 2022 surveys.

| Country | Vessel | Area | Dates | Period |
| :--- | :--- | :--- | :--- | :--- |
| Portugal | Vizconde de Eza | Portugal | Jan 23 | Feb 20 |

## Processing of samples

The analysis of the plankton and fecundity samples were carried out according to the sampling protocols as described in the WGMEGS Survey Manual (ICES, 2019a) \& Fecundity manual (ICES, 2019b).

A total of 2184 plankton samples were collected and sorted. Mackerel and horse mackerel eggs were identified and the egg development stages determined. Depending on the vessel facilities and the experience of the participants this was done either during the cruise or back in the national institutes.
Double micropipette samples and slices from ovaries of mackerel were taken during each survey. Additional samples were collected during periods 3 and 4 by participants in an effort to carry out DEPM analysis. Fecundity sampling for horse mackerel only took place during the expected peak spawning periods, 6 and 7. After each survey the ovary screening and fecundity samples were sent to different European research institutes for histological and whole mount analysis to determine the realised fecundity (potential fecundity minus atresia). Fecundity samples have to be analysed in the laboratory upon return from sea and the procedures for analyses are time consuming. The last samples were collected in July. Additional samples were collected during period 3 with the cooperation of the Netherlands PFA.

Horse mackerel is considered to be an indeterminate spawner and therefore since 2007 IPMA has adopted the DEPM methodology for southern horse mackerel in ICES division 9a. The egg survey design in the western area is directed at the AEP method for mackerel which produces an estimate of SSB. Fecundity samples for horse mackerel were taken during the survey in the western area in order to develop a modified DEPM approach for estimating the biomass of the southern horse mackerel stock

## Survey coverage and mackerel egg production by period

The 2022 MEGS was split into 6 survey periods, the start and end dates of which can be found in iError! No se encuentra el origen de la referencia.. For each of the six sampling periods, the following can be noted

Period 2 - Portugal started the 2022 survey series on January $23^{\text {rd }}$. This is a DEPM survey mainly targeting the southern horse mackerel stock and is designed for this purpose, however it also delivers mackerel egg abundance data. The survey is usually undertaken between Cadiz and Galicia and is confined to ICES division 9a (Figure 1).

Period 3 - Period 3 marks the commencement of the western area surveys as well as a continuation of sampling in the southern area. Sampling was undertaken by Ireland (West of Scotland, west of Ireland, Celtic Sea), Germany (Celtic Sea) and AZTI (northern Biscay). Further south the Bay of Biscay, Cantabrian Sea and Galicia were covered by Spain (IEO).
No eggs were found by Ireland in northern waters so after a number of days the vessel turned south and sampled in the Celtic Sea. Due to issues with COVID cases among the crew, the German survey was delayed in starting, however it was successful in linking with the lrish vessel. Both IEO and AZTI experienced difficulties with their vessels, and lost a number of sampling days, however full coverage was achieved (Figure 2).
Egg numbers were relatively low to the west of Ireland. In contrast, further south large numbers of eggs were found close to the 200 m contour line. In Biscay and the Cantabrian Sea AZTI and IEO recorded a number of stations with large numbers of mackerel eggs. 298 stations were sampled and there were only 13 interpolations. There were 52 replicate samples with the majority being completed in the Cantabrian Sea.

Period 4 - This period was covered by three surveys. Scotland sampled the area from the northwest of Ireland to the Shetland Islands. Germany surveyed west of Ireland, Celtic Sea and northern Biscay while IEO completed the survey coverage in southern Biscay and the Cantabrian Sea (Figure 3).
Due to difficulties in acquiring diplomatic clearance, the Scottish survey was unable to sample in Irish waters. As a result, Germany extended their survey area northwards to ensure continuity of survey coverage.
Once again moderate levels of eggs were recorded throughout the area, with the highest concentrations still being found close to the 200 m contour line. Large numbers of mackerel eggs were once again recorded to the west of Scotland, however, they were lower within this area and time period than those reported in 2019.327 stations were sampled and there were 46 interpolations. 52 replicate samples were taken and once again most of these were collected from the Cantabrian Sea.

Period 5 - In period 5, the entire spawning area from the Cantabrian Sea to the West of Scotland, and up to Faroese waters at around $61^{\circ} \mathrm{N}$ was surveyed by AZTI, the Netherlands, Scotland, and Faroes. Spawning in the Cantabrian Sea was tailing off with only low egg numbers being found. Throughout Biscay and into the southern Celtic Sea numbers were generally low to moderate (Figure 4). This pattern continued west of Ireland, to around $54^{\circ} \mathrm{N}$, with spawning remaining on and around the shelf edge. North of this, however, and similar to that noted in 2016 and 2019, spawning activity fanned out both west- and northwards. Due to the large area Scotland had to survey, their vessel was forced to restrict exploration of the western boundary around the SW of Rockall Bank. Egg counts recorded from the boundary stations within this area were lower than reported in 2019 so while the western boundary wasn't fully delineated, MEGS is happy that the survey has captured the majority of egg production in this area. North of this, the Faroese survey completed stations North of Hatton Bank and up towards the Icelandic coast. Some egg production was found to the north of Rockall, however the largest number of eggs were encountered west of the Shetland Islands. In total 444 stations were sampled and there were 214 interpolations. One replicate sample was undertaken.

Period 6 - During period 6 northern Biscay, northwards from $46^{\circ} \mathrm{N}$ and also the Celtic Sea were covered by the Netherlands while Ireland was to cover west of Ireland and also west of Scotland. Norway surveyed the area north of $59^{\circ} \mathrm{N}$ from the south of Iceland to the Norwegian coast, as well as carrying out four transects in the northern North Sea to assist England and Denmark in providing full coverage for the DEPM survey.
Ireland had planned to charter a research vessel from Northern Ireland to conduct the period 6 survey. One week prior to departure the vessel had to go to dry dock for emergency repairs. After much searching, a smaller Welsh RV was contracted as a replacement. Once at sea however it quickly became clear that the replacement vessel was wholly unsuitable and not up to the task. With only two stations successfully completed the decision was made to abandon the survey leaving the area from 53 N to 61 N unsampled. Norway and Netherlands both completed their survey sampling successfully.
Low levels of spawning were observed in Biscay and to the south to the West of Ireland and Porcupine bank (Figure 5). Similarly, in the northern area, spawning was observed at low levels, with the exception once again of the area west of the Shetland Islands. Due to an unavoidable reduction in the number of survey days available, Norway was unable to secure either the northwestern or northern boundary within the northern area, while Netherlands secured the western boundary in their area. 184 stations were sampled with 36 interpolations. No replicate stations were completed.

Period 7 - This period was covered entirely by Scotland sampling on alternate transects in the area from $47^{\circ} 15 \mathrm{~N}$ to north of the Hebrides and $59^{\circ} \mathrm{N}$ (Figure 6). Due to the lack of eggs encountered, the Scottish survey adhered very closely to the 200 m contour and 144 stations were sampled with 24 interpolations. Two replicate stations were completed. Only very low levels of spawning were observed and these were confined to the continental shelf and shelf edge with all spawning boundaries being delineated successfully.


Figure 1 Mackerel egg production by half rectangle for period 2 (Jan $23^{\text {rd }}-$ Feb $19^{\text {th }}$ ). Circle areas and colour scale represent mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 2 Mackerel egg production by half rectangle for period 3 (Mar 4th - Apr 8th). Circle areas and colour scale represent mackerel stage l eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 3 Mackerel egg production by half rectangle for period 4 (Apr 9th - 29th). Circle areas and colour scale represent mackerel stage l eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 4 Mackerel egg production by half rectangle for period 5 (Apr 30th - May 31st). Circle areas and colour scale represent mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 5 Mackerel egg production by half rectangle for period 6 (June 1st - 30th). Circle areas and colour scale represent mackerel stage l eggs $/ \mathrm{m} 2 /$ day by half rectangle. Crosses represent zero values.


Figure 6 Mackerel egg production by half rectangle for period 7 (July 1st -31 st). Circle areas and colour scale represent mackerel stage I eggs $/ \mathrm{m} 2 /$ day by half rectangle. Crosses represent zero values.

## Results - MACKEREL

## Stage I Egg production in the Western area

2010 provided an unusually large spawning event early in the spawning season, 2013 yielded an even larger spawning event indicating that spawning was probably taking place well before the nominal start date of $10^{\text {th }}$ February (day 42). In 2016 the first survey commenced on February $5^{\text {th }}$ which is five days prior to the nominal start date. That year however mackerel migration was later and slower than that recorded in the previous two surveys. The pattern in 2019 followed that of 2016 with no early peak spawning being recorded (Fig. 8 \& Table 2). In 2022 peak spawning was found to have taken place in period 4. The expansion observed in the western and north-western areas of the western area in the 5 and 6 periods of 2016 is reported again in the 2022 period., however egg numbers were not as large as in 2016. During period 5 the northern and northwestern boundaries were once again not delineated, however the exploratory egg surveys carried out in this region during both 2017 and 2018 provide significant evidence that while some spawning has been missed the loss of egg abundance is not sufficiently large to significantly impact the SSB estimate.

The cancelling of the Irish survey in period 6 was addressed by WGMEGS. The group estimated the spawning area that was missed, and also estimated mean daily egg production for the period. The survey area from $53^{\circ} \mathrm{N}$ to $61^{\circ} \mathrm{N}$, and $3.5^{\circ} \mathrm{W}$ to $21^{\circ} \mathrm{W}$ was looked at for the same time interval in the 2013 (period 5), 2016 (period 6) and 2019 (period 6) surveys. Positive stations were selected where stage I eggs were found in a rectangle on at least two occasions over these three surveys (Figure 7 , blue rectangles). WGMEGS estimated this amounted to 127 missed stations during the period. WGMEGS then calculated mean daily egg production from all the observed survey stations for period 6 in 2022 to be 19.58 stage I eggs $/ \mathrm{m}^{2} /$ day, and applied this figure to the 127 missed stations. jError! No se encuentra el origen de la referencia. 8 shows the spawning curve for 2022, with and without the correction for the lrish survey.


Figure 7 Area, blue colour, where it is estimated eggs would have been found during the Irish period 6 survey.

The nominal start and end dates of spawning are February $10^{\text {th }}$ and July $31^{\text {st }}$ respectively. These are the same dates that were used during previous survey years and the shape of the egg production curve for 2022 does not suggest that the chosen end date needs to be altered. The provisional total annual egg production (TAEP) for the western area in 2022 was calculated as $\mathbf{1 . 8 0}{ }^{*} \mathbf{1 0}^{\mathbf{1 5}}$ (Table 2). This is a $47 \%$ increase on the 2019 TAEP estimate which was $1.22 * 10^{15}$.


Figure 8: The daily mackerel egg production curves for 2010-2022 surveys with (right) and without (left) corrected estimates for period 6 of 2022 (black lines). The curves for 2010, 2013, 2016 and 2019 are included for comparison.

Table 2. Western area estimate of mackerel total stage I egg production by period using the histogram method for 2022

| Dates | Period | Days | Annual stage I egg <br> production * $10^{15}$ |
| :---: | :---: | :---: | :---: |
| Feb $5^{\text {th }}-\mathrm{Mar} 3^{\text {rd }}$ | Pre 3 | 31 | 0.087 |
| Mar $4^{\text {th }}-$ Apr $8^{\text {th }}$ | 3 | 36 | 0.319 |
| Apr $9^{\text {th }}-$ Apr $26^{\text {h }}$ | 4 | 18 | 0.120 |
| Apr $27^{\text {th }}-$ Apr $29^{\text {th }}$ | 4-5 | 3 | 0.043 |
| Apr $30^{\text {th }}-$ May $31^{\text {st }}$ | 5 | 32 | 0.863 |
| Jun $1^{\text {st }}-5^{\text {th }}$ | 5-6 | 5 | 0.068 |
| Jun $6^{\text {th }}-J$ un $22^{\text {nd }}$ | 6 | 17 | 0.211 |
| Jun $23^{\text {rd }}-$ Jul $4^{\text {th }}$ | 6-7 | 12 | 0.081 |
| Jul $5^{\text {th }}-\mathrm{Jul} 25^{\text {th }}$ | 7 | 21 | 0.007 |
| July $26^{\text {th }}$ - July $31^{\text {st }}$ | Post 7 | 6 | 0.0003 |
| Total |  |  | 1.7993 |

## Stage I Egg production in the Southern area

The start date for spawning in the southern area was the $\left.23^{\text {rd }}\right\rfloor$ January (iError! No se encuentra el origen de la referencia.3). Portugal surveyed in Period 2 in division 9a. Sampling in the Cantabrian Sea where the majority of spawning occurs within the Southern area commenced on the $18^{\text {th }}$ March. The same end of spawning date of July $17^{\text {th }}$ was used again this year and the spawning curve suggests that there is no reason for this to change (iError! No se encuentra el origen de la referencia.9). As in 2022 the survey periods were not completely contiguous and this has been accounted for (iError! No se encuentra el origen de la referencia.3). The total annual egg production (TAEP) for the southern area in 2022 was calculated as $\mathbf{2 . 9 3} * \mathbf{1 0}^{\mathbf{1 4}}$ (iError! No se encuentra el origen de la referencia.3). This is a 30\% decrease on the 2019 TAEP estimate which was 4.23 * $10^{14}$.


Figure 9: Annual egg production curve for mackerel in the southern spawning area for 2022, black line). The curves for 2010, 2013, 2016 and 2019 are included for comparison.

Table 3: Estimate of the 2022 total mackerel stage I egg production by period for the southern compo-nent using the histogram method.

| Dates | Period | Days | Annual stage I egg production $\times 10^{14}$ |
| :---: | :---: | :---: | :---: |
| $J a n 25^{\text {th }}-\mathrm{Feb} 17^{\text {th }}$ | 2 | 24 | 0.012 |
| Feb $18{ }^{\text {th }}$ - Mar $17^{\text {th }}$ | 2-3 | 28 | 1.213 |
| Mar $18^{\text {th }}-$ Apr $2^{\text {nd }}$ | 3 | 16 | 1.274 |
| Apr $3^{\text {rd }}$ | 3-4 | 1 | 0.052 |
| Apr $4^{\text {th }}-25^{\text {th }}$ | 4 | 22 | 0.327 |
| Apr $26^{\text {th }}$ - May $1^{\text {st }}$ | 4-5 | 6 | 0.036 |
| May $2^{\text {nd }}-11^{\text {th }}$ | 5 | 10 | 0.011 |
| May $12^{\text {th }}-$ May $31^{\text {st }}$ | Post 5 | 20 | 0.009 |
| Total |  |  | 930 |

## Total egg production (Western and Southern areas)

Total annual eggs production (TAEP) for both the western and southern areas combined in 2022 is 2.093*10 ${ }^{\text {15 }}$ (Fig. 10). This is an increase in production of $\mathbf{2 8 \%}$ compared to 2019, 1.64*10 ${ }^{15}$ (Figure 10).


Figure 10 Combined mackerel TAEP estimates (*1013)-1992-2022.

## Fecundity estimates

## Results Mackerel AEPM - Fecundity

During the 2022 survey, 6784 adult mackerel were collected from 170 trawl hauls between $36.30^{\circ} \mathrm{N}$ and $61.75^{\circ} \mathrm{N}$ during periods $2-7$. This year there has been a significant increase in the number of samples collected (2312) compared to previous surveys thanks to the valuable collaboration of the pelagic fishing industry (PFA). Only females were chosen for reproductive studies, so in this section, the results refer to mackerel females. In all, 1811 ovary samples were used for AEPM (annual egg production method). Only $43 \%$ of the samples planned were collected. Deviation from the initial plan was observed in all periods; the interannual variability in the mackerel migration as well as the probability of successful fishing effort makes it difficult to fit into the original sampling scheme.

## Adult Parameters

## Screening

From the 1813 ovary samples, a total of 1717 samples were screened by histology and classified as described in the manual (ICES, 2019a; for DEPM samples see DEPM section). 1699 were valid for further analysing. samples could be used for potential fecundity analyses, and $51 \%$ of the females were classified as spawning. Spawning females were only used for analyses of atresia.

From the histological screening, a total of 492 samples qualified for potential fecundity analysis, which represents $\sim 29 \%$ of the total samples screened by histology ( $N=1699$ ). After the whole mount screening, this number decreased to 396 samples. The reasons for disqualifications after the whole
mount screening was mostly due to the detection of spent ovaries and the presence of hydrated oocytes. According to the manual (ICES 2019a, SISP 5), ovaries with a leading cohort smaller than 400 $\mu \mathrm{m}$ are considered to be not fully recruited yet; not all oocytes that are going to be spawned may have reached the $185 \mu \mathrm{~m}$ threshold that is used to classify oocytes as maturing.

Different to previous years WGMEGS decided for the fecundity analysis samples that had signs of stage 4 oocytes (migratory nucleus stage) should also be included. Stage 4 oocytes can be regarded as a marker for imminent spawning. This was an extra measure taken to assure that no spawning fish were included in the potential fecundity estimate. More details in ICES, 2023

## Potential fecundity

For all mackerel females used for potential fecundity analysis an initial check (Figure $11 \mathrm{a}-\mathrm{d}$ ) was done on the distribution of fish length, weight, Fulton's condition factor ( $100 \times$ weight/length 3 ), and gonad-somatic index (GSI; $100 \times$ Ovary weight/Fish weight).


Figure 11 Histogram of the distribution of a) Fish length b) Fish weight, c) FultonK index and d) GSI of individuals analysed for fecundity.

Similar to the previous surveys only fish with condition factor between 0.5 and 1.2 , and GSI between 1 and 25 were included (ICES 2014) in the fecundity and atresia estimates. In 2022, as in 2019, no females were excluded from the analysis based on these biological parameters.
Relative potential fecundity in 2022 ranged from 34 to 3212 oocytes/g fish, with a median value of 1313 oocytes/g fish (Figure 12). In surveys prior to 2013, values below $300 \mu \mathrm{~m}$ and above $2100 \mu \mathrm{~m}$ were excluded. Since the 2013 survey (ICES 2014) it was agreed not to delete them however, but instead replace the use of arithmetic mean by median. The median is considered to be a more robust
measure than the arithmetic mean.. In 2019, this issue was discussed again, and it was agreed to test a trimmed mean as an alternative to the median. WGMEGS analysed the time-series (ICES 2021) and found that the median estimates were close to the mean and trimmed mean estimates. In 2022 the trimmed mean, removing $10 \%$ of the data, was 1311 oocytes/g fish, practically similar to the median and mean estimations. For consistency with previous years, we continue to use the median and calculate realised fecundity based on the detailed calculation manual completed with corresponding STATA and R-code by the working group in 2020.


Figure 12 Relative fecundity values of 2022 mackerel samples. The red and black dotted lines are the median and mean values of relative potential fecundity ( 1313 n oocytes/g fish and 1319 n oocytes/g fish, respectively).

As previously mentioned about the fecundity estimate, this year we also included ovary samples with stage 4 oocytes. Oocyte stage 4 samples represented about half the total number of fecundity samples and had almost similar median fecundity value as the other samples; 1314 versus 1302 ( $\mathrm{n} / \mathrm{g}$ ). This difference was not significant ( $P=0.726$, Pearson median test)

In 2022 the Netherlands pelagic commercial fleet contributed significantly to the sampling of adult fish. WGMEGS was concerned about the impact of this new supplier on the size distribution of fish. As described above, fecundity is size-dependent, and the fecundity in 2022 was $10 \%$ higher than in 2019. Thus, if the samples provided by the commercial fleet had come from a different fraction of the population, i.e., considering that commercial fishing gear selects larger fish compared to scientific gear, this could have affected fecundity. The mean and scatter plot of fish length distribution by sampling source indicates no difference between the length distribution of fish caught in scientific or commercial surveys.

## Atresia and realized fecundity

The samples used for the analysis of atresia were collected from the entire survey area and during all periods. Of the 858 fish which were classified as spawning, 243 showed signs of early alpha atresia, which resulted in a prevalence of $28 \%$ (Table 4). This value was similar to the value obtained in 2019 (Table 4). For the 2022 surveys, 74 samples were analysed for intensity of atresia. The geometric mean for the intensity of atresia for 2022 was slightly higher than found for 2019 ( $20 \mathrm{vs} 19 \mathrm{n} / \mathrm{g}$ ) as
was also the loss per day ( $45 \mathrm{vs} 43 \mathrm{n} / \mathrm{g}$ ) and total loss during the spawning period ( $45 \mathrm{vs} 43 \mathrm{n} / \mathrm{g}$ ). However, since the potential fecundity was higher for 2022 compared to 2019 and previous years back to 1998 , the percentage loss was the lowest recorded; $3 \%$ for 2022 versus $4 \%$ for 2019 , which was the previous lowest.

By subtracting the atretic loss from the potential fecundity, a realized fecundity of 1268 (oocytes/g fish) was obtained. From 1998 to 2019 (Table 4), the final estimates of realized fecundity ranged from 1002 to 1209 (grand mean $=1076, S D=71$ ). The 2022 estimate of realized fecundity ( 1268 oocytes $/ \mathrm{g}$ fish) is $18 \%$ higher than the mean for those years. Furthermore, the number of samples analysed in 2022 was the highest in the series, indicating that the fecundity estimate sufficiently reflects the stock fecundity of 2022. Therefore, it is considered suitable for calculating the 2022 SSB.

Table 1. Time series of estimated parameters for adults..

| Parameter | Y1998 | Y2001 | Y2004 | Y2007 | Y2010 | Y2013 | Y2016 | Y2019 | Y2022 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fecundity samples $(\mathrm{n})$ | 96 | 187 | 205 | 176 | 74 | 132 | 97 | 62 | 396 |
| Prevalence of atresia $(\mathrm{n})$ | 112 | 290 | 348 | 416 | 511 | 735 | 713 | 252 | 243 |
| Intensity of atresia $(\mathrm{n})$ | 112 | 290 | 348 | 416 | 511 | 56 | 66 | 64 | 74 |
| Relative potential fecundity $(\mathrm{n} / \mathrm{g})$ | 1206 | 1097 | 1127 | 1098 | 1140 | 1257 | 1159 | 1191 | 1313 |
| Prevalence of atresia | 0.55 | 0.2 | 0.28 | 0.38 | 0.33 | 0.22 | 0.3 | 0.28 | 0.28 |
| Ceometric mean intensity of arresia $(\mathrm{n} / \varepsilon$ | 46 | 40 | 33 | 30 | 26 | 27 | 30 | 19 | 20 |
| Potential fecundity lost per day $(\mathrm{n} / \mathrm{g})$ | 3.37 | 1.07 | 1.25 | 1.48 | 1.16 | 0.8 | 1.2 | 0.71 | 0.75 |
| Potential fecundity lost $(\mathrm{n} / \mathrm{g})$ | 202 | 64 | 75 | 89 | 70 | 48 | 72 | 43 | 45 |
| Relative potential fecundity lost $(\%)$ | 17 | 6 | 7 | 9 | 6 | 4 | 6 | 4 | 3 |
| Realised fecundity $(\mathrm{n} / \mathrm{g})$ | 1002 | 1033 | 1052 | 1009 | 1070 | 1209 | 1087 | 1147 | 1268 |

## Biomass estimation for NEA mackerel

Total spawning stock biomass (SSB) was estimated using the realised fecundity estimate of 1268 oocytes/gr female, a sex ratio of 1:1 and a raising factor of 1.08 (ICES, 1987). According to the Annual Egg production Method (AEPM) the spawning stock biomass (SSB) was estimated as shown below:

$$
S S B=\frac{T A E P}{F^{\prime}} * s * c f
$$

Where $\mathrm{F}^{\prime}$ represents realized fecundity,s means a sex ratio of 1:1 and cf is a fixed raising factor at 1.08 to convert pre-spawning fish to spawning fish.. Spawning stock biomass was thus calculated :

- 3.065 million tonnes for western area (2019: 2.301).
- 0.499 million tonnes for southern area (2019: 0.792).
- 3.565 million tonnes for western and southern areas combined (2019:3.093)

This is an increase in SSB estimate of 18\% compared to the 2019 SSB estimate (Figure 13)


Figure 13. SSB estimates for NEA mackerel. 1992-2022

Table 5. Total annual egg production and SSB for the western and southern areas and for combined area. Temporal series

| Area | Year | TAEP <br> $\left({ }^{*} \mathbf{1 0}^{\mathbf{1 2}}\right)$ | SSB (kt) |
| :--- | ---: | ---: | ---: |
| Combined | 1992 | 2570 | 3874 |
| Combined | 1995 | 2230 | 3766 |
| Combined | 1998 | 2020 | 4199 |
| Combined | 2001 | 1670 | 3234 |
| Combined | 2004 | 1500 | 3107 |
| Combined | 2007 | 1770 | 3783 |
| Combined | 2010 | 2380 | 4811 |
| Combined | 2013 | 2700 | 4832 |
| Combined | 2016 | 1770 | 3524 |
| Combined | 2019 | 1640 | 3088 |
| Combined | 2022 | 2090 | 3563 |
| Southern | 1992 | 336 | 507 |
| Southern | 1995 | 186 | 370 |
| Southern | 1998 | 479 | 883 |
| Southern | 2001 | 318 | 417 |
| Southern | 2004 | 138 | 309 |
| Southern | 2007 | 348 | 745 |
| Southern | 2010 | 459 | 926 |
| Southern | 2013 | 506 | 904 |
| Southern | 2016 | 225 | 447 |
| Southern | 2019 | 423 | 797 |
| Southern | 2022 | 293 | 499 |
| Western | 1992 | 2230 | 3367 |
| Western | 1995 | 2050 | 3396 |
| Western | 1998 | 1540 | 3316 |
| Western | 2001 | 1350 | 2816 |
|  |  |  |  |


| Western | 2004 | 1360 | 2798 |
| :--- | :--- | :--- | :--- |
| Western | 2007 | 1420 | 3038 |
| Western | 2010 | 1920 | 3884 |
| Western | 2013 | 2200 | 3928 |
| Western | 2016 | 1550 | 3077 |
| Western | 2019 | 1220 | 2291 |
| Western | 2022 | 1800 | 3064 |

## HORSE MACKEREL

## Western horse mackerel egg production by period

Period 3 - In period 3 horse mackerel spawning started in the Cantabrian Sea and southern Biscay, but numbers of eggs found were very low. Higher spawning took place in the Celtic Sea but again numbers were still low (Figure 14).

Period 4 - Horse mackerel spawning continued in the Cantabrian Sea, extending into southern Biscay. Eggs were again found in the Celtic Sea but numbers were lower than in period 3 (Figure 15)

Period 5 - Horse mackerel spawning continues in the Cantabrian Sea, Celtic Sea and northern Bay of Biscay, but still in low numbers. Some eggs were also found south and west of Ireland (Figure 16).

Period 6 -Spawning continued in northern Biscay, the Celtic Sea and to the southwest of Ireland. For the first time in a number of years large numbers of eggs were reported in a number of stations close to the 200 m contour. Peak spawning took place during this period (Figure 17)

Period 7 - Eggs were found from northern Biscay to west of Scotland, being concentrated off the southwest of Ireland. In general egg numbers were low but occasional stations with moderate to high counts were observed (Figure 18).


Figure 14. Western horse mackerel egg production by half rectangle for period 3 (March 4th - April 8 th). Circle areas and colour scale represent horse mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 15 Western horse mackerel egg production by half rectangle for period 4 (April 9th - 29th). Circle areas and colour scale represent horse mackerel stage l eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 16 Western horse mackerel egg production by half rectangle for period 5 (Apr 30th - May 31st). Circle areas and colour scale represent horse mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.


Figure 17 Western horse mackerel egg production by half rectangle for period 6 (June 1st -30 th). Circle areas and colour scale represent horse mackerel stage leggs/m2/day by half rectangle. Crosses represent zero values.


Figure 18 Horse mackerel egg production by half rectangle for period 7 (July 1st - July 31st). Circle areas and colour scale represent horse mackerel stage I eggs/m2/day by half rectangle. Crosses represent zero values.

## TAEP results - Western Horse Mackerel

Period number and duration are the same as those used to estimate the western mackerel stock, as are the dates defining the start and end of spawning (Table 5). The shape of the egg production curve does not suggest that those dates should be altered for 2022 (ICES, 2023). An exercise, similar to the one carried out for mackerel in period 6, was not carried out for horse mackerel as WGMEGS feel that the Netherlands period 6 survey adequately delineated the northern boundary of horse mackerel spawning during this period. The total annual egg production was estimated at $5.51 \times 10^{14}$. This is more than a threefold increase on 2019 which at $1.78 \times 10^{14}$ was the lowest estimate of annual egg production ever recorded for this species. (Figure 20, Table 6)


Figure 19: Egg production by period for the western horse mackerel stock since 2007.


Figure 20: The total annual horse mackerel egg production for 1992 - 2022 for the western stock.

Table 5: Estimate of Western horse mackerel total stage I egg production by period using the histogram method for 2022

| Dates | Period | Days | Annual stage I egg production * $10^{15}$ |
| :---: | :---: | :---: | :---: |
| Feb $1^{\text {st }}-$ Mar $3^{\text {rd }}$ | Pre 3 | 31 | 0.015 |
| Mar $4^{\text {th }}-$ Apr $8^{\text {th }}$ | 3 | 36 | 0.055 |
| Apr $9^{\text {th }}-26^{\text {th }}$ | 4 | 18 | 0.016 |
| Apr $27^{\text {th }}-29^{\text {th }}$ | 4-5 | 3 | 0.003 |
| Apr 30 ${ }^{\text {th }}$ - May 31 ${ }^{\text {st }}$ | 5 | 32 | 0.040 |
| Jun $1^{\text {st }}-5^{\text {th }}$ | 5-6 | 5 | 0.047 |
| Jun $6^{\text {th }}-22^{\text {nd }}$ | 6 | 17 | 0.246 |
| Jun $23^{\text {rd }}-$ Jul $4^{\text {th }}$ | 6-7 | 12 | 0.010 |
| Jul $5^{\text {th }}-25^{\text {th }}$ | 7 | 21 | 0.028 |
| July $26^{\text {th }}$ - July $31^{\text {st }}$ | Post 7 | 6 | 0.001 |
| Total |  |  | 0.551 |

Table 6. Western horse mackerel. The time series of Total Annual Egg Production (TAEP) estimates ( $10^{12} \mathrm{eggs}$ ).

| Year | TAEP $\left({ }^{*} \mathrm{e}^{12}\right)$ |
| :---: | :---: |
| 1992 | 2094 |
| 1995 | 1344 |
| 1998 | 1242 |
| 2001 | 864 |
| 2004 | 884 |
| 2007 | 1486 |
| 2010 | 1033 |
| 2013 | 366 |
| 2016 | 331 |
| 2019 | 178 |
| 2022 | 551 |

## Daily Egg Production Method for western horse mackerel.

Horse mackerel eggs at stage 1a were used to estimate daily egg production. Peak spawning for western horse mackerel is expected to occur in period 6 Peak spawning for western horse mackerel is expected to occur in period 6 . Average daily egg production and total spawning area were estimated for calculating egg production (Table 7) (ICES 2023).

Table 7. Daily egg production estimate for the 2022 using DEPM

| Period | Po <br> $\left(\mathrm{eggs} / \mathrm{m}^{2} /\right.$ day $)$ | Spawning area <br> $\left(\mathbf{m}^{2}\right)\left(\times 10^{12}\right)$ | Ptot <br> $(\mathrm{eggs} /$ day $)\left(\times 10^{13}\right)$ |
| :---: | :---: | :---: | :---: |
| $\mathbf{6}$ | 133.857 | 1.390603962 | 0.186 |

## Fecundity investigations

Fecundity samples for western horse mackerel were taken during the expected peak spawning period in survey in order to develop a modified DEPM approach for estimating the biomass of the horse mackerel stock. The DEPM requires an intensive sampling to estimate adult parameters and it has proved to be difficult to achieve the planned number of samples. During the peak spawning, the probability of finding valid samples increases and therefore adult sampling should be targeted to the months when peak spawning is expected.During the 2022 MEGS surveys, it was planned to collect samples of 1840 female horse mackerel in the peak of spawning (periods 6 and 7). In total 1658 horse mackerel were caught from 41 trawl hauls between $46.70^{\circ} \mathrm{N}$ and $60.70^{\circ} \mathrm{N}$. Of these fish, 587 were females and these ovaries were used for the application of the DEPM. Regarding achieving the sampling goal, only $32 \%$ of the planned samples were collected. Difficulties in reaching the adult sampling target are common during WGMEGS surveys, so deviation from the initial plan was perceived in both periods. Nevertheless, this year the number of females collected increased considerably ( 587 compared to 182 in 2019) thanks to the sampling contribution of the commercial fleet.

Preliminary results of the adult DEPM parameters for western horse mackerel are summarised in Table 8. (ICES 2023)

Table 8 Preliminary mean parameter estimates for 2022 adult horse mackerel period 6 (spawning peak)

| Period | Parameter | mean | var |
| :---: | :--- | :---: | :---: |
| 6 | Female Mean weight $(\mathrm{g})$ | 275.4 | 124.9 |
| 6 | Mean Sex Ratio | 0.508 | 0 |
| 6 | Mean Spawning Fraction | 0.187 | 0 |
| 6 | Mean batch fecundity <br> (N.eggs) | 59860 | 4173054 |
| 6 | Relative mean batch fecundity <br> (eggs/g female) | 219.5 | 2.56 |
| 6 | Daily fecundity <br> (eggs/g female day) | 20.9 | NA |

SSB is then calculated as a ratio of total daily egg production to daily fecundity (DF, eggs day/g female) during period 6. Preliminary results are shown in Table 9).

Table 9. Preliminary SSB estimate for the 2022 Western horse mackereI DEPM

| Period | Daily fecundity <br> (eggs/g female day) | Ptot <br> (eggs/day)( $\mathbf{x 1 0 1 3}$ ) | SSB (kt) |
| :---: | :---: | :---: | :---: |
| 6 | 20.865 | 0.186 | 891.445 |

Horse mackerel length varied from 19.9 to 42 cm in period 6 and from 22 to 40.5 cm in period 7 . The mode was slightly higher in period $6(31.6 \mathrm{~cm})$ than in period $7(30.2 \mathrm{~cm})$. Total weight ranged from $70-548 \mathrm{~g}$ in period 6 and from 96.5 to 550 g in period 7. As for length, the mode for weight was 253 and 230 g for period 6 and period 7 respectively.

## Daily Egg Production Method analyses for mackerel in the western, North Sea and Southern areas

## The western and Southern areas

Following the recommendation of WKMSPA (ICES, 2012b) to compare the Annual Egg Produc-tion Method (AEPM) and the Daily Egg Production Method (DEPM), during 2022 the DEPM was implemented again next to the AEPM, as has been done in last MEGS surveys since 2013 for-ward. Daily egg production has been estimated using stage 1 a mackerel eggs. The peak spawn-ing for mackerel in the western and southern spawning areas is expected to be in Periods 3 and 4 (March and April) (ICES 2021).

Consequently, Total Daily Egg Production (Ptot) was estimated as $1.98 * 10^{13} \mathrm{eggs} /$ day for period 3 and $1.09 * 10^{13}$ eggs/day for period 4. Therefore, $P 0$ tot for period 3 is around twice as high as that for period 4

The results for 2022 DEPM fecundity parameters should be considered preliminary, as an extensive review of the application of the daily method in mackerel will be carried out next autumn 2023 at the Workshop on Mackerel Daily Egg production (WKMADE). The workshop will look into the calculations of spawning fraction and batch fecundity, and derive a Daily Egg Production Method (DEPM) based estimates of Spawning Stock Biomass (SSB)

## The North Sea area

The egg survey in the North Sea has been designed for utilizing the Daily Egg Production Method (DEPM). In 2022 Denmark, England and Norway conducted the North Sea mackerel egg survey in June (period 6). The samples were collected and analysed according to the WGMEGS manuals (ICES 2019a, ICES,2019b)..

The North Sea mackerel survey was carried out from 5th June to 24th June. During this period the
spawning area between 530 N and 620 N was surveyed once, receiving a single coverage (Figure 21).
The survey is designed to cover the entire spawning area with samples collected every half ICES statistical rectangle (ICES 2014).

The spatial stage la egg distribution is shown in Figure 22. Egg distributions are comparable to 2021, however egg numbers seemed to be more evenly distributed throughout the survey area this year. The egg production was calculated for the North Sea between $53^{\circ} \mathrm{N}$ and $62^{\circ} \mathrm{N}$ and bounded by the relevant coastlines to the east and west. No clear pattern in the distribution of egg densities can be observed (Figure 22).

Due to technical reasons, allied to the sampling, the majority of the stations along the transects between 53 and $54^{\circ} \mathrm{N}$ do not have valid quantitative data, however qualitative data describing the mackerel stage la and lb egg abundance are available to interpret the overall egg distribution in this area.(ICES, 2023)

The total area sampled in 2022 was slightly smaller than the area sampled in 2021, the first full transect was started at $54^{\circ} 15^{\prime} \mathrm{N}$ compared to $53^{\circ} 15^{\prime} \mathrm{N}$ in 2021 . The spawning area estimated for 2022 is 371126 km 2 and the average Daily egg production (P0) using stage la mackerel eggs is 18.6 eggs/m2/day. The total Daily egg production (Ptot) was calculated for the total investigated area (Table 10). The total Daily egg production for 2022 was $0.6909^{*} 10^{13} \mathrm{eggs} /$ day. This is a $50 \%$ decrease in egg numbers reported in 2021 (Table 11).

Table 10. Daily egg production estimate for mackerel (stage 1a)(PO) and spawning area in the North Sea in 2022.

| Year | P0 <br> (eggs/m2/day) | Spawning area <br> (km2) | Ptot <br> (eggs/day)(*10 $\mathbf{1 3}^{\mathbf{1 3})}$ |
| :---: | :---: | :---: | :---: |

Table 11 Comparison of Total Daily Egg production (Ptot) between 2022 and 2021

|  | $\mathbf{2 0 2 2}$ | $\mathbf{2 0 2 1}$ |
| :---: | :---: | :---: |
| Ptot * $^{10}{ }^{13}$ | 0.69 | 1.28 |



Figure 21. Survey coverage for the North Sea, 2022.Norway (Green), England (Blue), Denmark (Red).


Figure 22 Heat map of Stage la mackerel egg production (eggs. m2. day-1) by half rectangle for the North Sea, 2022. Grey circles represent observed values, crosses represent observed zeros.

The results for the DEPM fecundity parameters should be considered very preliminary, since a full review of the application of the daily method to mackerel will be undertaken at the WKMADE next autumn.

## References

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[^0]:    ICES INTERNATIONAL COUNCIL FOR THE EXPLORATION OF THE SEA CIEM CONSEIL INTERNATIONAL POUR L'EXPLORATION DE LA MER

[^1]:    *incomplete

[^2]:    * the values of catches inside/outside NEAFC RA have been estimated based on the ICES Preliminary Catch Statistics.

[^3]:    *assuming GM(1996-2022) recruitment in 2024.

[^4]:    * Percentage based on ICES estimate with regards to age samples.
    **provided only length distributions
    ***provided length distributions not incl. in InterCatch
    ****age samples processed by the Netherlands

[^5]:    ${ }^{1}$ Quotas as reported to NEAFC by each party for 2023 and include exchange to other parties.
    ${ }^{2}$ Expected uptake in 2023 excluding any exchange/transfer to other parties.
    ${ }^{3}$ The uptake of mackerel in 2023 was estimated to follow a similar development in the fishery for the remainder of the 2023 season as in the last two years.
    ${ }^{4}$ Previous years estimate.
    ${ }^{5}$ No estimates of banking from 2023 to 2024 are available.

[^6]:    * Missing value of mean weight-at-age per component are replaced by component mean value in the calculation of the stock weights

[^7]:    $4 \quad 0.282 \quad 0.278 \quad 0.283 \quad 0.293 \quad 0.289 \quad 0.281 \quad 0.306$
    $\begin{array}{llllllll}5 & 0.298 & 0.308 & 0.314 & 0.311 & 0.348 & 0.323 & 0.340\end{array}$
    $\begin{array}{llllllll}6 & 0.340 & 0.308 & 0.327 & 0.346 & 0.363 & 0.352 & 0.358\end{array}$
    $7 \quad 0.368 \quad 0.338 \quad 0.346 \quad 0.3650 .3760 .392 \quad 0.398$
    $8 \quad 0.385 \quad 0.377 \quad 0.364 \quad 0.3710 .394 \quad 0.416 \quad 0.408$
    $\begin{array}{llllllll}9 & 0.404 & 0.394 & 0.389 & 0.397 & 0.400 & 0.423 & 0.431\end{array}$
    $\begin{array}{llllllll}10 & 0.424 & 0.426 & 0.419 & 0.428 & 0.423 & 0.446 & 0.439\end{array}$
    $\begin{array}{llllllll}11 & 0.440 & 0.430 & 0.437 & 0.431 & 0.445 & 0.458 & 0.442\end{array}$
    $\begin{array}{llllllll}12 & 0.473 & 0.499 & 0.491 & 0.482 & 0.488 & 0.494 & 0.484\end{array}$

[^8]:    50.9991 .0001 .0001 .0001 .0001 .0001 .000
    $6 \quad 1.0001 .0001 .0001 .0001 .0001 .0001 .000$
    $7 \quad 0.9990 .9990 .9991 .0001 .000 \quad 0.9991 .000$
    $8 \quad 1.0001 .0001 .0001 .0001 .0001 .0001 .000$
    $9 \quad 1.000 \quad 1.000 \quad 1.000 \quad 1.000 \quad 1.000 \quad 1.000 \quad 1.000$
    $\begin{array}{llllllll}10 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000 & 1.000\end{array}$
    $111.0001 .0001 .0001 .000 \quad 1.000 \quad 1.000 \quad 1.000$
    121.0001 .0001 .0001 .0001 .0001 .0001 .000

[^9]:    50.3430 .3270 .3120 .2960 .3120 .3290 .345
    $6 \quad 0.3430 .3270 .312 \quad 0.296 \quad 0.312 \quad 0.329 \quad 0.345$
    $\begin{array}{llllllll}7 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329 & 0.345\end{array}$
    $8 \quad 0.3430 .327 \quad 0.312 \quad 0.296 \quad 0.312 \quad 0.329 \quad 0.345$
    $\begin{array}{lllllllll}9 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329 & 0.345\end{array}$
    $\begin{array}{lllllllll}10 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329 & 0.345\end{array}$
    $\begin{array}{lllllllll}11 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329 & 0.345\end{array}$
    $\begin{array}{lllllll}12 & 0.343 & 0.327 & 0.312 & 0.296 & 0.312 & 0.329\end{array} 0.345$

[^10]:    $50.2250 .245 \quad 0.267 \quad 0.264$
    $6 \quad 0.238 \quad 0.269 \quad 0.324 \quad 0.333$
    $\begin{array}{lllll}7 & 0.185 & 0.283 & 0.281 & 0.371\end{array}$
    $8 \quad 0.185 \quad 0.283 \quad 0.281 \quad 0.371$
    $9 \quad 0.185 \quad 0.283 \quad 0.281 \quad 0.371$
    $\begin{array}{lllll}10 & 0.185 & 0.283 & 0.281 & 0.371\end{array}$
    $\begin{array}{lllll}11 & 0.185 & 0.283 & 0.281 & 0.371\end{array}$
    $\begin{array}{lllll}12 & 0.185 & 0.283 & 0.281 & 0.371\end{array}$

[^11]:    *In 2011 the northern part of the Norwegian was not surveyed, hence the index for that year is not representative of mackerel stock size. See IESSNS 2011 cruise report for details.

[^12]:    ${ }^{1}$ Institute of Marine Research, Bergen, Norway
    ${ }^{2}$ DTU-Aqua, Denmark
    ${ }^{3}$ Marine and Freshwater Research Institute, Hafnarfjordur, Iceland
    ${ }^{4}$ Faroe Marine Research Institute, Tórshavn, Faroe Islands
    ${ }^{5}$ Wageningen Marine Research, Netherlands
    ${ }^{6}$ CEFAS, United Kingdom

[^13]:    Figure 2 Total NSSH catches in Q4

[^14]:    * End of Russian participation, ^ excluding Spanish effort due to temporal mismatch.

[^15]:    «ICES defines stakeholders as those who affect or are affected by a decision, process, or action of ICES, including scientists and knowledge providers operating within ICES network. Because of the central role that scientists in the network play in framing and conducting the work of ICES, scientists have a unique role as "internal stakeholders" from within the organization, and this strategy generally reserves the word "stakeholders" to denote all other stakeholders. " Extracted from ICES (2023).

[^16]:    ${ }^{2}$ Arrêté du 12 février 2015 créant un régime national de gestion pour la pêcherie de la sole commune (Solea solea) dans le golfe de Gascogne (divisions CIEM VIII a et b)

